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# Comprehensive Genomic and Transcriptomic Analysis of Polycyclic Aromatic Hydrocarbon Degradation by a Mycoremediation Fungus, Dentipellis sp. KUC8613

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#### 33 Abstract

The environmental accumulation of polycyclic aromatic 34 hydrocarbons (PAHs) is of great concern due to potential carcinogenic and 35 mutagenic risks, as well as their resistance to remediation. While many 36 fungi have been reported to break down PAHs in environments, the details 37 of gene-based metabolic pathways are not yet comprehensively 38 understood. Specifically, the genome-scale transcriptional responses of 39 fungal PAH-degradation have rarely been reported. In this study, we report 40 the genomic and transcriptomic basis of PAH-bioremediation by a potent 41 fungal degrader, *Dentipellis* sp. KUC8613. The genome size of this fungus 42 was 36.71 Mbp long encoding 14,320 putative protein-coding genes. The 43 strain efficiently removed more than 90 % of 100 mg/liter concentration of 44 PAHs within ten days. The genomic and transcriptomic analysis of this 45 46 white rot fungus highlights that the strain primarily utilized nonligninolytic enzymes to remove various PAHs, rather than typical 47 ligninolytic enzymes known for playing important roles in PAH-48 degradation. PAH-removal by non-ligninolytic enzymes was initiated by 49 both different PAH-specific and common overexpression of P450s, followed 50 by downstream PAH-transforming enzymes such as epoxide hydrolases, 51 dehydrogenases, FAD-dependent monooxygenases, dioxygenases and 52 glycosyl- or glutathione transferases. Among the various PAHs, 53 phenanthrene induced a more dynamic transcriptomic response possibly 54 due its greater cytotoxicity, leading to highly upregulated genes involved 55 in the translocation of PAHs, a defense system against reactive oxygen 56 species, and ATP synthesis. Our genomic and transcriptomic data provide 57 a foundation of understanding regarding the mycoremediation of PAHs 58 and the application of this strain for polluted environments. 59

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# 67 Keywords

- 68 PAH (Polycyclic aromatic hydrocarbon), Mycoremediation, *Dentipellis* sp.
- 69 KUC8613, White rot fungus, Genomics, Transcriptomics

#### 70 **INTRODUCTION**

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Fungi play a major role as the decomposers of recalcitrant organic 72 matters in nature (de Boer et al. 2005). The potent degrading abilities of 73 fungi can be attributed to low specificity of catabolic enzymes and the 74 75 formation of mycelial networks that can enhance chemical bioavailability (Harms, Schlosser and Wick 2011). Mycoremediation, which is based on 76 the use of fungi and mushrooms for the bioremediation of polluted areas, 77 78 is a promising method to remove many hazardous chemicals of environmental and public health concern (Deshmukh, Khardenavis and 79 Purohit 2016). 80

Polycyclic aromatic hydrocarbons (PAHs) are a group of chemicals 81 with two or more fused aromatic rings of carbon and hydrogen atoms 82 83 (Kadri et al. 2017). Among the various toxic materials, PAHs has been one of the major targets for mycoremediation due to their possible 84 carcinogenic and mutagenic risks (Mastrangelo, Fadda and Marzia 1996, 85 Schutzendubel et al. 1999). The high hydrophobicity and chemical stability 86 make them persistent in environments and cause bioaccumulation. While 87 PAHs can be naturally formed by forest fires or volcanic eruptions, the 88 incomplete combustion of organic materials during industrial and other 89 human activities account for the majority of PAH formation (Johansson and 90 van Bavel 2003). Several physical and chemical treatment methods 91 including incineration, UV oxidation, fixation, and solvent extraction have 92 93 been developed to remove PAHs, but they have many drawbacks since they are not cost-effective and environment-friendly (Gan, Lau and Ng 94 95 2009). The biodegradation of PAHs using fungi and other microorganisms 96 has been regarded as an alternative method to remove PAHs without causing significant ecological damages (Abe et al. 1995, Gran-Scheuch et 97 al. 2017). 98

It has been reported that many wood-degrading fungi can
efficiently degrade a wide variety of PAHs (Field et al. 1992). Both
ligninolytic and non-ligninolytic fungi can degrade PAHs with initial

oxidation of substrates but the metabolic pathways they use may differ 102 (Pozdnyakova 2012, Marco-Urrea, Garcia-Romera and Aranda 2015). 103 Ligninolytic fungi can produce extracellular ligninolytic enzymes including 104 manganese peroxidase (MnP), lignin peroxidase (LiP), and laccase to 105 produce quinone intermediates. Non-ligninolytic degradation of PAHs may 106 use intracellular cytochrome P450 monooxygenases (CYPs) to produce 107 unstable arene oxides which can be subsequently converted into phenols 108 or trans-dihydrodiols. It was previously reported that even some 109 ligninolytic fungi, including Phanerochaete and Pleurotus can use CYPs 110 instead of ligninolytic enzymes for PAH-degradation (Ghosal et al. 2016). 111 112

Although great efforts have been made to demonstrate PAH degradation in wood-degrading fungi through proteomic and metabolic
 assays, whole genome and transcriptome studies have rarely been
 conducted. A genome scale survey of potential PAH-responsive genes and
 their regulation at transcriptional level can contribute to the
 comprehensive understanding of complex genetic networks in a fungal
 PAH-degradation system and identification of novel PAH-responsive genes
 encoded in the genome.

120 Dentipellis sp. KUC8613, one of white rot fungi in the order Russulales, has previously been screened by high tolerance to various 121 PAHs and regarded as a potential host for mycoremediation (Lee et al. 122 2014). In this study, we report the efficient PAH-removal by the KUC8613 123 strain over various PAHs. In order to expand our knowledge of the genetic 124 125 basis for the PAH-removal capability of this fungus, we sequenced and analyzed the genome of *Dentipellis* sp. KUC8613 as part of the 1000 126 Fungal Genomes Project (1k FGP) (Grigoriev et al. 2014) at the US 127 Department of Energy Joint Genome Institute (JGI) 128 (http://jgi.doe.gov/fungi). From the genome, we investigated genetic 129 130 repertoires potentially involved in the metabolism of PAHs. We further performed transcriptomic analysis to observe differential expression of 131 these genes by four different PAHs and to identify novel PAH-responsive 132

133 genes. Our genomic and transcriptomic analysis to elucidate PAH-removal

- 134 by KUC8613 will give us a more comprehensive understanding of PAH-
- 135 degradation by wood-degrading fungi.

#### 138 Materials and Methods

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#### 140 Fungal strains, chemicals, and media

*Dentipellis* sp. KUC8613 was originally isolated from South Korea and identified by the Korea University Culture collection (KUC). Anthracene (ANT), fluoranthene (FLU), phenanthrene (PHE), pyrene (PYR), and all solvents including ethyl acetate, chloroform, and acetonitrile were purchased from Sigma (Sigma-Aldrich). Malt extract (ME) medium contained 20 g malt extract for 1 liter distilled water. 15 g agar was additionally added for malt extract agar (MEA) medium.

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#### 149 **Determination of tolerance to PAHs**

Four different PAHs (ANT, FLU, PHE, and PYR) were used to 150 151 determine tolerance of the fungus to PAHs. Fungal mycelia were cultured on MEA containing 100 mg/liter of individual PAH. PAHs were first 152 dissolved in acetone prior to being added into the culture medium. 153 Acetone was evaporated from the medium by letting it stand for a few 154 days. An actively growing fungal disk was inoculated at the center of a 155 culture plate and incubated at 25 °C. Each sample was prepared in 156 triplicate and the growth rate of fungal mycelia was determined by 157 measuring the radius of the mycelia at between 0 and 10 days. 158

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#### 160 Liquid culture condition and determination of mycelial growth

Fungal mycelium was grown on MEA at 25 °C and maintained for 161 less than a week at 5 °C before use. Liquid ME containing 100 mg/liter of 162 individual PAH was prepared. PAHs were first dissolved in acetone and 163 164 added into ME medium. Acetone was evaporated from the medium by letting it stand for a few days before fungal inoculation. An actively 165 growing mycelium disc (4 cm diameter) was homogenized for 10 seconds 166 in a sterile blender cup, containing 50 ml of ME medium. 2 ml of the 167 homogenate was used to inoculate 50 ml ME medium in a 250 ml 168 Erlenmeyer flask. Mycelium was grown at 25 °C in a rotary shaker with 169

shaking at 180 rpm. Heat-killed mycelia (400 mg in dry weight) and
cultures without fungal inoculation were also used for the determination of
PAH-adsorption to mycelia and natural degradation. Cultures were
incubated for between 0 and 10 days.

For the determination of mycelial growth during liquid culture, mycelium was harvested after incubation by centrifugation and washed twice with 50 ml distilled water to a final volume of 10 ml. The sample was then and filtered through pre-weighted dried filter paper filter in a vacuum filtration apparatus. The mycelium and filter paper were oven-dried at 65 °C for one day, and dry weight was measured.

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#### 181 HPLC analysis of PAH-removal

The dried extracts were dissolved in 1 ml chloroform. For the 182 quantitative analysis of residual PAHs, Waters high-pressure liquid 183 chromatography (HPLC) system (Milford, MA, USA) equipped with Waters 184 2487 UV detector was used. Data acquisition was carried out using the 185 Waters Empower 2 software. Ten microliters of each extract was injected 186 onto the Waters C18 column with a flow rate of 1 ml/min. The mobile 187 188 phase was acetonitrile : water (80:20) and the PAHs were detected at 254 nm. 189

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#### 191 Genomic DNA extraction and sequencing

The fungus was cultured on solid MEA media at 25 °C in the dark. 192 Genomic DNA was extracted from mycelium using a CTAB-based fungal 193 DNA isolation protocol (Fulton, Chunwongse and Tanksley 1995). The 194 concentration of prepared DNA was determined using Qubit fluorometer 195 (Invitrogen). For the whole genome sequencing of Dentipellis sp. 196 KUC8613, two Illumina libraries with insert sizes 5.5 kb and 370 bp were 197 prepared and sequenced using  $2 \times 150$  bp reads from HiSeq-1TB. The 198 sequence data was filtered so as to remove low quality reads and 199 subsequently assembled with AllPathsLG release version R44008 (Gnerre 200 et al. 2011). The genome was annotated using the IGI Annotation Pipeline 201

202 (Grigoriev et al. 2014) and FunGAP (Min, Grigoriev and Choi 2017). From
203 JGI-MycoCosm repository
204 (https://genome.jgi.doe.gov/Densp1/Densp1.home.html), we could
205 retrieve the genome wide annotation data such as InterPro, CAZy, KOG,
206 and KEGG for functional analysis.

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#### 208 RNA extraction and mRNA sequencing

For transcriptomic analysis, RNA from 5 days of culture were used. 209 The grown mycelium was harvested from the culture medium by 210 centrifugation and washed twice with 50 ml distilled water. The mycelium 211 212 pellet was immediately frozen under liquid nitrogen and homogenized using a mortar and a pestle. Total RNA was isolated using an RNeasy plus 213 mini kit (Qiagen, the Netherlands) according to the manufacturer's 214 protocol. RNA concentration was determined using Qubit fluorometer 215 (Invitrogen). RNA library was constructed using TruSeg RNA sample 216 preparation kit (Illumina, San Diego, CA, US). The sequencing of the RNA 217 library was carried out in MiSeq ( $2 \times 200$  cycles). Low-quality reads were 218 219 filtered using the Trim Galore program. The filtered RNA-seg data was 220 mapped against the assembled genome using CLC Genomics Workbench software v.12.0 (Qiagen, the Netherlands). Read alignment was performed 221 with the following parameters: minimum length fraction and minimum 222 similarity fraction = 0.8, strand specific = both, maximum number of hits 223 for a read = 3. Reads Per Kilobase of transcript per Million mapped reads 224 (RPKM) value was generated using default settings of CLC genomics and 225 further used for the analysis of differential expression of genes. To check if 226 the RNA-seq data from different conditions correlate each other, a 227 Principal Component Analysis (PCA) was performed by an internal routine 228 of CLC Genomics Workbench. 229

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#### **Gene Ontology enrichment analysis of upregulated genes**

Upregulated genes by PAHs were identified according to the following criteria: fold change > 2 and p-value < 0.01. To further reveal

- the enriched functions among upregulated genes, the Gene Ontology (GO)
  enrichment analyses of upregulated genes were performed using the R
  package clusterProfiler (version 3.2.1) (Yu et al. 2012). The threshold for
  the enrichment analysis was p-value < 0.01.</li>

#### 239 Results

240

#### 241 **PAH tolerance and removal capability by** *Dentipellis* **sp. KUC8613**

We measured the PAH tolerance of the fungus according to the 242 radial growth of mycelia in solid media (Fig. 1A). ANT, FLU, PHE, and PYR 243 were used as the representative PAHs having two to four fused aromatic 244 rings. The average growth rate of the fungus at 25 °C was 4.7 mm/day 245 without PAH. When four different PAHs were individually added in the 246 growth media, KUC8613 showed high tolerance to all four PAHs with 247 different degrees. In more detail, no detectable growth inhibition by ANT, 248 249 FLU, and PYR was observed at any time point within ten days. A moderate decrease (14 %) in growth was observed only in PHE-added media. No 250 significant change in mycelial morphology or pigmentation by any of PAH 251 was recognized (Data not shown). 252

In order to verify PAH-removal by this fungus, the removal of PAHs 253 254 in liquid culture was examined (Fig. 2). HPLC-mediated guantification showed that 44 % (ANT), 49 % (FLU), 33 % (PHE), and 46 % (PYR) of PAHs 255 were removed after five days. The removal of 90.1 % (ANT), 99 % (FLU), 256 94.3 % (PHE), and 94.4 % (PYR) of PAHs were observed after 10 days, 257 while 17.4 % (ANT), 25 % (FLU), 24.4 % (PHE), and 27.7 % (PYR) were 258 removed by autonomous chemical decomposition. We also observed that 259 less than 10 % of PAHs could be removed by adsorption to heat killed 260 mycelia. Fungal biomass (dry weight) in liquid culture was also measured 261 (Fig. S1). Only PHE induced a moderate growth retardation, which was 262 congruent with our observation in PAH tolerance test at solid media. 263

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#### 265 Genome properties of Dentipellis sp. KUC8613

In order to elucidate genetic contents for PAH-removal in this fungus, whole genome sequencing was carried out under the 1k FGP program. The genome sequence was 36.71 Mbp long, comprising 1,184 contigs and 425 scaffolds (Table 1). The GC content of the genome was 55.4 %. In total, 14,320 protein-coding genes were predicted in the 271 genome with an average gene length of 1,737 bp. Both the genome size 272 and the number of gene models of KUC8613 were typical among the 273 compared white rot fungi available from JGI mycocosm (Grigoriev et al. 274 2014). Among 14,320 gene models, 6,305 (44 %) and 7,538 (52.6 %) were 275 assigned to different GO (Gene Ontology) terms and KOG (euKaryotic 276 Ortholog Groups) classes, respectively.

A genomic repository of carbohydrate-active enzymes (CAZymes) 277 was first surveyed in order to analyze the pattern of carbohydrate 278 metabolism in this fungus. The genome harbored a total of 417 CAZymes 279 including of 71 auxiliary activities (AA), 49 carbohydrate-binding modules 280 281 (CBM), and 180 glycoside hydrolases (GH) (Table S1). The cellulolytic CAZyme composition in typical white rot fungi is generally represented by 282 the presence of lytic polysaccharide monooxygenases (LMPOs) (AA9), 283 cellobiohydrolases (GH6 and GH7), a single cellobiose dehydrogenase 284 (AA3-1), frequent CBM1-containing proteins, and ligninolytic enzymes 285 (Riley et al. 2014). Indeed, this strain revealed 11 copies of LMPOs that 286 potentially carry out the oxidative cleavage of polysaccharide chains 287 (Table S2). It also contained four cellobiohydrolases, 22 genes encoding 288 289 proteins carrying a CBM1 family module, and a single gene encoding cellobiose dehydrogenase. The CAZyme-pattern suggests that *Dentipellis* 290 sp. KUC8613 has the classical wood decay mode of previously well-291 characterized white rot fungi. 292

The genome of KUC8613 was annotated with 11 genes encoding ligninolytic enzymes, including seven fungal class II peroxidases (PODs; AA2) and four laccases (AA1\_1). PODs were further classified into six MnPs and a single LiP by sequence comparison. When we compared the KUC8613 genome with eight other white rot fungal genomes, KUC8613 showed less than the average number of both PODs and laccase genes among the compared genomes (Fig. 3).

While only a limited number of genes encoding ligninolytic enzymes were found in the genome of KUC8613, we could instead identify other genes encoding non-ligninolytic type of enzymes that can potentially

mediate the initial oxidation of aromatic ring structures. The presence of 154 putative P450s were revealed from the fungal genome (Table S3). Using the Kyoto Encyclopedia of Genes and Genomes (KEGG) database analysis, we could further sort out 77 P450s potentially involved in the metabolism of xenobiotics, specifically those from CYP1, CYP2, and CYP3 families (Lewis 2003).

Additional genes that are potentially involved in the downstream 309 steps for PAH-transformation were also investigated (Table S4). We 310 identified 19 putative epoxide hydrolases that might catalyze a reaction to 311 produce trans-dihydrodiols from arene oxides. Oxidoreductase enzymes 312 313 including 16 alcohol dehydrogenases, 17 aldehyde dehydrogenases, one trans-1,2-dihydrobenzene-1,2-diol dehydrogenase, and 31 FAD-dependent 314 monooxygenases may catalyze a series of oxidation reactions to produce 315 metabolic intermediates. Multicomponent dioxygenases comprising eight 316 ferredoxin reductases, 317 ferredoxins. four two hydroxyquinol 1.2dioxygenases, and two aromatic ring-opening dioxygenases were also 318 found in the fungal genome. Aromatic intermediates formed by above 319 enzymes then can act as substrates for additional ring cleavage or 320 321 conjugation steps (Casillas et al. 1996, Habe and Omori 2003). Glutathione S-transferases (GSTs), sulfotransferases, or 322 glycosyltransferases (GTFs) can catalyze the addition of glutathione, 323 sulfate, or glycosyl donors to PAH molecules, respectively, thus making 324 them less toxic and more water-soluble. In the genome of KUC8613, ten 325 putative GSTs, two sulfotransferases, and 13 GTFs were encoded for the 326 potential conjugation of PAH intermediates. 327

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#### 329 Genome-wide transcriptomic responses during PAH removal

In order to further investigate which of the potential genes are actually upregulated during PAH-removal in this fungus, a genome-wide transcriptomic analysis was performed. Mycelial discs grown for five days in liquid ME media supplemented with ANT, FLU, PHE, and PYR (100 mg/liter) were subjected to RNA-seq. Principal component analysis (PCA) of the gene expression data clearly separated PAH-added groups from the control group (Fig. 4A). The PCA data also showed a separation within PAH samples. While the ANT, FLU, and PYR samples were clustered together, the PHE samples formed an independent group distinct from other PAH samples, suggesting that the gene expression profile of PHE sample is unlike those of other PAHs.

Of the 14,320 gene models predicted in the genome, we identified 341 1,922 genes whose expression was upregulated by at least one of four 342 different PAHs (Fold change > 2 and P-value < 0.01) (Table S5). The 343 numbers of total genes upregulated by ANT, FLU, PHE, and PYR were 915, 344 345 950, 1539, and 1,095, respectively (Fig. 4B). Among the upregulated genes, the molecular function of 772 genes could not be predicted. 346 Upregulated genes were assigned to 23 specific KOG classes and 347 dominantly (54.9 % of 1,041 KOG-assigned genes) categorized into five 348 KOG classes including posttranslational modification, signal transduction, 349 eneray production and conversion, carbohydrate transport and 350 metabolism, and transcription (Table S6). As illustrated by a Venn-diagram 351 352 in Figure 4B, upregulated genes showed common and different PAH-353 specific overexpression during PAH-removal. A total of 550 genes were commonly overexpressed by all four PAHs. The numbers of PAH-specific 354 genes by ANT, FLU, PHE, and PYR were 67, 72, 548, and 59, respectively. 355 The number of PHE-induced genes were significantly higher than genes 356 induced by other PAHs. A heatmap plotted by log<sub>2</sub>RPKM of up-regulated 357 gene expression showed a similar pattern as well (Fig. 4C). The majority of 358 up-regulated genes were commonly overexpressed in all PAH-samples 359 while PHE showed a distinct gene expression pattern compared to those of 360 other three PAHs. 361

We first investigated the expression profiles of ligninolytic enzymes (LiP, MnP, and laccase) and P450s for initial aromatic-ring oxidation. No ligninolytic enzymes were upregulated during PAH-removal (Table S7). On the other hand, P450s showed both common and different PAH-specific overexpression. Among 154 putative P450 genes found in the KUC8613

genome, transcription of 15 genes was induced by one or more PAHs
(Table 2). Six P450 genes were constitutively upregulated by two or three
PAHs, while the remaining nine genes showed different PAH-specific
overexpression. These nine genes could be further categorized into six
PHE-inducible (ProtID 548471, 772453, 879478, 840982, 841965, and
873894) and one for each ANT-, FLU- or PYR-inducible genes (ProtID
861037, 798081, and 832704 for ANT, FLU, and PYR, respectively).

We also identified a total of 27 other PAH-responsive genes 374 potentially responsible for the downstream steps for PAH-transformation. 375 The expression patterns of these genes showed that most of them were 376 377 commonly overexpressed regardless of PAH-type (Fig. S2). These 27 PAHresponsive genes were three epoxide hydrolases (ProtID 581771, 840197, 378 and 871191), three ferredoxins (ProtID 226033, 839077, and 887248), one 379 ferredoxin reductase (ProtID 832353), five alcohol dehydrogenases (ProtID 380 843873, 859290, 867273, 712202, and 886660), three aldehyde 381 dehydrogenases (ProtID 837250, 884278, and 857188), Five FAD-382 dependent monooxygenases (ProtID 820609, 772612, 838248, 870155, 383 and 879414), two dioxygenases (ProtID 841503 and 842583), three GSTs 384 385 (ProtID 847321, 697729, and 830936), one GTF (ProtID 879011), and one sulfotransferase (ProtID 838043). 386

In addition to the PAH-transforming genes, the 550 commonly 387 overexpressed genes by all four PAHs included 41 genes involved in lipid 388 transport and metabolism; these include a lipase (ProtID 835364), a sterol 389 desaturase (ProtID 864816), a methyltransferase (ProtID 833162), and a 390 phospholipid/glycerol acyltransferase (ProtID 39743) (Table S8). This 391 observation is consistent with previous studies in which the bioavailability 392 of hydrophobic PAHs can be enhanced by the cellular production of lipid 393 biosurfactants to emulsify PAHs and promote solubility (Cao et al. 2015). 394 395 Since our genomic and transcriptomic analysis suggests that this

fungus utilizes the intracellular non-ligninolytic type of enzymes instead of
 extracellular ligninolytic enzymes for PAH-transformation, translocation of
 PAHs and metabolic intermediates across the cell membranes and

organelles should be controlled by membrane bound transporters. It is 399 well known that major facilitator superfamily (MFS) transporters and ATP-400 binding cassette (ABC) transporters are two major transporter families 401 that mediate the import and export of drugs and xenobiotics (leong et al. 402 2017, Carmona et al. 2009). We observed the common upregulation of 14 403 MFS and five ABC transporters by all four PAHs (Table S9). The active 404 transcription of MFS transporters was particularly noticeable by PHE. The 405 more than ten-fold elevated expression of four MFS transporters (ProtID 406 787334, 857773, 800757, and 328771) by PHE suggests an important role 407 of these transporters for cellular response to PHE. 408

409 While a significant number of genes were commonly overexpressed by all four PAHs, certain genes also showed PAH-specificity within different 410 PAH-samples. Among four PAHs, PHE induced 35 % more genes than the 411 average number of upregulated genes by other PAHs. This resulted in the 412 PHE-specific overexpression of 548 genes (Fig. 4B). According to Gene 413 Ontology (GO) functional enrichment analysis, enriched GO terms among 414 PHE-specific genes were ATP synthesis coupled proton transport and 415 different types of cytochrome c oxidases (Fig S3). We also identified 10 416 417 additional MFS and two ABC transporters, suggesting their restricted involvement in the translocation of PAHs other than PHE (Table S10). PAHs 418 are known to increase the production of reactive oxygen species (ROS), 419 leading to oxidative stress (Alkio et al. 2005). Since only PHE showed 420 detectable cytotoxicity during fungal growth, significant upregulation of 421 422 ROS scavenging enzymes is expected. Indeed, we observed the PHEspecific upregulation of genes such as manganese and iron superoxide 423 dismutase (ProtID 648712, and 860732), and alkyl hydroperoxide 424 reductase (ProtID 194063), which are potentially involved in enzymatic 425 antioxidant defense mechanisms. 426

#### 428 **Discussion**

PAH-transformation by wood-degrading fungi might be a ubiguitous 429 phenomenon (Field et al. 1992, Mao and Guan 2016). In this study, we 430 demonstrated that *Dentipellis* sp. KUC8613 have a high capability (> 90 % 431 of 100 mg/L PAH) of removing four different types of PAHs within a short 432 period of time. Our genomic and transcriptomic analysis to further 433 elucidate PAH-transforming system encoded in the genome revealed a 434 total of 1,922 upregulated genes many of which have unknown molecular 435 functions. To our knowledge, genome-scale transcriptomic responses by a 436 wood-degrading fungus during PAH-transformation have not been 437 previously described. 438

While ligninolytic system in white rot fungi was often described as 439 the key for PAH-transformation (Ghosal et al. 2016), our data showed that 440 441 KUC8613 was capable of removing PAHs using P450s instead of ligninolytic enzymes. Ligninolytic enzymes may not be essential for PAH-442 transformation in this fungi, but they still may take part in PAH-443 transformation at certain growth conditions. It has been previously shown 444 that the production of ligninolytic enzymes by white rot fungi can be 445 dependent on carbon and nitrogen concentration in the growth media 446 (Ben Hamman, de La Rubia and Martinez 1997, D'Souza, Merrit and Reddy 447 1999). It would be worthwhile to check PAH-removal pattern by this 448 fungus under different conditions that favor the production of ligninolytic 449 enzymes. Taken together, these data strongly suggest that the screening 450 451 of strong mycoremdiation hosts primarily based on ligninolytic enzyme activities should be reconsidered. 452

Our transcriptomic analysis revealed that P450s might be responsible for the initial oxidation of aromatic rings in this fungus. A previous genome-scale identification of P450 genes in *Phanerochaete chrysosporium* showed PAH-oxidizing activity of these enzymes with varying PAH specificity (Syed et al. 2010). Similarly, upregulated P450s in this fungus showed both PAH-specific and common overexpression patterns. PHE induced the highest number of P450 genes, suggesting

metabolic complexity of this compound compared to other three PAHs. In 460 addition to P450s, we observed upregulation of many other genes 461 encoding potential PAH-transforming activities such as epoxide hydrolase, 462 aldehyde 463 alcohol dehydrogenase, dehydrogenase, FAD-dependent monooxygenase, dioxygenase, GTF, GST and sulfotransferase. The 464 biotransformation of PAHs by dioxygenases is mainly known to occur in 465 many PAH-degrading bacteria (Peng et al. 2008), but in some Trichoderma 466 involvement of dioxygenase species, was also reported during 467 phenanthrene degradation (Hadibarata, Tachibana and Itoh 2007). The 468 shared upregulation of many potential PAH-transforming genes suggests 469 470 some overlap between the metabolic pathways of different PAHs. The elucidation of the exact molecular mechanisms associated with the key 471 enzymes discovered in this study requires further study. In addition, 472 upregulation of many genes with unknown molecular functions also gives 473 a clue for future research into the identification of novel genes involved in 474 PAH-transformation. 475

The distinct transcriptomic response induced by PHE in this fungus 476 was explained by a large number of PHE-specific genes involved in 477 478 translocation of PAHs, defense against ROS, and ATP synthesis. The reason why PHE induced the transcription of a significantly larger number of 479 genes is not clear but out experimental evidences suggested that it might 480 be associated with PHE-induced cellular cytotoxicity. A similar observation 481 that PHE upregulated a larger number of genes than other PAHs was also 482 previously reported in soil fungus (Gao et al. 2019). 483

Based on the global gene expression pattern, we constructed a 484 tentative pathway for the transformation and detoxification of PAHs in 485 Dentipellis sp. KUC8613 (Fig. 5). The pathway includes the initial ring-486 oxidation step followed by downstream transformation step as well as 487 transporters for the translocation of PAHs and their metabolites, lipases 488 for the enhanced solubility of PAHs, and antioxidant enzymes for reduced 489 oxidative stress. Our data based on whole-genomic and transcriptomic 490 analysis will provide a strong insight into the complex gene regulation of 491

- 492 wood-degrading fungi for PAH-degradation, and may serve as a guide for
- the efficient screening and utilization of fungal hosts for mycoremediation.

# 495 Accession numbers and availability of the genome sequence and 496 the strain

This whole genome sequence of *Dentipellis* sp. KUC8613 has been 497 DDBJ/EMBL/GenBank under the 498 deposited at accession number NSIX00000000. In addition, the genome assembly and annotation are 499 available at DOE JGI Genome Portal MycoCosm (Grigoriev et al. 2014) 500 (http://genome.jgi.doe.gov/Densp1). The strain is available from Korean 501 Collection for Type Cultures (KCTC) with the accession number KCTC46678 502 (http://kctc.kribb.re.kr/) 503

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#### 518 **References**

- Abe, A., A. Inoue, R. Usami, K. Moriya & K. Horikoshi (1995) Degradation of
   polyaromatic hydrocarbons by organic solvent-tolerant bacteria from
   deep sea. *Biosci Biotechnol Biochem*, 59, 1154-6.
- 522 Alkio, M., T. M. Tabuchi, X. C. Wang & A. Colon-Carmona (2005) Stress 523 responses to polycyclic aromatic hydrocarbons in Arabidopsis 524 include growth inhibition and hypersensitive response-like 525 symptoms. *Journal of Experimental Botany*, 56, 2983-2994.
- Ben Hamman, O., T. de La Rubia & J. Martinez (1997) Effect of carbon and
  nitrogen limitation on lignin peroxidase and manganese peroxidase
  production by Phanerochaete flavido-alba. *Journal of Applied Microbiology*, 83, 751-757.
- Cao, J., Q. Lai, J. Yuan & Z. Shao (2015) Genomic and metabolic analysis of
  fluoranthene degradation pathway in Celeribacter indicus P73T. *Sci Rep*, 5, 7741.
- Carmona, M., M. T. Zamarro, B. Blazquez, G. Durante-Rodriguez, J. F.
  Juarez, J. A. Valderrama, M. J. Barragan, J. L. Garcia & E. Diaz (2009)
  Anaerobic catabolism of aromatic compounds: a genetic and
  genomic view. *Microbiol Mol Biol Rev*, 73, 71-133.
- Casillas, R. P., S. A. Crow, T. M. Heinze, J. Deck & C. E. Cerniglia (1996)
  Initial oxidative and subsequent conjugative metabolites produced
  during the metabolism of phenanthrene by fungi. *Journal of Industrial Microbiology*, 16, 205-215.
- D'Souza, T. M., C. S. Merrit & C. A. Reddy (1999) Lignin-modifying enzymes
   of the white rot basidiomycete Ganoderma lucidum. *Applied and Environmental Microbiology*, 65, 5307-5313.
- de Boer, W., L. B. Folman, R. C. Summerbell & L. Boddy (2005) Living in a
  fungal world: impact of fungi on soil bacterial niche development. *Fems Microbiology Reviews*, 29, 795-811.
- Deshmukh, R., A. A. Khardenavis & H. J. Purohit (2016) Diverse Metabolic
  Capacities of Fungi for Bioremediation. *Indian J Microbiol*, 56, 24764.

- Field, J. A., E. de Jong, G. Feijoo Costa & J. A. de Bont (1992)
  Biodegradation of polycyclic aromatic hydrocarbons by new isolates
  of white rot fungi. *Appl Environ Microbiol*, 58, 2219-26.
- Fulton, T. M., J. Chunwongse & S. D. Tanksley (1995) Microprep Protocol for
   Extraction of DNA from Tomato and Other Herbaceous Plants. *Plant Molecular Biology Reporter*, 13, 207-209.
- Gan, S., E. V. Lau & H. K. Ng (2009) Remediation of soils contaminated
  with polycyclic aromatic hydrocarbons (PAHs). *Journal of Hazardous Materials*, 172, 532-549.
- Gao, R., D. C. Hao, W. L. Hu, S. Song, S. Y. Li & G. B. Ge (2019)
  Transcriptome profile of polycyclic aromatic hydrocarbon-degrading
  fungi isolated from Taxus rhizosphere. *Current Science*, 116, 12181228.
- 563 Ghosal, D., S. Ghosh, T. K. Dutta & Y. Ahn (2016) Current State of 564 Knowledge in Microbial Degradation of Polycyclic Aromatic 565 Hydrocarbons (PAHs): A Review. *Front Microbiol*, **7**, 1369.
- Gnerre, S., I. MacCallum, D. Przybylski, F. J. Ribeiro, J. N. Burton, B. J.
  Walker, T. Sharpe, G. Hall, T. P. Shea, S. Sykes, A. M. Berlin, D. Aird,
  M. Costello, R. Daza, L. Williams, R. Nicol, A. Gnirke, C. Nusbaum, E.
  S. Lander & D. B. Jaffe (2011) High-quality draft assemblies of
  mammalian genomes from massively parallel sequence data. *Proceedings of the National Academy of Sciences of the United*States of America, 108, 1513-1518.
- Gran-Scheuch, A., E. Fuentes, D. M. Bravo, J. C. Jimenez & J. M. PerezDonoso (2017) Isolation and Characterization of Phenanthrene
  Degrading Bacteria from Diesel Fuel-Contaminated Antarctic Soils. *Front Microbiol*, 8, 1634.
- Grigoriev, I. V., R. Nikitin, S. Haridas, A. Kuo, R. Ohm, R. Otillar, R. Riley, A.
  Salamov, X. Zhao, F. Korzeniewski, T. Smirnova, H. Nordberg, I.
  Dubchak & I. Shabalov (2014) MycoCosm portal: gearing up for
  1000 fungal genomes. *Nucleic Acids Res*, 42, D699-704.
- 581 Habe, H. & T. Omori (2003) Genetics of polycyclic aromatic hydrocarbon
  - 23

- 582 metabolism in diverse aerobic bacteria. *Biosci Biotechnol Biochem*,
  583 67, 225-43.
- Hadibarata, T., S. Tachibana & K. Itoh (2007) Biodegradation of
  phenanthrene by fungi screened from nature. *Pak J Biol Sci*, 10,
  2535-43.
- Harms, H., D. Schlosser & L. Y. Wick (2011) Untapped potential: exploiting
  fungi in bioremediation of hazardous chemicals. *Nat Rev Microbiol*,
  9, 177-92.
- Jeong, C. B., D. H. Kim, H. M. Kang, Y. H. Lee, H. S. Kim, I. C. Kim & J. S. Lee
  (2017) Genome-wide identification of ATP-binding cassette (ABC)
  transporters and their roles in response to polycyclic aromatic
  hydrocarbons (PAHs) in the copepod Paracyclopina nana. *Aquat Toxicol*, 183, 144-155.
- Johansson, I. & B. van Bavel (2003) Levels and patterns of polycyclic
  aromatic hydrocarbons in incineration ashes. *Sci Total Environ*, 311,
  221-31.
- Kadri, T., T. Rouissi, S. Kaur Brar, M. Cledon, S. Sarma & M. Verma (2017)
  Biodegradation of polycyclic aromatic hydrocarbons (PAHs) by
  fungal enzymes: A review. *J Environ Sci (China)*, 51, 52-74.
- Lee, H., Y. Jang, Y. S. Choi, M. J. Kim, J. Lee, H. Lee, J. H. Hong, Y. M. Lee, G.
  H. Kim & J. J. Kim (2014) Biotechnological procedures to select white
  rot fungi for the degradation of PAHs. *J Microbiol Methods*, 97, 56-62.
- Lewis, D. F. V. (2003) Human cytochromes P450 associated with the phase 1 metabolism of drugs and other xenobioties: A compilation of substrates and inhibitors of the CYP1, CYP2 and CYP3 families. *Current Medicinal Chemistry*, 10, 1955-1972.
- Mao, J. & W. W. Guan (2016) Fungal degradation of polycyclic aromatic
  hydrocarbons (PAHs) by Scopulariopsis brevicaulis and its
  application in bioremediation of PAH-contaminated soil. Acta
  Agriculturae Scandinavica Section B-Soil and Plant Science, 66, 399405.
- 613 Marco-Urrea, E., I. Garcia-Romera & E. Aranda (2015) Potential of non-
  - 24

- 614 ligninolytic fungi in bioremediation of chlorinated and polycyclic 615 aromatic hydrocarbons. *New Biotechnology*, 32, 620-628.
- Mastrangelo, G., E. Fadda & V. Marzia (1996) Polycyclic aromatic
   hydrocarbons and cancer in man. *Environmental Health Perspectives*, 104, 1166-1170.
- Min, B., I. V. Grigoriev & I. G. Choi (2017) FunGAP: Fungal Genome
  Annotation Pipeline using evidence-based gene model evaluation. *Bioinformatics*, 33, 2936-2937.
- Peng, R. H., A. S. Xiong, Y. Xue, X. Y. Fu, F. Gao, W. Zhao, Y. S. Tian & Q. H.
  Yao (2008) Microbial biodegradation of polyaromatic hydrocarbons. *FEMS Microbiol Rev*, 32, 927-55.
- Pozdnyakova, N. N. (2012) Involvement of the ligninolytic system of white rot and litter-decomposing fungi in the degradation of polycyclic
   aromatic hydrocarbons. *Biotechnol Res Int*, 2012, 243217.
- Riley, R., A. A. Salamov, D. W. Brown, L. G. Nagy, D. Floudas, B. W. Held, A. 628 Levasseur, V. Lombard, E. Morin, R. Otillar, E. A. Lindquist, H. Sun, K. 629 M. LaButti, J. Schmutz, D. Jabbour, H. Luo, S. E. Baker, A. G. 630 Pisabarro, J. D. Walton, R. A. Blanchette, B. Henrissat, F. Martin, D. 631 632 Cullen, D. S. Hibbett & I. V. Grigoriev (2014) Extensive sampling of basidiomycete genomes demonstrates inadequacy of the white-rot/ 633 brown-rot paradigm for wood decay fungi. Proc Natl Acad Sci U S A, 634 111, 9923-8. 635
- Schutzendubel, A., A. Majcherczyk, C. Johannes & A. Huttermann (1999) 636 Degradation of fluorene, anthracene, phenanthrene, fluoranthene, 637 and pyrene lacks connection to the production of extracellular 638 by Pleurotus ostreatus and Bjerkandera adusta. enzymes 639 International Biodeterioration & Biodegradation, 43, 93-100. 640
- Syed, K., H. Doddapaneni, V. Subramanian, Y. W. Lam & J. S. Yadav (2010)
  Genome-to-function characterization of novel fungal P450
  monooxygenases oxidizing polycyclic aromatic hydrocarbons (PAHs). *Biochemical and Biophysical Research Communications*, 399, 492497.

Yu, G. C., L. G. Wang, Y. Y. Han & Q. Y. He (2012) clusterProfiler: an R
 Package for Comparing Biological Themes Among Gene Clusters.
 *Omics-a Journal of Integrative Biology*, 16, 284-287.

#### 650 TABLES AND FIGURES LEGENDS

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Table 1. Genome features of *Dentipellis* sp. KUC8613 compared with other white rot fungi

Table 2. Common and different PAH-specific upregulation of P450 genes during PAH-removal

Figure 1. Radial growth of the fungal mycelia at the presence of different PAHs. ANT, anthracene; FLU, fluoranthene; PHE, phenanthrene; PYR, pyrene.

Figure 2. PAH-removal efficiency by *Dentipellis sp. KUC8613*. The residual amount of PAH during the fungal culture was measured by HPLC. ANT, anthracene; FLU, fluoranthene; PHE, phenanthrene; PYR, pyrene; Control 1, 10 days of incubation without fungal inoculation; Control 2, 10 days of incubation with dead mycelia.

Figure 3. Comparison of relative tolerance to mixed PAHs (Lee et al., 2014) and the number of ligninolytic enzymes in white rot fungi. The species tree was built based on single-copy orthologs including *A. nidulans* as an out-group. The bootstrap-based branch supports and the scale bar that represents the mean number of amino acid substitutions per site are shown. Gene numbers are shaded white or red based on the Z-scores computed for each column.

Figure 4. Transcriptomic analysis during removal of four different PAHs. (A) Principal component analysis (PCA) of transcriptomics data. (B) Venndiagram representing the number of upregulated genes by four different PAHs. (C) Heatmap showing transcription pattern (log<sub>2</sub>RPKM) of 1,922 upregulated genes by four different PAHs. Columns represent different genes and rows represent different PAH-treated groups. ANT, anthracene; FLU, fluoranthene; PHE, phenanthrene; PYR, pyrene.

Figure 5. A schematic view of PAH-transformation in *Dentipellis* sp.
 KUC8613. Blue and Red arrows represent the common and PAH-specific 27

overexpression of PAH-transforming genes, respectively. The PAHtransformation pathway was divided into 1) initial ring-oxidation, and 2)
downstream transformation steps based on key enzymatic reactions. ANT,
anthracene; FLU, fluoranthene; PHE, phenanthrene; PYR, pyrene.

(Table 1)	( fa	b	le	1)	
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	Dontinollia	Н.	Т.	P. gigantea	B. adusta	Р.	Р.	S. paradoxa
Assembly statistics	Dentipellis	annosum	versicolor			brevispora	ostreatus	KUC8140
	sp.					HHB-7030	PC15	
	KUC8613					SS6		
Genome size (Mbp)	36.71	33.6	44.79	30.14	42.73	49.96	35.6	44.41
Number of contigs	1,184	630	1,443	1,195	1,263	3,178	3,272	1,374
Number of scaffolds	425	39	283	573	508	1,645	572	1,291
Annotation								
statistics								
Number of gene	14,320	13,405	14,296	11,891	15,473	16,170	11,603	17,098
models Average gene length	1,737	1,601	1,790	1,714	1,703	1,627	1,772	1,735
(bp) Average exon length	263	561	241	230	248	225	217	246
(bp) Average intron length	69	82	83	69	71	78	76	68
(bp) Average protein length	405	379	422	411	406	400	445	413
(aa) Number of exons per		5.39	5.81	6	5.59	5.66	6.4	5.78
gene	5.47							

ProtID	Fold change	P-value ANT/Contr	Fold change	P-value FLU/Contro	Fold change	P-value PHE/Contr	Fold change	P-value PYR/Contro
	ANT/ Control	ol	FLU/Contro I	I	PHE/Contr ol	ol	PYR/Contro I	I
859292	2.39*	1.00E-05	1.95	9.00E-04	2.59*	1.36E-06	2.63*	4.92E-07
882466	2.92*	8.40E-03	3.65*	1.30E-03	-	-	-	-
926550	-	-	2.13*	3.10E-03	-2.43	3.80E-03	2.03*	4.80E-03
833741	1.90	1.27E-05	2.13*	2.57E-07	2.78*	2.37E-12	2.05*	8.29E-07
818401	1.93	8.20E-03	3.04*	5.26E-06	2.62*	8.53E-05	2.27*	7.00E-04
875659	-	-	-	-	2.61*	1.99E-07	2.16*	2.54E-05
861037	2.19*	5.0E-04	1.91	5.10E-03	-	-	1.98	1.70E-03
798081	-	-	2.28*	3.30E-3	-	-	-	-
548471	-	-	-	-	2.55*	1.83E-11	-	-
772453	-	-	-	-	2.59*	3.60E-03	-	-
879478	1.73	2.60E-03	1.91	4.00E-04	3.17*	2.06E-10	1.91	4.00E-04
840982	-	-	-		2.69*	1.90E-03	-	-
841965	-	-	1.92	3.90E-3	2.99*	7.74E-07	-	-
873894	-	-			3.49*	9.90E-03	-	-
832704	-	-	1.73	2.00E-04	1.91	8.50E-06	2.07*	3.86E-07

688 Gene expression values with P-value > 0.01 were substituted by a dash (-). Fold changes of higher than 2 were bolded and denoted 689 with an asterisk (\*).











