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Authors
Pérez Castro, Sherlynette
Esch, Ellen H
Eviner, Valerie T
et al.

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Title: Exotic herbaceous species interact with severe drought to alter soil N cycling in a semi-arid shrubland

Authors: Sherlynette Pérez Castro¹,², Ellen H. Esch³, Valerie T. Eviner⁴, Elsa E. Cleland⁵, David A. Lipson¹

¹Department of Biology, San Diego State University, 5500 Campanile Drive San Diego, CA 92182-4614
²Ecosystem Center/Bay Paul Center, Marine Biological Laboratory, Woods Hole, MA 02543
³Department of Integrative Biology, University of Guelph, Guelph, Ontario, Canada
⁴Department of Plant Sciences, University of California Davis, One Shields Ave. Davis, CA 95616
⁵Ecology, Behavior & Evolution Section, University of California San Diego, 9500 Gilman Dr. #0116, La Jolla, CA USA 92093-0116

*Corresponding author contact: sperezcastro@mbl.edu, (+1) 2394401240;
ORCID: 0000-0002-2257-0966

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Abstract

Mediterranean-type ecosystems are increasingly threatened by climate change and exotic annual species, jeopardizing the native communities and their global biodiversity. In these systems, soil nitrogen (N) limits net primary production, and its availability can be influenced by both of these stressors. To understand the interactive effects of droughts and exotic herbaceous species on soil N, we monitored the temporal variability of soil inorganic N, net N mineralization, net nitrification, and N leaching under native- and exotic-dominated stands exposed to rainfall manipulation plots in a Mediterranean-type shrub-dominated community. Increasing drought severity resulted in the accumulation of soil $NH_4^+$ and $NO_3^-$, with a more pronounced increase in exotic-dominated plots. Increased net N mineralization and net nitrification and reduced leaching losses were observed as mechanisms of inorganic N accumulation. In comparison to soils under native plants, soils under exotic plants had enhanced leaching losses upon soil rewetting. We propose that distinct traits of exotic annual herbaceous species associated with higher N inputs, faster turnover, and reduced temporal uptake determine the changes in N cycling in response to droughts. Severe droughts and exotic plants may produce a larger, more vulnerable pool of N that is prone to losses while providing a competitive advantage to promote exotic growth in these N-limited ecosystems.

Introduction

Nitrogen (N) cycling is being altered globally more than any other essential nutrient (Fields, 2004). Agriculture and fossil fuel combustion are increasing reactive N in the
atmosphere (Galloway et al., 2008), droughts are increasing the ecosystems’ N losses (Homyak et al., 2017), and invasive plants are altering ecosystems’ N pools and fluxes (Ehrenfeld, 2003; Liao et al., 2008). These N alterations represent a major challenge particularly in ecosystems where N limits net primary production (Delgado-Baquerizo et al., 2014), and its availability is being threatened by multiple co-occurring global change drivers (Chapin et al., 2000; Sala et al., 2000). For example, semi-arid shrublands are currently experiencing invasions by exotic herbaceous species (Minnich and Dezzani, 1998; Yu et al., 2016), more intense and prolonged droughts (Spinoni et al., 2014; Swain et al., 2018), and elevated N deposition (Valliere et al., 2017), constituting multiple stressors that may have large impacts on N cycling.

In semi-arid ecosystems, N’s limited accessibility is controlled by the temporal dynamics of water availability (Reichmann et al., 2013; Shen et al., 2008). During wet periods, N becomes accessible for microbial and plant assimilation and transformations, which contrasts to natural periods of drought when decreases in diffusion and biological uptake result in the accumulation of immobile N replenished from dead biomass and dry deposition (Cregger et al., 2014). The synchronized immobilization by microbes and plants during wet periods coupled with limited diffusivity during dry periods promotes a closed N cycle (recycling) in these systems (Homyak et al., 2016; Joergensen and Wichern, 2018). However, an asynchrony/decoupling between soil N availability and plant and microbial processes creates N leaks, particularly when dry soils rewet outside of the plant growing season (Homyak et al., 2016). Drought-tolerant microbial transformations and a lack of plant demand during ephemeral periods with mobile N availability favors N losses via leaching and gaseous pathways (Dijkstra et al., 2012). However, losses due to
this open N cycling have been reported to be low or limited to intermittent rain pulses in semi-arid ecosystems (Evans and Burke, 2012; Jongen et al., 2013).

More intense and prolonged droughts due to climate change are expected to promote open N cycling in semi-arid shrublands via sustained N processing during periods of low plant uptake (Evans and Burke, 2012; Homyak et al., 2016). Invasions by exotic herbaceous species have also been reported to increase soil N availability and cycling rates (Blank, 2008; Hawkes et al., 2005; Wolkovich et al., 2010) by possessing traits associated with faster N turnover (Castro-Díez et al., 2014). In semi-arid ecosystems this excess of N availability has been associated with detrimental environmental effects, such as increasing emissions of nitrogenous greenhouse gases (Homyak et al., 2016), eutrophication, fire frequency (Fenn et al., 2003), and competitive advantages for exotic plant species (Launganí and Knops, 2009; Liu et al., 2018). While soil N cycling alterations in response to droughts and plant invasions in semi-arid shrublands have been evaluated, a mechanistic understanding of the interactive effects of these co-occurring global change drivers is not only limiting our ability to model ecosystem responses to multiple global change drivers but also limits improvements to conservation efforts aimed at maintaining ecosystem services (Alba et al., 2017; Launganí and Knops, 2009; Valliere, 2016). Levels of uncertainty arise from challenges in the evaluation of temporally dynamic abiotic and biotic drivers (Alba et al., 2017) and the intrinsic variability of dryland ecosystems that are characterized by distinct plant community types (Fenn et al., 2003) and hydrologic regimes (D’Odorico and Bhattachan, 2012).
Here we present the results of a four-year rainfall manipulation field study in a coastal sage scrub community (CSS), an understudied semi-arid Mediterranean-type, shrub-dominated ecosystem distributed from coastal central California, USA to northern Baja California, Mexico, that has been invaded by exotic herbaceous species (Minnich and Dezzani, 1998) and experienced multi-year extreme drought conditions (Swain et al., 2018). Established exotic annual grasses and forbs species differ from dominant CSS native perennial semi-deciduous shrubs not only in growth form and phenology but also in distinct drought strategies (Balachowski et al., 2016; Puritty et al., 2019; Wainwright et al., 2012; Wolkovich et al., 2010). While CSS native shrubs are adapted to persist through seasonal droughts by partial senescence of mature foliage that is released late in the dry season, exotic annual species exhibit faster growth and earlier complete senescence before experiencing dehydration (Balachowski et al., 2016). The monitoring of temporal variability of soil inorganic N, net N mineralization, net nitrification, and \(NO_3^-\) leaching under native- and exotic-dominated rainfall manipulated plots allowed us to assess interactive mechanisms of soil N alterations. Combining field manipulations and functional trait approaches has become a powerful tool to improve our mechanistic understanding of ecosystem responses and feedbacks to multiple global change drivers (Boyd et al., 2018; Laungani and Knops, 2009). We hypothesized that the earlier senescence of exotic species, and their more easily decomposable litter compared to native species, would enhance soil N availability and result in more open cycling in response to droughts (Liao et al., 2008).

Materials and Methods
This study was conducted from 2013 – 2016 at the Santa Margarita Ecological Reserve (SMER) rainfall manipulation experiment (33.44128 N, 117.1644 W), located at the Riverside-San Diego county line, California, USA. The region has a semi-arid Mediterranean-type climate with cool wet winters and hot dry summers with a compared 30-year average of 358 mm yr⁻¹ of rainfall. The site was a grazed pasture prior to the establishment of the Reserve in 1962. Currently, the area is a heterogeneous matrix of native mature shrubs such as *Artemisia californica* and *Salvia mellifera* interspersed with exotic annual species such as *Centaura melitensis*, *Hirschfeldia incana*, *Erodium cicutarium* and *Bromus madritensis*. All plots were located on similar soil, classified as Las Posas loam, which are well drained with an argillic horizon under the top soil (http://websoilsurvey.sc.egov.usda.gov). Soil analyses at the beginning of the experiment showed no differences in physicochemical properties between plots (Table 1). We are therefore confident that the effects observed in this experiment did not arise from pre-experimental variability among plots. The field experiment consisted of 30 3 m x 3 m rainfall shelters covered with a clear, polycarbonate roof permitting light transmission. Half of the plots were dominated by the native mature shrubs, while the remaining plots were dominated by exotic annual species. To implement the rainfall manipulation, rainfall collected from the rain shelters was re-applied to plots using sprinkler heads at three rainfall levels (in each vegetation type, 5 replicates of 50%, 100%, and 150% of ambient rainfall) after rainfall events. During the sampling period, the region experienced below average annual rainfall with the ambient rainfall plots receiving 157, 125, 139, and 156 mm yr⁻¹ during water years 2013 – 2016, respectively. Due to the lack of rainfall, the 50% and 100% rainfall treatments represented severe drought and drought conditions, respectively, and the 150% treatment corresponded to an average water year.
Water year 2015 received major Pacific storms and summer rains associated with the 2015 – 2016 El Niño event.

*N pools*

To evaluate temporal variability in soil inorganic N, we sampled soils from all plots during the winter (rainy season: October-February), spring (peak plant biomass: March-April), and summer (dry season: May-September) over the 4-year sampling period, resulting in 12 sampling events. Soil was sampled to a depth of 15 cm using a 5 cm diameter x 15 cm deep slide hammer soil core sampler. Samples were stored in a cooler for transportation and processed within 24 h of collection. All visible stones, roots, and litter were removed with forceps prior to analysis. Dissolved $NH_4^+$ and $NO_3^-$ were determined in 0.5M potassium sulfate extracts using the indophenol method and by vanadium (III) reduction, respectively, and measured on a spectrophotometer (Miranda et al., 2001).

*Net N mineralization, net nitrification and leaching*

Net N mineralization and nitrification were assessed using in situ soil core incubations with 4 cm diameter and 10 cm depth PVC tubes (DiStefano and Gholz, 1986). Two cores were inserted into the soil inside the plots of 3 out of the 5 replicates for each plant community type and each precipitation manipulation ($n_{natives} = 9$ vs. $n_{exotics} = 9$) at the beginning of the winter.
seasons. One of the cores was immediately removed for analyses ($T_0$) and the other core, which
was fitted with an ion exchange resin bag at the bottom of the tube, was removed during the
summer season ($T_F$). Dissolved $NH_\text{4}^+$ and $NO_\text{3}^-$ were analyzed in the potassium sulfate extracts
from soils in the cores and in 2 M KCl extracts from the ion exchange beads in the resin bags.
Annual N cycling measurements included: (1) N mineralization, calculated as the difference in
DIN ($NH_\text{4}^+ + NO_\text{3}^-$) concentrations between $T_F$ and $T_0$ plus the DIN in resin bags, (2)
nitrification, calculated as the difference in $NO_\text{3}^-$ concentrations between $T_F$ and $T_0$ in resin
bags, and (3) leaching as the amount of DIN in resin bags ($T_F$). Although resin bags may not
retain all $NH_\text{4}^+$ and $NO_\text{3}^-$ flowing down the core, the amount of DIN in resin bags provides a
reliable estimate of N leaching (Bhogal et al., 1999). To compliment leaching measurements,
immediately after the 2015 summer storms, we collected soil samples to a depth of 50 cm and
measured dissolved $NO_\text{3}^-$ on water extracts using ion-chromatography (Thermo ICS-5000+).

Statistical analyses

Data analysis was performed in R version 3.3.3 (R Core Team, 2017). To determine the
effects of rainfall treatment, plant community type, season, and year and their interactions on soil
$NH_\text{4}^+, NO_\text{3}^-, N$ mineralization, nitrification, and leaching, mixed-effects repeated measures
analysis of variance (RM-ANOVA) was employed using the lme function within the ‘nlme’
package (Pinheiro et al., 2018). Linear model assumptions of normality and homoscedasticity
were tested using Shapiro-Wilk and Levene’s test and observed visually using Q-Q plots and
plotting residuals against fitted values, respectively. When significant effects were found,
multiple comparisons between treatments, seasons, and years were performed using the glht
function within the ‘multcomp’ package (Hothorn et al., 2008). P-values less than 0.05 were considered significant.

**Results**

**Nitrogen pools**

We detected significant differences in soil NH\(_4^+\) and NO\(_3^-\) concentrations between years \((p_{NH} = 0.0002 \text{ and } p_{NO} < 0.0001),\) seasons \((p_{NH} = 0.001 \text{ and } p_{NO} =0.01),\) rainfall treatments \((p_{NH} = 0.001 \text{ and } p_{NO} = 0.006),\) and plant community types \((p_{NH} = 0.02 \text{ and } p_{NO} =<0.0001; \text{ Table 1}).\) Mean annual soil NH\(_4^+\) and NO\(_3^-\) concentrations increased over the 4-year sampling period under both plant community types, with a more pronounced increase in exotic-dominated plots (Fig. 1a-b). In 2016, annual average soil NH\(_4^+\) was 3.1 and 2.5 times higher in exotic- and native-dominated plots, respectively, when compared to 2013 (Fig. 1a). Annual average soil NO\(_3^-\) was 6.0 and 8.4 times higher in native- and exotic-dominated plots, respectively, in 2016 when compared to 2013 (Fig. 1b). Increases in soil NH\(_4^+\) and NO\(_3^-\) concentrations were also observed during the winter season (Fig. 2a-b) and in the 50% rainfall treatment under both plant communities (Fig. 3a-b). Soil NH\(_4^+\) in the 50% treatment (severe drought) was 2.0 and 1.3 times higher in exotic- and native-dominates plots, respectively, when compared to the 150% treatment (Fig. 3a). Soil NO\(_3^-\) was 2.0 and 1.8 times higher in exotic- and native-dominates plots, respectively, in the 50% treatment when compared to the 150% treatment (Fig. 3b). Soil NH\(_4^+\) concentrations were increased in exotic plots by approximately 43% based...
on average values across the entire time of observations (estimated mean ± SD, native-dominated: 2.4 ± 3.6 μg N g⁻¹ soil vs. exotic-dominated: 3.4 ± 4.2 μg N g⁻¹ soil). Soil \( NO_3^- \) concentrations were increased in exotic plots by approximately 219% (estimated mean ± SD, native-dominated: 2.2 ± 3.2 μg N g⁻¹ soil vs. exotic-dominated: 7.0 ± 8.2 μg N g⁻¹ soil).

Net N mineralization, net nitrification and leaching

Net N mineralization was influenced by year (p = 0.03), and there was a moderately significant interaction between rainfall treatment and plant community type (p = 0.07; Table 1). Annual net N mineralization was more variable in exotic-dominated plots (Fig. 1c). In exotic plots, net N mineralization rates increased from 2014 to 2015 and decreased from 2015 to 2016. Highest net N mineralization occurred in 2015 in exotic-dominated plots (Fig. 1c). During 2016, net N mineralization was negative (i.e. net immobilization) in exotic plots (Fig. 1c). Net N mineralization differed between the two community types in response to drought. In exotic-dominated plots net N mineralization increased with increasing drought intensity, while in native plots the opposite pattern was present with the highest levels observed in the 150% treatment (Fig. 3c).

While we did not detect significant effects of rainfall treatment, plant community type, or year on net nitrification (Table 2), similar patterns to net N mineralization rates were observed (Figure 1d & 3d). Highest net nitrification rates occurred in 2015 in exotic-dominated plots, and
negative net nitrification was observed in 2016 (Fig. 1d). The response of net nitrification to drought differed in direction between the two community types, although this was non-significant. In exotic-dominated plots net nitrification increased with drought intensity, and in native plots highest levels of net nitrification were found in the 150% treatment (Fig. 3d).

Leaching (DIN in resin bags) was significantly influenced by the interaction between plant community type and year ($p < 0.0001$). In soils under native plants, leaching was higher during 2015 when compared with 2013, 2014, and 2016 (Fig. 4a). In soils under exotic plants, leaching was higher during 2015 and 2016 when compared with 2013 and 2014 (Fig. 4a). During 2015 and 2016 leaching was 1.9 and 4.1 times higher in exotic-dominated plots when compared to native plots, respectively. After the 2015 summer rains we measured a 1.9, 3.9, 5.0, 6.7, and 16.4 times increase in soil dissolved $NO_3^-$ concentrations in exotic plots when compared to native plots at 0-10, 10-20, 20-30, 30-40, and 40-50 cm soil depths, respectively (Fig. 4b).

**Discussion**

**Soil inorganic N accumulation under drier conditions**

Consistent with the generalization that drier conditions facilitate inorganic N accumulation (Evans and Burke, 2012; Homyak et al., 2017, 2016), in this study we found higher levels of soil $NH_4^+$ and $NO_3^-$ concentrations in response to increasing drought severity due both to rainfall manipulations and over the 4-year period of below average annual rainfall under both
plant community types. We observed several patterns that provide insight into the mechanisms of 
soil inorganic N accumulation. First, annual rates of net N mineralization and nitrification rates 
were observed to be resistant to drought severity on average (rainfall treatments, Table 2), and 
net nitrification did not differ significantly during the 4 consecutive years of below average 
annual rainfall, suggesting statistically unaltered contributions to the supply of $NH_4^+$ and $NO_3^-$ 
concentrations during dry periods. The resistance of microbial processes to low water potentials 
has been reported particularly in dryland ecosystems due to their natural exposure to seasonal 
droughts (Canarini et al., 2017; Drigo et al., 2017). However, it is also possible that there were 
seasonal changes in N cycling rates that were not detected with our annual measures. Increased 
inorganic N in response to drought may have also been caused by other drought-tolerant 
microbial N sources, such as asymbiotic N fixation, (Boring et al., 1988), or N deposition, 
particularly dry deposition (Valliere et al., 2017). In the location of our study, N deposition 
occurs at levels ranging from 9-11 kg N ha$^{-1}$yr$^{-1}$ (Fenn et al., 2010). Secondly, we observed a 
significant increase in leaching during 2015 (wettest year) when compared to previous years 
under both plant community types. This pattern suggests inorganic N accumulates through 
reduced leaching losses during drier conditions. Nitrate is highly mobile, and its diffusivity has 
been previously reported limited under drought conditions (Evans and Burke, 2012). Lastly, 
differences in soil inorganic N concentrations and leaching between exotic- and native-
dominated plots suggest that plant-derived mechanisms influence N accumulation.

Soil inorganic N accumulation as affected by invasion
In comparison to soils under native plants, soils under exotic herbaceous species were observed to have higher soil inorganic N accumulation. We propose that exotic herbaceous species increased soil inorganic N availability by making N cycling more rapid. Litter N release may have been faster from the exotic herbaceous species because of their higher litter quality and their earlier and complete senescence of biomass. This explanation is consistent with the generalization that invasive plants increase N pools by possessing traits associated with faster N turnover (Castro-Díez et al., 2014; Liao et al., 2008). Additionally, invasion by exotic species may accelerate N release to soils via increased litter susceptibility to UV photodegradation compared with native litter (Esch et al., 2019a). Vegetation structure could also play a role, aboveground cover by exotic herbaceous species when compared with native shrubs could enhance the effects of solar radiation on litter decomposition. Plant traits associated with N-fixation have been proposed as another mechanism by which invasive plants enhance N pools (Castro-Díez et al., 2014). While not evaluated in this study, higher asymbiotic N fixation activity may be a mechanism by which the non N-fixing exotic species could have increased soil N inputs (Castro-Díez et al., 2014; Eisele et al., 1989; Mack et al., 2001). An increase in the relative abundance of Cyanobacterial DNA sequences has been found in our exotic-dominated plots (Pérez Castro et al., 2019), which are associated with N fixation (Green et al., 2008).

The interaction of invasion and dry conditions

Contrary to soils under native-dominated plots, we observed a moderate positive relationship between drought severity and N accumulation and net transformations under exotic-dominated plant communities. We proposed that accelerated N cycling by exotic herbaceous
species coupled with drought-induced increases in N accumulation generated a larger inorganic N pool. However, significant decreases in soil $NH_4^+$ from 2014 to 2015 (the relatively wet year), and increases in N leaching and higher deep soil $NO_3^-$ levels during 2015 in soils under exotic species when compared to soils under native species, suggest that soils under exotic plants did not retain the accumulated N. We speculate that the shorter phenology of the exotic annual species (Esch et al., 2019b) and reduced N uptake (demand) during water inputs increased its vulnerability to leaching losses. This is consistent with the hypothesis that drought-induced soil inorganic N accumulation is more vulnerable to losses upon rewetting (Evans and Burke, 2012).

Conclusions

In this experimental semi-arid shrubland, we found that plant functional traits are key factors in providing mechanistic explanations of soil responses and feedbacks to environmental disturbances. In comparison to soils under native woody plants, we observed that soils under exotic herbaceous plants had enhanced inorganic N accumulation and leaching losses after experiencing severe drought conditions. Our findings indicate that distinct morphological, physiological, and phenological traits of annual herbaceous species, associated with N higher inputs, faster turnover, and reduced temporal uptake, play a major role in N pools and fluxes. Nitrogen alterations under severe droughts and exotic plants likely produce a larger pool of N that is more vulnerable to increased leaching and may provide a competitive advantage to promote exotic growth and invasiveness success in these N-limited ecosystems (Liu et al., 2018; Vourlitis, 2017). While this study suggests N transformation rates as mechanisms of inorganic N accumulation, future efforts that measure gross N production are needed to confirm this speculation and better understand microbially mediated N-cycling processes.
Acknowledgments

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Liu, G., Yang, Y.B., Zhu, Z.H., 2018. Elevated nitrogen allows the weak invasive plant Galinsoga quadriradiata to become more vigorous with respect to inter-specific competition. Scientific Reports 8, 1–8. doi:10.1038/s41598-018-21546-z


Table 1. Soil physicochemical properties (mean) in native- and exotic-dominated plots by rainfall treatment prior the beginning of the experiment.

<table>
<thead>
<tr>
<th>Plant community type</th>
<th>Native</th>
<th>Exotic</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Rainfall treatment</strong></td>
<td>50%</td>
<td>100%</td>
</tr>
<tr>
<td>Sand (g g⁻¹)</td>
<td>0.3</td>
<td>0.3</td>
</tr>
<tr>
<td>Silt (g g⁻¹)</td>
<td>0.5</td>
<td>0.5</td>
</tr>
<tr>
<td>Clay (g g⁻¹)</td>
<td>0.2</td>
<td>0.2</td>
</tr>
<tr>
<td>K (mg g⁻¹)</td>
<td>0.2</td>
<td>0.1</td>
</tr>
<tr>
<td>Ca (mg g⁻¹)</td>
<td>3.1</td>
<td>2.6</td>
</tr>
<tr>
<td>Mg (mg g⁻¹)</td>
<td>1.3</td>
<td>1.3</td>
</tr>
<tr>
<td>Al (mg g⁻¹)</td>
<td>2.7</td>
<td>2.2</td>
</tr>
<tr>
<td>pH (saturation paste)</td>
<td>6.1</td>
<td>6.2</td>
</tr>
<tr>
<td>Soil organic matter (g g⁻¹)</td>
<td>0.1</td>
<td>0.1</td>
</tr>
</tbody>
</table>
Table 2. Summary of p-values for the linear mixed model for effects of rainfall treatments, plant community type, season, year and the interaction on soil N parameters (there were no significant 4-way interaction).

<table>
<thead>
<tr>
<th>Factor(s)</th>
<th>NH4</th>
<th>NO3</th>
<th>Net Min</th>
<th>Net Nit</th>
<th>Leaching</th>
</tr>
</thead>
<tbody>
<tr>
<td>Rainfall treatment</td>
<td>0.0013*</td>
<td>0.0059*</td>
<td>0.2667</td>
<td>0.3516</td>
<td>0.2499</td>
</tr>
<tr>
<td>Plant community type</td>
<td>0.0168*</td>
<td>&lt;0.0001*</td>
<td>0.1316</td>
<td>0.2969</td>
<td>&lt;0.0001*</td>
</tr>
<tr>
<td>Season</td>
<td>0.0006*</td>
<td>0.0010*</td>
<td>N.A</td>
<td>N.A</td>
<td>N.A</td>
</tr>
<tr>
<td>Year</td>
<td>0.0002*</td>
<td>&lt;0.0001*</td>
<td>0.0265*</td>
<td>0.1548</td>
<td>&lt;0.0001*</td>
</tr>
<tr>
<td>Rain:Plant</td>
<td>0.1983</td>
<td>0.2178</td>
<td><strong>0.0698</strong></td>
<td>0.3176</td>
<td>0.4442</td>
</tr>
<tr>
<td>Rain:Season</td>
<td>0.2038</td>
<td>0.1680</td>
<td>N.A</td>
<td>N.A</td>
<td>N.A</td>
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<tr>
<td>Plant:Season</td>
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<td>0.7013</td>
<td>N.A</td>
<td>N.A</td>
<td>N.A</td>
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<tr>
<td>Rain:Year</td>
<td>0.1617</td>
<td>0.2553</td>
<td>0.7984</td>
<td>0.7497</td>
<td>0.3474</td>
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<tr>
<td>Plant:Year</td>
<td>0.1145</td>
<td>&lt;0.0001*</td>
<td>0.2107</td>
<td>0.1816</td>
<td>&lt;0.0001*</td>
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<tr>
<td>Season:Year</td>
<td><strong>0.0863</strong></td>
<td><strong>0.0029</strong>*</td>
<td>N.A</td>
<td>N.A</td>
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<tr>
<td>Rain:Plant:Season</td>
<td><strong>0.0365</strong>*</td>
<td>0.1335</td>
<td>N.A</td>
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<tr>
<td>Rain:Plant:Year</td>
<td>0.7893</td>
<td>0.6098</td>
<td>0.7927</td>
<td>0.8729</td>
<td>0.5234</td>
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<tr>
<td>Rain:Season:Year</td>
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<td>0.5563</td>
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<td>N.A</td>
<td>N.A</td>
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<tr>
<td>Plant:Season:Year</td>
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<td><strong>0.0523</strong>.</td>
<td>N.A</td>
<td>N.A</td>
<td>N.A</td>
</tr>
</tbody>
</table>
Figure Legends

Figure 1. Inter-annual variation of soil: (a) $NH_4^+$, (b) $NO_3^-$, (c) N mineralization, and (d) nitrification for each plant community type from 2013 to 2016 across rainfall treatments (no plant community type x rainfall treatment interaction). Yellow bars = exotic plots; green bars = native plots. Error bars represent the standard error of the mean.

Figure 2. Intra-annual variation of (a) $NH_4^+$ and (b) $NO_3^-$ for each plant community type from winter (Wi) to summer (Su). Yellow bars = exotic plots; green bars = native plots. Error bars represent the standard error of the mean.

Figure 3. Drought severity effects on (a) $NH_4^+$, (b) $NO_3^-$, (c) N mineralization, and (d) nitrification for each plant community type. Yellow bars = exotic plots; green bars = native plots. Error bars represent the standard error of the mean.

Figure 4. Inter-annual variation of leaching (DIN in resin bags) and deep soil $NO_3^-$ concentrations during the 2015 summer season. Error bars represent the standard error of the mean.
Figure 1

(a) NH$_4$ (µg N g$^{-1}$ soil) with significant effects:
- Plant type: $P = 0.0168$
- Year: $P = 0.0002$
- Plant x Year: $P = 0.1145$

(b) NO$_3$ (µg N g$^{-1}$ soil) with significant effects:
- Plant type: $P < 0.0001$
- Year: $P < 0.0001$
- Plant x Year: $P < 0.0001$

(c) Net N mineralization (g N m$^{-2}$) with significant effects:
- Plant type: $P = 0.1316$
- Year: $P = 0.0265$
- Plant x Year: $P = 0.2107$

(d) Net nitrification (g N m$^{-2}$) with significant effects:
- Plant type: $P = 0.2969$
- Year: $P = 0.1548$
- Plant x Year: $P = 0.1816$
Figure 2

(a) Ammonium (NH₄⁺) concentrations (µg N g⁻¹ soil) show a significant difference between plant types (P = 0.0168) and seasons (P = 0.0006). The interaction between plant type and season is not significant (P = 0.5998).

(b) Nitrate (NO₃⁻) concentrations (µg N g⁻¹ soil) also show significant differences between plant types (P < 0.0001) and seasons (P = 0.0010). The interaction between plant type and season is not significant (P = 0.7013).
Figure 3

- **Figure 3a**: Plant type: $P = 0.0168$
  Rain: 0.0013
  Plant x Rain: $P = 0.1983$

- **Figure 3b**: Plant type: $P < 0.0001$
  Rain: 0.0059
  Plant x Rain: $P = 0.2178$

- **Figure 3c**: Plant type: $P = 0.1316$
  Rain: 0.2667
  Plant x Rain: $P = 0.0698$

- **Figure 3d**: Plant type: $P = 0.2969$
  Rain: 0.3516
  Plant x Rain: $P = 0.3176$
Figure 4
Highlights

- Reduced leaching and microbial N cycling promote soil N accumulation during droughts
- Droughts and exotic plants result in the enhanced accumulation of soil N
- Exotic plants facilitate increases in N leaching upon soil rewetting
Title: Exotic herbaceous species interact with severe drought to alter soil N cycling in a semi-arid shrubland

Authors: Sherlynette Pérez Castro¹², Ellen H. Esch³, Valerie T. Eviner⁴, Elsa E. Cleland⁵, David A. Lipson¹

¹Department of Biology, San Diego State University, 5500 Campanile Drive San Diego, CA 92182-4614
²Ecosystem Center/Bay Paul Center, Marine Biological Laboratory, Woods Hole, MA 02543
³Department of Integrative Biology, University of Guelph, Guelph, Ontario, Canada
⁴Department of Plant Sciences, University of California Davis, One Shields Ave. Davis, CA 95616
⁵Ecology, Behavior & Evolution Section, University of California San Diego, 9500 Gilman Dr. #0116, La Jolla, CA USA 92093-0116

*Corresponding author contact: sperezcastro@mbl.edu, (+1) 2394401240;

ORCID: 0000-0002-2257-0966

Keywords: Climate change, Exotic species invasion, Coastal sage scrub; Soil nitrogen
Abstract

Mediterranean-type ecosystems are increasingly threatened by climate change and exotic annual species, jeopardizing the native communities and their global biodiversity. In these systems, soil nitrogen (N) limits net primary production, and its availability can be influenced by both of these stressors. To understand the interactive effects of droughts and exotic herbaceous species on soil N, we monitored the temporal variability of soil inorganic N, net N mineralization, net nitrification, and \( \text{NO}_3^- \) leaching under native- and exotic-dominated stands exposed to rainfall manipulation plots in a Mediterranean-type shrub-dominated community. Increasing drought severity resulted in the accumulation of soil \( \text{NH}_4^+ \) and \( \text{NO}_3^- \), with a more pronounced increase in exotic-dominated plots. Increased net N mineralization and net nitrification and reduced leaching losses were observed as mechanisms of inorganic N accumulation. In comparison to soils under native plants, soils under exotic plants had enhanced leaching losses upon soil rewetting. We propose that distinct traits of exotic annual herbaceous species associated with higher N inputs, faster turnover, and reduced temporal uptake determine the changes in N cycling in response to droughts. Severe droughts and exotic plants may produce a larger, more vulnerable pool of N that is prone to losses while providing a competitive advantage to promote exotic growth in these N-limited ecosystems.

Introduction

Nitrogen (N) cycling is being altered globally more than any other essential nutrient (Fields, 2004). Agriculture and fossil fuel combustion are increasing reactive N in the
atmosphere (Galloway et al., 2008), droughts are increasing the ecosystems’ N losses (Homyak et al., 2017), and invasive plants are altering ecosystems’ N pools and fluxes (Ehrenfeld, 2003; Liao et al., 2008). These N alterations represent a major challenge particularly in ecosystems where N limits net primary production (Delgado-Baquerizo et al., 2014), and its availability is being threatened by multiple co-occurring global change drivers (Chapin et al., 2000; Sala et al., 2000). For example, semi-arid shrublands are currently experiencing invasions by exotic herbaceous species (Minnich and Dezzani, 1998; Yu et al., 2016), more intense and prolonged droughts (Spinoni et al., 2014; Swain et al., 2018), and elevated N deposition (Valliere et al., 2017), constituting multiple stressors that may have large impacts on N cycling.

In semi-arid ecosystems, N’s limited accessibility is controlled by the temporal dynamics of water availability (Reichmann et al., 2013; Shen et al., 2008). During wet periods, N becomes accessible for microbial and plant assimilation and transformations, which contrasts to natural periods of drought when decreases in diffusion and biological uptake result in the accumulation of immobile N replenished from dead biomass and dry deposition (Cregger et al., 2014). The synchronized immobilization by microbes and plants during wet periods coupled with limited diffusivity during dry periods promotes a closed N cycle (recycling) in these systems (Homyak et al., 2016; Joergensen and Wichern, 2018). However, an asynchrony/decoupling between soil N availability and plant and microbial processes creates N leaks, particularly when dry soils rewet outside of the plant growing season (Homyak et al., 2016). Drought-tolerant microbial transformations and a lack of plant demand during ephemeral periods with mobile N availability favors N losses via leaching and gaseous pathways (Dijkstra et al., 2012). However, losses due to
this open N cycling have been reported to be low or limited to intermittent rain pulses in semi-arid ecosystems (Evans and Burke, 2012; Jongen et al., 2013).

More intense and prolonged droughts due to climate change are expected to promote open N cycling in semi-arid shrublands via sustained N processing during periods of low plant uptake (Evans and Burke, 2012; Homyak et al., 2016). Invasions by exotic herbaceous species have also been reported to increase soil N availability and cycling rates (Blank, 2008; Hawkes et al., 2005; Wolkovich et al., 2010) by possessing traits associated with faster N turnover (Castro-Díez et al., 2014). In semi-arid ecosystems this excess of N availability has been associated with detrimental environmental effects, such as increasing emissions of nitrogenous greenhouse gases (Homyak et al., 2016), eutrophication, fire frequency (Fenn et al., 2003), and competitive advantages for exotic plant species (Laungani and Knops, 2009; Liu et al., 2018). While soil N cycling alterations in response to droughts and plant invasions in semi-arid shrublands have been evaluated, a mechanistic understanding of the interactive effects of these co-occurring global change drivers is not only limiting our ability to model ecosystem responses to multiple global change drivers but also limits improvements to conservation efforts aimed at maintaining ecosystem services (Alba et al., 2017; Laungani and Knops, 2009; Valliere, 2016). Levels of uncertainty arise from challenges in the evaluation of temporally dynamic abiotic and biotic drivers (Alba et al., 2017) and the intrinsic variability of dryland ecosystems that are characterized by distinct plant community types (Fenn et al., 2003) and hydrologic regimes (D’Odorico and Bhattachan, 2012).
Here we present the results of a four-year rainfall manipulation field study in a coastal sage scrub community (CSS), an understudied semi-arid Mediterranean-type, shrub-dominated ecosystem distributed from coastal central California, USA to northern Baja California, Mexico, that has been invaded by exotic herbaceous species (Minnich and Dezzani, 1998) and experienced multi-year extreme drought conditions (Swain et al., 2018). Established exotic annual grasses and forbs species differ from dominant CSS native perennial semi-deciduous shrubs not only in growth form and phenology but also in distinct drought strategies (Balachowski et al., 2016; Puritty et al., 2019; Wainwright et al., 2012; Wolkovich et al., 2010). While CSS native shrubs are adapted to persist through seasonal droughts by partial senescence of mature foliage that is released late in the dry season, exotic annual species exhibit faster growth and earlier complete senescence before experiencing dehydration (Balachowski et al., 2016). The monitoring of temporal variability of soil inorganic N, net N mineralization, net nitrification, and NO$_3^-$ leaching under native- and exotic-dominated rainfall manipulated plots allowed us to assess interactive mechanisms of soil N alterations. Combining field manipulations and functional trait approaches has become a powerful tool to improve our mechanistic understanding of ecosystem responses and feedbacks to multiple global change drivers (Boyd et al., 2018; Laungani and Knops, 2009). We hypothesized that the earlier senescence of exotic species, and their more easily decomposable litter compared to native species, would enhance soil N availability and result in more open cycling in response to droughts (Liao et al., 2008).

Materials and Methods
This study was conducted from 2013 – 2016 at the Santa Margarita Ecological Reserve (SMER) rainfall manipulation experiment (33.44128 N, 117.1644 W), located at the Riverside-San Diego county line, California, USA. The region has a semi-arid Mediterranean-type climate with cool wet winters and hot dry summers with a compared 30-year average of 358 mm yr\(^{-1}\) of rainfall. The site was a grazed pasture prior to the establishment of the Reserve in 1962. Currently, the area is a heterogeneous matrix of native mature shrubs such as *Artemisia californica* and *Salvia mellifera* interspersed with exotic annual species such as *Centaura melitensis*, *Hirschfeldia incana*, *Erodium cicutarium* and *Bromus madritensis*. All plots were located on similar soil, classified as Las Posas loam, which are well drained with an argillic horizon under the top soil (http://websoilsurvey.sc.egov.usda.gov). Soil analyses at the beginning of the experiment showed no differences in physicochemical properties between plots (Table 1). We are therefore confident that the effects observed in this experiment did not arise from pre-experimental variability among plots. The field experiment consisted of 30 3 m x 3 m rainfall shelters covered with a clear, polycarbonate roof permitting light transmission. Half of the plots were dominated by the native mature shrubs, while the remaining plots were dominated by exotic annual species. To implement the rainfall manipulation, rainfall collected from the rain shelters was re-applied to plots using sprinkler heads at three rainfall levels (in each vegetation type, 5 replicates of 50%, 100%, and 150% of ambient rainfall) after rainfall events. During the sampling period, the region experienced below average annual rainfall with the ambient rainfall plots receiving 157, 125, 139, and 156 mm yr\(^{-1}\) during water years 2013 – 2016, respectively. Due to the lack of rainfall, the 50% and 100% rainfall treatments represented severe drought and drought conditions, respectively, and the 150% treatment corresponded to an average water year.
Water year 2015 received major Pacific storms and summer rains associated with the 2015 – 2016 El Niño event.

To evaluate temporal variability in soil inorganic N, we sampled soils from all plots during the winter (rainy season: October-February), spring (peak plant biomass: March-April), and summer (dry season: May-September) over the 4-year sampling period, resulting in 12 sampling events. Soil was sampled to a depth of 15 cm using a 5 cm diameter x 15 cm deep slide hammer soil core sampler. Samples were stored in a cooler for transportation and processed within 24 h of collection. All visible stones, roots, and litter were removed with forceps prior to analysis. Dissolved $\text{NH}_4^+$ and $\text{NO}_3^-$ were determined in 0.5M potassium sulfate extracts using the indophenol method and by vanadium (III) reduction, respectively, and measured on a spectrophotometer (Miranda et al., 2001).

Net N mineralization, net nitrification and leaching

Net N mineralization and nitrification were assessed using in situ soil core incubations with 4 cm diameter and 10 cm depth PVC tubes (DiStefano and Gholz, 1986). Two cores were inserted into the soil inside the plots of 3 out of the 5 replicates for each plant community type and each precipitation manipulation ($n_{\text{natives}}=9$ vs. $n_{\text{exotics}}=9$) at the beginning of the winter
seasons. One of the cores was immediately removed for analyses ($T_0$) and the other core, which was fitted with an ion exchange resin bag at the bottom of the tube, was removed during the summer season ($T_F$). Dissolved $NH_4^+$ and $NO_3^-$ were analyzed in the potassium sulfate extracts from soils in the cores and in 2 M KCl extracts from the ion exchange beads in the resin bags. Annual N cycling measurements included: (1) N mineralization, calculated as the difference in DIN ($NH_4^+ + NO_3^-$) concentrations between $T_F$ and $T_0$ plus the DIN in resin bags, (2) nitrification, calculated as the difference in $NO_3^-$ concentrations between $T_F$ and $T_0$ in resin bags, and (3) leaching as the amount of DIN in resin bags ($T_F$). Although resin bags may not retain all $NH_4^+$ and $NO_3^-$ flowing down the core, the amount of DIN in resin bags provides a reliable estimate of N leaching (Bhogal et al., 1999). To compliment leaching measurements, immediately after the 2015 summer storms, we collected soil samples to a depth of 50 cm and measured dissolved $NO_3^-$ on water extracts using ion-chromatography (Thermo ICS-5000+).

Statistical analyses

Data analysis was performed in R version 3.3.3 (R Core Team, 2017). To determine the effects of rainfall treatment, plant community type, season, and year and their interactions on soil $NH_4^+$, $NO_3^-$, N mineralization, nitrification, and leaching, mixed-effects repeated measures analysis of variance (RM-ANOVA) was employed using the lme function within the ‘nlme’ package (Pinheiro et al., 2018). Linear model assumptions of normality and homoscedasticity were tested using Shapiro-Wilk and Levene’s test and observed visually using Q-Q plots and plotting residuals against fitted values, respectively. When significant effects were found, multiple comparisons between treatments, seasons, and years were performed using the glht
function within the ‘multcomp’ package (Hothorn et al., 2008). P-values less than 0.05 were considered significant.

Results

Nitrogen pools

We detected significant differences in soil $NH_4^+$ and $NO_3^-$ concentrations between years ($p_{NH_4^+} = 0.0002$ and $p_{NO_3^-} < 0.0001$), seasons ($p_{NH_4^+} = 0.001$ and $p_{NO_3^-} = 0.01$), rainfall treatments ($p_{NH_4^+} = 0.001$ and $p_{NO_3^-} = 0.006$), and plant community types ($p_{NH_4^+} = 0.02$ and $p_{NO_3^-} < 0.0001$; Table 1). Mean annual soil $NH_4^+$ and $NO_3^-$ concentrations increased over the 4-year sampling period under both plant community types, with a more pronounced increase in exotic-dominated plots (Fig. 1a-b). In 2016, annual average soil $NH_4^+$ was 3.1 and 2.5 times higher in exotic- and native-dominated plots, respectively, when compared to 2013 (Fig. 1a). Annual average soil $NO_3^-$ was 6.0 and 8.4 times higher in native- and exotic-dominated plots, respectively, in 2016 when compared to 2013 (Fig. 1b). Increases in soil $NH_4^+$ and $NO_3^-$ concentrations were also observed during the winter season (Fig. 2a-b) and in the 50% rainfall treatment under both plant communities (Fig. 3a-b). Soil $NH_4^+$ in the 50% treatment (severe drought) was 2.0 and 1.3 times higher in exotic- and native-dominates plots, respectively, when compared to the 150% treatment (Fig. 3a). Soil $NO_3^-$ was 2.0 and 1.8 times higher in exotic- and native-dominates plots, respectively, in the 50% treatment when compared to the 150% treatment (Fig. 3b). Soil $NH_4^+$ concentrations were increased in exotic plots by approximately 43% based
on average values across the entire time of observations (estimated mean ± SD, native-dominated: 2.4 ± 3.6 μg N g⁻¹ soil vs. exotic-dominated: 3.4 ± 4.2 μg N g⁻¹ soil). Soil NO₃ concentrations were increased in exotic plots by approximately 219% (estimated mean ± SD, native-dominated: 2.2 ± 3.2 μg N g⁻¹ soil vs. exotic-dominated: 7.0 ± 8.2 μg N g⁻¹ soil).

**Net N mineralization, net nitrification and leaching**

Net N mineralization was influenced by year (p = 0.03), and there was a moderately significant interaction between rainfall treatment and plant community type (p = 0.07; Table 1). Annual net N mineralization was more variable in exotic-dominated plots (Fig. 1c). In exotic plots, net N mineralization rates increased from 2014 to 2015 and decreased from 2015 to 2016. Highest net N mineralization occurred in 2015 in exotic-dominated plots (Fig. 1c). During 2016, net N mineralization was negative (i.e. net immobilization) in exotic plots (Fig. 1c). Net N mineralization differed between the two community types in response to drought. In exotic-dominated plots net N mineralization increased with increasing drought intensity, while in native plots the opposite pattern was present with the highest levels observed in the 150% treatment (Fig. 3c).

While we did not detect significant effects of rainfall treatment, plant community type, or year on net nitrification (Table 2), similar patterns to net N mineralization rates were observed (Figure 1d & 3d). Highest net nitrification rates occurred in 2015 in exotic-dominated plots, and
negative net nitrification was observed in 2016 (Fig. 1d). The response of net nitrification to drought differed in direction between the two community types, although this was non-significant. In exotic-dominated plots net nitrification increased with drought intensity, and in native plots highest levels of net nitrification were found in the 150% treatment (Fig. 3d).

Leaching (DIN in resin bags) was significantly influenced by the interaction between plant community type and year (p <0.0001). In soils under native plants, leaching was higher during 2015 when compared with 2013, 2014, and 2016 (Fig. 4a). In soils under exotic plants, leaching was higher during 2015 and 2016 when compared with 2013 and 2014 (Fig. 4a). During 2015 and 2016 leaching was 1.9 and 4.1 times higher in exotic-dominated plots when compared to native plots, respectively. After the 2015 summer rains we measured a 1.9, 3.9, 5.0, 6.7, and 16.4 times increase in soil dissolved $NO_3^-$ concentrations in exotic plots when compared to native plots at 0-10, 10-20, 20-30, 30-40, and 40-50 cm soil depths, respectively (Fig. 4b).

**Discussion**

**Soil inorganic N accumulation under drier conditions**

Consistent with the generalization that drier conditions facilitate inorganic N accumulation (Evans and Burke, 2012; Homyak et al., 2017, 2016), in this study we found higher levels of soil $NH_4^+$ and $NO_3^-$ concentrations in response to increasing drought severity due both to rainfall manipulations and over the 4-year period of below average annual rainfall under both
plant community types. We observed several patterns that provide insight into the mechanisms of soil inorganic N accumulation. First, annual rates of net N mineralization and nitrification rates were observed to be resistant to drought severity on average (rainfall treatments, Table 2), and net nitrification did not differ significantly during the 4 consecutive years of below average annual rainfall, suggesting statistically unaltered contributions to the supply of $NH_4^+$ and $NO_3^-$ concentrations during dry periods. The resistance of microbial processes to low water potentials has been reported particularly in dryland ecosystems due to their natural exposure to seasonal droughts (Canarini et al., 2017; Drigo et al., 2017). However, it is also possible that there were seasonal changes in N cycling rates that were not detected with our annual measures. Increased inorganic N in response to drought may have also been caused by other drought-tolerant microbial N sources, such as asymbiotic N fixation, (Boring et al., 1988), or N deposition, particularly dry deposition (Valliere et al., 2017). In the location of our study, N deposition occurs at levels ranging from 9-11 kg N ha$^{-1}$ yr$^{-1}$ (Fenn et al., 2010). Secondly, we observed a significant increase in leaching during 2015 (wettest year) when compared to previous years under both plant community types. This pattern suggests inorganic N accumulates through reduced leaching losses during drier conditions. Nitrate is highly mobile, and its diffusivity has been previously reported limited under drought conditions (Evans and Burke, 2012). Lastly, differences in soil inorganic N concentrations and leaching between exotic- and native-dominated plots suggest that plant-derived mechanisms influence N accumulation.

Soil inorganic N accumulation as affected by invasion
In comparison to soils under native plants, soils under exotic herbaceous species were observed to have higher soil inorganic N accumulation. We propose that exotic herbaceous species increased soil inorganic N availability by making N cycling more rapid. Litter N release may have been faster from the exotic herbaceous species because of their higher litter quality and their earlier and complete senescence of biomass. This explanation is consistent with the generalization that invasive plants increase N pools by possessing traits associated with faster N turnover (Castro-Díez et al., 2014; Liao et al., 2008). Additionally, invasion by exotic species may accelerate N release to soils via increased litter susceptibility to UV photodegradation compared with native litter (Esch et al., 2019a). Vegetation structure could also play a role, aboveground cover by exotic herbaceous species when compared with native shrubs could enhance the effects of solar radiation on litter decomposition. Plant traits associated with N-fixation have been proposed as another mechanism by which invasive plants enhance N pools (Castro-Díez et al., 2014). While not evaluated in this study, higher asymbiotic N fixation activity may be a mechanism by which the non N-fixing exotic species could have increased soil N inputs (Castro-Díez et al., 2014; Eisele et al., 1989; Mack et al., 2001). An increase in the relative abundance of Cyanobacterial DNA sequences has been found in our exotic-dominated plots (Pérez Castro et al., 2019), which are associated with N fixation (Green et al., 2008).

The interaction of invasion and dry conditions

Contrary to soils under native-dominated plots, we observed a moderate positive relationship between drought severity and N accumulation and net transformations under exotic-dominated plant communities. We proposed that accelerated N cycling by exotic herbaceous
species coupled with drought-induced increases in N accumulation generated a larger inorganic N pool. However, significant decreases in soil $NH_4^+$ from 2014 to 2015 (the relatively wet year), and increases in N leaching and higher deep soil $NO_3^-$ levels during 2015 in soils under exotic species when compared to soils under native species, suggest that soils under exotic plants did not retain the accumulated N. We speculate that the shorter phenology of the exotic annual species (Esch et al., 2019b) and reduced N uptake (demand) during water inputs increased its vulnerability to leaching losses. This is consistent with the hypothesis that drought-induced soil inorganic N accumulation is more vulnerable to losses upon rewetting (Evans and Burke, 2012).

Conclusions

In this experimental semi-arid shrubland, we found that plant functional traits are key factors in providing mechanistic explanations of soil responses and feedbacks to environmental disturbances. In comparison to soils under native woody plants, we observed that soils under exotic herbaceous plants had enhanced inorganic N accumulation and leaching losses after experiencing severe drought conditions. Our findings indicate that distinct morphological, physiological, and phenological traits of annual herbaceous species, associated with N higher inputs, faster turnover, and reduced temporal uptake, play a major role in N pools and fluxes. Nitrogen alterations under severe droughts and exotic plants likely produce a larger pool of N that is more vulnerable to increased leaching and may provide a competitive advantage to promote exotic growth and invasiveness success in these N-limited ecosystems (Liu et al., 2018; Vourlitis, 2017). While this study suggests N transformation rates as mechanisms of inorganic N accumulation, future efforts that measure gross N production are needed to confirm this speculation and better understand microbially mediated N-cycling processes.
Acknowledgments

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Table 1. Soil physicochemical properties (mean) in native- and exotic-dominated plots by rainfall treatment prior the beginning of the experiment.

<table>
<thead>
<tr>
<th>Plant community type</th>
<th>Native</th>
<th>Native</th>
<th>Native</th>
<th>Exotic</th>
<th>Exotic</th>
<th>Exotic</th>
</tr>
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<tbody>
<tr>
<td></td>
<td>50%</td>
<td>100%</td>
<td>150%</td>
<td>50%</td>
<td>100%</td>
<td>150%</td>
</tr>
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<td>Sand (g g(^{-1}))</td>
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<td>0.3</td>
<td>0.3</td>
<td>0.2</td>
<td>0.3</td>
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<tr>
<td>Silt (g g(^{-1}))</td>
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<td>0.5</td>
<td>0.6</td>
<td>0.6</td>
<td>0.5</td>
<td>0.6</td>
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<tr>
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<td>0.2</td>
<td>0.2</td>
<td>0.2</td>
<td>0.2</td>
<td>0.2</td>
</tr>
<tr>
<td>K (mg g(^{-1}))</td>
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<td>0.1</td>
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<td>0.2</td>
<td>0.2</td>
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</tr>
<tr>
<td>Ca (mg g(^{-1}))</td>
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<td>2.9</td>
<td>3.0</td>
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</tr>
<tr>
<td>Mg (mg g(^{-1}))</td>
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<td>1.3</td>
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<td>6.1</td>
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<tr>
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<td>0.1</td>
<td>0.1</td>
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Table 2. Summary of p-values for the linear mixed model for effects of rainfall treatments, plant community type, season, year and the interaction on soil N parameters (there were no significant 4-way interaction).

<table>
<thead>
<tr>
<th>Factor(s)</th>
<th>NH4</th>
<th>NO3</th>
<th>Net Min</th>
<th>Net Nit</th>
<th>Leaching</th>
</tr>
</thead>
<tbody>
<tr>
<td>Rainfall treatment</td>
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<td>0.0059*</td>
<td>0.2667</td>
<td>0.3516</td>
<td>0.2499</td>
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<td>0.1316</td>
<td>0.2969</td>
<td>&lt;0.0001*</td>
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<td>Season</td>
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<td>0.0010*</td>
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<td>N.A</td>
<td>N.A</td>
</tr>
<tr>
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<td>0.0002*</td>
<td>&lt;0.0001*</td>
<td>0.0265*</td>
<td>0.1548</td>
<td>&lt;0.0001*</td>
</tr>
<tr>
<td>Rain:Plant</td>
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<td><strong>0.0698</strong></td>
<td>0.3176</td>
<td>0.4442</td>
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<td>N.A</td>
<td>N.A</td>
<td>N.A</td>
</tr>
<tr>
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<td>N.A</td>
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<td>&lt;0.0001*</td>
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<td><strong>0.0029</strong>*</td>
<td>N.A</td>
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<td>Rain:Plant:Season</td>
<td><strong>0.0365</strong>*</td>
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<tr>
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<td>0.8729</td>
<td>0.5234</td>
</tr>
<tr>
<td>Rain:Season:Year</td>
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<td>0.5563</td>
<td>N.A</td>
<td>N.A</td>
<td>N.A</td>
</tr>
<tr>
<td>Plant:Season:Year</td>
<td>0.7303</td>
<td><strong>0.0523</strong></td>
<td>N.A</td>
<td>N.A</td>
<td>N.A</td>
</tr>
</tbody>
</table>
Figure Legends

**Figure 1.** Inter-annual variation of soil: (a) $NH_4^+$, (b) $NO_3^-$, (c) N mineralization, and (d) nitrification for each plant community type from 2013 to 2016 across rainfall treatments (no plant community type x rainfall treatment interaction). Yellow bars = exotic plots; green bars = native plots. Error bars represent the standard error of the mean.

**Figure 2.** Intra-annual variation of (a) $NH_4^+$ and (b) $NO_3^-$ for each plant community type from winter (Wi) to summer (Su). Yellow bars = exotic plots; green bars = native plots. Error bars represent the standard error of the mean.

**Figure 3.** Drought severity effects on (a) $NH_4^+$, (b) $NO_3^-$, (c) N mineralization, and (d) nitrification for each plant community type. Yellow bars = exotic plots; green bars = native plots. Error bars represent the standard error of the mean.

**Figure 4.** Inter-annual variation of leaching (DIN in resin bags) and deep soil $NO_3^-$ concentrations during the 2015 summer season. Error bars represent the standard error of the mean.
**Figure 1**

- **Figure 1a**: Plant type: $P = 0.0168$
  
  Year: $P = 0.0002$
  
  Plant x Year: $P = 0.1145$

- **Figure 1b**: Plant type: $P < 0.0001$
  
  Year: $P < 0.0001$
  
  Plant x Year: $P < 0.0001$

- **Figure 1c**: Plant type: $P = 0.1316$
  
  Year: $P = 0.0265$
  
  Plant x Year: $P = 0.2107$

- **Figure 1d**: Plant type: $P = 0.2969$
  
  Year: $P = 0.1548$
  
  Plant x Year: $P = 0.1816$
Figure 2

(a) NH$_4$ (µg N g$^{-1}$ soil)

- Plant type: P = 0.0168
- Season: 0.0006
- Plant x Season: P = 0.5998

(b) NO$_3$ (µg N g$^{-1}$ soil)

- Plant type: P < 0.0001
- Season: 0.0010
- Plant x Season: P = 0.7013

Wi, Sp, Su
**Figure 3**

(a) Plant type: $P = 0.0168$
Rain = 0.0013
Plant x Rain: $P = 0.1983$

(b) Plant type: $P < 0.0001$
Rain = 0.0059
Plant x Rain: $P = 0.2178$

(c) Plant type: $P = 0.1316$
Rain = 0.2667
Plant x Rain: $P = 0.0698$

(d) Plant type: $P = 0.2969$
Rain = 0.3516
Plant x Rain: $P = 0.3176$
Figure 4
Declaration of interests

☒ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

☒ The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: