Climate and plant trait strategies determine tree carbon allocation to leaves and mediate future forest productivity.
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Abstract

Forest leaf area has enormous leverage on the carbon cycle because it mediates both forest productivity and resilience to climate extremes. Despite widespread evidence that trees are capable of adjusting to changes in environment across both space and time through modifying carbon allocation to leaves, many vegetation models use fixed carbon allocation schemes independent of environment, which introduces large uncertainties into predictions of future forest responses to atmospheric CO2 fertilization and anthropogenic climate change. Here, we develop an optimization-based model, whereby tree carbon allocation to leaves is an emergent property of environment and plant hydraulic traits. Using a combination of meta-analysis, observational datasets, and model predictions, we find strong evidence that optimal hydraulic–carbon coupling explains observed patterns in leaf allocation across large environmental and CO2 concentration gradients. Furthermore, testing the sensitivity of leaf allocation strategy to a diversity in hydraulic and economic spectrum physiological traits, we show that plant hydraulic traits in particular have an enormous impact on the global change response of forest leaf area. Our results provide a rigorous theoretical underpinning for improving carbon cycle predictions through advancing model predictions of leaf area, and underscore that tree-level carbon allocation to leaves should be derived from first principles using mechanistic plant hydraulic processes in the next generation of vegetation models.

Keywords
aridity gradient, carbon allocation, climate change, CO2 fertilization, leaf area, plant hydraulic traits, sapwood area, vegetation model

1 | INTRODUCTION

Forest leaf area mediates both terrestrial ecosystem productivity and drought-driven tree mortality during climate extremes (Jump et al., 2017; Myneni et al., 2001; Nemani et al., 2003; Zhu et al., 2016). Tree allocation to leaf area is fundamental to forest climate responses because the water lost through the canopy cannot exceed the water supplied by the sapwood (tree water
transport tissue), thus the ratio of plant leaf area ($A_L$) to sapwood area ($A_S$) constrains whole-plant photosynthesis (Sperry & Love, 2015). During climatic extremes, such as droughts, the increased evaporative demand of the forest canopy relative to supply capacity of the sapwood drives plants to reduce transpiration through stomatal closure to avoid strong tensions and hydraulic failure in the xylem, which can induce metabolic stress, hydraulic damage, and even mortality (Jump et al., 2017; Martinez-Vilalta et al., 2009; Trugman et al., 2018). Thus, the optimal canopy (i.e., tree leaf area) for a tree to support, or the $A_L:A_S$ ratio that maximizes net primary productivity (NPP) (commonly used as a proxy for plant fitness; Franklin et al., 2012), is dependent on local water availability and atmospheric conditions.

Coordination of tree leaf and sapwood area has been observed to be a major physiological mechanism through which trees adjust to changes in water availability and moderate internal plant water stress (Carter & White, 2009; Rosas et al., 2019). Within species, $A_L:A_S$ exhibits greater plasticity across environmental gradients than other common physiological traits that are important for predicting tree water stress, such as the water potential at which 50% loss of stem hydraulic conductance occurs (PS50; Rosas et al., 2019). Observations of coordinated leaf area adjustment with water availability have been extensively documented across plant physiological scales and environmental gradients ranging from hydraulic adjustment of $A_L:A_S$ at the branch- or tree-level (DeLucia, Maherali, & Carey, 2000; Martinez-Vilalta et al., 2009; Mencuccini & Bonosi, 2001; Mencuccini & Grace, 1994; Pinol & Sala, 2000; Rosas et al., 2019), to ecosystem-level trends in forest leaf area due to regional differences in water availability (Baldocchi & Xu, 2007; Eagleson, 1982; Gholz, 1982; Joffre & Ramball, 1993). Importantly, trees are responsive to changes in water supply not just over space but also over annual-scale time periods through changes in both the rate of leaf production and leaf turnover, mechanisms through which trees decrease $A_L:A_S$ with increasing water stress (Limosin et al., 2012).

The concepts of acclimation and adaptation of $A_L:A_S$ to changing climate conditions are crucial when considering future forest productivity because anthropogenic climate change has a strong potential to alter tree allocation to $A_L:A_S$ and forest leaf area, which greatly impact terrestrial ecosystem productivity. However, the sum of the effects of different climate change drivers is unclear. For example, projected increases in atmospheric vapor pressure deficit (VPD) with warmer temperatures (Williams et al., 2012) have the potential to drive decreases in $A_L:A_S$ and forest leaf area. In contrast, increases in plant water use efficiency associated with higher concentrations of atmospheric CO$_2$ may increase the amount of leaf area that forest ecosystems can support under a fixed climatic water supply. Thus, anticipating future forest carbon allocation and terrestrial productivity requires a better understanding of how leaf area adjustment is mediated by plant and environmental factors, including the role of physiological traits that are important to plant water stress (e.g., PS50), and the net effect of VPD and CO$_2$ fertilization as competing environmental drivers.

Here, we used observational data of hydraulic adjustment of $A_L:A_S$ across environmental gradients, a meta-analysis of the CO$_2$ fertilization literature documenting adjustment of $A_L:A_S$, and an optimization-based model of tree gas exchange, hydraulic transport, and carbon allocation (Figure S1; Trugman et al., 2018) to ask: (a) Do geographic patterns of $A_L:A_S$ within- and across species match optimality-based predictions? (b) What environmental factors are most important to variability in $A_L:A_S$? (c) What physiological drivers or functional traits are responsible for variability in $A_L:A_S$? (d) Which functional traits are most important for understanding regional trends in $A_L:A_S$ in response to global environmental change?

## 2 | MATERIALS AND METHODS

### 2.1 | Measured response of $A_L:A_S$ across a climate gradient in Western Australia and Tasmania

We measured $A_L:A_S$ for terminal twigs (cut at the first branching point) for one Acacia and seven different Eucalyptus species (Acacia acuminata, Eucalyptus amygdalina, Corymbia calophylla, Eucalyptus marginata, Eucalyptus salmonophloia, Eucalyptus obliqua, Eucalyptus ovata, Eucalyptus viminalis) across a large aridity gradient (Midsummer (January) VPD ranging from 460 Pa to 2,460 Pa) in Western Australia and Tasmania in October 2014 and February 2016, respectively. For each species, we collected samples from four to five sites, covering as much of their aridity range as possible. Each site comprised of three plots that were located >500 m but generally <5 km apart. At each plot, average $A_L:A_S$ was calculated for five trees based on three branch samples per tree from the sun exposed, north-facing canopy. We looked at how $A_L:A_S$ varied with January VPD using plot location and ~0.86 km$^2$ resolution monthly climate data from the WorldClim Global Climate Data (Fick & Hijmans, 2017).

### 2.2 | Measured response of $A_L:A_S$ in Panamanian seasonally dry tropical forests

Observational measurements for six species with distinct traits and drought phenologies of seasonal leaf area were conducted in two seasonally dry forests in Panama, the Parque Natural Metropolitano (8°59’N, 79°32’W) with annual rainfall of 1,800 mm, and the Eugene Eisenmann Reserve (8°31’N, 79°53’W) with annual rainfall of 1,590 mm (Wolfe, Sperry, & Kursar, 2016). Both are mature secondary forests that experience dry seasons from mid-December to May. Tree species and phenology type included in this study are as follows: Annona hayesii (brevi-deciduous), Genipa americana (deciduous), Bursera simaruba (deciduous), Cavanillesia platanifolia (deciduous), Cojoba rufescens (evergreen), Astronium graveolens (evergreen). Full methodological details regarding observational data collection are available from Wolfe et al. (2016).

### 2.3 | Literature search for the $A_L:A_S$ response to CO$_2$

We compiled data drawing from two meta-analyses of CO$_2$ fertilization experiments (Ainsworth & Long, 2005; Gielen & Ceulemans,
2001), a Google Scholar search of studies documenting the impacts of CO₂ fertilization on woody species, and a CO₂ fertilization experiment of *Pinus halepensis* (Supplementary Methods). Studies that satisfied the following constraints were included: (a) Woody species were included in the experiment; (b) some metric analogous to sapwood area (including basal area, tree diameter at breast height (dbh), or sapwood area) was recorded; and (c) some metric analogous to leaf area (including leaf biomass, LAI, or leaf area) was documented. These criteria rely on the following assumptions: (a) tree sapwood area is linearly related to tree basal area and the sapwood to basal area ratio is invariant with tree size (a simplification that is broadly consistent with reports in the literature; Meir, Goldstein, & Andrade, 2001), (b) leaf area is linearly related to leaf biomass (and this relationship is invariant with tree size), and (c) leaf area can be approximated by LAI. Although these approximations are imperfect, they are necessary given the limited number of studies that document changes in A:S, with CO₂ fertilization. This led to the identification of 11 published studies in addition to our experiment using *Pinus halepensis* that span a CO₂ gradient of 360 ppm to 870 ppm for 19 different species derived from both field and closed chamber experiments on both seedlings and larger trees (Table S1).

We quantified the sensitivity of A:S to CO₂ using linear mixed effects models. We included ΔA:S as the response variable, ΔCO₂ as a fixed effect, and study as a random effect. We computed the relative change ΔA:S as:

\[
\Delta A:S = \frac{(A:S)_e - (A:S)_c}{(A:S)_c}
\]

where "E" signifies A:S at elevated CO₂ and "C" signifies A:S at control CO₂ concentrations (both of which vary by study, Table S1). We experimented with including fixed effects to account for additional treatment type (nitrogen, water stress, ozone tolerance) and angiosperm versus gymnosperm classification; however, these fixed effects were not significant and increased the AIC, so we opted for the most parsimonious model relating ΔA:S to ΔCO₂ (Table S2). We then used results from our mixed effects model analysis to project ΔA:S with an approximate doubling of atmospheric CO₂ concentrations from 400 ppm to 800 ppm. Response coefficients in the mixed effects models were estimated using a maximum likelihood Laplace approximation with the *fitglme* function in MATLAB.

### 2.4 Tree model

We use a simple tree model that couples plant hydraulics to photosynthetic carbon gain (the Hydraulic Optimization Theory for Tree and Ecosystem Resilience or HOTTER model). The HOTTER model uses a single resistor to represent whole-plant hydraulic transport up to the substomatal cavity and a hydraulic optimization-based stomatal conductance model (Trugman et al., 2018; Wolf, Anderegg, & Pacala, 2016). While the model contains some necessary simplifications, it is broadly consistent with the Ohm’s law analogy for hydraulic elements in series and the observed responses of gas exchange to changes in leaf-specific hydraulic conductance (Hubbard, Ryan, Stiller, & Sperry, 2001; Sperry, 2000). HOTTER optimizes A:S to maximize NPP (Figure S1) given the following environmental inputs: VPD (a metric of atmospheric dryness), soil water content, and atmospheric CO₂ concentration. Tree biological parameters that influence whole-plant photosynthesis in HOTTER include hydraulic traits (maximum stem water conductivity, K<sub>max</sub>, and P50), A:S, tree size, as well as a number of other physiological traits (Table S3). HOTTER assumes that photosynthesis is not significantly light or nutrient limited, all leaves experience the same VPD and CO₂, and all fine roots experience the same soil water potential. We provide a description of the HOTTER model below and a full derivation in the Supplementary Information.

Water transport within the soil/plant continuum is represented in HOTTER by the pipe model of a tree (Shinozaki, Yoda, Hozumi, & Kira, 1964). Flow from the soil to the plant roots, stem, and out of the stomata is driven by soil water potential and VPD, and it is regulated by plant physiological traits. Water storage within the plant is not represented, an assumption that recent work suggests is justified for reasonably long recovery timescales (Huang et al., 2017). The flow, F, throughout a plant element is computed by integrating the hydraulic conductivity per unit of xylem area (K) from one end of the pipe continuum with water potential ψ<sub>r</sub> to the other with water potential ψ<sub>L</sub> (Sperry, Adler, Campbell, & Comstock, 1998), which can be expressed by the differences in the Kirchhoff transforms as:

\[
F = a \int_{\psi_r}^{\psi_L} K(\psi) d\psi = a \int_{\psi_1}^{\psi_2} (\phi_2 - \phi_1). \tag{1}
\]

where a is the xylem area of the element and L is the pipe length. The element conductivity (K) decreases as stem water potential falls as a result of embolism. A logistic function is used to represent the loss of conductivity as water potential becomes more negative.

Water flow from the roots to the stem, leaves, and into the atmosphere is modeled as:

\[
\Delta \psi_{\text{root}} (\phi_{\text{soil}} - \phi_{\text{root}}) = \frac{\Delta a_{\text{stem}}}{L_{\text{stem}}} (\phi_{\text{stem}} - \phi_{\text{stem}}) = \frac{\Delta a_{\text{petiole}}}{L_{\text{petiole}}} (\phi_{\text{petiole}} - \phi_{\text{leaf}}) = a_{\text{leaf}} g_s D. \tag{2}
\]

where Δa<sub>stem</sub>, Δa<sub>petiole</sub>, and a<sub>leaf</sub> are the surface area of the tree roots, cross-sectional area of the xylem, and cross-sectional xylem area within a given petiole summed over the tree, and leaf area, respectively. L<sub>root</sub>, L<sub>stem</sub>, and L<sub>petiole</sub> are the path length from the soil to the base of the stem, the tree height, and the length of the petiole, respectively. g<sub>s</sub> is stomatal conductance and D is the VPD. As a simplification, this formulation represents a tree canopy as comprised of a single leaf layer and assumes that under average growth conditions the two limiting photosynthetic rates (i.e., the photosynthetic rate limited by the maximum rate of Rubisco carboxylation and the photosynthetic rate limited by the electron transport rate for the regeneration of ribulose-1,5-bisphosphate) should be equal according to Smith et al. (2019).
Stomatal conductance, \( g_s \), is modeled following a modified version of the Leuning model that incorporates the effect of soil water potential on plant leaf water potential (Leuning, 1995; Wolf et al., 2016),

\[
x_c = \left( \frac{g_s}{A_n} \right) \left( \frac{C_1}{C_2 - \Gamma} \right)^\beta \left( \frac{\psi_L}{\psi_{\text{max}}} \right).
\]

In Equation 3, \( C_2 \) is the atmospheric CO\(_2\) concentration, \( C_1 \), \( D_o \), and \( \Gamma \) are empirical constants from the Leuning model (Table S3), \( A_n \) is net photosynthesis, \( \psi_L \) is leaf water potential. The function \( \beta \) represents the stomatal response to leaf water potential and serves to downregulate photosynthesis under water-stressed conditions. The \( \beta \) function is determined by the carbon cost of sustaining negative water potential and loss of conductivity in the stem. For simplicity we assumed that \( \beta \) increases linearly with the integral of the conductivity of the petiole from the Kirchhoff transform

\[
\beta = \frac{\phi_{\text{leaf}}}{\phi_{\text{max}}},
\]

where \( \phi_{\text{max}} \) is the integral of maximum hydraulic conductivity of the xylem (Table S3). \( \beta \) varies between 1 (leaf at full hydration) and 0 (leaf under full water stress) and captures the monotonic decrease in \( \beta \) with more negative leaf water potentials (Wolf et al., 2016). Here, \( \beta \) broadly conforms to the solution for the Leuning model, but with a more mechanistic representation of soil moisture stress through soil water potential's effect on leaf water potential.

We assume the classic photosynthetic model of CO\(_2\) demand (Farquhar, Caemmerer, & Berry, 1980),

\[
A_n = V_c \left( \frac{C_i - \Gamma^*}{C_i - k_m} - R \right).
\]

In Equation 5, \( C_i \) is the interstitial CO\(_2\) concentration, \( R \) is the daytime leaf respiration, \( V_c \) is an estimate of the effect of absorbed light on maximum rate of carboxylation based on the assumption that the electron transport and Rubisco-limited rates of photosynthesis are colimiting under typical daytime conditions (Smith et al., 2019; Wang et al., 2017; Supplementary Methods). \( \Gamma^* \) is the CO\(_2\) compensation point in the absence of mitochondrial respiration, and \( k_m \) is the Michaelis constant (ppm) for the Farquhar model (Farquhar et al., 1980). We rewrite Fick’s Law for diffusion of CO\(_2\) from the atmosphere through the stomata in terms of \( x_c \) using Equation 3 as

\[
x_c = \left( C_2 - \frac{1}{\psi_{\text{L}}} \right).
\]

Given the solution for \( x_c \) and the expression for photosynthesis in terms of \( x_c \) (see full derivation in the Supplementary Information), we relate \( A_n \) to tree size, functional leaf and xylem biomass, and hydrological and atmospheric drivers. Finally, we calculate NPP for the whole plant (including all respiration costs) as

\[
\text{NPP} = (1 - \epsilon) \left( A_n - R_{\text{root}} - R_{\text{phloem}} - R_{\text{xylem}} - R_{\text{dark}} \right).
\]

In Equation 7, NPP includes growth and maintenance respiration costs, \( \epsilon \) is the growth respiration fraction, \( R_{\text{root}} \) is the root respiration, \( R_{\text{phloem}} \) is the combined respiration rate of the phloem and cambium (which remains proportional to tree size regardless of drought-induced hydraulic damage to the xylem), \( R_{\text{xylem}} \) is the respiration rate of the xylem (which is proportional to functional xylem biomass and decreases with decreased functional xylem resulting from drought-induced hydraulic damage), and \( R_{\text{dark}} \) is the dark respiration rate of the leaves. Although growth respiration can vary appreciably with age (Mäkelä & Valentine, 2001), we treat it as constant in the interest of parsimony and in accordance with a number of other widely utilized vegetation models (Medvigy, Wofsy, Munger, Hollinger, & Moorcroft, 2009).

### 2.5 HOTTER Model predictions for the \( A_L: A_S \) response to environmental conditions

We compared trends in model-predicted \( A_L: A_S \) that maximized NPP (Figure S1) along an atmospheric moisture gradient to observed trends in \( A_L: A_S \) across Western Australia and Tasmania. HOTTER plant traits were representative of Australian species (Table S3 note). Atmospheric CO\(_2\) (=400 ppm) and soil water potential (≈1 MPa) were kept constant (due to a lack of site-specific soil moisture data). We varied VPD across environmentally relevant values ranging from 500 Pa to 2,500 Pa. In this comparison, we assumed that plants behave as pipe models such that the \( A_L: A_S \) is conserved along the height of the tree (Shinozaki et al., 1964). Although a better understanding of \( A_L: A_S \) variability within trees would allow for more rigorous scaling techniques from branch-level measurements to trees, data are currently limiting, and a number of studies suggest that our scaling assumption is reasonable (Mencuccini, Manzoni, & Christoffersen, 2018).

Next, we compared measurements of observed leaf area for six species with distinct traits and drought phenologies documented by Wolfe et al. (2016) with model-predicted optimal leaf areas that maximized NPP, given site-specific climate and species-specific hydraulic and photosynthesis traits (Figure S2, Table S4). Climate data used to force HOTTER were derived from site-specific daily mean VPD. We applied a smoothing low-pass filter to VPD spanning 10% of the measurement period to avoid the impacts of daily variability in VPD because we were interested in capturing the effects of seasonal trends in water availability on predicted changes in \( A_L \). Site-specific soil moisture data were not available. However, predawn leaf water potentials are generally representative of soil water potentials and were measured by Wolfe et al. (2016). Thus, we reconstructed soil moisture by gap-filling mean predawn leaf water potential measurements at each site across all species except for Bursera simaruba and Cavanillesia platanifolia (which disconnect from the soil before the soil dries, and so do not necessarily have predawn leaf water potentials that are representative of the soil water content; Wolfe, 2017) to obtain daily-level soil moisture forcing datasets for HOTTER model predictions for both the Parque Natural Metropolitano and the Eugene Eisenmann Reserve. Because measurements for canopy CO\(_2\) were
not available, we assumed a constant \(CO_2 = 400\) ppm. Model predictions of the seasonal dynamics of relative \(A_L\) (calculated as current \(A_L\) relative to maximum \(A_L\) during the 2011–2013 measurement period) given site-specific climate and species-specific traits were compared to observed seasonal dynamics of relative \(A_L\) for each species at each site.

Finally, we used HOTTER to predict trends in \(A_L:A_S\) that maximized NPP along a \(CO_2\) gradient ranging from preindustrial levels (280 ppm) to an approximate doubling of current \(CO_2\) concentrations (800 ppm). We assumed an 8% depression in SLA for trees exposed to elevated \(CO_2\) concentrations in accordance with Ainsworth and Long (2005). All other plant traits (Table S3), atmospheric VPD (\(=1.200\) Pa), and soil water potential (\(=-1\) MPa) were kept constant. We then compared the model-predicted sensitivity of \(A_L:A_S\) to the sensitivity of the meta-analysis observed \(A_L:A_S\) by projecting the change in \(A_L:A_S\) with a doubling in \(CO_2\) concentrations from 400 ppm to 800 ppm using both the mixed effects model response coefficient (Table S2) and the HOTTER model.

2.6 | HOTTER estimates of the sensitivity of \(A_L:A_S\) to climate change and global variation in traits

To understand the sensitivity of \(A_L:A_S\) to changes in VPD and \(CO_2\) with anthropogenic climate change, we ran factorial simulations forced with VPD, soil moisture, and atmospheric \(CO_2\) concentrations. We derived model forcing from the Coupled Model Intercomparison Project Phase 5 (CMIP5) for the steepest \(CO_2\) emissions scenario, the Representative Concentration Pathway (RCP) 8.5. We ran simulations using average climate conditions over historical (1981–2000) and future (2080–2099) climates (Figure S3). Factorial simulations were as follows: (a) historical soil moisture and VPD and globally constant \(CO_2 = 370\) ppm; (b) historical soil moisture, constant \(CO_2 = 370\) ppm, and RCP 8.5 projected VPD; (c) historical soil moisture and VPD and RCP 8.5 projected globally constant \(CO_2 = 925\) ppm; and (d) RCP 8.5 projected soil moisture and VPD and globally constant \(CO_2 = 925\) ppm. We found the soil moisture effect between historical and RCP 8.5 to be minimal compared to the VPD and \(CO_2\) effects and thus did not include it in our analysis. We used the multipanel median soil water potential and atmospheric VPD to understand the sensitivity of the HOTTER model allocation predictions to predicted changes in mean climate (Figure S3). We assumed that tree size remained constant between current and future projections and initialized the model with the tree size dataset from Simard, Pinto, Fisher, & Baccini (2011). We converted tree height to dbh (required for the idealized tree model input) assuming the allometric relationship from Table S3. Overall, these simulations were designed to understand the sensitivity of \(A_L:A_S\) to mean changes in VPD and \(CO_2\) rather than to predict absolute changes in forest leaf area globally, given assumptions of constant tree height and the lack of competition in the HOTTER model.

To understand the sensitivity of \(A_L:A_S\) to plant trait strategy and how trait strategy interacts with climate, we varied biome-specific traits globally including specific leaf area (SLA), maximum stem water conductivity (\(K_{\text{max}}\)), and P50 based on Anderegg, (2015) and Oleson et al. (2010) (Table 55). \(K_{\text{max}}\) was not included in these databases so we varied \(K_{\text{max}}\) proportionally with \(V_{c,\text{max}}\). All other traits remained constant (Table S3). We then used a 0.5° resolution MODIS land cover map from year 2000 (http://glcf.umd.edu/data/lc/) upscaled to 1° for the purposes of determining grid cell-specific traits (Channan, Collins, & Emanuel, 2014; Friedl et al., 2010). We included all forest types in our simulations with possible woody plant cover including Evergreen Needleleaf forest, Evergreen Broadleaf forest, Deciduous Needleleaf forest, Deciduous Broadleaf forest, Mixed forest, Closed shrubland, Open shrubland, Woody savanna, Savanna (code values 1–9). For grid cells of mixed forest type, we assumed each species type occupied 50% of the grid area and ran two simulations examining the sensitivity of \(A_L:A_S\) one with each species trait type and took the average \(A_L:A_S\) prediction derived from these two simulations. We ran an additional set of five simulations, forced the model with RCP 8.5 projected VPD, soil moisture, and \(CO_2\) to quantify the relative importance of SLA, \(V_{c,\text{max}}, K_{\text{max}}\), and P50 in determining \(A_L:A_S\). We held fixed one trait, either P50, SLA, \(V_{c,\text{max}}\) or \(K_{\text{max}}\), and varied all others. We then quantified the trait effect on \(\Delta A_L:A_S\) by looking at the percent difference between the fixed trait simulation minus the globally varying trait simulation quantity divided by the globally varying trait simulation. These simulations were designed to understand the relative importance of different physiological traits on influencing tree \(A_L:A_S\) and to place this sensitivity in the context of spatial variation in global climate. Although many of these physiological traits covary in reality due to physiological trade-offs (and this covariation would impact the sensitivity of modeled \(A_L:A_S\)), this modeling analysis allowed us to isolate individual trait effects in a manner that would be impossible to do in the field.

3 | RESULTS

The optimality-based HOTTER model predicted a negative relationship between \(A_L:A_S\) and increased water stress (Figure 1a). Thus, given an expected atmospheric dryness or soil water availability, the optimal strategy for a tree is to decrease \(A_L:A_S\) as water availability decreases. Indeed, the HOTTER model captured the observed decrease in \(A_L:A_S\) and the rate of change in \(A_L:A_S\) with increasing VPD across a large climate gradient in Australia and Tasmania (ranging from 300 to 1,475 mm in mean annual precipitation and 460 to 2,460 Pa in midsummer VPD). As VPD increased, allocation to \(A_S\) relative to \(A_L\) decreased asymptotically. At low to moderate VPD levels, this corresponded to a rapid downregulation of \(A_L:A_S\) with increased VPD. However, the adjustment in \(A_L:A_S\) slowed at more extreme VPD stresses. Interestingly, substantial intraspecific variability in \(A_L:A_S\) existed in the observations at a fixed VPD that was not predicted by the HOTTER model, potentially due to a disconnect between VPD and soil water availability at different sites (Novick et al., 2016). Given a lack of site-specific soil moisture, it was not possible to test how access to groundwater impacted observed \(A_L:A_S\) responses or incorporate soil water
responses in HOTTER predictions. However, site-specific VPD alone was sufficient to predict broad allocational trends in $A_{L:S}$ across species.

Plant hydraulic traits influenced the optimal $A_{L:S}$. Specifically, plants with a more resistant xylem, corresponding to a more negative P50, were able to support more leaf area and maintain photosynthesis at all VPD levels (Figure 1b). Similarly, plants with a high maximum stem water conductivity ($K_{max}$) were able to support more leaf area at all VPD levels (Figure 1c) because higher conductivities decreased internal water stress on plant tissues, provided sufficient soil water is available to maintain a higher conductivity. Although hydraulic traits were integral in predicting the optimal $A_{L:S}$ with a given water availability, the $A_{L:S}$ response to increasing VPD (i.e., the shape of the curve) was robust regardless of plant trait strategy (Figure 1b,c).

We then considered whether optimality theory could predict seasonal variations in $A_{L:S}$ given intra-annual variations in environmental conditions. Seasonally dry tropical forests provide the ideal biome in which to test the model because plant hydraulic traits are informative of temporal variations in $A_{L:S}$ (i.e., evergreen or deciduous phenological strategy). Indeed, when we compared the observed seasonal phenology of six different tropical dry forest tree species to HOTTER, parameterized with species-specific hydraulic traits and forced with observed daily-level VPD and soil moisture, the model predicted the phenology of drought deciduous trees (and to an extent the phenology of brevi-deciduous trees) by allocating to leaves to maximize NPP (Figure 2a–d). In contrast, HOTTER was not able to predict the seasonal leaf phenology of evergreen species. However, hydraulic traits alone were predictive of some leaf retention in the evergreen species (which are more tolerant to decreased water availability) during the dry season (Figure 2e,f), despite the lack of light-driven competition dynamics in HOTTER that are influential in tropical dry forest evergreen phenology.

In addition to water availability, $A_{L:S}$ has been shown to be sensitive to increased atmospheric CO$_2$ concentrations (see Tables S1 and S2 for studies included in this meta-analysis) because high atmospheric CO$_2$ increases tree water use efficiency, potentially increasing $A_{L:S}$ for a fixed climatic water availability. HOTTER predictions showed an increase in optimal $A_{L:S}$ by ~70% with a doubling of CO$_2$ from 400 to 800 ppm, with all other climate conditions and plant traits held constant (Figure 3). We performed a meta-analysis of CO$_2$ fertilization studies derived from field and closed chamber experiments that are representative of both seedlings and larger trees and used mixed effects models to project $\Delta A_{L:S}$ with an approximate doubling of atmospheric CO$_2$ concentrations from 400 to 800 ppm. The model-predicted fertilization response fell well within the meta-analysis observed response range (Figure 3b; Tables S1 and S2).

Taken together, the optimal allocation model HOTTER appears to explain plant allocation patterns across both space—among and within species—and time and in response to the critical drivers of water availability and CO$_2$ concentrations. Thus, we applied HOTTER to understand the sensitivity of plant carbon allocation to both future climate conditions and physiological trait strategies. HOTTER predicted that the increased water use efficiency associated with strong atmospheric CO$_2$ fertilization will likely outweigh the increased water stress associated with predicted mean increases in VPD, resulting in a potential increase in $A_{L:S}$ under mean climate conditions circa 2100 for most locations around the globe (Figure 4). Notable exceptions to this mean climate response included portions of Brazil in
South America due to the large predicted increase in VPD in a region comprised predominantly of wet-adapted species (Figures 4c and 5).

Of the traits tested, $A_L:As$ responses were most sensitive to plant hydraulic traits. Compared to traits including SLA, $V_{c,max}$ and $K_{max}$, interspecific variation in P50 resulted in the largest change in allocation strategy to $A_L:As$ between predictions where P50 was held at a constant global mean value (Materials and Methods) versus predictions where P50 varied by biome and vegetation type (Figure 5). The P50 effect was strongest in wet tropical forests where model predictions with a constant P50 predicted a much larger $A_L:As$ compared to the predictions with biome-specific traits due to an underestimation of the hydraulic vulnerability of moist tropical forests in the fixed trait scenario.

4 | DISCUSSION

This study combines multiple observational datasets of leaf area adjustment along temporal and spatial environmental gradients
in diverse biomes, a meta-analysis of the CO₂ fertilization effects on leaf area, and a tractable, rigorous model of optimal plant carbon allocation based on known physiological mechanisms of hydraulic gas exchange coupling. We find strong evidence that observed allocation strategies to leaf area over time and space and across multiple biomes can be explained by our optimality approach. With projected increases of atmospheric CO₂ and VPD, our optimality approach suggests that, for mean conditions, increased water use efficiency associated with increasing atmospheric CO₂ concentrations overcompensates for increased water stress associated with higher atmospheric VPD, potentially driving increases in leaf area globally. Exceptions include parts of wet tropical forests where strong VPD stress combined with wet-adapted plant hydraulic trait strategies inhibits increases in A₁:Aₛ. From a theoretical/first principles perspective, the moderating effect that increased atmospheric CO₂ concentrations have on maintaining or increasing A₁:Aₛ in spite of increased VPD are not unexpected because increased atmospheric CO₂ increases plant water use efficiency and decrease water demand per unit leaf area, so there is significant potential for trees to support a greater leaf area with less sapwood area, even with substantial temperature-driven increases in atmospheric VPD.

By evaluating the sensitivity of these allocation trends to other plant traits, we find that hydraulic traits in particular have an enormous impact on modeled forest leaf area globally. Given that allocation strategy to A₁:Aₛ is most sensitive to plant hydraulic traits, and hydraulic traits have been shown to be critical in explaining both tree productivity and mortality responses during drought (Anderegg et al., 2016, 2018), these results further emphasize the importance of simulating A₁:Aₛ not as a fixed allometric trait in large-scale vegetation models, but rather as an adaptive property of plant traits and environment.

A potentially moderating factor to the predicted increase in A₁:Aₛ is increased light limitation with growth (Luo et al., 2004), a process which is not currently included in HOTTER. Specifically, if vegetation productivity and leaf area index (LAI) increases, more competition for light could lead to more allocation to stems to outgrow neighboring competitors, particularly in light-limited regions such as the tropics. Furthermore, competition for water and nutrients may stimulate increased root allocation relative to leaves. Although HOTTER did not explain the lack of intra-annual variation in A₁:Aₛ in the tropical dry forest evergreen species (Figure 2e,f), HOTTER predicted that the more resistant xylem found in the evergreen species (realized through xylem vulnerability curves in the model) supplied sufficient water to the tree crown to merit some leaf retention during the dry season. In the observations, the absence of any leaf shedding in the evergreen trees likely reflects a strategy that capitalizes on lower competition for light during the dry season (Detto, Wright, Calderón, & Muller-Landau, 2018; Wright & van Schaik, 1994), and the additional nutrient and carbon constraints associated with growing new leaves on an annual basis. Given that light competition and nutrient constraints are not processes directly incorporated in the HOTTER model (see Materials and Methods), it is not surprising that HOTTER did a poor job predicting seasonal variations in evergreen A₁:Aₛ. However, these results highlight that plant hydraulic traits are critical to the tropical dry forest evergreen tree strategy that capitalizes on dry season productivity.

Finally, additional experimental studies targeting CO₂-driven changes in A₁:Aₛ and the scaling of A₁:Aₛ from branch to tree to ecosystem are needed to refine our understanding of tree structural allocation strategies. Despite these caveats, the insights from HOTTER scale tissue-level properties to tree-level photosynthetic responses and trade-offs, providing an important mechanistic foundation to understanding how plant traits and environment jointly constrain carbon allocation strategy.

The predicted allocational sensitivity, A₁:Aₛ, due to changes in mean climate and increased atmospheric CO₂ has important implications for forest productivity. Specifically, fixed allocation strategies used in many vegetation models that are independent of climate (De Kauwe et al., 2014) would tend to underestimate both future productivity (through suboptimal allocation to A₁:Aₛ, e.g., Figure S1), and potentially ecosystem vulnerability to catastrophic mortality events. While it is unclear how these two compensating processes will interact to influence the accuracy of the magnitude

**FIGURE 4** (a) Projected percentage change in canopy leaf area relative to tree sapwood area (A₁:Aₛ) for the 2080–2099 climatological mean relative to 1981–2000. (b) Projected percentage change in canopy leaf area relative to tree sapwood area due to CO₂ fertilization alone. (c) Projected percentage change in canopy leaf area relative to tree sapwood area due to increased vapor pressure deficit (VPD) alone.
and sign of projections for the terrestrial carbon sink, it is clear that hydraulic constraints are a fundamental process governing plant carbon allocation strategy (Figures 1 and 5), productivity (Anderegg et al., 2018), and mortality (Jump et al., 2017; Trugman et al., 2018).

Here, we explicitly couple plant hydraulics with carbon metabolism to demonstrate that hydraulic functional traits and environment are prognostic of tree carbon allocation strategy. Given that vegetation models are now largely functional trait based (Fisher et al., 2018), and substantial advances in vegetation models are being made so that now many regional-scale models resolve plant hydraulics (Kennedy et al., 2019; Xu, Medvigy, Powers, Becknell, & Guan, 2016), it is timely and feasible to implement dynamic carbon allocation schemes into these models. Our optimization-based model provides the mechanistic underpinning motivating (a) explicit representation of plant hydraulics in vegetation models and (b) the representation of allocation to \( A_L : A_S \) as an emergent property of environment and plant hydraulic traits.

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**CONFLICT OF INTEREST**

The authors declare no conflict of interest.

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SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section at the end of the article.

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