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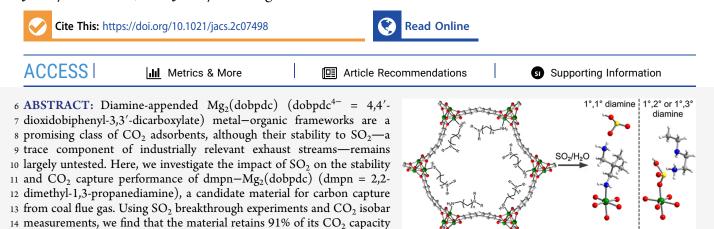


Article

¹ Evaluation of the Stability of Diamine-Appended Mg₂(dobpdc) ² Frameworks to Sulfur Dioxide

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¹⁶ levels of CO₂ and SO₂, highlighting the robustness of this framework to ¹⁷ SO₂ under realistic CO₂ capture conditions. Initial SO₂ cycling experiments suggest dmpn-Mg₂(dobpdc) may achieve a stable ¹⁸ operating capacity in the presence of SO₂ after initial passivation. Evaluation of several other diamine-Mg₂(dobpdc) variants reveals ¹⁹ that those with *primary,primary* (1°,1°) diamines, including dmpn-Mg₂(dobpdc), are more robust to humid SO₂ than those ²⁰ featuring *primary,secondary* (1°,2°) or *primary,tertiary* (1°,3°) diamines. Based on the solid-state ¹⁵N NMR spectra and density ²¹ functional theory calculations, we find that under humid conditions, SO₂ reacts with the metal-bound primary amine in 1°,2° and ²² 1°,3° diamine-appended Mg₂(dobpdc) to form a metal-bound bisulfite species that is charge balanced by a primary ammonium ²³ cation, thereby facilitating material degradation. In contrast, humid SO₂ reacts with the free end of 1°,1° diamines to form ²⁴ ammonium bisulfite, leaving the metal-diamine bond intact. This structure-property relationship can be used to guide further ²⁵ optimization of these materials for CO₂ capture applications.

26 INTRODUCTION

27 Carbon dioxide emissions derived from fossil fuel combustion 28 have been steadily rising over the last several decades and are 29 widely acknowledged to be the leading cause of global climate 30 change.¹ In tandem with increasing reliance on low- and zerocarbon energy technologies, restricting global warming to <2 31 32 °C above preindustrial levels will require a targeted reduction $_{33}$ in anthropogenic CO₂ emissions within the current energy 34 infrastructure. Coal-fired power plants account for approx-35 imately 30% of CO_2 emissions worldwide, ^{1,2} and, as fossil fuels, 36 are projected to continue supplying the majority of global 37 energy in the short term. Post combustion capture from these 38 point sources stands as a critical mitigation strategy.³ Aqueous 39 amine solutions represent the most mature carbon capture 40 technology, although their wide-spread implementation has 41 been stymied due to their intrinsically high regeneration 42 energies, low CO₂ capacities, and operational issues related to 43 their corrosivity and volatility.^{3,4}

15 after saturation with a wet simulated flue gas containing representative

⁴⁴ In recent years, amine-functionalized solid adsorbents— ⁴⁵ including porous carbons, silicas, and metal—organic frame-⁴⁶ works (MOFs)—have gained increasing attention as promising alternatives for energy-efficient carbon capture, owing to their 47 potentially low regeneration energy, selectivity for CO_2 at low 48 concentrations relevant to flue streams, and cycling stabil- 49 ity. ^{5–14} Among these materials, MOFs of the type diamine— 50 Mg₂(dobpdc) (dobpdc^{4—} = 4,4'-dioxidobiphenyl-3,3'-dicar- 51 boxylate)—and recently tetraamine—Mg₂(dobpdc)¹⁵—have 52 attracted significant interest, given their high adsorption 53 capacities, selectivities for CO_2 over N₂, and stability in the 54 presence of water.^{16–23} These materials reversibly adsorb CO_2 55 via a cooperative mechanism involving CO_2 insertion into the 56 framework metal—amine bonds to form chains of ammonium 57 carbamate that propagate down the framework channels. As a 58 result of this unique behavior, CO_2 uptake occurs in a step- 59 wise fashion, such that initial adsorption and subsequent 60

diamine-Mg2(dobpdc)

Mg-amine

bond intact

Mq-amine

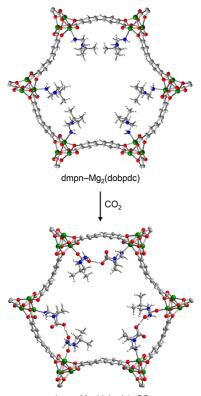
bond broken

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61 saturation occur within a narrow pressure or temperature 62 range. These adsorbents can therefore achieve very high CO_2 63 working capacities with relatively small temperature or 64 pressure swings compared to typical Langmuir-type adsorb-65 ents, which represent the vast majority of other amine-66 functionalized materials known to date.

⁶⁷ Importantly, the characteristic CO_2 adsorption threshold of ⁶⁸ diamine $-Mg_2(dobpdc)$ can be tuned by changing the ⁶⁹ appended amine, allowing optimization for a specific capture ⁷⁰ application.^{18–20} For example, dmpn $-Mg_2(dobpdc)$ (dmpn = ⁷¹ 2,2-dimethyl-1,3-propanediamine) exhibits a CO_2 adsorption ⁷² step at 15 mbar and 40 °C, ideally positioned for 90% CO_2 ⁷³ capture from coal flue gas.¹⁹ This material is also unusual in ⁷⁴ that it adsorbs CO_2 via a mixed mechanism to form carbamic ⁷⁵ acids and ammonium carbamate chains that propagate along ⁷⁶ the framework channels at adjacent vertices and interact ⁷⁷ through hydrogen bonding (Figure 1).^{19,24} Recent modeling of

f1



dmpn-Mg₂(dobpdc)-CO₂

Figure 1. Structures of dmpn–Mg₂(dobpdc) before (upper) and after (lower) uptake of CO₂, generated from vdW-corrected DFT calculations, illustrating the formation of a 1:1 mixture of ammonium carbamate to carbamic acid chains.²⁴ Green, gray, red, blue, and white spheres represent Mg, C, O, N, and H atoms, respectively. Views are down the *c* axis.

⁷⁸ the performance of dmpn $-Mg_2(dobpdc)$ in various temper-⁷⁹ ature-swing adsorption processes revealed that it may be cost-⁸⁰ competitive with state-of-the-art monoethanolamine solutions ⁸¹ for CO₂ capture from coal flue gas if the heat generated upon ⁸² adsorption can be recycled efficiently.²⁵

⁸³ Further advancement of diamine-appended $Mg_2(dobpdc)$ ⁸⁴ materials toward carbon capture applications requires the ⁸⁵ rigorous evaluation of their stabilities and properties under ⁸⁶ practical working conditions. While some of these frameworks ⁸⁷ have been shown to maintain excellent capture performance in

the presence of humidity and even $O_{2\prime}^{19,20,22,23,26,27}$ their 88 performance in the presence of SO_2 , another key component ⁸⁹ of coal flue gas, remains largely unexplored.^{22,23} This is in 90 contrast to the growing number of literature studies that have 91 examined the stability of other solid adsorbents to SO2, such as 92 zeolitic imidazolate frameworks and amine-appended sili- 93 cas.^{28–32} Although coal flue gas is treated to remove a portion 94 of the SO₂ to comply with environmental regulations, 8,33,34 a 95 non-negligible concentration remains in the outlet stream. For 96 example, based on tabulated data for CO2 and SO2 emissions 97 from coal-fired power plants in the United States and assuming 98 an average flue gas CO_2 concentration of 13.5%,³⁵ the flue gas 99 streams of established plants contain an estimated SO₂ 100 concentration of 80 ppm (see Section S2 of the Supporting 101 Information for details).³⁶ Even with new regulations enacted 102 in 2010, the flue streams of new or remodeled power plants 103 can contain as much as 20 to 50 ppm $\mathrm{SO}_2.^{37,38}$ Modifying 104 current desulfurization processes to remove the residual SO2 105 would increase the costs associated with an already expensive 106 process. $^{33,39-43}$ As such, candidate CO₂ capture adsorbents 107 must be robust and maintain their capture performance in the 108 presence of trace quantities of SO_{22}^{7} particularly in the 109 presence of water, which can constitute as much as 5 to 15% of 110 the coal flue gas.^{8,44,45} While it is a growing area of interest, few 111 studies to date have investigated the effect of SO2 on the 112 performance of CO₂ capture materials and even fewer under 113 relevant concentrations of SO₂ in the presence of humidity. 114

To the best of our knowledge, only two studies have 115 reported preliminary data on the stability and CO₂ capture 116 performance of diamine-Mg₂(dobpdc) following exposure to 117 humid SO₂.^{22,23} In brief, Mg_2 (dobpdc) appended with the 118 primary, primary (1°,1°) diamines ethylenediamine (en), 1-119 methylethylenediamine (men), and 1,1-dimethylethylenedi- 120 amine (dmen) and primary, secondary $(1^{\circ}, 2^{\circ})$ N-isopropylethy- 121 lenediamine (i-2) and N-ethylethylenediamine (e-2) were 122 exposed to 500 ppm of SO₂ with 100% humidity for periods of 123 up to 2 h. The men-, dmen-, e-2-, and i-2-appended variants 124 were reported to retain >90% of their CO₂ capacity under all 125 conditions, whereas the capacity of en-Mg₂(dobpdc) 126 diminished by more than 30% upon exposure for more than 127 30 min.^{22,23} To rationalize the superior performance of men-, 128 dmen-, i-2-, and e-2-Mg₂(dobpdc) relative to en- 129 $Mg_2(dobpdc)$, it was proposed that sterically encumbering 130 substituents on the amines could provide enhanced stability in 131 the presence of SO₂. However, this hypothesis was not tested 132 further, and a meaningful analysis of the data is complicated by 133 the fact that $en-Mg_2(dobpdc)$ partially decomposes in the 134 presence of water alone.^{22,23,27} Additionally, complete 135 saturation of the materials with SO2 was not measured or 136 confirmed. Thus, further investigation is needed to develop a 137 detailed understanding of how SO₂ exposure impacts the 138 performance of these materials over extended exposure times, 139 the nature of any interaction between the framework and SO_{21} 140 and the role of the diamine structure in dictating relative 141 stability.

Here, we present an in-depth study of the stability of dmpn- 143 $Mg_2(dobpdc)$ to humid SO_2 using SO_2 breakthrough experi- 144 ments and CO_2 adsorption measurements. We find that the 145 CO_2 capacity of the material decreases following exposure to 146 humid SO_2 , although this capacity loss is substantially 147 attenuated when the sample is exposed to a simulated flue 148 gas (30 ppm SO_2 , 14% CO_2 , ~2% H₂O, balance N₂ at 40 °C). 149 Using solid-state NMR spectroscopy and van der Waals 150

151 (vdW)-corrected density functional theory (DFT), we identify 152 that in the presence of humidity, SO₂ reacts with the free end 153 of the appended diamine in dmpn-Mg₂(dobpdc) to form 154 ammonium bisulfite. Analysis of several other Mg₂(dobpdc) 155 variants appended indicates that ammonium bisulfite for-156 mation upon SO₂ exposure is a generalizable phenomenon in 157 these materials. However, in materials featuring 1° , 2° and 1° , 3° 158 diamines, SO₂ preferentially reacts with the metal-bound 159 amine, rendering these variants far less stable to the gas than 160 those appended with 1° , 1° amines. These results provide 161 unprecedented insight into the factors governing the stability 162 of diamine-appended Mg₂(dobpdc) materials to humid SO₂ 163 and will support their iterative design for real-world 164 applications.

165 RESULTS AND DISCUSSION

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Breakthrough Experiments. The stability of dmpn-166 167 Mg₂(dobpdc) to SO₂ was examined by carrying out SO₂ 168 breakthrough experiments followed by analysis of the CO₂ 169 capture performance of the material using isobar measure-170 ments. Breakthrough adsorption measurements were conducted using a custom-built instrument designed to enable 171 172 exposure of the material to controlled gas mixtures containing 173 SO₂ and the detection of SO₂ outlet concentrations over time 174 (see Section S3 of the Supporting Information and Figure S3). 175 An initial breakthrough experiment was carried out with a 176 stream containing 30 ppm of dry SO₂ in a balance of N₂ at 40 177 °C and atmospheric pressure. Under these conditions, SO2 178 breakthrough occurred within minutes, and negligible uptake 179 was measured (Figure S5 and Table S1). A CO₂ adsorption 180 isobar collected for dmpn-Mg₂(dobpdc) after this experiment 181 revealed no change in the capacity of the material (Figure 2a). 182 Furthermore, the material retained more than 90% of its CO₂ 183 capacity after exposure to concentrations of SO₂ up to 300 184 ppm in N₂ (Figure S7). In all, dmpn–Mg₂(dobpdc) is among 185 only a handful of other amine-functionalized adsorbents in the 186 literature found to be robust to such levels of dry 187 SO_2 .

We next examined the stability of dmpn-Mg₂(dobpdc) in 188 189 the presence of SO₂ and humidity by carrying out a 190 breakthrough measurement using an inlet stream containing 191 30 ppm of SO₂ and \sim 2% water in N₂ at atmospheric pressure $_{192}$ and 40 $\,^{\circ}\text{C}.$ Under these conditions, SO_2 breakthrough took 193 more than 2 days, significantly longer than under dry 194 conditions. The framework also adsorbed significantly more 195 SO₂ in the presence of water than under dry conditions (4.5 \pm 196 0.3 mmol/g; see Figure S5 and Table S1). This capacity 197 corresponds to an uptake of approximately one SO₂ molecule 198 per diamine, based on a diamine loading of 98% (as 199 determined from solution ¹H NMR analysis of the digested 200 framework). This approximate 1:1 stoichiometry was also 201 confirmed by elemental analysis, which revealed a S/N mass 202 ratio of 1.08, corresponding to 0.95 \pm 0.1 SO₂ molecules per 203 diamine (Table S3). While dmpn-Mg2(dobpdc) still exhibited 204 a step-shaped CO2 uptake at 40 °C following saturation with 205 humid SO₂, it retained only 75% of its original CO₂ capacity, 206 considerably less than that retained after dry SO₂ exposure 207 (Figure 2a). It is clear that humidity promotes at least partial 208 deactivation of dmpn-Mg₂(dobpdc) in the presence of SO_{24} 209 consistent with previous studies of other amine-functionalized 210 solids.^{8,30}

Interestingly, when pristine dmpn $-Mg_2(dobpdc)$ was exposed to a simulated flue gas stream containing 30 ppm

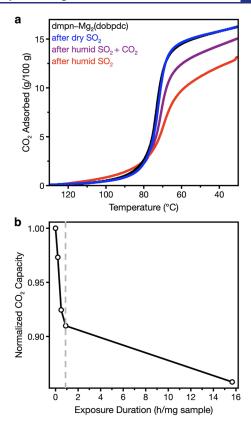


Figure 2. (a) Pure CO₂ adsorption isobars obtained at atmospheric pressure for pristine dmpn-Mg₂(dobpdc) before (light blue) and after saturation with the following gas streams at 40 °C and atmospheric pressure: dry 30 ppm SO₂ in N₂ (dark blue), 30 ppm SO₂ with 2% water in N₂ (red curve), and 30 ppm, 2% water, and 14% CO₂ in N₂ (purple curve). (b) Total CO₂ capacity retained by dmpn-Mg₂(dobpdc) based on isobar data obtained following exposure to a simulated coal flue gas (30 ppm SO₂, 14% CO₂, ~2% H₂O in N₂ at 40 °C and atmospheric pressure) under breakthrough conditions for various time durations. The gray dashed line represents the point at which SO₂ saturation occurs under these conditions.

SO₂, ~2% water, and 14% CO₂ in a balance of N₂ at 213 atmospheric pressure and 40 °C, SO₂ breakthrough occurred 214 in only 15 h, less than half the time required in the absence of 215 CO2. The SO2 uptake under these conditions was also 216 significantly less, only 1.4 \pm 0.3 mmol/g or 0.4 \pm 0.1 SO₂/ 217 diamine (Figure S5 and Table S1), revealing that the presence 218 of CO₂ inhibits SO₂ binding significantly. Based on pure CO₂ 219 isobar data collected following this breakthrough experiment, 220 the framework retained 91% of its original CO₂ capacity at 40 221 °C (Figure 2a, purple curve). In addition, dmpn-222 Mg₂(dobpdc) retained 86% of its initial CO₂ capacity after 223 exposure to the simulated flue gas stream for a duration 224 corresponding to approximately 16 sequential breakthrough 225 experiments (Figure 2b). This robust performance further 226 highlights the potential of dmpn-Mg₂(dobpdc) as a candidate 227 for CO₂ capture from coal flue gas streams. 228

To investigate dmpn-Mg₂(dobpdc) deactivation in the 229 presence of humid SO₂, we used solution ¹H NMR 230 spectroscopy and CHNS elemental analysis to analyze samples 231 following humid SO₂ exposure and subsequent regeneration. 232 Sample regeneration at 130 °C under dry, flowing N₂ was 233 found to be optimal after testing a variety of conditions, 234 including different temperatures and purge gas streams 235 (Figures S8-S10). Analysis of ¹H NMR spectra collected for 236

С

237 digested framework samples following humid SO₂ exposure 238 before and after regeneration revealed that diamine loss 239 (~10%) occurs only upon regeneration. Elemental analysis of 240 dmpn-Mg₂(dobpdc) following humid SO₂ exposure and 241 regeneration revealed an N/C ratio consistent with this 242 diamine loss (Table S4) and an S/N ratio of 0.10, 243 corresponding to slightly less than 0.1 molecules of SO₂ per 244 diamine. Altogether, these results indicate that the diminished 245 CO₂ capacity of dmpn-Mg₂(dobpdc) following SO₂ exposure 246 is due to diamine loss upon regeneration and the persistence of 247 a small quantity of bound sulfur species.

Importantly, exposure of dmpn-Mg2(dobpdc) to humid 248 249 SO₂ does not cause irreversible degradation of the underlying 250 framework. Indeed, powder X-ray diffraction analysis of a sample of dmpn $-Mg_2(dobpdc)$ following humid SO₂ exposure 251 252 and subsequent diamine stripping with methanol revealed a powder pattern consistent with that obtained for pristine 253 $_{254}$ Mg₂(dobpdc) (Figure S15). Further, N₂ adsorption data 255 obtained at 77 K for both samples revealed almost identical $_{256}$ surface areas of ~4200 m²/g (Figure S16). This indicates that 257 the grafted amines protect the Mg₂(dobpdc) framework from 258 H₂O- and SO₂-induced poisoning at the metal sites.⁵⁰ 259 Importantly, this opens the possibility that the material could 260 potentially be occasionally regenerated in situ through 261 replacement of the diamines to regain capacity losses due to extensive SO_2 exposure. 262

We next investigated the stability of dmpn-Mg₂(dobpdc) to 263 264 repeated cycling under the simulated flue gas stream (30 ppm 265 SO₂, 14% CO₂, \sim 2% H₂O, balance N₂ at atmospheric pressure 266 and 40 °C). Four cycles were carried out in total, and each 267 cycle consisted of exposure of the framework to the gas stream 268 until full breakthrough of SO2 was observed, followed by 269 regeneration for 30 min under flowing N2 at 130 °C and 270 atmospheric pressure (see Figure S6 and Table S2). After each cycle, pure CO_2 adsorption isobars were collected (Figure 3a). 271 272 Over the course of the four cycles, the CO₂ capacity at 40 °C 273 was reduced by 21%, with the greatest capacity loss (10%) 274 occurring after the first cycle (Figure 3). The progressively 275 smaller losses in CO₂ capacity with each cycle number suggest 276 that this material may reach a stable CO₂ operating capacity in 277 the presence of SO₂ after initial passivation. This observation is also supported by the diminished SO₂ uptake with increasing 278 279 cycle number (Figure 3b). Similar results were obtained from 280 SO₂ cycling experiments using a gas stream consisting of 30 ppm SO₂ and ~2% H_2O in N_2 at atmospheric pressure and 40 281 °C, indicating that, independent of CO₂, the framework affinity 282 283 for SO₂ uptake decreases with repeated cycling (see Figures 284 S12 and S13; Table S7). Analogous behavior has also been 285 reported for amine-silica materials cycled repeatedly with ppm 286 levels of dry SO₂, with and without CO₂.^{28,29,51} In all, these 287 results further highlight the potential of $dmpn-Mg_2(dobpdc)$ 288 for coal flue gas CO₂ capture applications. However, more 289 extended cycling tests are needed to substantiate this 290 performance along with evaluation of the CO₂ capture 291 performance in the presence of SO2 during simulated flue 292 gas exposure.

Spectroscopic Characterization of Humid SO₂ Ad-294 **sorption.** Infrared spectroscopy was used as an initial probe of 295 the species formed upon humid SO₂ uptake in dmpn– 296 Mg₂(dobpdc). Following saturation of the material with humid 297 SO₂, new peaks appeared at 913 and 864 cm⁻¹ (Figure S14), 298 closely corresponding to the S–O stretching frequency of the 299 sulfite ion and/or bisulfite ion (984 to 917 cm⁻¹)^{31,32,48,52–57}



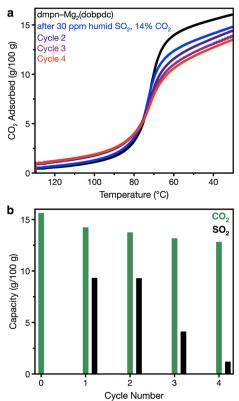


Figure 3. (a) Pure CO₂ isobars collected at atmospheric pressure for dmpn-Mg₂(dobpdc) before (black curve) and after (colored curves) successive cycles of saturation with SO₂ using a gas stream containing 30 ppm SO₂, 2% water, and 14% CO₂ in N₂ at 40 °C and atmospheric pressure. Prior to isobar collection, the sample was regenerated under a flow of dry N₂ at 130 °C for 30 min. The material was then heated in the breakthrough apparatus to 100 °C under flowing N₂ at atmospheric pressure before the next cycle. (b) Comparison of the SO₂ capacities measured from SO₂ breakthrough experiments using a simulated coal flue gas and the CO₂ capacity at 40 °C determined from isobaric measurements following each SO₂ saturation cycle.

and the S-OH vibration of bisulfate or bisulfite (870 to 850 300 cm⁻¹), respectively.^{31,32,55-57} Together, the appearance of 301 these new peaks suggests that SO_2 adsorbs in dmpn- 302 Mg₂(dobpdc) to form bisulfite (HSO₃⁻). Liquid chromatog- 303 raphy-mass spectrometry analysis of the liquid solution 304 generated upon stripping the amines from dmpn- 305 Mg₂(dobpdc) following exposure to a humid 30 ppm SO₂ 306 stream revealed the presence of the sulfite ion, which was not 307 present in a control sample prepared by stripping the amines 308 from pristine dmpn-Mg₂(dobpdc) (Figure S17). Because the 309 charge-inducing nature of ESI mass spectrometry does not 310 enable one to distinguish between sulfite and bisulfite ions, we 311 turned to solid-state magic angle spinning (MAS) NMR 312 spectroscopy and DFT calculations to make a more conclusive 313 assignment of the species formed upon SO₂ uptake in dmpn- 314 $Mg_2(dobpdc)$. 315

The room-temperature solid-state ¹⁵N NMR spectra for a 316 sample of dmpn-Mg₂(dobpdc) before and after saturation 317 with humid 30 ppm SO₂ are shown in Figure 4. The pristine 318 f4 framework exhibits a single ¹⁵N resonance at 12 ppm, as 319 determined previously.²⁴ While at any given time, there will be 320 two chemically distinct nitrogen atoms in the framework, 321 corresponding to the metal-bound and free amines, symmetric 322 diamines are known to exhibit only one ¹⁵N resonance at room 323

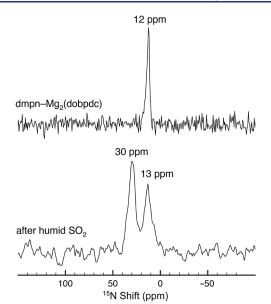


Figure 4. Room-temperature ¹⁵N solid-state MAS NMR spectra of pristine dmpn $-Mg_2(dobpdc)$ (upper, 16.4 T, 15 kHz MAS rate) and dmpn $-Mg_2(dobpdc)$ after saturation with a gas stream containing 30 ppm SO₂ and 2% water in N₂ (lower, 11.7 T, 10 kHz MAS rate). The spectrum for pristine dmpn $-Mg_2(dobpdc)$ is reprinted with permission from ref 24. Copyright 2018, Journal of the American Chemical Society.

324 temperature due to rapid exchange of the two sites.^{24,58} 325 Following saturation of the material with humid SO₂, the 326 original amine resonance persists and a new ¹⁵N resonance 327 appears at 30 ppm, which was assigned to an ammonium 328 species based on previous solid-state ¹⁵N NMR character-329 ization of ammonium formed in the material following CO₂ 330 dosing.²⁴ Although these experiments were not quantitative, 331 the retention of the original peak indicates that unreacted 332 amines remain in the material following dosing with SO₂. 333 Considering the previously quantified loading of one SO₂ for 334 every diamine and the need to maintain charge balance, this 335 result supports the formation of ammonium bisulfite ($-NH_3^+$ 336 HSO₃⁻) species.

We next turned to vdW-corrected DFT to further elucidate 337 338 the chemical species formed upon uptake of humid SO₂ in 339 dmpn-Mg₂(dobpdc) (see Section S12 of the Supporting 340 Information for details). Based on the spectroscopic data, we 341 initially considered an optimized ammonium bisulfite structure 342 featuring bisulfite charge balanced by ammonium generated 343 from the free end of the diamine (Figure 5a). This amine is 344 expected to be more basic and less sterically encumbered than 345 the metal-bound amine and therefore more likely to interact 346 with SO₂. The predicted ¹⁵N shifts for this species are 27 and 347 12 ppm, which correspond closely with the experimentally 348 observed shifts of 30 and 13 ppm. The predicted chemical 349 shifts for this structure were also the closest match to the 350 experiment when compared with the shifts for several other 351 hypothetical structures, including neutral SO₂ adducts and an 352 analogue of the ammonium carbamate structure, with CO₂ 353 replaced by SO₂₁ featuring direct bonds from an amine 354 nitrogen to sulfur (Figure 5 and Table 1). We also considered 355 a second ammonium bisulfite structure featuring a metal-356 bound bisulfite charge balanced by ammonium generated from 357 the metal-bound amine (Figure 5b), although the calculated

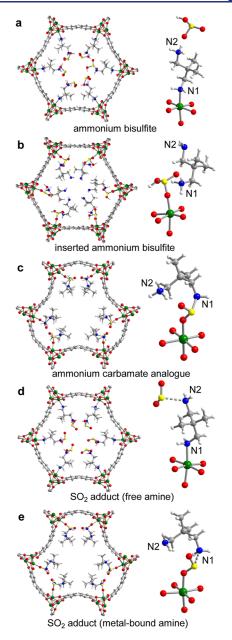


Figure 5. Structures of candidate products formed upon humid SO_2 uptake in dmpn-Mg₂(dobpdc) generated from vdW-corrected DFT calculations: ammonium bisulfite formed upon reaction of SO_2 at the (a) free amine and (b) metal-bound amine, (c) SO_2 analogue of the ammonium carbamate structure formed in dmpn-Mg₂(dobpdc)²⁴ and an adduct of SO_2 , and (d) free amine and (e) metal-bound amine. Green, gray, red, blue, yellow, and white spheres represent Mg, C, O, N, S, and H atoms, respectively.

¹⁵N shifts for this species (32 and 5 ppm) show poorer 358 agreement with the experimental data. 359

Owing to the broad nature of the experimental ¹⁵N NMR ³⁶⁰ resonances, it is not possible to exclude with confidence the ³⁶¹ latter structure based on these data alone. However, binding ³⁶² energies computed for the various SO₂-derived products using ³⁶³ vdW–DFT further support ammonium bisulfite formation at ³⁶⁴ the free amine. Indeed, this structure is the most stable of all ³⁶⁵ examined products, and it is more stable by ~50 kJ/mol ³⁶⁶ compared to the metal-bound bisulfite structure (Table 1). It ³⁶⁷ should be noted that our DFT approach for calculating binding ³⁶⁸ energies neglects quantum zero-point and finite-temperature ³⁶⁹

t1

	calculated ¹⁵ N shift (ppm)		
SO ₂ -derived product	N1	N2	calculated binding energy (kJ/mol)
ammonium bisulfite	12	27	-148
inserted ammonium bisulfite	32	5	-98
ammonium carbamate analogue	128	28	-63
SO ₂ adduct (free amine)	14	69	-70
SO ₂ adduct (metal-bound amine)	81	19	-53

370 corrections associated with the nuclei. Based on prior studies 371 of CO₂ uptake in various diamine $-Mg_2(dobpdc)$ materials, the 372 predicted binding energies deviate from isosteric heats of 373 adsorption measured in the experiment; however, they do 374 accurately capture measured trends.^{15,20,59}

Effect of Diamine Structure on MOF Stability to SO₂. To further investigate the stability of alkyldiamine-appended Mg₂(dobpdc) materials to SO₂, we evaluated the CO₂ capture performance of several analogues featuring 1°,1°-, 1°,2°-, or 1°,3°-alkyldiamines (including dmen, i-2, and e-2 studied previously^{22,23}) following saturation with a humid stream (2% 881 H₂O) containing 30 ppm of SO₂ in N₂ (see Table 2 and Figure S18). Notably, materials functionalized with 1°,1°-diamines— 382 dmpn-Mg₂(dobpdc) and dmen-Mg₂(dobpdc)-were found 383 to be the most stable to humid SO_2 based on the CO_2 capacity 384 and diamine loading that were retained following SO₂ exposure 385 and subsequent regeneration. Both of these materials are stable 386 to water in contrast to the previously studied en- 387 Mg₂(dobpdc),^{22,23,27} which also features a 1°,1° amine. 388 These preliminary results indicate that, contrary to what was 389 proposed previously,²³ the presence of bulky substituents is 390 not alone sufficient to rationalize the stability of diamine- 391 appended Mg₂(dobpdc) to SO₂. However, of the 1°,2°- and 392 1°,3°-alkylamine-appended variants investigated here, those 393 with larger secondary or tertiary amine substituents, namely, i- 394 $2-Mg_2(dobpdc)$ and $ee-2-Mg_2(dobpdc)$ (ee-2 = N,N- 395 diethylethylenediamine), were found to be more stable than 396 their counterparts with smaller N-alkyl substituents. Thus, for 397 diamines in the same class, the size of the alkyl substituents 398 may also be an important factor, as discussed further below. 399

It was previously established that $1^{\circ},2^{\circ}$ - and $1^{\circ},3^{\circ}$ -diamines 400 bind to the open metal sites of Mg₂(dobpdc) through the 401 primary amine,¹⁸ leaving the secondary or tertiary amine 402 dangling into the framework pore. Separately, multiple studies 403 of other amine-appended adsorbents have found that humid 404 SO₂ often reacts irreversibly with primary amines and to a 405 lesser degree with secondary amines, while it tends to bind 406 reversibly to materials featuring tertiary amines.^{30,39,60,61} We 407 hypothesized that in $1^{\circ},2^{\circ}$ - and $1^{\circ},3^{\circ}$ -diamine-appended 408 Mg₂(dobpdc), humid SO₂ may preferentially react with the 409

Table 2. Results of Stability Analysis for Diamine $-Mg_2(dobpdc)$ after Saturation with a Gas Stream Containing 30 ppm SO₂ and 2% Water in N₂ at 40 °C and Atmospheric Pressure

Diamine (Abbreviation)	Structure	Class	Diamine Loading (%) ^a			CO ₂
			Pristine	After Satu- ration with SO ₂	After SO ₂ and Regeneration ^c	Capacity Retained (%) ^c
2,2-dimethyl-1,3-propanediamine (dmpn)	H ₂ N NH ₂	1°,1°	98	98	89	77
1,2-diamino-2-methylpropane (dmen)	H ₂ N NH ₂	1°,1°	97	91	75	75
N-ethylethyelenediamine (e-2)	NH ₂ NH ₂	1°,2°	98	92	52	39
N-isopropyl-ethylenediamine (i-2)	NH ₂	1°,2°	101	89	70	64
<i>N,N</i> -dimethylethylenediamine (mm-2)	NH ₂	1°,3°	92	68	59	34
<i>N,N-</i> dimethylpropoanediamine (mm-3)	NH ₂	1°,3°	97	75	68	38
<i>N,N</i> -diethylethylenediamine (ee-2)		1°,3°	97	66	65	53

^{*a*}Determined from solution ¹H NMR data collected for digested samples. An error of $\pm 2\%$ was consistently determined for all samples based on standard deviation of three distinct diamine proton signals determined from two identically prepared samples. ^{*b*}Regeneration at 130 °C under flowing N₂ for 30 min. ^CBased on pure CO₂ isobars obtained after framework saturation with SO₂ and regeneration at 130 °C under flowing N₂ for 30 min. The error on each value is $\pm 2\%$ based on standard deviation of CO₂ capacity from three identically prepared samples.

410 metal-bound primary amine over the exposed but sterically 411 hindered amines. This disruption of the metal–amine bond 412 would in turn promote more facile degradation of the material 413 relative to 1° , 1° -diamine-appended variants, where SO₂ is 414 expected to react preferentially with the free amine. To 415 evaluate this possibility further, we collected solid-state MAS 416 ¹⁵N NMR spectra for e-2–Mg₂(dobpdc) and ee-2– 417 Mg₂(dobpdc) as representative 1° , 2° - and 1° , 3° -diamine-418 appended frameworks before and after saturation with humid 419 30 ppm SO₂ in N₂ (Figure 6). The spectra of the pristine 420 frameworks both feature two resonances, which were assigned 421 to the chemically distinct primary and secondary/tertiary 422 amines in the materials.²⁴ Saturation with humid SO₂ resulted 423 in disappearance of the primary amine resonance for e-2– 424 Mg₂(dobpdc) (Figure 6a), whereas in the case of ee-2–

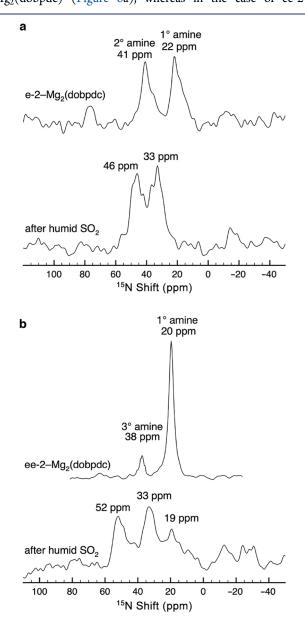


Figure 6. Room-temperature ^{15}N MAS NMR spectra of e-2– $Mg_2(dobpdc)$ (a) and ee-2– $Mg_2(dobpdc)$ (b) before and after saturation with a gas stream containing 30 ppm of SO₂ and 2% water in N_2 . Experiments were conducted at 11.4 T with a MAS rate of 10 kHz.

Mg₂(dobpdc), this resonance was diminished but not absent 425 (Figure 6b). For both materials, SO₂ exposure resulted in the 426 appearance of a strong resonance at 33 ppm, which was 427 assigned to a primary ammonium species, based on similar 428 chemical shifts identified for primary ammonium species 429 formed upon CO₂ uptake in diamine-appended 430 Mg₂(dobpdc).²⁴ These results support protonation of the 431 initially metal-bound primary amine in both materials 432 following exposure to humid SO₂. The secondary amine 433 peak for e-2-Mg₂(dobpdc) shifts from 41 to 46 ppm upon 434 saturation of the material with humid SO₂, possibly a result of 435 weak interactions between this amine and bisulfite. Similar 436 interactions may lead to a more dramatic shift of the tertiary 437 amine peak for ee-2-Mg₂(dobpdc) from 38 to 52 ppm after 438 humid SO₂ saturation, although further experiments would be 439 needed to make definitive assignments. Finally, following 440 saturation with humid SO₂, the infrared spectra of both 441 materials feature two new bands that support the presence of 442 bisulfite (Figure S19). 443

We again turned to vdW-DFT to help elucidate the species 444 formed upon saturation of e-2-Mg_(dobpdc) and ee-2- 445 $Mg_2(dobpdc)$ with humid SO_2 . Both materials experience 446 substantial diamine loss following humid SO₂ exposure and 447 regeneration (Table 2). However, e-2–Mg₂(dobpdc) retains a 448 much higher diamine loading following humid SO₂ exposure, 449 and as such the structure of this material with adsorbed SO2 is 450 presumed to be more uniform than that of ee-2- 451 $Mg_2(dobpdc)$. Accordingly, we used e-2- $Mg_2(dobpdc)$ as 452 the model framework for our calculations. Based on the ¹⁵N 453 NMR and IR data for e-2-Mg₂(dobpdc) and the evidence of 454 ammonium bisulfite formation (Figures 6 and S19), we 455 evaluated two possible structures formed upon SO2 uptake, 456 featuring ammonium bisulfite formed at the site of the 457 dangling amine or bisulfite inserted into the metal-amine 458 bond and stabilized by a primary ammonium cation (see 459 Figure S23 and Table S8). Interestingly, the predicted 460 chemical shifts for the structure resulting from protonation 461 of the primary metal-bound amine and its displacement by 462 bisulfite (34 and 52 ppm) are in better agreement with the 463 experimental shifts than those for the structure with bisulfite 464 and a dangling secondary ammonium (17 and 48 ppm). This 465 result contrasts with the reactivity determined for dmpn- 466 Mg₂(dobpdc), where bisulfite forms an ion pair with 467 ammonium generated from the free amine. In all, these results 468 support the hypothesis that $1^{\circ}, 2^{\circ}$ - (and by extension $1^{\circ}, 3^{\circ}$ -) 469 diamine-appended Mg₂(dobpdc) materials are less robust to 470 humid SO2 than 1°,1°-diamine-appended variants, owing to 471 the preferential reaction of humid SO₂ with the metal-bound 472 primary amine and disruption of the critical metal-nitrogen 473 linkage in these materials. As noted above, the trends in Table 474 2 additionally indicate that bulkier diamines within the same 475 class are more stable to humid SO2. Among diamines that bind 476 through a primary amine, heavier analogues are known to be 477 more thermally stable and volatilize at higher temperatures.¹⁸ 478 This stability may translate to decreased diamine loss during 479 SO_2 exposure and subsequent regeneration. 480

CONCLUSIONS

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We have carried out an in-depth investigation of the effect of $_{482}$ humid SO₂ exposure on the stability and CO₂ adsorption $_{483}$ properties of the MOF dmpn-Mg₂(dobpdc), a promising $_{484}$ candidate for CO₂ capture from coal flue gas.¹⁹ Using a $_{485}$ custom-built SO₂ breakthrough apparatus, we found that this $_{486}$

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487 material exhibits excellent stability to dry and humid SO₂. In 488 particular, the material exhibits no loss in CO₂ capacity 489 following saturation with a stream containing 30 ppm SO_2 in 490 $N_{\rm 2}$ and only a 25% loss in $\rm CO_2$ capacity following saturation ⁴⁹¹ with a stream of humid 30 ppm SO_2 in N_2 . Importantly, this 492 capacity loss is reduced to under 10% when dmpn-493 Mg₂(dobpdc) is exposed to humid 30 ppm SO₂ in the 494 presence of 14% CO₂. Further, preliminary cycling experiments 495 performed using a simulated flue gas (30 ppm SO_2 , 14% CO_2 , 496 ~2% H₂O in N₂) indicate that the CO₂ capacity may stabilize 497 around 80% after initial SO₂-induced passivation. In all, these 498 results are promising compared to the few studies that have 499 been conducted on other amine-appended adsorbents under 500 relevant SO₂ concentrations, and they further highlight the 501 potential of dmpn-Mg₂(dobpdc) for CO₂ capture applica-502 tions.

Characterization of dmpn-Mg2(dobpdc) following humid 503 504 SO₂ exposure using IR and ¹⁵N NMR spectroscopy, supported 505 by vdW-DFT calculations, revealed that SO₂ reacts with the 506 free end of the metal-bound diamine to form ammonium 507 bisulfite species. To the best of our knowledge, these results 508 afford the most comprehensive evidence to date of the 509 formation of ammonium bisulfite in amine-functionalized 510 materials upon reaction with humid SO₂. Spectroscopic and 511 DFT analyses of several other diamine-appended 512 Mg₂(dobpdc) variants revealed that SO₂ preferentially reacts 513 with the metal-bound primary amine in variants featuring $514 1^{\circ}, 2^{\circ}$ - or $1^{\circ}, 3^{\circ}$ -diamines—due to weaker interactions between 515 SO2 and the pore-dwelling N-alkylated amines-forming a 516 metal-bound bisulfite charge balanced by a primary ammonium 517 cation. This disruption of the metal-amine bond is likely what 518 renders these materials significantly less stable to SO₂ than the 519 1°,1°-diamine-appended variants studied here, including 520 dmpn-Mg₂(dobpdc). Knowledge of this critical structure-521 property relationship will support the design of more robust 522 amine-appended frameworks for carbon capture from coal flue 523 gas and, more generally, for the capture of CO₂ from other 524 industrially relevant exhaust streams (e.g., from cement 525 manufacturing and biofuel combustion), which contain similar 526 SO₂ concentrations to those investigated here.^{62,63}

527 **EXPERIMENTAL SECTION**

General Procedures. All solvents and reagents were purchased from commercial sources and used without further purification. The sol linker H₄dobpdc was purchased from Hangzhou Trylead Chemical sources and used without further purification. The adsorption measurements, with custom mixtures purchased from sources provide X-ray diffraction (PXRD) patterns were collected on source operating at 40 kV, 40 mA. Elemental analysis was conducted so at the Microanalytical Facility at UC Berkeley. Attenuated total source Fourier transform infrared spectra were collected in air on source operating at 40 kV. Fourier transform IR equipped with a source previous transform infrared spectra were collected in air on source flation for the fourier transform IR equipped with a source Pike GladiATR module.

Synthesis of Mg₂(dobpdc). The framework Mg₂(dobpdc) was 541 produced following a previously reported solvothermal synthesis.¹⁸ 542 Briefly, Mg(NO₃)₂ (5.75 g, 22.5 mmol, 1.24 equiv) and H₄dobpdc 543 (4.95 g, 18.0 mmol, 1.00 equiv) were added to 100 mL of a 55:45 544 methanol/*N*,*N*-dimethylformamide (v/v) solution. This mixture was 545 sonicated until all solids completely dissolved and then gravity filtered 546 to ensure removal of any particulates. The solution was then 547 transferred to a glass pressure vessel with a stir bar and heated to 120 548 °C in an oil bath for 20 h under constant stirring. The resulting white 549 powder was then isolated by vacuum filtration and washed three times 550 at 60 °C for at least 3 h each in 300 mL of fresh *N*,*N*- dimethylformamide per wash. A solvent exchange was then performed 551 with three subsequent washes at 60 °C for at least 3 h each in 300 mL 552 of fresh methanol per wash. The washed material was then isolated 553 again through vacuum filtration and stored in fresh methanol. The 554 framework was desolvated prior to characterization by heating under 555 flowing N₂ at 180 °C for a minimum of 3 h, followed by heating at 556 180 °C under vacuum overnight. Langmuir surface area and PXRD 557 data were consistent with previously published values^{17,18} and are 558 presented in Figures S1 and S2.

Synthesis of Diamine-Appended Mg₂(dobpdc). Diamines 560 were appended onto the open metal sites of the Mg2(dobpdc) 561 framework following an adaptation of a previously reported 562 procedure.¹⁹ Approximately, 30 mg of Mg₂(dobpdc) was isolated 563 from methanol using vacuum filtration and left to dry on the filter 564 over vacuum until the powder was no longer moist, typically no more 565 than 5 min. This powder was then washed with 20 mL of fresh 566 toluene before being transferred to a scintillation vial charged with a 567 solution of 1 mL of diamine in 4 mL of toluene. After 16 h at room 568 temperature, the powder was isolated by vacuum filtration and 569 washed with 20 mL of fresh toluene over the filter paper. The 570 resulting diamine-appended material was then activated at 130 °C 571 under flowing N2 for 30 min to remove the solvent and excess 572 diamine before use. Diamine loading was then assessed using ¹H 573 NMR spectroscopy and collected on a Bruker AVQ-400 instrument 574 (see Table 2). In these measurements, roughly 5 mg of diamine- 575 appended Mg₂(dobpdc) material was suspended in 1 mL of DMSO- 576 d_6 and dissolved by the subsequent addition of 20 μ L of 35 wt % DCl 577 in D₂O solution. These solutions were sealed in a 4 mL vial and 578 heated at 60 °C until complete dissolution of solids was achieved. 579

Thermogravimetric Analysis. Thermogravimetric CO_2 isobars 580 were measured at atmospheric pressure using a TA Instruments TGA 581 Q5000. An aliquot (~5 mg) of the material was activated in the 582 instrument at 130 °C under flowing N₂ for 30 min before isobaric 583 measurements were conducted under 100% CO_2 at a temperature 584 ramp rate of 1 °C/min. Thermal decompositions were conducted 585 under pure N₂ at a 1.5 °C/min ramp rate. All CO_2 capacity values are 586 based on the gravimetric uptake in a pure, dry CO_2 isobar at 40 °C, 587 the target coal flue gas capture temperature. 588

SO₂ Exposure. A custom-built corrosive gas exposure system (see 589 Figure S3 and Section S3 of the Supporting Information) was used for 590 saturation of samples with humid SO2 under dilute and precisely 591 controllable conditions. A sample of activated diamine-Mg2(dobpdc) 592 was loaded into the bed in air as a loose powder and heated in the 593 exposure system under 50 sccm of dry N_2 at 100 $^{\circ}\mathrm{C}$ for 1 h to remove 594 any gases adsorbed during air exposure. The bed was then cooled to 595 40 °C under N₂, and the gas stream was then switched to 50 sccm of 596 30 ppm humid (~2% H₂O) or dry SO₂ in N₂, as specified for each 597 experiment. All exposure experiments were conducted using a flow 598 rate of 50 sccm at 40 °C and a total pressure of 1 atm. The custom gas 599 mixture was flowed continuously over the adsorbent bed until the 600 Tango TX-1 SO₂ detector measured the outlet concentration of SO₂ 601 as equivalent to the inlet concentration, indicating full material 602 saturation with SO2. The adsorbent bed was then purged with 50 603 sccm of dry N₂ to remove any excess free SO₂ while cooling to room 604 temperature before the bed was opened, and the sample was 605 recovered for subsequent testing. 606

The incorporation of CO_2 is possible via a three-way valve 607 connecting a gas cylinder containing 15% CO_2 in N₂ to the inert gas 608 mass flow controller. Upon blending this mixture with the SO_{2^-} 609 containing stream, the final test conditions were 30 ppm SO_2 , 14% 610 CO_2 , ~2% H₂O, balance N₂ at the same temperature and flow rate. 611 The bubbler water was exchanged following any change in the gas 612 stream composition to eliminate contamination from dissolved gases. 613 The SO₂ detector was used to ensure that the bubbler was completely 614 saturated with SO₂ before starting any exposure tests. The inlet gas 615 stream is humidified by a single 20-ounce bubbler with a fritted 616 dispersion tube for all humid tests. This stream is assumed to be 617 saturated with water at room temperature (20 °C), yielding a ~2.3% 618 H₂O content and 32% RH at 40 °C. Stainless steel (316 grade) and 619

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620 corrosion-resistant coatings were used in all parts of this apparatus to 621 ensure stability to dry and humid SO₂.

All SO₂ exposure experiments were conducted under 30 ppm of 622 623 SO2 unless otherwise stated. Dry SO2 exposure experiments at 624 concentrations above 30 ppm were conducted using the same system 625 without the outlet SO2 detector due to its incompatibility with the 626 elevated concentrations. Instead, the flow time was calculated based 627 on a 1 or 2 SO2 per diamine ratio using the sample mass, stream 628 concentration, and flow rate (Figure S7).

The SO₂ capacity was calculated from the measured breakthrough 629 630 curves following analysis presented in Section S3 of the Supporting 631 Information.

Diamine Stripping from Diamine-Mg2(dobpdc). Diamine 632 633 molecules were removed from the Mg2(dobpdc) framework by 634 soaking in methanol at room temperature for 2 h, followed by 635 isolation of the framework powder through vacuum filtration. 636 Complete removal of diamines was confirmed by ¹H NMR 637 spectroscopy.

Mass Spectrometry. The recovered methanol solution from 638 639 diamine stripping was analyzed using an HR-LCMS, the Agilent 6545 640 AJS-ESI QToF. A 100% acetonitrile mobile solvent was used with a 641 reverse-phase C18 column at a flow rate of 0.5 mL/min. A 1 μ L 642 sample was injected at 1.25 g/L. The solution recovered from a humid 643 SO2-exposed dmpn-Mg2(dobpdc) sample was analyzed. A control 644 solution was also produced by stripping the diamines from an 645 untreated dmpn-Mg2(dobpdc) sample. The mass spectra of the 646 control and treated solutions were compared using Mass Hunter qualitative analysis. 647

DFT Methods. DFT methods are reported in Section S12 of the 648 649 Supporting Information.

Solid-State NMR Methods. Solid-state NMR methods are 650 651 reported in Section S13 of the Supporting Information. All samples 652 were studied using the natural ¹⁵N isotopic abundance and were not 653 further enriched.

ASSOCIATED CONTENT 654

Supporting Information 655

656 The Supporting Information is available free of charge at 657 https://pubs.acs.org/doi/10.1021/jacs.2c07498.

658 Characterization of Mg2(dobpdc); estimation of SO2 concentration in coal flue gas; SO2 breakthrough 659 apparatus and results; exposure of dmpn-Mg2(dobpdc) 660 to various concentrations of dry SO2; CHNS elemental 661 analysis; evaluation of optimal regeneration conditions 662 for dmpn-Mg2(dobpdc) following humid SO2 expo-663 sure; humid SO2 cycling and extended exposure with 664 dmpn-Mg2(dobpdc); spectroscopic characterization of 665 dmpn-Mg2(dobpdc) before and after humid SO2 666 saturation; characterization of Mg2(dobpdc) following 667 SO2 saturation and diamine stripping; mass spectrom-668 etry; CO2 isobar profiles for diamine-Mg2(dobpdc) 669 obtained after saturation with humid SO2; density 670 functional theory calculations; and solid-state NMR 671 methods (PDF) 672

Computationally generated hypothetical SO₂ product 673 structures used for predicted ¹⁵N NMR calculations 674 (ZIP) 675

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Notes

The authors declare the following competing financial 741 interest(s): The authors declare the following competing 742financial inter-est: J.R.L. has a financial interest in Mosaic 743

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