Lawrence Berkeley National Laboratory

LBL Publications

Title

The transformation of U(VI) and V(V) in carnotite group minerals during dissimilatory respiration by a metal reducing bacterium

Permalink

<https://escholarship.org/uc/item/0wq788dk>

Authors

Glasauer, Susan Fakra, Sirine C Schooling, Sarah [et al.](https://escholarship.org/uc/item/0wq788dk#author)

Publication Date 2022-03-01

DOI 10.1016/j.chemgeo.2022.120726

Copyright Information

This work is made available under the terms of a Creative Commons Attribution-NonCommercial License, available at <https://creativecommons.org/licenses/by-nc/4.0/>

Peer reviewed

22 *For Chemical Geology*

23 **ABSTRACT**

24 Recent results from laboratory and field studies support that dissimilatory metal reducing 25 (DMR) bacteria influence the fate and transport of uranium in anaerobic subsurface 26 environments. To date, most research efforts have focused on the reduction of soluble U(VI) by 27 DMR bacteria to form insoluble uraninite (UO_2) . Subsurface environments harbor, however, 28 large reservoirs of U(VI) in solid or mineral form. Uranium that is structure-bound in minerals is 29 expected to be more refractory to microbial reduction than soluble U, based on analogy with Fe 30 respiration. The reducibility of U(VI) could impact the fate of U(IV) by controlling mineral 31 precipitation reactions, which has implications for the long-term immobilization of U in 32 subsurface environments. We studied anoxic cultures of *Shewanella putrefaciens* CN32 incubated 33 with natural carnotite-group minerals by X-ray diffraction, electron microscopy, scanning 34 transmission X-ray microscopy (STXM). Near-edge X-ray absorption fine structure (NEXAFS) 35 spectroscopy measurements at U-N_{4.5}, V-L_{2.3}, and O-K edges on cultures incubated up to 10 36 months show that $V(V)$ was reduced to $V(IV)$, whereas U was not reduced. In contrast, $V(V)$ and 37 U(VI) in solution were both completely reduced to lower oxidation states by CN32, as 38 precipitates within the exopolymer surrounding the bacteria. Assays for the toxicity of U and V to 39 CN32 showed that biofilm formation was stimulated at 0.001 M U(VI), and growth was inhibited 40 at concentrations of U(VI) greater than 0.001 M. Vanadium did not inhibit growth or stimulate 41 biofilm formation at any concentration tested. Investigations of the bacteria-mineral and bacteria-42 metal interface at the nanometer and molecular scales provide new insights into the co-respiration 43 of V and U that help explain their biogeochemical cycling and have implications for subsurface 44 bioremediation of these elements.

- 45
- 46

47 **1.0 INTRODUCTION**

48 The most important ore minerals of the Colorado Plateau uranium deposits belong to the 49 uranyl vanadate group, also known as carnotite group minerals (CGM). These include carnotite 50 [K₂(UO₂)₂(V₂O₈)(H₂O)₃] and tyuyamunite [Ca(UO₂)₂(V₂O₈)(H₂O)₈], which consist of uranyl 51 divanadite $(V_2O_8)^{6}$ -layer complexes that are analogous to layer silicates and contain cations in the 52 interlayer positions (Evans and White, 1987). The ability of uranyl vanadate minerals to 53 accommodate different interlayer cations explains the wide range of possible compositions 54 (Evans and Garrels, 1958; Evans and White, 1987; Finch and Murakami, 1999). The Colorado 55 Plateau ores were processed for radium, vanadium and uranium since the late $19th$ century (Thews 56 and Heinle, 1923; Weeks, 1961), leaving legacies of tailings. Although they are considered 57 insoluble under the slightly alkaline, oxidizing conditions that dominate Colorado Plateau surface 58 environments today (Evans and Garrels, 1958; Langmuir, 1978), extraction activity has resulted 59 in areas of elevated concentrations in Utah and Colorado as well as in other locations where U 60 was processed.

61 The reduction of soluble U(VI) by some bacteria is well documented (e.g., Lovely et al., 62 1991; Haas and DiChristina, 2002; Lloyd et al., 2002; Lloyd and Renshaw, 2005; reviewed in 63 Kolhe et al., 2018). Bioremediation strategies that center on manipulating the activities of 64 dissimilatory metal reducing (DMR) bacteria have been explored for the removal of U(VI) from 65 groundwater, perhaps most notably at Rifle, CO (e.g., Xu et al., 2017; Bargar et al., 2013; 66 Williams et al. 2011; Zhuang et al., 2012; Li 2010; Vrionis et al., 2005; Ortiz Bernad et al., 67 2004b; Anderson et al., 2003). Subsurface bioremediation strategies are based on the reduction of 68 U(VI) species to relatively insoluble hydroxylated uranate complexes (Langmuir, 1978; Bargar et 69 al., 2013; Stylo et al., 2013). In contrast to studies with soluble U(VI), the ability of DMR 70 bacteria to reduce U(VI) that is present in natural minerals has received less attention. This is a

71 striking gap in the understanding of how bacteria may transform U, given that solid mineral 72 phases are the largest reservoirs of metals in weathering environments, as well as the ultimate 73 sinks. Several synthetic U(VI) mineral analogs have been shown to be reducible by dissimilatory 74 bacteria, including: metaschoepite [UO3·2H2O] (Fredrickson et al., 2000), uramphite 75 [(NH4)(UO2)(PO4)·H2O] (Khijniak et al., 2005), synthetic U(VI) borate and boronate crystals 76 (Yang et al., 2014), and natural boltwoodite [HK(UO2)(SiO4)·H2O (Liu et al, 2006; Liu et al., 77 2009). In one study, the U(VI) contained in meta-autunite $(Ca[(UO₂)(PO₄)](H₂O)₆)$ could not be 78 reduced by DMR bacteria (Smeaton et al., 2008). The U(VI) minerals in these studies are, 79 however, much less widespread than for CGM, which are associated with roll front deposits in 80 the Colorado Plateau and in Australia (reviewed in Cumberland, 2016).

81 The V(V) contained in CGM is another possible electron acceptor. Given its favorable 82 solution and redox chemistry (Wehrli and Stumm, 1989; Huang et al., 2016), vanadium should 83 compete effectively with U(VI) for electrons produced during respiration by DMR bacteria. The 84 aqueous chemistry of vanadium is complex due to multiple oxidation states and strong tendencies 85 to hydrolyze and polymerize (Macara, 1980; Rehder, 2008). Vanadate ions $(H_2VO_4$ and HVO_4) 86 are relatively stable under oxidizing conditions, exhibiting chemical behavior similar to 87 phosphate. Vanadyl (VO^{2+}) species are found in reducing environments and are typically more 88 insoluble than vanadate ions (Eckstrom et al., 1983; Premovic et al., 1986). Trivalent V occurs in 89 complexes of low solubility under strongly reducing, i.e., sulfidic, conditions (Wehrli and 90 Stumm, 1989). Several bacterial species have been shown to reduce soluble $V(V)$ to $V(IV)$, 91 including *Shewanella oneidensis* and *Geobacter metallireducens* (Lyalikova and Yurkova, 1992; 92 Carpentier et al., 2003, 2005; Ortiz-Bernad et al., 2004a), as well as a native microbial 93 community (Hao et al., 2018). There is just one report of reduction to V(III) (Li et al., 2007). To

94 the best of our knowledge, the bacterial reduction of solid or mineral-bound V has not been 95 investigated.

96 Bacteria introduce complexity to geochemical reactions by their ability to establish micro 97 and nanoscale chemical gradients (Hunter and Beveridge, 2005). To investigate the bioreactivity 98 of mineral U and V, we incubated anoxic cultures of *Shewanella putrefaciens* CN32 with natural 99 CGM contained in U ore associated with sandstone and examined the products using nanoscale 100 synchrotron-based X-ray spectromicroscopy techniques. *S. putrefaciens* CN32 is known to 101 reduce U as well as Fe and other transition metals (reviewed in DiChristina et al., 2005). Because 102 CN32 was originally isolated from Colorado Plateau deposits, it is a relevant model organism for 103 investigating biological contributions to the terrestrial cycling of metals in this environment.

104 Our investigations were centered on testing: 1) whether the presence of $V(V)$, an alternate 105 electron acceptor, will inhibit the reduction of U(VI) by DMR bacteria; 2) whether the chemical 106 phase of U and V, i.e., solid vs. soluble, will affect bacterial dissimilatory reduction and 107 associated mineral products, and 3) bacterial growth responses to U and V. To characterize the 108 samples, we used soft X-ray scanning transmission X-ray microscopy (STXM), electron 109 microscopy, X-ray diffraction and wet chemical techniques. Elemental mapping (C, K, Ca, O, 110 Mn, Fe, V, U and Ba) and NEXAFS spectroscopy at U 4f, V 2p and O1s edges were performed 111 using STXM on minerals and bacteria samples over the 10-month incubation period, in both dry 112 and hydrated sample conditions. The combination of bulk analyses with spectromicroscopic 113 techniques at the nanoscale allowed us to capture the small-scale heterogeneity induced by active 114 bacteria as well as the relative magnitude of the observed changes in mineralogy.

115

116 **2.0 MATERIALS AND METHODS**

117 **2.1 Materials**

118 Material from a stockpile of unprocessed uranium ore was obtained from southeastern 119 Utah, where it is associated with roll front and tabular deposits of the Colorado Plateau (Weeks, 120 1961; Finch and Murakami, 1999). The uranium minerals occur as coatings on consolidated 121 sandstone that is porous and friable, which we identified using XRD. The metal concentrations of 122 solids were determined using an adaptation of EPA SW 846 Method 3050B. The mineral sample 123 (between 10 and 40 mg) was digested with three ml of HCl and 1 ml concentrated HNO₃ (trace 124 metal grade) in a Teflon bomb overnight followed by 110 °C for 3 hours in an oven. The sample 125 was filtered (Whatman #42) and diluted to 50 ml with deionized water, followed by analysis 126 using inductively coupled plasma spectroscopy (ICP-OES; Varian Vista Pro) or atomic 127 absorption spectroscopy with graphite furnace (Varian GTA100Z). All mineral assays were 128 performed in triplicate and replicates had standard deviations of 5% or less. The mineral-coated 129 sandstone contained 14.6 g V kg⁻¹ and 111 g U kg⁻¹. In order to obtain a concentrated sample of 130 the U-bearing minerals for XRD, the fine mineral fraction was separated from the sand grains by 131 agitating and sonicating in deionized water; decanting, centrifuging and drying the separated fine 132 minerals. These were digested and analyzed as described above and characterized also by XRD 133 and SEM-EDS to identify the minerals.

134

135 **2.2 Culture experiments**

136 **2.2.1 Cultivation of CN32 with U and V as electron acceptors**

137 The cultures of *S. putrefaciens* CN32 that we used were originally isolated from the 138 Morrison Formation in New Mexico (Fredrickson et al., 1998). It is, therefore, a terrestrial rather 139 than a marine isolate and occurs in the same geological setting as the U-ore minerals. Cultures 140 were maintained as frozen stocks in our lab and were revived from frozen stock for each 141 experiment. The defined culture medium (DM)contained 10 mM sodium lactate and 1mM

142 phosphate , added as Na2HPO4, as previously described in Glasauer et al., 2003. Cultures reached 143 the stationary growth phase under oxic conditions after around 24 hours. For the incubation 144 experiments with the U-ore, 1.6 g of CGM-sandstone was added to 80 ml of minimal medium + 145 lactate in serum bottles, degassed with N_2 , sealed and autoclaved. Bottles were inoculated in the 146 glove box (Coy; 3% H2/97% Ar) where they remained throughout the experiments. For the 147 treatments with soluble U and V, preparation and incubation conditions were identical, except U 148 and V were added from stock solutions prepared from uranyl acetate and sodium vanadate, 149 respectively. Final concentrations for U and V for these treatments were 1 mM each. 150 CN32 cultures were prepared for inoculation as previously described (Glasauer et al., 151 2003). After conditioning the bacteria to grow on the defined medium (DM), the final pellet was 152 resuspended in DM to form a slurry of bacteria. The slurry was transferred to the glove box (no-153 vacuum mode) and inoculated to achieve an initial concentration of around 10^8 cfu/ml, 154 determined by protein assay (Glasauer et al., 2001). All treatments were performed in triplicate. 155

156 **2.2.2 Growth inhibition assay**

157 For the growth inhibition assay, we selected conditions to model exposure of bacteria to 158 U(VI) and V(IV), the oxidation states of U and V that dominated in the incubation treatments 159 with CGM. Conditions were oxic to maintain the oxidation state of the metals and to facilitate the 160 assay procedures. A modified micro-dilution method (Wiegand et al. 2008) was used to assess 161 soluble U(VI) and V(IV) (0.001-1 mM) for microbial growth inhibition. In order to compare the 162 response of CN32 to *E. coli* K-12, a well characterized bacterial strain, bacteria were cultured in 163 one-tenth strength trypticase soy broth (10 % TSB) that has been shown to be compatible with 164 metal and mineral studies (French, 2013a; Hunter & Beveridge 2005). The volume was set to 200 165 µl/well. Stock solutions containing 0.02 M UO₂(CH₃COOH)₂, 0.1 M VCl₄, or 0.1 M CaCl₂ were

166 deoxygenated by bubbling oxygen-scrubbed N2 gas (30 min/100 ml and 10 min for headspace 167 degassing). VCl4 was prepared in HCl and solutions were pH-adjusted with NaOH. Degassed 168 solutions were transferred to an anaerobic chamber and sterilized by syringe filtration through a 169 sterile 0.22 µm filter into acid-washed autoclaved serum bottles. The stock solutions were 170 removed from the anaerobic chamber, stored in the dark, and visually inspected for precipitates, 171 flocs or colour change prior to use. Aliquots were aseptically removed from the stock solutions, 172 as required, using a syringe and fine gauge needle. *S. putrefaciens* CN32 and *E. coli* K-12 were 173 inoculated into 10 % TSB and grown for 16-20 h (room temperature, 60 rpm). 10 % TSB was 174 inoculated from these starter cultures at 15 % (vol/vol); stirred at 300 rpm (stir plate) at room 175 temperature (*S. putrefaciens* CN32) or 37 °C (*E. coli* K-12), grown to an OD₆₀₀ of 0.4-0.5 units 176 and adjusted to a final in-assay concentration of 5×10^5 CFU/ml. Negative growth control wells 177 contained sterile medium; positive growth control wells contained inoculated medium. The plates 178 were incubated for 20 h at optimum conditions (room temperature for *S. putrefaciens* CN32 or 37 179 °C for *E. coli* K-12), then removed and observations on the presence/absence of visible growth 180 were recorded. The MIC was the lowest concentration which resulted in optically clear wells 181 denoting no cell growth. Assays were done as triplicate replicates and repeated as independent 182 duplicate experiments $(n = 6)$.

183

184 **2.2.3 Biofilm formation assay**

185 The influence of soluble U(VI) and V(IV) (0.001-1 mM) on biofilm formation was 186 performed as for the growth inhibition assay except that the biofilms were grown on the inner 187 wall of sterile glass tubes which contained 2 ml volumes of 10% TSB with U, V or Ca added. 188 Following incubation for 20 h at room temperature (*S. putrefaciens* CN32) or 37 ⁰C (*E. coli* K-189 12), the tubes were removed and 0.1 ml of Hucker's crystal violet was added. After 15 min

190 incubation (room temperature), the contents were gently decanted, excess unretained stain was 191 removed by washing with distilled water, and the tubes were air dried. Retained stain was 192 solubilized with 33% acetic acid and the absorbance (600 nm) was measured using a Bio-Tek 193 EL800 plate reader. Assays were done as triplicate replicates and repeated as independent 194 duplicate experiments $(n = 6)$.

195

196 **2.3 Assessment for biotransformed elements**

197 **2.3.1 Electron microscopy**

198 Samples for transmission electron microscopy (TEM) were prepared as previously 199 described, for whole mount and thin section preparation (Glasauer et al., 2001). No metal stains 200 were used, so that all observed contrast was imparted to the bacteria by the metals (chiefly U and 201 V) present in the culture medium. Observations were made using a Philips CM10 TEM operating 202 at 80 kV, using an EDAX Sapphire detector and Genesis software. Scanning electron microscopy 203 was performed on untreated CGM-ore, and after 4 months incubation with CN32 on a Hitachi 204 S4500 field emission SEM. Secondary electron (SE) images were obtained with a 5 kV electron 205 beam. The samples were sputtered with gold to alleviate charging problems during SEM 206 examination.

207

208 **2.3.2 X-ray diffraction**

209 Biotransformation of the CGM appeared to have ceased by 10 months. At this time, the 210 mineral solids were separated into their component fractions, identified as the sand or fine 211 fraction. Separation was carried out in the glove box (3% H2/97% Ar). Suspensions were shaken 212 and the suspension bearing the fine clay fraction was decanted. This process was repeated until 213 the wash solution remained clear. All washes were combined in one centrifuge tube, which was

214 sealed, removed from the glove box, and centrifuged (5000 x g). The resulting pellet was dried in 215 the glove box and lightly crushed with a mortar for analysis.

- 216 XRD data were acquired with a custom X-ray diffractometer at the Department of
- 217 Physics, University of Guelph. The X-ray diffractograms were recorded from 5 to 80 °20 with a
- 218 step width of 0.0125 °2 Θ and five seconds counting time. The applied Cu wavelength was
- 219 created by a rotating anode. Analysis of the XRD data was carried with the help of the evaluation
- 220 program EVA15.0 and LeBail (LeBail et al., 1988) with TOPAS4-2 by Bruker AXS.
- 221

222 **2.3.3 Scanning Transmission X-ray Microscopy (STXM)**

223 STXM analyses were conducted at the Molecular Environmental Science Beamline 11.0.2 224 (90 – 2000 eV) of the Advanced Light Source at Lawrence Berkeley National Laboratory (Bluhm 225 et al. 2005). Culture samples were enclosed between a pair of 100 nm thick $Si₃N₄$ membranes. A 226 micro-liter droplet of culture was deposited onto a Si_3N_4 window (Silson Ltd.), air-dried and then 227 sandwiched with another window and hermetically sealed with glue. Another batch of treatments 228 were analyzed in hydrated conditions using the same protocol, except the wet droplet was 229 sandwiched immediately and the assembly sealed with glue. Uranium-bearing powder standards 230 were deposited onto a $Si₃N₄$ window using a standard inspection microscope at 20x 231 magnification. A 2 cm human hair fiber fixed to a tantalum wire was used to transfer the 232 radioactive powder particles onto the window, then sandwiched and hermetically sealed with 233 glue. Standards measured were 238 UO₂ obtained from Alfa-Aesar, and mineral samples were 234 obtained from Excalibur Mineral Corporation (New York). Small metal foil containers were 235 inserted into the STXM prior to use with radioactive materials so that radioactive material could 236 be captured in case of membrane failure.

237 STXM measurements were performed using a Fresnel zone plate lens (35nm outer zones)

238 to focus a monochromatic X-ray beam onto a 2D-scanned sample to record images in 239 transmission mode using a scintillator-photomultiplier detector assembly. The imaging contrast is 240 based on core electron excitation by X-ray absorption. X-ray images recorded at energies just 241 below and at the relevant absorption edges were converted into optical density (OD) images and 242 used to derive elemental maps. The optical density (OD) can be expressed for a given X-ray 243 energy by the Beer-Lambert law as OD=-ln(*I/Io*)=*µρt* where I is the transmitted flux through the 244 sample, I_0 is the incident flux, μ is the mass absorption coefficient, ρ is the density and t is the 245 sample thickness. NEXAFS measurements were performed at the V 2p, O 1s and U 4f edges and 246 obtained from image sequences (i.e., stacks) collected at energies spanning the relevant element 247 absorption edges (508-555 eV at V 2p, O 1s edges; 715-800eV at U 4f edges, unless otherwise 248 specified). A minimum of two different sample regions were analyzed for each element and two 249 different batches of samples were analyzed. The STXM was pumped-purged with He to avoid 250 decompressing the Si3N4 sandwiched samples. The theoretical spectral and spatial resolutions 251 during our measurements were +/- 100 meV and 40 nm respectively. The photon energy was 252 calibrated at the C 1s edge using the 3p Rydberg peak of gaseous $CO₂$ at 292.74 eV, at the O1s 253 using the O 1s \rightarrow 3s transition at 538.9 eV of gaseous CO₂ and at the U 4f edges using gaseous 254 Neon transition at 867.3eV. All data processing was carried out using IDL aXis2000 software 255 (Hitchcock, 2019).

256

257 **3.0 RESULTS**

258 **3.1 Characterization of the sandstone ore minerals**

259 The U- and V- containing mineral components occurred as bright yellow coatings of fine 260 basal plates (1-2 μ m) on sand grains (grain size 200-500 μ m) (Figure 1). The U-coated sandstone 261 contained around 11 % U and 1.5 % V by weight. Based on the fit of the diffraction peaks, the

286 change, and Ba was slightly decreased in the post-reduction solids. The increase in K and 287 decrease in Ca of the solids is consistent with the observed transformation from tyuyamunite to 288 carnotite ([K2(UO2)2(V2O8)⋅3H2O]).

289 Bacteria were still viable after 5 months at a density of around 5×10^5 colony forming 290 units (cfu) ml⁻¹. The population reached a plateau around 10^5 cfu ml⁻¹ that was maintained 291 between 6 d and at least 138 d. The cell density was around 10^3 cfu ml⁻¹ after 8 months. Cell 292 numbers could be underestimated due to the tendency of the bacteria to form flocs; other methods 293 for determining cell numbers (i.e., fluorescent probes, light scattering) could not be used due to 294 the interference of the minerals with light. Bacteria were closely associated with the mineral 295 particles throughout the incubation (Figure 3). Over time, biofilms with cells embedded in a 296 matrix of exopolymeric substances (EPS) decreased in abundance, and bacteria were observed 297 mainly as single or few cells by four months (Figure 3 and Appendix B).

298 For comparison with the CGM treatments, we inoculated anoxic media containing 299 dissolved $V(V)$ and $U(VI)$ at 1 mM concentration each with CN32 cultures. These 300 concentrations are below the minimal inhibitory concentrations (MIC) of V(V) and U(VI) for *E.* 301 *coli,* a Gram-negative model bacterium (Nies, 2007); MIC data for CN32 is not available. The 302 cultures were monitored for V, U and cell concentrations for one week, at which point cell counts 303 had declined to less than 20 cfu ml⁻¹ from an initial density after inoculation of around 10^8 cfu ml⁻¹ 304 $^{-1}$, and no further chemical changes were observed. Vanadium reduction began within 20 minutes 305 of inoculation as shown by the rapid development of blue color in the medium and V in solution 306 stabilized at around 0.1 mM by 30 hours (Figure 2). The concentration of U in solution quickly 307 decreased and remained low; this decrease corresponded to the formation of a dark, fine grained 308 precipitate visible in the culture bottles. Uranium in solution also decreased rapidly in the

309 bacteria-free control, which corresponded to the appearance of white precipitates, likely uranium 310 phosphate minerals based on SEM imaging and EDS analysis (Appendix C).

311

312 **3.3 Changes in chemistry and mineralogy during incubation of carnotite group minerals**

313 Oxygen K-edge NEXAFS is an excellent probe for the covalency of actinide-oxide bonds 314 (Wu, 1999; Minasian et al., 2013; Wen et al., 2014). Initial STXM investigations of uranium 315 oxides have shown that the $4d_{5/2}$ edge is the most useful absorption edge for STXM in the soft X-316 ray region above \sim 100 eV. In the case of uranium, this edge has a reproducible charge state shift 317 of ~1.3 eV from uranium (IV) dioxide to uranium (VI) trioxide (Kalkowski et al, 1987; Nilsson et 318 al, 2005). There have been several studies of metallic uranium compounds at the 4d edges 319 (Kalkowski, et al. 1987, Van der Laan 2004). A useful comparison is to the actinide metallic 320 counterparts of the dioxides at the 4d5/2 edge from these studies. The differences in band 321 structures for actinide metal versus actinide dioxide can lead to $4d_{5/2}$ edge peaks with different 322 widths. This has been observed in the NEXAFS of transition metals and transition metal oxides 323 (de Groot, 1991). For vanadium, the 2p $(L_2,3)$ absorption spectra are useful to detect oxidation 324 states of V(V), V(IV) and V(III) (Cressy et al., 1993; Abbate, 1994; Maganas et al., 2014). Using 325 STXM, we identified $V(V)$ and $V(IV)$ during the incubation with the U-sandstone. In support of 326 our initial visual observations, V(IV) was detected 1 d after inoculation and was the dominant 327 oxidation state detected in association with the bacteria, although V(III) was also detected during 328 the 10 month period. Some V persisted as V(V) throughout, supporting that chemical 329 transformation of the U-bearing minerals was incomplete. In contrast, U was detected only as 330 U(VI) (Figure 4). Differences in the spatial distribution of U and V developed over time, as 331 observed on regions of the solid material (Figure 5), and which is consistent with the detection of

332 V(IV). Fine precipitates formed that appeared to coat larger mineral grains and bacteria (Figure 333 3d).

334 The concentration of Fe was relatively low in the fine fractions (Table 1). The iron 335 valency appeared to be stable as Fe(III) throughout the incubation; STXM analysis showed areas 336 of high Fe concentration as particulates. We also examined the sand fraction and fine fractions 337 that formed distinct layers during the incubation using XRD. The sand fraction contained quartz, 338 with traces of cristobalite and feldspar (sanidine) (Figure 1). In the fine fraction, refinement of the 339 data indicated a mixture of K-carnotite, which belongs to the monoclinic space group, and meta-340 tyuyamunite. This fraction was enriched in K and depleted in Ca compared to the initial yellow 341 fine fraction (Table 1), consistent with the appearance of carnotite. Fritzscheite was not detected 342 even though Mn and Ba concentrations in the fine fraction were relatively unchanged after 343 incubation (Table 1). Observations using SEM showed mineral particles that had a platey 344 structure similar to that seen for the initial U-bearing minerals, but the plates were thinner and 345 formed rosettes (Appendix C). SEM-EDS indicated elevated concentrations of P in association 346 with U for these structures, suggesting autunite $\lbrack Ca(\text{UO}_2)_2(\text{PO}_4)_2\cdot 10^{-12}(\text{H}_2\text{O})\rbrack$, which can also 347 accommodate Na and K. Autunite could not be confirmed by bulk XRD likely because of very 348 low abundance, but it is supported by STXM results (Figure 4). Distinct particles of vanadium 349 oxide (VO₂) were observed in association with the bacteria by STXM (discussed below) but were 350 also not detected by bulk XRD. There was no chemical, mineralogical, or visual evidence for a 351 distinct $UO₂$ component.

352

353 **3.4 Associations between bacteria, U and V**

354 **3.4.1 Carnotite group mineral ore**

355 STXM images show the bacteria were closely associated with nm and micron-sized 356 particles of the CGM. (Figure 5). Wet samples imaged using STXM after one week revealed 357 bacteria embedded in an extensive exopolymer matrix that contained protein, lipids, and 358 polysaccharides , consistent with biofilm formation and growth, with distinct mineral particles 359 distributed heterogeneously within the biofilm matrix (Appendix B). Over time, the bacteria 360 appeared to associate preferentially as individuals or as clusters of few cells with the CGM. 361 Sparse, fine (2-3 nm) precipitates accumulated on the bacteria (Figure 3d). We did not observe U 362 or V precipitates in the periplasm or cytoplasm of CN32 during the incubation of CGM, as 363 assessed by TEM observations on thin sections. This supports that metal reduction took place 364 outside the cell, consistent with the location of metal reducing enzymes at the outer leaflet of the 365 outer membrane (Myers and Myers, 1992).

366 Elemental maps of the solid phase after 8 months show the close correspondence between 367 V, U, Ba, Ca and K (Appendix A).

368

369 **3.4.2 Soluble U and V**

370 After inoculation with CN32, the medium became light blue within two hours, indicating 371 reduction of $V(V)$ to $V(V)$. Fine precipitates accumulated in the periplasm of the bacteria and 372 external to the cell during the incubation, as observed by TEM on thin sections and reported by 373 others for dissimilatory U(VI) reduction (e.g., Lloyd et al., 2002). Bacteria formed flocs 374 containing an abundant matrix of exopolymeric substances (EPS) during reducing conditions. 375 The EPS contained proteins, lipids and carbohydrates, and is chemically similar to the biofilms 376 observed for the CGM treatment, as indicated by STXM (data not shown). Uranium and 377 vanadium appeared to concentrate in the extracellular matrix (Figure 6). After 24 hours, we

- 378 detected U(IV) and V(III) within the matrix of bacteria, exopolymer and mineral solids. Uranium 379 (VI), V(V) and V(IV) were not detected after 24 hours (Figure 6).
- 380
-

381 **3.5 Growth inhibition and biofilm assays**

382 Because of the longevity of CN32 in the U-ore treatments compared to those with soluble 383 U and V, we tested growth inhibition over a range of U(VI) and V(IV) concentrations. Vanadyl 384 [V(IV)] was included rather than V(V) because it was the form of V present in association with 385 bacteria during the incubation with the U-ore, whereas U(VI) was the only observed oxidation 386 state of U in the same treatment. Vanadyl is stable in solution against oxidation for at least 24 387 hours (French et al., 2013a). To maintain these oxidation states of U and V, the assays were 388 conducted under aerobic conditions, with the understanding that CN32 adjusts cell wall 389 biochemistry in response to oxygen presence (French et al., 2013a). The results should, therefore, 390 be interpreted qualitatively and to compare relative responses to different metal concentrations. 391 *Escherichia coli*, well studied as a model Gram-negative organism, was included as a reference 392 strain (K-12). Calcium was included as a cation which, at the tested concentrations, would not 393 have negative impact on cell growth. Calcium ions have important roles in maintaining cell wall 394 health and affect fundamental processes such as bacterial adhesion (Ilangovan et al., 2001; Naik 395 et al., 2006). Calcium and V(IV) did not inhibit growth at any of the tested concentrations, for 396 both bacterial species. *S. putrefaciens* was 100-fold more sensitive to the presence of U(VI) than 397 was *E. coli*, with growth suppressed at concentrations of 0.01 and 1 mM respectively. *E. coli* 398 strains in general have shown tolerance to a wide range of metals (Nies, 2007).

399 The influence of these soluble metals on the formation and growth of biofilms was also 400 investigated. As noted earlier, cells surrounded by EPS were intimately associated with the 401 carnotite group minerals or occurred as flocs following incubation with U and V species; both are 402 consistent with descriptions of bacterial growth as a biofilm. To assess biofilm development in 403 the presence of dissolved V and U, biofilms were assessed as staining of biomass attached to 404 glass test tubes, shown by the band thickness and density (Figure 7). Dense biofilms were formed 405 at all concentrations tested of Ca and V for CN32, but only at the lowest concentration of U, 406 0.001 mM. At this concentration, the amount of biofilm was increased by 30% relative to the 407 control. The stimulation of biofilm growth by sub-inhibitory concentrations of U is reminiscent of 408 bacterial responses to environmental stresses such as sub-inhibitory concentrations of antibiotics 409 (Andersson and Hughes, 2014). In contrast, *E. coli* did not show increased biofilm formation in 410 response to sub-inhibitory U(VI) and formed less abundant biofilms overall (Figure 7)*.* The outer 411 membrane of *S. putrefaciens* is known to be perturbed by uranium (French, 2013a), with the 412 possible induction of stress responses (reviewed in Kolhe et al., 2018). Vanadium stimulated 413 biofilm growth of *S. putrefaciens* at 1 mM only, with values increasing by 60 % relative to the 414 control; lower concentrations of V yielded values similar to the control. Calcium had little impact 415 on biofilm growth.

416

417 **4.0 DISCUSSION**

418 **4.1 U and V chemical transformations**

419 The reduction of naturally occurring, solid phase U(VI) has been demonstrated to date 420 only for U(VI) precipitates located between larger lithic fragments (Liu et al., 2009). We did not 421 observe a net chemical reduction of U(VI) in CGM, in agreement with other studies that did not 422 observe U(VI) reduction when different U-containing minerals were incubated with metal 423 reducing bacteria (Ilton et al. 2006; Smeaton et al. 2008). This outcome is likely due to steric and 424 bonding considerations (Stohl and Smith, 1981), similar to limits on Fe reduction observed under 425 nutrient limited conditions (Glasauer et al., 2003).

450 phosphate, readily form polynuclear complexes with phosphate as well as surface complexes with 451 hydrous oxides (Wehrli and Stumm, 1989). The redox transition for $V(V)-V(IV)$ occurs at E_H 452 values comparable to those for Mn(II)-Mn(IV), around $0.1 - 0.5$ V, which is characteristic for 453 sediment-water interfaces (Wehrli and Stumm, 1989). Vanadyl species in many natural waters are 454 predicted to hydrolyze, sorb strongly to mineral surfaces, and are considered relatively insoluble 455 (Wehrli and Stumm, 1989; Huang et al., 2015). The gradual disappearance of total V from 456 solution correlates with the accumulation of V(IV) and V(III) precipitates observed on the 457 bacteria, likely as $VO(OH)_2$ and V_2O_3 . Vanadium(III) species are highly insoluble except in 458 acidic conditions (below pH 2) and in the absence of O_2 (Macara, 1980).

459

460 **4.2 Mineral Transformations**

461 The minerals in the U-sandstone were transformed by two main processes: 1) bacterial 462 dissimilatory reduction of V(V), and 2) exchange of cations contained in the interlayer of the 463 carnotite group minerals. The incubation studies with CGM illustrate how biotic and abiotic 464 processes can simultaneously affect mineral transformation. Given that U(VI) and V(V) only 465 coexist in mineral form when both are oxidized (Langmuir, 1978), the bacterial reduction of V(V) 466 would necessarily release both metals. The separation of V and U consequent to V reduction is 467 supported by the appearance of V oxide particles on the bacteria and the appearance of distinct 468 solids containing U and P. Soluble U was not detected during the transformation of CGM. The 469 high affinity of U for adsorption to organic matter and minerals would scavenge dissolved U, 470 which could help foster the precipitation of U phases. Precipitates of fine-grained V- and U-471 containing phases that formed during the incubation could only be indirectly assessed for mineral 472 properties, i.e., using SEM, TEM and STXM. These nanometer-scale particles can be directly

473 linked to the microbial transformation of V(V) contained in the U-sandstone minerals. STXM 474 elemental maps revealed regions where U appeared to be distinct from V, suggesting separation. 475 Mineral formation was likely fostered by the controlled conditions of our experiments, 476 e.g., the concentration of K in the culture medium likely favored CGM over fritzscheite. The 477 limits to cation substitution for the interlayer of the UO_2VO_4 sheets have not been established. 478 The elements that comprise the distinct variants include K, Pb, Ba, Mn, Cs, Ca, Cu, Na, and Al 479 (no Fe variant is known) (Finch and Murakami, 1999), although it is feasible that additional 480 substituted cations could occur in low concentrations. In addition, structural water in the 481 interlayer can vary. For example, hydration-dehydration for tyuyamunite is reversible for water 482 contents ranging from 3-8.5 H2O, with meta-tyuyamunite at the lower end of that range. (Stern et 483 al.1956). The optical properties and XRD patterns of the CGM variants are distinct, analogous 484 to the behavior of swelling clays such as montmorillonite that respond similarly to hydration or 485 exposure to cations having different radii. Although these transformations were a consequence of 486 experimental conditions, they illustrate the flexible response of the minerals to changes in their 487 chemical surroundings, on short time scales, which can have implications for mineral solubility. 488 For example, carnotite is less soluble than tyuyamunite under some conditions (Hostetler and 489 Garrels, 1962, Weeks, 1961). Carnotite group members typically occur together in complex 490 assemblages which cannot be physically separated and respond readily to environmental 491 conditions (Stern et al., 1956, Finch and Murakami, 1999).

492 In addition, the microscopy studies – both SEM and STXM – support that autunite formed 493 during the incubation, suggesting that U(VI) that was released from the U-ore minerals was 494 immobilized by precipitation with phosphate.

495

496 **4.3 Reactions at bacteria-mineral and bacteria-metal interfaces**

497 The lack of periplasmic precipitates during the dissimilatory reduction of V in CGM 498 supports that reduction took place at the interface between the cell wall and the extracellular 499 environment. For Gram negative bacteria, the outer and plasma membranes sandwich the 500 periplasm, a gel-like region that contains proteins involved in shuttling chemicals and electrons 501 between the outer membrane and the cytoplasm. Shuttling factors include soluble cytochrome 502 proteins that can reduce soluble, oxic forms of metals during anaerobic respiration. If the reduced 503 form of the metal is insoluble, nano-sized precipitates form in the periplasm, as documented for U 504 and Tc (e.g., Lloyd et al., 2002) and observed during the reduction of soluble V and U in our 505 treatments.

506 It is unknown how bacteria that respire metals maintain critical chemical gradients and 507 membrane fluidity, and continue to uptake nutrients, when bulk conditions favor the sorption of 508 metals and minerals to the cell envelope (French et al., 2013b). Bacteria are highly interactive 509 with dissolved metal ions due to a high surface-to-volume ratio and a high density of metal-510 reactive functional groups in the cell wall (Beveridge, 1989). They develop extracellular 511 gradients by actively and passively taking up and expelling metals and other chemical species. As 512 a result, the interface between the cell wall and the immediate extracellular environment differs in 513 metal composition and concentration from the bulk suspending fluid. Metal speciation is also a 514 factor; for example, the rate of $Fe³⁺$ reduction has been shown to depend on the form of the metal 515 that is present: soluble, complexed, sorbed or mineral (Urrutia et al., 1998; Zachara et al., 1998; 516 Haas and DiChristina, 2002; Glasauer et al., 2003). Cell respiration may contribute to keeping 517 metals in solution near bacteria through H^+ efflux. For example, more acidic pH values were 518 observed proximal to bacteria in a biofilm, relative to the bulk exopolymer, which was suggested 519 to increase metal solubility (Hunter and Beveridge, 2005). Bacteria have biochemical responses 520 to environmental change that may help them resist the impacts of soluble metals on the cell wall.

521 In earlier research, we observed that *S. putrefaciens* CN32 altered lipid chemistry in response to 522 uranium and vanadium, as well as in response to oxygen, which may impact the accumulation of 523 these elements (French et al., 2013a). In the case of the treatments with soluble U and V, the 524 accumulation of these elements in the exopolymer matrix and their exclusion from the bacteria 525 suggest distinct microenvironments, although this remains speculative.

526 Our results support that biofilms and flocs of CN32 immobilize U, as shown by others for 527 U in the presence of microbes in controlled studies (reviewed in Cao et al., 2011; Cologgi, 2014; 528 Stylo et al, 2015). Immobilization has been demonstrated in a field study of natural biofilms 529 (Amano et al., 2017) and in natural organic matter associated with surface and subsurface 530 sediments (Bone et al., 2017; Bone et al., 2020). In particular, Cao et al. (2011) showed that 531 adsorption of U(VI) is competitive between EPS and cells of *Shewanella* HRCR-1, with a higher 532 proportion of U(VI) associated with EPS when U(VI) concentrations were lower. This study (Cao 533 et al., 2011) also showed that the presence of EPS did not affect the reduction efficiency of U(VI) 534 to U(IV), with around 60% of U(VI) reduced at a concentration of 1 mM, identical to the 535 concentration we used to investigate reduction of soluble $U(VI)$ and $V(V)$. Microorganisms 536 respond to environmental stresses such as nutrient limitation as well as to antimicrobial stress by 537 producing EPS (e.g., Myska and Czaczyk, 2009; Andersson and Hughes, 2014). The response of 538 CN32, in terms of growth and proliferation of exopolymeric substances (EPS), may enhance 539 survival by keeping U species from interacting with membrane lipids, which decreases membrane 540 fluidity (French et al., 2013a). The role of EPS in binding potentially toxic elements is not well 541 understood, largely due to differences in methodology (reviewed in Butzen and Fein, 2019). We 542 speculate that the binding of U to high affinity sites in EPS, as observed for Cd (Butzen and Fein, 543 2019), would favor bacterial survival. We cannot, however, infer that the response of CN32 to the 544 lowest concentration of U in our study of biofilm formation is unequivocally a defense

545 mechanism. It is nevertheless remarkable that CN32 responds in a way that appears to keep U 546 away from the cell wall. Our observation that this occurs particularly at sub-inhibitory 547 concentrations of U suggests that EPS formation and reduced toxicity are linked. A link between 548 EPS and reduced toxicity to bacteria may also help to explain why the response to V occurred 549 only at much higher concentrations, given the relatively low toxicity of V.

550 A change from oxidizing to reducing conditions, such as at redox transition zones, may 551 mobilize V in the short term from carnotite-type minerals. Uranium (VI) species that are released 552 consequently will sorb to minerals and precipitate at low ion activity; however, it should be kept 553 in mind that the affinity of U species for organic matter is particularly high (reviewed in 554 Cumberland et al., 2016). In one study, this affinity had a greater impact on U mobility than did 555 complexation by carbonate species, despite thermodynamic predictions (Yang et al, 2012). In our 556 studies, U did not become soluble when the CGM were transformed under reducing conditions, 557 indicated by the lack of U in solution or associated with cells in this treatment. If V and U remain 558 in pore water and adsorbed to solids, a return to oxidizing conditions could induce the 559 precipitation of uranyl vanadate minerals (Tokunaga et al., 2009; Tokunaga et al., 2012). The 560 strong interaction of U(VI) with organic matter is speculated to have immobilized and 561 concentrated U in the Colorado Plateau environment (Cumberland et al., 2016; Spirakis, 1996; 562 Hansley and Spirakis, 1992) with eventual precipitation.

563

564 **4.4 Implications for ore formation**

565 The formation of carnotite-type deposits is controversial. In the Colorado Plateau, it has 566 been hypothesized that cycles of reducing and oxidizing conditions have created the present roll 567 front and tabular structures that characterize the deposits (Weeks,1961; Hostetler and Garrels, 568 1962). V(III) and U(IV) existing separately in primary reduced minerals would have mobilized

569 upon exposure to moderately oxidizing conditions in weathering environments. Transport as 570 U(VI) and V(IV) species under slightly reducing (-0.1 V), alkaline conditions is one proposed 571 scenario (Evans and Garrels, 1958; Weeks, 1961). These redox conditions are similar to those 572 created during the anoxic incubation of CN32 with the CGM. It is likely that U and V were 573 transformed during cycles of oxidizing and reducing conditions; reducing conditions would have 574 prevailed, for example, in the organic-rich Triassic deposits which may have been infiltrated by 575 U-bearing fluids during the Jurassic period (Hansley and Spirakis, 1992; Spirakis, 1996). 576 Subsequent oxidizing conditions would have led to the mobilization and ultimate coprecipitation 577 of V and U as carnotite-type minerals. Bacterial activity may continue to exert an important 578 control on the mobility of both U and V in the subsurface environment of the Colorado Plateau 579 today.

580

581

582 **5.0 CONCLUSIONS**

583 Our results support previous studies in demonstrating the resistance of mineral U(VI) to 584 bacterial reduction. In contrast to those studies, however, the carnotite group minerals include an 585 alternate oxidized element, V(V), that readily served as an electron acceptor in our experiments. 586 Reduction of mineral-bound V(V) did not liberate U into solution; instead, there was evidence 587 that autunite was formed. The initial carnotite group mineral fritzscheite was transformed to K-588 carnotite through the replacement of interlayer cations, indicating that mineral changes were 589 induced through biotic and abiotic pathways. In contrast to our investigations with mineral U and 590 V, soluble V(V) and U(VI) were readily reduced in the presence of metal-respiring bacteria. The 591 abundant exopolymer matrix which surrounded the bacteria during respiration of the soluble 592 electron acceptors appeared to accumulate these elements in preference to bacterial surfaces. This 593 suggests that the biofilm matrix helped to reduce the exposure of bacteria in particular to U, 594 which is highly toxic. Understanding the role of biofilms in ameliorating toxicity is important 595 given that exopolymeric substances are produced by bacteria in response to environmental

596 conditions.

597 Natural environments contain many possible electron acceptors for bacteria that can adapt 598 readily to challenging conditions. Understanding the relative availability of electron acceptors 599 from the microbial perspective is key to interpreting element solubility and mineral 600 transformation reactions in the present – and perhaps in the past.

601

602 **6.0 AKNOWLEDGEMENTS**

603

604 The authors are grateful to Brian S. Fairchild of LBNL EH&S for his help with the STXM 605 experiments. We thank Farhana Islam and Christine Cousins for assistance with the culture 606 experiments at the University of Guelph and Dr. Susan Koval (University of Western Ontario: 607 London, ON, Canada) for provision of *Escherichia coli* K-12 (AB264 lineage). This research was 608 supported by the Natural Science and Engineering Research Council (NSERC) through a 609 Discovery grant to S. Glasauer. Parts of this research, the ALS, and the ALS-MES Beamline 610 11.0.2 were supported by the Director, Office of Science, Office of Basic Energy Sciences, 611 Division of Chemical Sciences, Geosciences, and Biosciences and Materials Sciences Division of 612 the U.S. Department of Energy at the Lawrence Berkeley National Laboratory under Contract 613 No. DE-AC02-05CH11231 (SF, DKS, TT).

614 None of the authors has a competing interest for this research.

615

616 **REFERENCES**

- 618 Abbate M (1994). The O 1s and V 2p X-ray absorption spectra of vanadium oxides. Braz. J.
- 619 Phys. 24: 785-795. 620
- 621 Amano Y, Iwatsuki T, Naganuma T (2017) Characteristics of naturally grown biofilms in deep
- 622 groundwaters and their heavy metal sorption property in a deep subsurface environment. 623 *Geomicrobiol. J*. 34: 769-783.
-
- 624
625 625 Andersson DI, Hughes D (2014) Microbiological effects of sublethal levels of antibiotics. *Nat.* 626 *Rev. Microbiol.* 12: 465-478.
- 627
- 628 Anderson RT, Vrionis HA, Ortiz-Bernad I, Resch CT, Long PE, Dayvault R, Karp K, Marutzky
- 629 S, Metzler DR, Peacock A, White, DC, Lowe M, Lovely DR (2003) Stimulating the in situ
- 630 activity of Geobacter species to remove uranium from the groundwater of a uranium-
- 631 contaminated aquifer. *Appl. Environ. Microbiol.* 69: 5884-5891.
- 632
- 633 Bargar JR, Williams KH, Campbell KM, Long PE, Stubbs JE, Suvorova EI, Lezama-Pacheco JS,
- 634 Alessi DS, Stylo M, Webb SM, Davis JA, Giammar DE, Blue LY, Bernier-Latmani R (2013)
- 635 Uranium redox transition pathways in acetate-amended sediments. *Proc. Natl. Acad. Sci* 110: 636 4506-4511.
- 637
- 638 Beveridge TJ (1989). Metal ions and bacteria. In Metal Ions and Bacteria, TJ Beveridge and RJ 639 Doyle, eds. John Wiley and Sons, New York.
- 640
- 641 Bluhm H., Andersson K., Araki T., Benzerara K., Brown G.E., Dynes J.J., Ghosal S., Gilles
- 642 M.K., Hansen H.-Ch., Hemminger J.C., Hitchcock A.P., Ketteler G., Kilcoyne A.L.D., Kneedler
- 643 E., Lawrence J.R., Leppard G.G., Majzlam J., Mun B.S., Myneni S.C.B., Nilsson A., Ogasawara
- 644 H., Ogletree D.F., Pecher K., Salmeron M., Shuh D.K., Tonner B., Tyliszczak T., Warwick T.,
- 645 Yoon T.H. (2006). Soft X-ray microscopy and spectroscopy at the molecular environmental
- 646 science beamline at the Advanced Light Source, *J. of Electron Spectroscopy and Related*
- 647 *Phenomena*:150: 86-104 648
- 649 Bone SE, Cahill MR, Jones ME, Fendorf S, Davis KHW, Bargar JR (2017). Oxidative uranium 650 release from anoxic sediments under diffusion-limited conditions. *Environ. Sci. Tech.* 51: 11039- 651 11047.
- 652
- 653 Bone SE, Cliff J, Weaver K, Takacs CJ, Roycroft S, Fendorf S, Bargar JR (2020). Complexation 654 by organic matter controls uranium mobility in anoxic sediments. *Environ. Sci. Tech.* 54: 1493- 655 1502.
- 656
- 657 Butzen ML and Fein JB (2019) Influence of extracellular polymeric substances on the adsorption
- 658 of cadmium onto three bacterial species. *Geomicrobiol. J.* 36: 412-422.
- 659

660 Carpentier W, Sandra K, De Smet I, Brigé A, De Smet L and Van Beeumen J (2003) Microbial reduction and precipitation of vanadium by Shewanella oneidensis. Appl. Environ. Microbiol. 6 661 reduction and precipitation of vanadium by Shewanella oneidensis. *Appl. Environ. Microbiol.* 69: 3636-3639. 663 664 Carpentier W, De Smet ., Van Beumen and Brigé A. (2005) Respiration and growth of 665 Shewanella oneidensis MR-1 using vanadate as the sole electron acceptor. *J. Bacteriol*. 187: 666 3293-3301. 667 668 Cao B, Ahmed B, Kennedy DW, Wang Z, Shi L, Marshall MJ, Fredrickson JK, Isern NG, Majors 669 PD, Beyenal H (2011) Contribution of extracellular polymeric substances from Shewanella sp. 670 HRCR-1 biofilms to U(VI) immobilization. *Environ. Sci. Technol*. 45: 5483-5490. 671 672 Cologgi DL, Speers AM, Bullard BA, Kelly SD, Reguera G (2014) Enhanced uranium 673 immobilization and reduction by Geobacter sulfurreducens biofilms. *Appl. Environ. Microbiol.* 674 80: 6638-6646. 675 676 Cressey G, Henderson CMB, van der Laan G (1993) Use of L-edge X-ray absorption 677 spectroscopy to characterize multiple valence states of 3d transition metals; a new probe for 678 mineralogical and geochemical research. Phys. Chem. Minerals 20:111-119. 679 680 Cumberland SA, Douglas G, Grice K, Moreau JW (2016) Uranium mobility in organic matter-681 rich sediments: A review of geological and geochemical processes. *Earth-Science Reviews* 159: 682 160-185. 683 684 de Groot, FMF (1991) Ph.D. Thesis, X-ray Absorption of Transition Metal Oxides, University of 685 Nijmegen, Nijmegen, Netherlands 686 687 DiChristina TJ, Fredrickson JK and Zachara J. (2005) Enzymology of electron transport: Energy 688 generation with geochemical consequences; Reviews in Mineralogy and Geochemistry vol. 59 689 (eds. J. F. Banfield, J. Cervini-Silva and K. M. Nealson), pp. 27-52. Mineralogical Society of 690 America, Washington, D.C. 691 692 Eckstrom A, Fooks CJ., Hambley T, Loeh HJ, Miller SA and Taylor JC (1983) Determination of 693 the crystal structure of a porphyrin isolated from oil shale. *Nature* 306: 173-174. 694 695 Evans HT and Garrels R. (1958) Thermodynamic equilibria of vanadium in aqueous systems as applied to the interpretation of the Colorado Plateau ore deposits. *Geochim. Cosmochim. Acta* 15 applied to the interpretation of the Colorado Plateau ore deposits. *Geochim. Cosmochim. Acta* 15: 697 131-149. 698 699 Evans HT and White JS (1987) The colorful vanadium minerals: A brief review and a new 700 classification. *The Mineralogical Record* 18: 333-340. 701 702 Finch R and Murakami T (1999) Systematics and paragenesis of uranium minerals. In Uranium: 703 Mineralogy, Geochemistry and the Environment; Reviews in Mineralogy vol. 38 (eds. P. C. 704 Burns and R. Finch), pp. 91-180. 705

706 Fredrickson JK, Zachara JM, Kennedy DW, Dong H, Onstott TC, Hinman NW, Li S (1998)
707 Biogenic iron mineralization accompanying the dissimilatory reduction of hydrous ferric oxi 707 Biogenic iron mineralization accompanying the dissimilatory reduction of hydrous ferric oxide by
708 a groundwater bacterium. Geochim. Cosmochim. Acta 62, 3239-3257. 708 a groundwater bacterium. *Geochim. Cosmochim. Acta* 62, 3239-3257. 709 710 Fredrickson JK, Zachara JM, Kennedy DW, Duff MC, Gorby YA, Li S-M, Krupka KM (2000) 711 Reduction of U(VI) in goethite suspensions by a dissimilatory metal-reducing bacterium. 712 *Geochim. Cosmochim. Acta* 64: 3085-3098. 713 714 French S, Fakra S, Trevors ., Glasauer S (2013a) Changes in *Shewanella putrefaciens* CN32 715 membrane lipid chemistry and fluidity in the presence of soluble Mn(II), $V(IV)$, and $U(VI)$.
716 Geomicrobiology Journal 30: 245-254. 716 *Geomicrobiology Journal* 30: 245-254. 717 718 French S, Puddephatt D, Habash M, Glasauer S (2013b) The dynamic nature of bacterial 719 surfaces: Implications for metal-membrane interaction. *Critical Reviews in Microbiology* 39: 196- 217. 721 722 Glasauer S, Langley S, Beveridge TJ (2001) Sorption of Fe(hydr)oxides to the surface of 723 Shewanella putrefaciens: cell-bound fine-grained minerals are not always formed de novo. 724 *Applied and Environmental Microbiology* 67: 5544-5550. 725 726 Glasauer S, Weidler PG, Langley S, Beveridge T J (2003) Controls on Fe reduction and mineral 727 formation by a subsurface bacterium. *Geochim. Cosmochim. Acta* 67: 1277-1288. 728 Haas JR and DiChristina TJ (2002) Effects of Fe(III) chemical speciation on dissimilatory Fe(III) 730 reduction by *Shewanella putrefaciens. Environ. Sci. Technol*. 36: 373-380. 731 732 Hansley PL, Spirakis CS. 1992. Organic-matter diagenesis as the key to a unifying theory for the 733 genesis of tabular uranium-vanadium deposits in the Morrison Formation, Colorado Plateau. 734 *Economic Geology* 87: 352-365. 735 736 Hao L, Zhang B, Feng C, Zhang Z, Lei Z, Shimizu K, Xuelong C, Liu H, Liu H (2018) Microbial 737 vanadium (V) reduction in groundwater with different soils from vanadium ore mining areas. 738 Chemosphere 202: 272-279. 739 740 Hitchcock A. 2019. An IDL-based analytical package. http://unicorn.mcmaster.ca 741 742 Huang JH, Huang F, Evans L, Glasauer S (2015) Vanadium: Global (bio)geochemistry. *Chemical* 743 *Geology* 417: 68-89. 744 745 Hunter RC and Beveridge TJ (2005) Application of a pH-sensitive fluoroprobe (C-SNARF-4) for 746 pH microenvironment analysis in *Pseudomonas aeruginosa* biofilms. (2005) *Appl. Environ.* 747 *Microbiol.* 71: 2501-2510. 748 749 Hostetler PB, Garrels RM. 1962. Transportation and precipitation of uranium and vanadium at 750 low termperatures with special reference to sandstone-type uranium deposits. *Economic Geology* 751 57: 137-167. 752

- 753 Ilangovan U, Ton-That H, Iwahara J, Schneewind O, Clubb RT (2001) Structure of sortase, the transpeptidase that anchors proteins to the cell wall of Staphylococcus aureus. *Proc. Natl. Acad.*
- 754 transpeptidase that anchors proteins to the cell wall of Staphylococcus aureus. *Proc. Natl. Acad.* 755 *Sci.* 98: 6056-6061.
- 756
- 757 Ilton ES, Liu C, Yantasee W, Wang Z., Moore DA, Felmy AR, Zachara JM (2006) The
- 758 dissolution of synthetic Na-boltwoodite in sodium carbonate solutions *Geochimica et* 759 *Cosmochimica Acta* 70: 4836-4849.
- 760
- 761 Kalkowski G, Kaindl G, Brewer W D, Krone W (1987) Near edge X-ray absorption fine structure 762 in uranium minerals. *Phys Rev B* 35:2667-2677
- 763
- 764 Khijniak TV, Slobodkin A, Coker V, Renshaw JC, Livens FR, Bonch-Osmolovskaya EA,
- 765 Birkeland NK, Medvedeva-Lyalikova NN, Lloyd JR (2005) Reduction of uranium (VI)
- 766 phosphate during growth of the thermophilic bacterium Thermoterrabacterium ferrireducens. 767 *Appl. Environ. Microbiol.* 71: 6423-6426.
- 768

774

- 769 Kolhe N, Zinjarde S, Acharya C (2018) Responses exhibited by various microbial groups relevant 770 to uranium exposure. *Biotechnology Advances* 36: 1828-1846. 771
- 772 Langmuir D. (1978) Uranium solution-mineral equilibria at low temperatures with applications to 773 sedimentary ore deposits. *Geochim. Cosmochim. Acta* 42: 547-569.
- 775 Le Bail A., Duroy H, Fourquet J. (1988). Ab-initio structure determination of LiSbWO₆ by 776 X-ray powder diffraction. *Math. Res. Bull*. 23: 447-452.
- 778 Lee J. D. (1992) Concise Inorganic Chemistry, 4th ed. Chapman and Hall, London.
- 779 780 Li L, Steefel CI, Kowalsky MB, Englert ., Hubbard SS (2010) Effects of physical and
- 781 geochemical heterogeneities on mineral transformation and biomass accumulation during 782 biostimulation experiments at Rifle, Colorado. *J. Contam. Hydrol.* 112: 45-63.
- 783
- 784 Li XS, Glasauer S, Le XC (2007) Speciation of vanadium in oilsand coke and bacterial culture 785 by high performance liquid chromatography inductively coupled plasma mass spectrometry.
- 786 *Analytica Chimica Acta* 602: 17-22 and 648: 128 (2009) 787
- 788 Liu C, Byong-Hun J, Zachara JM, Zheming W, Dohnalkova A and Fredrickson JK (2006) 789 Kinetics of microbial reduction of solid phase U(VI). *Environ. Sci. Technol*. 40: 6290-6296. 790
- 791 Liu C, Zachara JM, Zhong L, Heald SM, Wang Z, Byong-Hun J and Fredrickson JK (2009) 792 Microbial reduction of intragrain U(VI) in contaminated sediment. *Environ. Sci. Technol*. 43: 4928-4933.
- 794
- Lloyd JR, Chesnes J, Glasauer S, Bunker DJ, Livens FR and Lovley DR (2002) Reduction of 796 actinides and fission products by Fe(III)-reducing bacteria. *Geomicrobiol. J.* 19: 103-120.
- 797
- 798 Lloyd JR and Renshaw JC (2005) Bioremediation of radioactive waste: radionuclide-microbe
799 interactions in laboratory and field-scale studies. *Current Opinion in Biotechnology* 16: 254-20
- 799 interactions in laboratory and field-scale studies. *Current Opinion in Biotechnology* 16: 254-260. 800
- 801 Lovley DR, Phillips EJP, Gorby YA, Landa ER (1991) Microbial reduction of uranium. *Nature* 802 350: 413-416.
- 803
- 804 Lyalikova NN and Yurkova NA (1992) Role of microorganisms in vanadium concentration and 805 dispersion. *Geomicrobiology Journal* 10: 15-26.
- 806
- 807 Macara IG (1980) Vanadium an element in search of a role. *Trends in Biochemical Sciences* 5: 808 92-94.
- 809
- 810 Maganas D, Moemelt M, Weyhermuller T, Blume R, Havecker M, Knop-Gericke A, DeBeer S,
- 811 Schlogl R, Neese F. 2014. L-edge X-ray absorption study of mononuclear vanadium complexes and spectral predictions using a restricted open shell configuration interaction ansatz. Phys.
- and spectral predictions using a restricted open shell configuration interaction ansatz. Phys.
- 813 Chem. Chem. Phys. 16: 264-276.
-
- $\frac{814}{815}$
- 815 Minasian SG, Keith JM, Batista ER, Boland KS, Bradley JA, Daly SR, Sokaras D, Kozimor SA, 816 Lukens WW, Martin RL, Nordlund D, Seidler GT, Shuh DK, Tyliszczak T, Wagner G, Weng T-816 Lukens WW, Martin RL, Nordlund D, Seidler GT, Shuh DK, Tyliszczak T, Wagner G, Weng T-
817 C, Yang P, (2013) Covalency in metal-oxygen multiple bonds evaluated K-edge spectroscopy
- 817 C, Yang P. (2013) Covalency in metal-oxygen multiple bonds evaluated K-edge spectroscopy
818 and electronic structure theory. J. Amer. Chem. Soc. 135, 1864-1871 (2013). DOI:
- 818 and electronic structure theory. *J. Amer. Chem. Soc.* 135, 1864-1871 (2013). DOI: 819 10.1021/ja310223b 819 10.1021/ja310223b
- 820
- 821 Moskalyk RR and Alfantazi AM (2003) Processing of vanadium: a review. *Minerals* 822 *Engineering* 16: 793-805.
- 823
- 824 Myers CR and Myers JM (1992) Localization of cytochromes to the outer membrane of 825 anaerobically grown Shewanella putrefaciens MR-1. *Journal of Bacteriology* 174: 3429-3438.
- 826
827
- 827 Myszka K, Czaczky K (2009) Characterization of adhesive exopolysaccaride (EPS) produced by 828 Pseudomonas aeruginosa under starvation conditions. *Curr. Microbiol*. 58: 541-546.
- 829
- 830 Naik MT, Suree N, Ilangovan U, Liew CK, Thieu W, Campbell DO, Clemens JJ, Jung ME,
- 831 Clubb RT (2006) Staphylococcus aureus Sortase A transpeptidase Calcium promotes sorting
- 832 signal binding by altering the mobility and structure of an active site loop. *Journal of Biological* 833 *Chemistry* 281: 1817-1826.
- 834
- 835 Nies DH (2007) Bacterial transition metal homeostasis. In Molecular Microbiology of Heavy 836 Metals, DH Nies and S Silver (eds), Microbiology Monographs 6, pp 117-142.
- 837
- 838 Nilsson HJ, Tyliszczak T, Wilson RE, Werme L, Shuh DK (2005) Soft X-ray scanning
- 839 transmission X-ray microscopy (STXM) of actinide particles. *J. Anal. Bioanal. Chem*. 383: 41- 840 47 (2005). DOI: 10.1007/s00216-005-3355-5
- 841
- 842 Ortiz-Bernad I, Anderson RT, Vrionis HA and Lovley DR (2004a) Vanadium respiration by
- 843 Geobacter metallireducens: Novel strategy for in situ removal of vanadium from groundwater.
- 844 *Appl. Environ. Microbiol.* 70: 3091-3095.

845
846 846 Ortiz-Bernad I, Anderson RT, Vrionis HA and Lovley DR (2004b) Resistance of solid-phase
847 U(VI) to microbial reduction during *in situ* bioremediation of uranium-contaminated U(VI) to microbial reduction during *in situ* bioremediation of uranium-contaminated 848 groundwater. *Appl. Environ. Microbiol*. 70: 7558-7560. 849 850 Premovic PI, Pavlovic MS and Pavlovic NZ (1986) Vanadium in ancient sedimentary rock of 851 marine origin. *Geochim. Cosmochim. Acta* 50: 1923-1931. 852 853 Rehder D. Bioinorganic Vanadium Chemistry. John Wiley & Sons, Chichester, 2008. 854
855 Smeaton CM, Weisener CG, Burns PC, Fryer BJ, Fowle DA (2008) Bacterially enhanced 856 dissolution of meta-autunite. *American Mineralogist* 93: 1858-1864. 857 858 Spirakis, CS (1996) The roles of organic matter in the formation of uranium deposits in 859 sedimentary rocks. *Ore Geology Reviews* 11: 53-69. 860 861 Stern TW, Stieff LR, Girhard MN, Meyrowitz R (1956) The occurrence and properties of meta-862 tyuyamunite, Ca(UO2)2(VO4)2∙3-5H2O. *American Mineralogist* 41:187-201. 863 864 Stohl FV and Smith DK (1981) The crystal chemistry of the uranyl silicate minerals. *Am. Mineral* 865 66:610-625. 866 867 Stylo M, Alessi DS, Shao PP, Lezama-Pacheco JS, Bargar JR, Bernier-Latmani R (2013) 868 Biogeochemical controls on the product of microbial U(VI) reduction. *Environ. Sci. Technol*. 47: 869 12351-12358. 870 871 Thews, KB and Heinle FJ (1923) Extraction and recovery of radium, vanadium and uranium from 872 carnotite. Industrial and Engineering Chemistry 15: 1159-1161. 873 874 Tokunaga TK, Kim Y and Wan JM (2009) Potential remediation approach for uranium-875 contaminated groundwaters through potassium uranyl vanadate precipitation. *Environ. Sci.* 876 *Technol.* 43: 5467-5471. 877 878 Tokunaga TK, Kim Y, Wan JM, Yang L (2012) Aqueous uranium (VI) concentrations controlled 879 by calcium uranyl vanadate precipitates. *Environ. Sci. Technol*. 46: 7471-7477. 880 881 Urrutia MM, Roden EE, Fredrickson JK, Zachara JM. (1998) Microbial and surface chemistry 882 controls on reduction of synthetic Fe(III) oxide minerals by the dissimilatory iron-reducing 883 bacterium Shewanella alga. *Geomicrobiol. J.* 15: 269-291. 884 885 Van der Laan G, Moore K T, Tobin J G, Chung B W, Wall M A, Schwartz A J (2004) *Phys Rev* 886 *Lett* 93:097401 887 888 Vrionis HA, Anderson RT, Ortiz-Bernad I, O'Neill KR, Resch CT, Peacock AD, Dayvault R, 889 White DC, Long PE, Lovely DR (2005) Microbiological and geochemical heterogeneity in an *in* 890 *situ* uranium bioremediation field site. *Appl. Environ. Microbiol*. 71: 6308-6318. 891

- 892 Weeks, AD (1961) Mineralogy and geochemistry of vanadium in the Colorado Plateau. *Journal* 893 *of the Less-Common Metals* 3: 443-450.
- 894
- 895 Wehrli B and Stumm W (1989) Vanadyl in natural waters: Adsorption and hydrolysis promote 896 oxygenation. *Geochim. Cosmochim. Acta* 53: 69-77.
- 897
- 898 Wen X-D, Löble MW, Batista ER, Bauer E, Boland KS, Burrell AK, Conradson SD, Daly SR,
- 899 Kozimor SA, Minasian SG, Martin RL, McCleskey TM, Scott BL, Shuh DK, Tyliszczak T,
- 900 Electronic Structure and O K-edge XAS Spectroscopy of U₃O₈, *J. Electron Spectros. Rel.* 901 *Phenom*. 194: 81-87 (2014). DOI: 10.1016/j.elecspec.2014.03.005
- 902
903
- 903 Wiegand I., Hilpert K., Hancock, R. E. W. (2008) Agar and broth dilution methods to determine
904 the minimum inhibitory concentration (MIC) of antimicrobial substances. Nature Protocols 3 (2) 904 the minimum inhibitory concentration (MIC) of antimicrobial substances. *Nature Protocols* 3 (2): 905 163-175.
- 906
- 907 Williams K.H., Long P.E., Davis J.A., Wilkins M.J., N'Guessan A.L., Steefel C.I., Yang L.,
- 908 Newcomer D., Spane F.A., Kerkhof L.J., McGuinness L., Dayvault R. and Lovley D.R. (2011)
- 909 Acetate Availability and its Influence on Sustainable Bioremediation of Uranium-Contaminated
- 910 Groundwater, *Geomicrobiology Journa*l 28: 5-6, 519-539.
- 911
- 912 Wu ZY, Jollet F, Gota S, Thromat N, Gautier-Soyer M, Petit T (1999) X-ray absorption at the
- 913 oxygen K edge in cubic f oxides examined using a full multiple-scattering approach. *J. Phys.* 914 *Condens. Matter* 11: 7185-7194.
- 915
916
- 916 Xu J, Veeramani H, Qafoku NP, Singh G, Riquelme MV, Pruden A, Kukkadapu RK, Gartman
917 BN, Hochella MF (2017) Efficacy of acetate-amended biostimulation for uranium sequestration
- 917 BN, Hochella MF (2017) Efficacy of acetate-amended biostimulation for uranium sequestration:
- 918 Combined analysis of sediment/groundwater geochemistry and bacterial community structure.
- 919 *Appl. Geochem.* 78: 172-185.
- 920
- 921 Yang Y, Wang S, Albrecht-Schmitt TE (2014) Microbial dissolution and reduction of uranyl 922 crystals by *Shewanella oneidensis* MR-1. *Chemical Geology* 387: 59-65.
- 923
- 924 Yang U, Saiers JE, Xu N, Minasian SG, Tyliszczak T, Kozimor SA, Shuh DK, Barnett MO.
925 2012. Impact of natural organic matter on uranium transport through saturated geologic mate
- 2012. Impact of natural organic matter on uranium transport through saturated geologic materials: 926 from molecular to column scale. *Enviro. Sci. Technol.* 46: 5931-5938.
-
- 927
928 Yelton A.P., Williams K.H., Fournelle J., Wrighton K.C., Handley K.M., Banfield J.F. Vanadate
- 929 and Acetate Biostimulation of Contaminated Sediments Decreases Diversity, Selects for Specific 930 Taxa, and Decreases Aqueous V5+ Concentration. *Enviro. Sci. Technol.* 47: 6500–6509.
- 931
- 932 Zachara JM, Fredrickson JK, Li SM, Kennedy DW, Smith SC, Gassman PL (1998) Bacterial
- 933 reduction of crystalline Fe3+ oxides in single phase suspensions and subsurface materials. *Am.* 934 *Mineral.* 83:1426-1443.
- 935
- 936 Zhuang K., Ma E., Lovley DR, Mahadevan R. (2012) The design of long-term effective uranium
- 937 bioremediation strategy using a community metabolic model. *Biotechnology and Bioengineering*
- 938 109: 2475-2483.
- 939 940 941 Table 1. Element concentrations for separated fine yellow particles of uranium ore and fine mineral fraction after incubation with *S. putrefaciens.* The standard deviation for. triplicate samples was less than 5%.
- 942 943
- 944

947 **FIGURE CAPTIONS**

948
949

Figure 1. a) Grain of quartz sand coated with U- and V-containing minerals observed by scanning 950 electron microscopy and energy dispersive spectroscopy b) X-ray diffractograms of the separated 951 sand (upper), fine yellow U-containing solids (tyuyamunite; middle) and fine clay material 952 separated after incubation with *S. putrefaciens* CN32. Continuous lines represent results of whole

- 953 pattern fitting.
- 954

955 Figure 2. Changes in solution concentrations of V and U. Concentrations of dissolved vanadium 956 (a) and uranium (b) during anaerobic respiration of *S. putrefaciens* CN32 in the presence of 957 carnotite.Concentrations of dissolved vanadium (c) and uranium (d) during incubation with a
958 mixed solution containing soluble V(V) and U(VI) at 1 mM concentrations. Open diamonds mixed solution containing soluble $V(V)$ and $U(V)$ at 1 mM concentrations. Open diamonds 959 represent treatments inoculated with CN32; filled diamonds are bacteria-free control treatments.

- 960
961
- Figure 3. Transmission electron micrographs showing the association of CN32 with U-ore 962 minerals. a) and b) thin sections of cultures after 3 days; c) whole mount preparation of culture 963 after 3 days; d) whole mount preparation of 4-month culture. Arrows indicate bacteria; scale bars
- 964 are 500 μm (a, c, d) and 250 μm (b).
- 965

966 Figure 4. STXM-derived Vanadium 2p core absorption spectra (left) and Uranium 4d core spectra 967 (right) showing changes during incubation with *S. putrefaciens* CN32. V L_{2,3} absorption maxima 968 shift to lower energy values with reduction from V^{5+} to V^{4+} in the range from 514-520 eV 969 (dashed line, left) and 524-526 eV. U d absorption maxima shift around 1.3 eV to lower energy with reduction from U(VI) to U(IV), from 780 eV to 778.7 eV (N₄ edge), and from 738 eV to 970 with reduction from U(VI) to U(IV), from 780 eV to 778.7 eV (N₄ edge), and from 738 eV to 971 around 737 (N₅ edge) (dashed lines, right), which was not observed for the U-ore treatments. around 737 (N₅ edge) (dashed lines, right), which was not observed for the U-ore treatments. 972

973 Figure 5. Changes over time in the spatial distribution of uranium and vanadium during

974 incubation of *S. putrefaciens* with uranium ore, as shown by scanning transmission X-ray

975 microscopy. After 3 days: a) STXM image recorded at 518.5eV showing bacteria and particles b) 976 chemical map indicating colocalization of $V(V)$ and $U(V)$ and. After 4 months: c) STXM image

- 977 recorded at 518.5eV and d) corresponding chemical map showing distinct V(V) and U(VI) phases
- 978 and e) chemical map evidencing $V(V)$ and $V(IV)$, derived from stack fitting using larger pixel
- 979 size. U(VI) is mainly colocalized with V(IV) and not with V(V). f) STXM image recorded at 980 518.5eV at 4 months and g) corresponding V map (derived from a "stack") showing that bacter 518.5eV at 4 months and g) corresponding V map (derived from a "stack") showing that bacteria
- 981 contain vanadium. Uranium was not detected on the cells. See corresponding spectra in Fig. 4. 982 Arrows point to bacteria. Scale bars are 1 μm.
- 983

984 Figure 6. STXM-derived elemental distribution and chemical speciation of vanadium and

985 uranium in CN32 samples during the reduction of soluble V, U after three days. a) STXM image 986 recorded at 307 eV. Elemental maps showing b) carbon in red, vanadium in green, c) carbon in

987 red, uranium in green. d and e) U N_{4,5} edges NEXAFS spectra showing that U is present as U(IV). 988

989 Figure 7. Biofilm formation of *S. putrefaciens* CN32 and *E. coli* K-12 in response to U(VI) (a,d);

990 V(IV) (b, e) and Ca (c, f). The amount of retained crystal violet stain indicates the amount of

991 biomass adhered to the sides of the glass tubes after 20 h incubation in the presence/absence of

992 added elements.

-
- 994
995 995 Appendix A: STXM image recorded at 738 eV (U N5 -edge) and STXM-derived elemental maps
996 of 8 carnotite after incubation with CN32 for 8 months. Scale bars are 2 microns. of 8 carnotite after incubation with CN32 for 8 months. Scale bars are 2 microns.
- 997 998 Appendix B. STXM-derived elemental maps (at C K and U N4,5 edges) of hydrated samples after 999 1 week incubation with U ore, showing abundant C matrix with potassium associated with 1000 embedded bacterial cells (a). The bacteria are not strongly associated with U (b). Scale bars are
- 1001 500 μm. Arrows point to cells.
- 1002
- 1003 Appendix C. Minerals containing U and P formed during the incubation of CN32 with uranium sandstone ore. sandstone ore.
- 1005
1006
- Appendix D. STXM-derived VL2,3 and O-K edges NEXAFS spectra of CN32 samples during
- 1007 the reduction of soluble V, U after three days suggesting the presence of V(III).
- 1008
- 1009

Absorbance [a.u.]

 $a)$

Concentration of element (mM)