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24 Abstract:

As one of the major agricultural production areas in the world, the United States (U.S.) 25 26 Midwest plays a vital role in the global food supply and agricultural ecosystem services. Although significant efforts have been made in modeling the carbon cycle dynamics over this area, large 27 uncertainty still exists in the previous simulations in terms of reproducing individual components 28 29 of the carbon cycle and their responses to environmental variability. Here we evaluated the performance of an advanced agroecosystem model, *ecosys*, in simulating carbon budgets over the 30 31 U.S. Midwest, considering both the magnitude of carbon flux/yield and its response to the 32 environmental (climatic and soil) variability. We conducted model simulations and evaluations at 7 cropland eddy-covariance sites as well as over 293 counties of Illinois, Indiana, and Iowa in the 33 U.S. Midwest. The site-level simulations showed that *ecosys* captured both the magnitude and 34 seasonal patterns of carbon fluxes (i.e., net ecosystem carbon exchange, ecosystem gross primary 35 production (GPP), and ecosystem respiration), leaf area index, and dynamic plant carbon allocation 36 processes, with R² equal to 0.92, 0.87, 0.87, and 0.78 for GPP, NEE, Reco, and LAI, respectively 37 across all the sites compared with the observations. For regional scale simulations, ecosys 38 reproduced the spatial distribution and interannual variability of corn and soybean yields with the 39 constraints of observed yields and a new remotely sensed GPP product, with R² of multi-year 40 averaged simulated and observed yield equal 0.83 and 0.80 for corn and soybean, respectively. 41 42 The simulated responses of carbon cycle dynamics to environmental variability were consistent 43 with that from the empirical observations at both site and regional scales. Our results demonstrated 44 the applicability of *ecosys* in simulating the carbon cycle and soil carbon sequestration of the U.S. 45 Midwestern agroecosystems under different climate and soil conditions.

Keywords: carbon fluxes, crop yield, *ecosys*, agroecosystems, environmental variabilities, U.S.
Midwest

48

49 1. Introduction

The terrestrial carbon balance of agroecosystems plays an important role in the global 50 51 carbon cycle (Dold et al., 2017; Verma et al., 2005). Depending on the temporal and spatial scales used for accounting as well as the geographical regions, croplands can be either carbon sinks or 52 53 sources for the atmospheric CO₂ (Blanco-Canqui and Lal, 2004; Kimble et al., 1998). In the U.S. Midwest, about 30-50% of soil organic carbon (SOC) has been lost when compared with that 54 before cultivation for most croplands (Lal, 2002). Since SOC content is often positively related to 55 soil fertility, SOC loss may enhance crop yield loss risk under future climate conditions (Lal, 2011, 56 57 2004, 2001). Fortunately, with recommended management practices (RMPs, i.e., conservation tillage, cover crops, and biosolids and manure, etc.), prior studies show that U.S. croplands have 58 the potential to sequester about 75-208 Tg C/year, which may recover 50-70% of the depleted soil 59 carbon (Jarecki and Lal, 2003; Lal, 2011, 2007, 2002; Meena et al., 2020; Hutchinson et al., 2007; 60 Chambers et al., 2016). Hence, in order to help realize this carbon sequestration potential in U.S. 61 croplands, and meanwhile ensure global food security, it is critical to accurately quantify the 62 carbon balance of agroecosystems, including carbon fixation and emission. 63

The carbon inventory (West et al., 2013, 2010, 2008; West and Marland, 2002), derived from crop yield survey reports (Vogel, 2018), SOC measurements (van Wesemael et al., 2010), and observation-based gross primary production (GPP) estimations (Jiang et al., 2021), can provide several components of cropland carbon budget. Among these carbon inventory methods, soil-sampling-based SOC measurement is the most direct approach to investigate SOC change, but

it still has uncertainties associated with soil sampling strategies (i.e., sampling location, depth, and 69 time), measurement methods, and duration of measurements(Jandl et al., 2014; Schrumpf et al., 70 71 2011; VandenBygaart and Angers, 2006). More importantly, it is difficult to scale up the soil sampling due to its high labor and financial costs. In the framework of carbon balance, SOC change 72 can in principle be derived from the whole carbon mass balance, which requires different carbon 73 74 cycling components (i.e., GPP, ecosystem respiration, and harvest etc.). However, most of the carbon inventories cover only part of the carbon cycling components, such as agroecosystem 75 76 carbon input (i.e., GPP) or outputs (i.e., yield). Measurements of other key carbon cycling 77 components (i.e., respiration and litterfall) of agroecosystems are still difficult and insufficient, especially at large scales (e.g. U.S. Midwest) (Osborne et al., 2010). All these factors limit wide 78 and robust applications of carbon inventory to quantify agroecosystem carbon budgets and SOC 79 change. 80

Alternatively, we can use process-based models to quantify the cropland carbon budget 81 82 (Brilli et al., 2017; Huang et al., 2009; Wattenbach et al., 2010; Zhang et al., 2015). However, existing studies using process-based models for cropland carbon quantification have suffered from 83 one or few of the following limitations. First, very few model-based quantifications of the cropland 84 85 carbon budget have gone through rigorous model validation covering the whole agroecosystem carbon cycle (i.e., carbon fixation, carbon allocation, and respiration), especially at regional scales. 86 87 Most process-based modeling studies for agroecosystems evaluated and constrained their models 88 with a limited number of observational variables, such as crop yield (Gilhespy et al., 2014; Stehfest 89 et al., 2007) and/or measured SOC (Li et al., 1997; Liu et al., 2006; Shirato, 2005). This lack of 90 sufficient model constraint may cause simulations to be apparently right with wrong reasons (Peng 91 et al., 2018). For example, models can generate the same crop yield with higher carbon fixation

but lower harvest index compared to the correct ones, because errors in plant carbon fixation can
be reconciled by unconstrained fluxes of respiration and litterfall. Therefore, to ensure that the
model simulates carbon emission and sequestration correctly in both short and long terms, we need
to use more carbon-related observations with fine temporal resolution (i.e., daily GPP, NEE, Reco,
LAI, plant carbon allocation, and phenology) to sufficiently constrain and validate the carbon
cycling processes of the models.

Second, most existing model-based studies only calibrated and validated the models at a 98 99 few specific sites due to limited availability of observations. In general, models involve both 100 parameters that are site-specific (i.e., maturity group and climate zone) and parameters that are shared among sites at a regional scale (i.e., parameters controlling the temperature responses of 101 activity of RuBP carboxylase-oxygenase) (Kuppel et al., 2012; Mäkelä et al., 2007). Thus 102 optimizing a model at specific sites will tie the resultant model parameterization closely to the site 103 104 information (e.g. climate, soil, groundwater depth, field microtopography, and land management 105 practices etc.), so that the model may not be suitable for other sites and regions with different soil and climate conditions. To ensure the model parameterization can be robustly transferred to other 106 sites or regions, systematic evaluations are needed. Specifically, we need to constrain and evaluate 107 108 models under a wide range of soil and climate conditions, using diverse data such as large-scale carbon inventories (e.g. crop yield reports and crop progress reports) and satellite remote sensing 109 110 carbon-related observations (e.g. GPP and LAI).

Finally, current model calibrations and validations have generally focused on matching the magnitude or time series of the target variables (e.g., GPP and yield) (Gurung et. al., 2020; Wang et al., 2020; Jin et al., 2017), which is achieved by minimizing a cost function (which measures model-data discrepancy) that does not take into account the relationship between these target

variables and environmental drivers. From the perspective of Bayesian inference (Tarantola, 2013), 115 since only uncertainties in model parameters are constrained, such a practice leads to an 116 117 underestimation of the prior information associated with the environmental drivers. To make a more comprehensive use of the information contained in observations and model driving variables, 118 as well as to deliver more confident predictions of how agroecosystems will respond to 119 120 environment changes, we thus further need to verify relationships between environment variables and model predicted variables to test whether the model can simulate emergent responses of those 121 122 variables to environmental factors from empirical observations (Peng et al., 2020). The accurate 123 representation of the response of the target variables to environmental factors (i.e., climate variability and soil conditions) will help expand the models to broader soil and climate conditions. 124

Based on the above rationale, to demonstrate a new standard to achieve a comprehensive 125 constraint and evaluation of an agroecosystem simulator, in this study we used an advanced 126 127 ecosystem model, *ecosys*, to simulate surface carbon fluxes and corn/soybean yield in the U.S. 128 Midwest at both eddy-covariance sites and county scales for the three I states (Illinois, Iowa and Indiana). As one of the world's largest crop production areas, the U.S. Midwest produces about 129 85% of U.S. corn and soybean (USDA, 2020). The soil health and crop yield of the U.S. Midwest 130 131 in the future is vital to the global food supply and agricultural ecosystem services. To improve the quantification of carbon cycle dynamics in the U.S. Midwest, both the absolute values of the 132 133 simulated carbon fluxes and yield as well as the responses of those variables to the environmental 134 variabilities were evaluated. Through the evaluations, we aim to evaluate the capability of ecosys in conducting spatiotemporal extrapolations of agroecosystem carbon cycle by addressing the 135 136 following two questions: (1) To what extent can ecosys simulate agroecosystem carbon dynamics 137 at different individual sites as well as across the broader regions in the U.S. Midwest? (2) How

well can *ecosys* capture responses of carbon fluxes and crop yield to environmental variabilities?
Although we use *ecosys* as an example, the procedures for model evaluation described in this study
are applicable to many other agroecosystem models.

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142 **2. Data and method**

143 **2.1 The process-based model** *ecosys*

Ecosys is an advanced mechanistic ecosystem model developed to simulate water, energy, carbon, and nutrient cycles simultaneously for various ecosystems, including agroecosystems at the hourly step (Figure 1a) (Grant, 2001). It is one of the very few models that are formulated primarily based on biophysical and biochemical principles, with fully connected balances and interactions for water, energy, carbon and nutrient cycles in the soil-plant-atmosphere continuum, and has been extensively validated in various ecosystems ranging from agricultural (Grant et al., 2007, 2011; Mezbahuddin et al., 2020) to forest systems (Grant et al., 2010, 2006).

151 The *ecosys* model was built based on the strategy that pursues the mechanistic representations and model outputs as directly comparable to observations as possible to 152 realistically inform agricultural practices, by combining reactive transport modeling and state of 153 154 the art knowledge of biogeochemistry (Grant, 2001). For example, photosynthesis and plant hydraulics in *ecosys* are coupled through leaf osmotic pressure, and then turgor pressure and leaf 155 156 water potential that is linked to stomatal conductance (Grant, 1995; Grant and Flanagan, 2007), 157 rather than empirical stress functions (Van den Hoof et al., 2011; Liu et al., 2016; Yokohata et al., 158 2020), and all of which can be measured in the field (Salmon et al., 2020; Shekoofa et al., 2021; Xue et al., 2021). As it integrates the plant hydraulics closely with the plant photosynthesis (Grant 159 160 et al., 1999), the plant stomata conductance in ecosys is directly controlled by the balance between

photosynthetic carbon assimilation and plant water hydraulics calculated for the soil-plant-161 atmosphere continuum, which can properly resolve the plant response to drought (Mekonnen et 162 163 al., 2017). Due to the explicit simulation of plant hydraulic impacts on stomatal conductance, the empirical crop response to atmospheric vapor pressure deficit does not need be prescribed as in 164 many other models (Van den Hoof et al., 2011; Liu et al., 2016; Yokohata et al., 2020). In response 165 166 to soil water and plant carbon stress, ecosys also dynamically adjusts the plants' carbon and nutrient allocation strategies (Grant et al., 2001a), so that all plant organs will balance their 167 168 respective growth to help the plants survive the harsh growth conditions and flourish under 169 favorable conditions. In addition, the plant carbon and nutrient allocation is represented following the source-storage-sink balance approach, rather than the fixed allometric relationship approach 170 adopted by most existing models (Grant, 1989b; Drewniak et al., 2013; Liu et al., 2016). 171

Moreover, ecosys employs much more complete physical and chemistry theories in 172 simulating soil related processes. Specifically, ecosys mechanistically resolves the oxygen stress 173 174 throughout the soil and plant roots (Grant, 1998), such that a flood condition will suppress plants' growth and alter the soil carbon and nutrient cycling. In addition, ecosys explicitly includes 175 microbes' competitive and symbiotic nutrient interactions with plants (Grant and Pattey, 2003; 176 177 Grant et al., 2006; Grant and Pattey, 2008; Grant et al., 2016), enabling a nutrient-based analysis of how various management practices could affect plant productivity. Meanwhile, soil organic 178 179 carbon dynamics in *ecosys* are driven explicitly by microbial community dynamics that emerge 180 from the interactions between bacteria and fungi, and another five functional groups carrying out 181 fermentation, methane and nitrogen cycling (Grant, 2013; Grant and Rochette, 1994). Emergent 182 microbial population structure, e.g. bacteria to fungi ratio, can be directly evaluated with respect 183 to field measurements (Anderson and Domsch, 1975; Bardgett and McAlister, 1999). Moreover,

the partitioning of soil carbon in *ecosys* is amenable to the density fractionation that is often used 184 by empiricists to characterize soil organic matter. In addition, *ecosys* outputs profiles and fluxes 185 186 of many easily measurable chemicals, including different phase existences of CO₂, CH₄, N₂O, NH₃, NO₃, HPO₄⁽²⁻⁾, etc. Finally, *ecosys* resolves many common agricultural practices, such as mixed 187 cropping, depth dependent irrigation and tillage (Grant, 1997), banded vs broadcast fertilization 188 189 (Grant et al., 2001b), soil liming, manure application (Grant et al., 2001c), denitrification inhibitor (Grant et al., 2020), and tile-drainage system (Mezbahuddin et al., 2017) etc. Finally, ecosys 190 191 generally requires no calibration for the soil and hydrological processes due to its complete 192 mechanistics thus provides scalability to regional scale applications (Grant et al., 2012). All these features make ecosys stand out as an unique simulator as compared to many other models that tend 193 to lump processes into simplified representations. We here refer detailed information about the 194 processes represented in *ecosys* to the supplement of Grant et al. (2019), and the code of *ecosys* 195 can be obtained from the online repository (https://github.com/jinyun1tang/ECOSYS). Below we 196 197 only describe major carbon cycling processes of agroecosystems simulated in ecosys (Eq. 1 and 198 2).

$$-NEE = GPP - Reco = GPP - (Ra + Rh) = GPP - ((Rm + Rg) + Rh)$$
(Eq. 1)

200

$$NBP = -NEE - Yield - \varepsilon$$
 (Eq. 2)

where NEE is net ecosystem exchange, GPP is gross primary production, Ra is ecosystem autotrophic respiration, Rh is ecosystem heterotrophic respiration, Reco is ecosystem respiration,
Rm and Rg are plant maintenance and growth respiration, NBP is net biome productivity, Yield is harvested crop yield, and ε is the carbon losses caused by disturbances (e.g., fire) excluding harvest.
In *ecosys*, the change of SOC (△SOC) is equal to the difference between plant litter fall,
Rh, and ecosystem carbon leakage, including CH4 emission, dissolved organic (DOC) and

207	inorganic carbon (DIC) lenching, etc (Eq. 3a). For annual cropping system in most of the U.S.
208	Corn Belt regions, we can use NBP to approximate \triangle SOC at long term scales (\geq annual scale). By
209	using Eq. 3, most part of simulated cropland soil carbon balance can be directly backuped with the
210	eddy covariance measurements or carbon inventory data, which provided another approach to
211	evaluate and verify the model performance in carbon budget estimations (Baker and Griffis, 2005).
212	$\triangle SOC = Litter_Fall - Rh - \varepsilon $ (Eq. 3a)
213	= (GPP-Ra-Yield + Seed_C) - Rh – ε (≥annual scale for annual cropping systems)
214	(Eq. 3b)
215	= -NEE + Seed_C - Yield - ε = NBP + Seed_C - ε (\geq annual scale for annual cropping
216	systems) (Eq. 3c)
217	where Litter_Fall is the litter fall from plants, including leaf senescence, harvest residue, and root

carbon exudation, Seed_C is the seed mass at planting, ε is the carbon leakage through CH₄ emission, and DOC and DIC are leaching terms.



220

Figure 1. (a) Major processes represented in the *ecosys* model (revised from (Grant, 2004)), and

(b) locations of the seven flux towers and the three I states in the U.S. Midwest.

223 2.1.1 Photosynthesis (GPP)

The *ecosys* model uses a multiple-layer canopy module to simulate canopy light absorption 224 and carbon assimilation (Grant et al., 1989). Photosynthesis of each individual leaf is calculated 225 independently using the Farquhar model for C3 plants and explicitly considering the mesophyll-226 bundle sheath carbon exchange for C4 plants at hourly time step (Farquhar et al., 1980; Grant, 227 1989a) with specific azimuth, leaf inclination, exposure of light conditions (i.e., sunlit and shaded 228 229 leaves), and canopy height. The canopy stomatal resistance (r_c) is controlled by canopy turgor potential ($\psi_t = \psi_c - \psi_{\pi p}$, where ψ_t, ψ_c , and $\psi_{\pi p}$ represent canopy turgor potential, total water potential 230 and osmotic potential, respectively) and canopy photosynthesis (Eq. 4) (Grant, 1995; Grant et al., 231 232 1993). ψ_c is calculated through explicitly modeling the plant hydraulics, i.e., by balancing the root water uptake from different soil layers with that transferred from root to canopy, and transpired 233 from the canopy to the atmosphere (Grant, 1995). Canopy photosynthesis is calculated by 234 summing the photosynthesis of all individual leaves, and is coupled with the calculation of canopy 235 stomatal resistance as: 236

237 $r_{cmin} = 0.64 (C_b - C_i') / V_c'$ r_c driven by rates of carboxylation *vs.* diffusion (Eq. 4a) 238 $r_c = r_{cmin} + (r_{cmax} - r_{cmin}) e^{(-\beta \psi_t)}$ r_c constrained by water status (Eq. 4b) 239 where r_c is canopy stomatal resistance to vapor flux, r_{cmin} is the minimum r_c at $\psi_c = 0$ MPa, C_b is 240 the CO₂ concentration in canopy air, C_i' is the intercellular CO₂ concentration at $\psi_c = 0$ MPa, V_c' 241 is the potential canopy CO₂ fixation rate at $\psi_c = 0$ MPa, r_{cmax} is canopy cuticular resistance to vapor 242 flux, and β is the stomatal resistance shape parameter.

243 **2.1.2** Carbon allocation, crop yield, and autotrophic respiration (*Ra*)

Ecosys simulates phenologically-driven plant carbon allocation to shoot and root (Grant, 1989b, 1989c). The dynamic ratio of shoot and root carbon allocation are functions of the number of phyllochron intervals and of the water and nutrient status of the plant (Grant, 1989b). The

allocated carbohydrate will be first used for maintenance respiration (*Rm*) in both shoot and root, 247 which is calculated based on the canopy temperature (shoot)/soil temperature (root), shoot/root 248 dry biomass, and nutrient stoichiometry. If the allocated carbohydrate can not meet the 249 maintenance respiration, the unmet requirement is remobilized from the existing foliage 250 carbohydrate pool, driving leaf senescence. Remaining carbohydrate after subtracting the 251 252 maintenance respiration from total carbohydrate is used for growth respiration (Rg) and dry mass (DM) formation. For shoots, DM is partitioned to as many as seven organs, including leaf, sheath, 253 254 stalk, soluble reserves, husk, cob, and grain, with dynamic partitioning coefficients varying with 255 growth stages (Grant, 1989b). Before floral induction, the shoot DM only consists of leaf and sheath compartments. After floral induction and before anthesis, the shoot DM is allocated to all 256 seven compartments except grain, which begins after anthesis, with partition coefficients 257 calculated from organ growth curves (Grant, 1989b). The modelled yield upon harvest is 258 determined by the seed number and kernel mass set during pre- and post-anthesis growth stages. 259 260 The plant growth status during stem elongation and the length of post anthesis period together determine the seed number formulation. The kernel mass is determined by the seed growth during 261 the early grain filling stage, limited by the predefined maximum kernel mass (Grant et al., 2011). 262 263 The grain filling rate in *ecosys* is limited by canopy temperature, and soluble reserve carbon and reserve nutrients in the grain. 264

265 **2.1.3** Heterotrophic respiration (*Rh*) and soil carbon dynamics

Ecosys computes Rh with explicit microbial dynamics that considers the stoichiometric
interactions among carbon, nitrogen and phosphorus (Grant, 2013; Grant and Rochette, 1994).
Specifically, organic matter and their transformation occur in five organic matter-microbial
complexes, which are coarse woody litter, fine nonwoody litter (including root exudates), animal

manure (if applied), particulate organic matter (POM) and humus. Each complex has five organic 270 271 states, including solid organic matter, sorbed organic matter, microbial residue, dissolved organic 272 matter, and the decomposition agents (microbes), all of which are vertically resolved from the surface litter layer to the bottom of the soil column. The microbes include diverse functional groups, 273 such as obligate aerobes (bacteria and fungi), aerobic and facultative nitrifiers, facultative 274 275 anaerobes (denitrifiers), obligate anaerobes (fermenters), heterotrophic (acetotrophic) and autotrophic (hydrogenotrophic) methanogens, and aerobic and anaerobic heterotrophic 276 277 diazotrophs (non-symbiotic N₂ fixers). In computing the organic matter transformation, solid 278 organic matter is first decomposed by microbes as a function of active microbial biomass (as an approximation to the exoenzyme hydrolysis), the product (aka soluble organic matter) is then taken 279 up by microbes in the presence of mineral soil sorption to support microbial catabolic activity (i.e., 280 heterotrophic respiration), which drives microbial biomass growth and mortality. Mineralization 281 associated with heterotrophic respiration produces ammonium, CO_2 and inorganic phosphorus to 282 283 drive the metabolism of lithotrophic groups. To maintain the elemental stoichiometry, all microbial groups compete with plants for inorganic nutrients, such as ammonium, nitrate and dissolved 284 inorganic phosphorus. Besides, aerobic microbes also compete with plant roots for oxygen. 285 286 Therefore, the heterotrophic respiration simulated by *ecosys* comprehensively resolves important process constraints from microbial population dynamics, organic matter formation and 287 288 destabilization, nutrient limitation and plant-microbial interaction as influenced by the soil 289 physical conditions. Mechanistically, ecosys is well positioned to conduct a comprehensive assessment of SOC change and greenhouse gas budget of agroecosystems. More details on the soil 290 291 biogeochemistry in *ecosys* can be found at Grant (2014).

292 **2.2 Model setup**

293 **2.2.1** Site-scale simulation, calibration, and validation

We evaluated the performance of *ecosys* using seven agricultural sites from the AmeriFlux network (https://ameriflux.lbl.gov/) that span a wide range of climate and soil conditions (Figure 1b and Table 1) located in the U.S. Midwest. Among these sites, US-Ne1 planted corn during the study period, whereas other sites had corn-soybean rotations; US-Ne1 and US-Ne2 are irrigated sites, whereas other sites are rainfed. Ecosystem CO₂, water, and energy fluxes were measured using the eddy covariance technique at these sites (Baldocchi et al., 2001; Baldocchi, 2003).

300 The hourly gap-filled meteorological variables (i.e., air temperature, precipitation, downward shortwave radiation, humidity, and wind speed) from AmeriFlux and soil information 301 (i.e., bulk density (BD), field capacity (FC), wilting point (WP), soil texture, saturated hydraulic 302 conductivity (KSat), soil organic carbon (SOC), pH, and cation exchange capacity (CEC)) from 303 the Gridded Soil Survey Geographic Database (gSSURGO) at these sites were used to drive *ecosys*. 304 305 For US-Ne1, US-Ne2 and US-Ne3, detailed land management practices (including planting time 306 and density, irrigation and fertilizer time and amount, tillage time and intensity) from the site records were also available as inputs for the model. For other sites, we used 7.5 plants/m² and 37.1 307 plants/m² for corn and soybean with the planting date from the Risk Management Agency (RMA) 308 309 of United States Department of Agriculture (USDA) (Lobell et al., 2014), and applied 18g $N/m^2/year$ fertilizer before planting for corn years. 310

The time series of GPP, NEE, and Reco of US-Ne1-3 during 2001–2012 were obtained from the FLUXNET2015 Tier 1 dataset (<u>http://fluxnet.fluxdata.org/data/fluxnet2015-dataset/</u>), and the LAI and carbon allocation data at different growth stages for those three sites during 2003– 2012 were obtained from Carbon Sequestration Program (CSP) at University of Nebraska-Lincoln's Agricultural Research and Development Center (<u>http://csp.unl.edu/Public/sites.htm</u>). For other four sites, the gap-filled GPP, NEE, Reco, and LAI from the AmeriFlux website were used for the model evaluation. We fine tuned the rubisco carboxylation activity and plant maturity group parameters of corn and soybean to match the seasonal patterns and magnitude of GPP and LAI at US-Ne1, US-Ne2 and US-Ne3 sites. The tuned model was evaluated at US-Ne sites using NEE, Reco, and carbon allocation measurements, and at other sites using the observed GPP, NEE, Reco, and LAI data.

_	Site	Latit	Longit	MAT	MAP	Simulate	Site	Crop Types	References
		ude	ude	(°C)	(mm)	d Period	Condition		
	US- Ne1	41.17	-96.48	10.4	710	2001-2012	Irrigated	Continuous corn	(Suyker et al., 2005; Suyker and Verma, 2012; Verma et al., 2005)
	US- Ne2	41.16	-96.47	10.4	710	2001-2012	Irrigated	Soybean in even years before 2009, corn in other years	(Suyker et al., 2005; Suyker and Verma, 2012; Verma et al., 2005)
	US- Ne3	41.17	-96.44	10.4	710	2001-2012	Rainfed	Corn in odd years and soybean in even years	(Suyker et al., 2005; Suyker and Verma, 2012; Verma et al., 2005)
	US- Bo1	40.01	-88.29	11.5	821	2001-2008	Rainfed	Corn in odd years and soybean in even years	(Bernacchi et al., 2005; Meyers, 2004)
	US- Br1	41.69	-93.69	9.1	938	2005-2011	Rainfed	Corn in odd years and soybean in	(Hernandez-Ramirez et al., 2011)

Table 1. Site information of selected flux towers in the U.S. Midwest for model evaluation.

US- Ib1	41.86	-88.22	9.5	972	2005-2011	Rainfed	Corn in even years and soybean in odd years	(Allison et al., 2005)
US- Ro1	44.71	-93.09	7.7	764	2004-2012	Rainfed	Corn in odd years and soybean in even years	(Baker and Griffis, 2005; Griffis et al., 2008)

even years



325 **2.2.2 Regional-scale crop yield and GPP simulation, calibration, and validation**

For regional-scale simulations, we focused on the three I states (Illinois, Indiana, and Iowa), 326 327 which is the major corn and soybean production area of the U.S. We conducted simulations at each county within three I states from 2001 to 2018 using corn-soybean rotation without irrigation (the 328 major planting strategies within this area), using the North American Land Data Assimilation 329 330 System (NLDAS-2) hourly meteorological data and gSSURGO soil data as inputs. NLDAS-2 meteorological data is from the integration of observation-based and model reanalysis data, with 331 0.125° spatial resolution covering central North America. The county scale meteorological 332 variables were aggregated from the NLDAS-2 grids within that county. The National 2020 333 Cultivated Layer (based on 2016-2020 USDA Cropland Data Layer) (USDA, 2021) and 334 gSSURGO datasets were used to obtain the county-scale soil properties (i.e., BD, soil texture, WP, 335 FC, KSat, SOC, pH, and CEC) that correspond to the county-scale cropland majority soil type. 336 For regional scale simulations, corn and soybean were also planted with 7.5 plants/m² and 37.1 337 plants/m² at the county scale based on the RMA planting date (2001–2012) (Lobell et al., 2014) 338 and the state-scale/agricultural district-scale crop progress reports (2013-2018) depending on the 339

data availability (Figure S4), and all crops were harvested on October 31. The state-wise crop 340 specific fertilizer information provided by USDA (USDA, 2019) was applied in the simulations. 341 342 For model calibration and evaluation, we used county-scale rainfed corn and soybean yield from USDA National Agricultural Statistics Service (NASS), and a new 250m resolution daily 343 GPP estimation using MODIS-based soil-adjusted near-infrared reflectance of vegetation 344 345 (SANIRv) and photosynthetically active radiation (PAR) (Jiang et al., 2021). The fixed linear yield trend was calculated from the NASS crop yield data for corn and soybean respectively for each 346 347 county, and was used to adjust the simulated yield to year 2009 (the midpoint of 2001-2018). To constrain *ecosys* efficiently, we built the surrogate models for crop yield and GPP separately using 348 the Long Short Term Memory networks (LSTM) to predict daily GPP and end-of-seasonal crop 349 yield under different corn and soybean parameters. In these models, the daily climate 350 meteorological data, three layers soil parameters (i.e., 0-5, 5-30, and 30-100cm), crop type, corn 351 parameters, soybean parameters, fertilizer amount, planting and harvest date, and day of year 352 (DOY) were used as input, and GPP or crop yield were used as output, respectively. The RMSE 353 of the surrogate models are 13.5 gC/m² and 0.46 gC/m²/day for yield and GPP, respectively. The 354 parameters for soybean include rubisco carboxylation activity, plant maturity group, maximum 355 356 number of fruiting sites per reproductive node, and maximum rate of kernel filling, and for corn include fraction of leaf protein in bundle sheath chlorophyll, plant maturity group, maximum 357 358 number of fruiting sites per reproductive node, and maximum rate of kernel filling. We conducted 359 the parameter calibration for each county based on the surrogate models, and used data from even years during 2001 to 2018 for model constraint and those from odd years for model validation. In 360 361 applying the constraint, the difference between simulated and observed yield, accumulated

362 growing season (i.e., May to September) GPP, and monthly growing season GPP were minimized363 using the cost function in Eq. 5c.

364
$$L(soybean) = NRMSE_{yield}(soybean) + NRMSE_{GPP}(soybean) + NRMSE_{GPP_monthly}(soybean)$$
 (Eq. 5a)

$$L(corn) = NRMSE_{yield}(corn) + NRMSE_{GPP}(corn) + NRMSE_{GPP_monthly}(corn)$$
(Eq. 5b)

$$L = L(soybean) + L(corn)$$
(Eq. 5c)

where $NRMSE_{yield}(soybean)$ and $NRMSE_{yield}(corn)$ are normalized RMSE of model simulated and measured crop yield for corn and soybean, respectively based on Eq. 6; $NRMSE_{GPP}(soybean)$ and $NRMSE_{GPP}(corn)$ are normalized RMSE of model simulated and measured growing season accumulated GPP for corn and soybean, respectively based on Eq. 7; $NRMSE_{GPP_monthly}(soybean)$ and $NRMSE_{GPP_monthly}(corn)$ are normalized RMSE of model simulated and measured growing season monthly GPP for corn and soybean, respectively based on Eq. 8.

373
$$NRMSE_{yield} = \frac{\sqrt{\frac{1}{9}\sum_{year=even_years}(Yield_{sim(year)} - Yield_{obs(year)})^2}}{\frac{1}{9}\sum_{year=even_years}Yield_{obs(year)}}$$
(Eq. 6)

374
$$NRMSE_{GPP} = \frac{\sqrt{\frac{1}{9}\sum_{year=even_years}(GPP_{sim(year)} - GPP_{obs(year)})^2}}{\frac{1}{9}\sum_{year=even_years}GPP_{obs(year)}}$$
(Eq. 7)

375
$$NRMSE_{GPP_monthly} = \frac{\sqrt{\frac{1}{5}\sum_{month=5}^{9} (\frac{\sum year = even_{years} GPP_{sim(year,month)}}{9} - \frac{(\sum year = even_{years} GPP_{obs(year,month)})^2}{9}}{\frac{1}{45}\sum year = even_{years} \sum_{month=5}^{9} GPP_{obs(year,month)}} (Eq. 8)$$

where *Yield_{sim(year)}* and *Yield_{obs(year)}* are the simulated and observed yield, *GPP_{sim(year)}* and *GPP_{obs(year)}* are the simulated and observed growing season accumulated GPP, *GPP_{sim(year,month)}* and *GPP_{obs(year,month)}* are the simulated and observed GPP at certain month, *even_years* is the years used
for model constrain (i.e., 2002, 2004, ..., 2018).

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381

383 **3. Results**

384 **3.1** Site-scale validation of *ecosys* in simulating carbon dynamics

We compared observed and modelled GPP, NEE, Reco fluxes at 7 eddy-covariance sites 385 in the U.S. Midwest (Figure 1b). The results indicate that ecosys can capture both the magnitude 386 and seasonal patterns of these carbon fluxes with high accuracy at both daily and monthly scale 387 388 (i.e., Figure 2, 3, S1, S2, and Table 2). The simulated GPP is consistent with the observations for both corn and soybean throughout the growing season, and can reflect the magnitude difference 389 between corn and soybean during peak growing season. At the daily scale, R² and RMSE are 0.94 390 and 2.15 gC/m²/day for corn, and are 0.86 and 1.90 gC/m²/day for soybean at Ne3, respectively 391 (Figure 2a). The seasonal pattern and magnitude of Reco, which is high during summer and low 392 during winter in the U.S. Midwest, can be captured by *ecosys* for both corn and soybean with high 393 modeling skills (i.e., $R^2=0.86$ and RMSE=2.04 gC/m²/day for corn, and $R^2=0.80$ and RMSE=1.37 394 $gC/m^2/day$ for soybean at Ne3, Figure 2c). As for NEE, the magnitude, peaking time, and zero-395 crossing points in observations are all captured by ecosys with R²=0.89 and RMSE=1.73 396 $gC/m^2/day$ for corn, and R²=0.75 and RMSE=1.27 $gC/m^2/day$ for soybean, respectively, at Ne3 397 (Figure 2b). 398

The comparison of observed and modelled above ground biomass (AGB) and its partition showed that the dynamics of AGB and its allocation to leaf, stem, and reproductive organs can be reproduced by *ecosys* for both corn and soybean, ensuring the application of *ecosys* for crop yield simulation (Figure 2, S1, and S2). The R² between the measured and simulated AGB and its leaf, stem, and reproductive percentages are 0.95, 0.92, 0.60, and 0.94 at Ne3. In both observations and simulations, during the early growing season, the AGB increase appears mostly as leaves to increase photosynthesis; during the peak growing season, the AGB increase is mostly found in stem for plant structural support, and at the late stage, the AGB increase is mostly allocated to thereproductive organ for grain formulation.

We also compared the responses of the modelled and observed GPP, Reco, and NEE to the 408 air temperature (Ta) and vapor pressure deficit (VPD) at those eddy-covariance sites (Figure 4 and 409 S3). The results indicate that *ecosys* captured the responses of major carbon fluxes e.g. GPP, Reco 410 411 and NEE to variations in air temperature and VPD at the eddy-covariance sites reasonably well. Taking corn as an example, when Ta is less than 30°C, GPP increases quickly, but stays stable 412 413 when Ta becomes higher. Both observations and simulations show such a response, which is primarily controlled by the limitation of temperature on leaf rubisco activity. As for the response 414 of GPP to VPD, GPP increases when VPD is small, but decreases when VPD gets higher in both 415 observations and simulations, which reveals the emergent influence of VPD on crop stomatal 416 conductance. The observed Reco showed a strong response to Ta (i.e., increases quickly with 417 higher Ta when Ta is below the optimal value) and no significant response to VPD, which can also 418 419 be captured by the *ecosys* simulations. As for NEE, the balance of carbon fixation and respiration shows similar responses to Ta and VPD as that of GPP in both observations and simulations. The 420 reason that results in the similar response of NEE and GPP to environmental factors is that NEE 421 422 is dominated by crop photosynthesis during peak growing season. Similar responses of GPP, NEE, and Reco to Ta and VPD are also captured by *ecosys* simulations for soybean at the eddy-423 424 covariance sites (i.e., Figure S3).



Figure 2. Comparing *ecosys* simulated GPP, NEE, Reco and carbon allocation with site
observations at Mead Ne3 site in Nebraska for both corn (light yellow shaded) and soybean (light
blue shaded).



Figure 3. Comparison of simulated and observed carbon fluxes (monthly) and LAI at the flux
tower sites. Red dashed lines indicate the 1-to-1 line.



Figure 4. Responses of simulated and observed daily GPP, NEE, and Reco to air temperature and



		NEE			GPP		Reco			
Sites -	RMSE (gC/m²/day)	Bias (gC/m²/day)	R ²	RMSE (gC/m²/day)	Bias (gC/m²/day)	R ²	RMSE (gC/m²/day)	Bias (gC/m²/day)	R ²	
Ne1	1.96	-0.60	0.86	2.44	0.67	0.93	1.92	-0.07	0.87	
Ne2	1.67	-0.28	0.88	2.32	0.25	0.92	1.98	-0.03	0.83	
Ne3	1.51	-0.04	0.86	2.02	-0.18	0.91	1.72	-0.11	0.79	
Bo1	2.26	0.11	0.65	3.31	0.06	0.74	2.04	0.21	0.67	
Br1	2.34	0.04	0.59	2.66	-0.06	0.80	1.46	-0.01	0.77	
Ib1	1.90	-0.35	0.69	1.64	0.28	0.91	1.66	0.05	0.77	
Ro1	1.77	-0.15	0.69	2.12	-0.69	0.89	1.30	-0.78	0.89	

Table 2. Comparison statistics of *ecosys* simulated daily surface carbon fluxes with flux towers
observations.

442 **3.2 Regional-scale crop yield and gross primary productivity simulation**

443 **3.2.1 Regional-scale corn and soybean yield simulation**

The comparison between modelled and NASS reported crop yield shows that *ecosys* can reproduce the spatial distribution and interannual variability of crop yield over three I States for both corn and soybean (Figure 5, Figure S5-7). Modeled long-term (2001–2018) averaged crop yield and NASS ground truth shows similar spatial patterns over three I States during both calibration years and validation years for both corn and soybean. The R², RMSE, and bias between the spatial patterns of modelled and measured yield are 0.83, 8.23 Bu/Acre, and 2.72 Bu/Acre for corn, and 0.80, 2.39 Bu/Acre, and 0.07 Bu/Acre for soybean, respectively. Long-term averaged 451 corn and soybean yield in the northern part of three I States are higher than that of the southern part 452 in both observations and simulations, which may be caused by the differences in soil (i.e., higher 453 SOC in the northern part) and climate conditions (i.e., more frequent heat stress and extreme 454 precipitation events in the southern part). The temporal variation of simulated average yield during 455 2001 to 2018 is also consistent with the observations with R^2 of 0.83 and 0.63 for corn and soybean, 456 respectively.



Figure 5. Comparison of *ecosys* simulated crop yield and NASS reported crop yield. (a) Spatial patterns of simulated and observed multi-year averaged corn yield in calibration and validation years. (b) Density scatter plot of simulated and observed multi-year averaged corn yield. Different colors mean different normalized density (means the ratio of points density to maximum points density, similar for the density scatter plot in other figures). (c) Spatial patterns of simulated and observed multi-year averaged soybean yield in calibration and validation years. (d) Density scatter

plot of simulated and observed multi-year averaged soybean yield. (e) and (f) is the time series of
three I states averaged corn and soybean yield respectively. Light green shaded years in (e) and (f)
are calibration years, and grey shaded years are validation years.

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469 **3.2.2 Regional-scale corn and soybean GPP simulation**

470 We also compared the modeled long-term averaged GPP and a new satellite-based GPP estimation during the peak growing season (June to August). The spatial patterns of simulated and 471 NIRv-based peak growing season accumulated GPP are similar during calibration years and 472 473 validation years for both corn and soybean (Figure 6), which are consistent with the spatial patterns of yield (Figure 5). The R², relative RMSE between the spatial patterns of modelled and NIRv-474 based GPP are 0.83 and 3.7% for corn, and 0.85 and 4.6% for soybean, respectively. The seasonal 475 variation of GPP for both corn and soybean can also be captured by *ecosys* at regional scale when 476 benchmarked with NIRv-based GPP (Figure 7, Figure S8, S10). For example, GPP of corn and 477 478 soybean grows quickly from June to July, and peaks at July and August in both simulations and NIRv-based observations (Figure 7). 479



Figure 6. Comparison of *ecosys* simulated peak growing season accumulated (June to August)
GPP and NIRv-based GPP. (a) Spatial patterns of simulated and NIRv-based multi-year averaged
corn GPP in calibration and validation years. (b) Density scatter plot of simulated and NIRv-based
multi-year averaged corn GPP. (c) Spatial patterns of simulated and NIRv-based multi-year
averaged soybean GPP in calibration and validation years. (d) Density scatter plot of simulated
and NIRv-based multi-year average soybean GPP.



490 Figure 7. Comparison of multi-year averaged *ecosys* simulated and NIRv-based monthly GPP for corn and soybean in validation years. (a) Simulated multi-year averaged monthly corn GPP during 491 492 validation years. (b) NIRv-based multi-year averaged monthly corn GPP during validation years. (c) Comparison of simulated and NIRv-based multi-year averaged monthly corn GPP at 493 494 Champaign, IL during validation years. (d) Simulated multi-year averaged monthly soybean GPP 495 during validation years. (b) NIRv-based multi-year averaged monthly soybean GPP during validation years. (c) Comparison of simulated and NIRv-based multi-year averaged monthly 496 497 soybean GPP at Champaign, IL during validation years.

499 **3.3 Response of crop yield to environmental variability in the U.S. Midwest**

500 Besides comparing the absolute value of modelled and observed yield and GPP, we also 501 investigated the response of these variables to the environmental factors to evaluate whether the 502 model could capture such response (Figure 8, Figure 9, and Figure S12). The LOWESS (LOcally Weighted Scatterplot Smoothing) was used to fit the response of observed and modelled crop yield
to key environmental factors, including Ta, precipitation, VPD, soil water content (SWC), bulk
density, and SOC, in the U.S. Midwest for corn and soybean during the growing season (Figure
8).

We found that the trend and inflection points of the observation-based response curves can 507 508 be simulated by ecosys at the regional scale for most of the months for climate variables and different depths for soil properties, demonstrating the ability of *ecosys* in capturing the response 509 510 of crop yield to environmental variabilities in the U.S. Midwest. Both observations and simulations 511 show that yield increases with increasing Ta until an optimal Ta value, and then decreases with higher Ta. The yield~Ta response is caused by the plant enzyme and growth activity with 512 temperature, which is also reflected in the GPP~Ta response (Figure S12) during key growing 513 months (i.e., July and August). For precipitation, the yield increases with increasing precipitation 514 515 when precipitation is smaller and then decreases at higher precipitation, revealing the tradeoff 516 between water limitation and excessive precipitation on crop growth (Li et al., 2019). For VPD, the yield increases with VPD when VPD is low, and decreases when VPD is higher, confirming 517 the impacts of VPD on crop productivity in both photosynthesis (through the VPD control on 518 519 stomatal conductance) (Ball, 1988; Grant et al., 1993) and crop yield (Kimm et al., 2020; Lobell et al., 2014; Zhou et al., 2020). The response of yield~SWC is similar to other environmental 520 521 variables, revealing the tradeoff between water supply and oxygen stress at high soil moisture on 522 crop growth. For both observations and simulations, the multi-year averaged crop yield decreases 523 with larger bulk density and increases with larger SOC in the U.S. Midwest.

- 524
- 525



Figure 8. Fitted responses of *ecosys* simulated and observed crop yield to climate variables at three
I States for corn and soybean using LOWESS. The shaded regions are the 95% confidence intervals
of LOWESS.



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Figure 9. Fitted responses of *ecosys* simulated and observed crop yield to soil conditions within
different soil depths at three I States for corn and soybean using LOWESS. The shaded regions are
the 95% confidence intervals of LOWESS.

539 4. Discussion

In this study, we used an advanced agroecosystem model, *ecosys*, to thoroughly simulate carbon budget for the U.S. Midwestern agroecosystems at both the site and regional scales. To address the gap that most previous model-based cropland carbon balance quantification studies with insufficient validations that only cover a small part of the carbon cycle components, we evaluated the model performance across a more comprehensive range of carbon cycle components, including carbon fixation, carbon allocation, and ecosystem respiration at site scale. In particular,

we tested *ecosys* performance at seven majority cropland eddy-covariance sites (with 55 site-years 546 observations) across the U.S. Midwest regarding GPP, NEE, Reco, LAI, and carbon allocation. 547 548 The model validation results reveal that *ecosys* can simulate the seasonal cycle and magnitude of agroecosystem carbon dynamics at different individual sites with high accuracy. Across all the 549 sites, the R² of the simulated and observed value for GPP, NEE, Reco, and LAI were 0.92, 0.87, 550 551 0.87, and 0.78, respectively (Figure 3). In addition, the dynamics of above ground biomass and its allocation to leaf, stem, and reproductive can be reproduced by *ecosys* (Figure 3, S1, and S2). The 552 553 overall model performance at Ne1, Ne2, and Ne3 sites are better than that at the other 4 sites in 554 simulating GPP, NEE and Reco (Table 2), which may largely be attributed to the more accurate records of land management practice (i.e., planting date and planting density, tillage information, 555 and irrigation information) at the Mead Ne sites. 556

Since the crop cultivar (e.g., maturity group) and management practices (e.g., fertilizer rate, 557 planting date) may varies in spatial, and is hard to obtain the information at high resolution, we 558 559 calibrated the *ecosys* model using the existing observations from both USDA survey for yield and satellite-based novel GPP estimations in even years to take the spatial variation of cultivars and 560 management practices into account, and validated the model in odd years at the regional scale by 561 562 simulating over 293 counties in the three I States. The model validation results show that ecosys can capture the spatial and temporal variability of crop yield as well as the magnitude and seasonal 563 patterns of GPP for both corn and soybean across the broader regions in the U.S. Midwest. The R² 564 565 of the multi-year averaged simulated and observed yield for corn and soybean is 0.83 and 0.80, respectively, showing the advanced ability of *ecosys* in capturing the crop yield spatial variance. 566 567 Based on our best knowledge, such a high performance in simulating crop yield with a direct 568 benchmark with county-level NASS data has not been achieved before (Zhang et al., 2015, 2020),

which is a strong demonstration of the ability of *ecosys* in simulating the carbon cycle for agroecosystems. The interannual variability of the observed crop yield can also be matched by *ecosys* simulations, but with some deviations at some years (i.e., 2003) between the observations and simulations for soybean, which may be caused by abiotic stress, such as pest, diseases, or uncaptured environmental impacts (e.g. hail, wind storm).

574 To fill the gap that previous studies only focused on matching the magnitude of the simulated target variables with observations, we also corroborated the simulated and observed 575 576 responses of carbon-related variables to climate and soil variabilities in both the site scale and 577 regional scale simulations. The response of the carbon flux and crop yield to the environmental variabilities obtained from the observations can be captured well by ecosys. For the site scale 578 simulation, the responses of the modelled and observed GPP, Reco, and NEE to the ambient 579 climate conditions (i.e., temperature and VPD) at the eddy-covariance sites are consistent; For the 580 regional scale simulations, the responses of simulated crop yield/GPP to the environmental factors 581 582 were similar to those of the observations during the growing season (i.e., Figure 8, Figure 9, and Figure S12). These results indicate the ability of *ecosys* in simulating carbon fluxes and crop yield 583 across the border soil and weather conditions. 584

Through the comprehensive evaluation of the simulated carbon components with the observations (including GPP, Reco, and carbon allocation at eddy-covariance sites, and GPP and yield at regional scales), we are able to simulate the NBP at the regional scale (Figure 10). The simulated multi-year averaged NBP had higher negative correlation to the SOC and NEE in the same period with *r* of -0.88 and -0.52, respectively, which indicates that both SOC stock and NEE drives NBP dynamics across space. As indicated in Eq. 3, annually, the accumulated NBP is approximately equal to Δ SOC, assuming ε (carbon leakage through runoff and methane emission) is sufficiently small. Our simulation results confirmed that using the carbon mass balance approach, we can regiously predict \triangle SOC (Figure 10b). This means that our method has the potential to be applied for quantifying annual-scale soil carbon sequestration for agroecosystems. However, cautions are given that, in being able to ensure the carbon mass balance approach work or have a low uncertainty, rigorous tests of different carbon cycle components, i.e., GPP, Reco, and harvest carbon, all should be conducted - currently no existing modeling-based study has demonstrated such a capability except this current study.

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Figure 10. Simulated multi-year averaged corn-soybean rotation cropland NBP during 2001-2018, and its correlation with \triangle SOC, SOC content, NEE, and harvest over three I states. (a) Simulated multi-year averaged corn-soybean rotation cropland carbon budget over three I states during 2001

to 2018. (b) The scatter plot of simulated SOC change and the sum of seed mass at planting and
NBP. (c) The scatter plot of averaged simulated SOC and NBP. (d) The scatter plot of simulated
NEE and NBP. (e) The scatter plot of simulated harvest carbon and NBP. The black lines and
shaded regions in (b)-(e) are the fitted linear regression models and the corresponding 95%
confidence interval.

610 Although we had validated the ability of *ecosys* in simulating the carbon cycle processes for both crop yield and GPP, there are still some limitations in the regional scale carbon balance 611 612 simulation that need to be further addressed. Specifically, in current simulations, we only focus on 613 the case that with no tillage and no cover crop. In the U.S. Midwest, tillage and cover crop are the commonly adopted conservation practices (Deines et al., 2019; Seifert et al., 2019), and may 614 change the soil carbon sequestration rate compared with the no till and no cover crop situation 615 (Baker et al., 2007; Poeplau and Don, 2015). For the tillage practice, it may redistribute SOC 616 content in the soil profile, affect the crop growth by influencing soil minimization and soil water 617 618 content, and also affect ecosystem respiration especially Rh (Mehra et al., 2018). For cover crops, it may influence the SOC sequestration rate by increasing GPP in the winter period and competing 619 the water and nutrients with the main crops in the summer (Abdalla et al., 2019). Studying the 620 621 impacts of cover crop and tillage is beyond the scope of the current paper, but they are under active investigation in our other studies. 622

623

624 **5. Conclusion**

In conclusion, we evaluated an advanced agroecosystem model, *ecosys*, to thoroughly simulate carbon budget for the agroecosystems at 7 cropland eddy-covariance sites and 293 counties in the U.S. Midwest. Both the magnitude of simulated carbon flux/yield and their response

to the environmental variabilities had been compared with that from the observations. For site 628 scale simulation, the R² of the simulated GPP, NEE, Reco, and LAI is 0.92, 0.87, 0.87, and 0.78, 629 respectively. In addition, the dynamics of carbon allocation processes for both corn and soybean 630 can also be reproduced by ecosys. For the regional scale simulation, the spatial pattern and 631 interannual variance of crop yield are consistent with that from the USDA survey for both corn 632 and soybean. Specifically, the R^2 of the multi-year averaged simulated and observed yield is 0.83 633 and 0.80 for corn and soybean, respectively; while the R^2 of spatial-averaged simulated and 634 observed crop yield from 2001to 2018 is 0.83 and 0.80 for corn and soybean, respectively. This 635 study is a strong demonstration of the ability of ecosys in simulating the carbon cycle for 636 agroecosystems. The response of carbon cycle processes/yield to the environmental variabilities 637 obtained from the simulations is consistent with that from the observations at both site-scale and 638 regional scale simulations, revealing the applicability of *ecosys* in simulating the impacts of future 639 climate change on the carbon cycle of the U.S. Midwestern agroecosystems. In addition, by 640 641 evaluating and constraining the majority carbon cycle process (i.e., GPP and yield at regional scale), we are able to simulate the net biome productivity, which can be applied to quantify the 642 soil carbon sequestration of agroecosystems. The method and framework adopted in this study can 643 644 also be applied to other land surface models and terrestrial biosphere models to improve the accounting of ecosystem carbon budget by integrating the mechanism models, observations, and 645 646 advanced machine learning tools.

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663 **6. Reference**

- Abdalla, M., Hastings, A., Cheng, K., Yue, Q., Chadwick, D., Espenberg, M., Truu, J., Rees, R.M.,
- Smith, P., 2019. A critical review of the impacts of cover crops on nitrogen leaching, net
 greenhouse gas balance and crop productivity. Glob. Chang. Biol. 25, 2530–2543.
 https://doi.org/10.1111/gcb.14644
- Allison, V.J., Michael Miller, R., Jastrow, J.D., Matamala, R., Zak, D.R., 2005. Changes in Soil
 Microbial Community Structure in a Tallgrass Prairie Chronosequence. Soil Science Society
 of America Journal. https://doi.org/10.2136/sssaj2004.0252
- Anderson, J.P.E., Domsch, K.H., 1975. Measurement of bacterial and fungal contributions to
 respiration of selected agricultural and forest soils. Can. J. Microbiol. 21, 314–322.
 https://doi.org/10.1139/m75-045

- Baker, J.M., Griffis, T.J., 2005. Examining strategies to improve the carbon balance of
 corn/soybean agriculture using eddy covariance and mass balance techniques. Agricultural
 and Forest Meteorology. https://doi.org/10.1016/j.agrformet.2004.11.005
- Baker, J.M., Ochsner, T.E., Venterea, R.T., Griffis, T.J., 2007. Tillage and soil carbon
 sequestration—What do we really know? Agric. Ecosyst. Environ. 118, 1–5.
 https://doi.org/10.1016/j.agee.2006.05.014
- Baldocchi, D.D., 2003. Assessing the eddy covariance technique for evaluating carbon dioxide
 exchange rates of ecosystems: past, present and future. Global Change Biology.
 https://doi.org/10.1046/j.1365-2486.2003.00629.x
- Baldocchi, D., Falge, E., Gu, L., Olson, R., Hollinger, D., Running, S., Anthoni, P., Bernhofer, C.,
- Davis, K., Evans, R., Fuentes, J., Goldstein, A., Katul, G., Law, B., Lee, X., Malhi, Y.,
- 685 Meyers, T., Munger, W., Oechel, W., Paw, K.T., Pilegaard, K., Schmid, H.P., Valentini, R.,
- 686 Verma, S., Vesala, T., Wilson, K., Wofsy, S., 2001. FLUXNET: A New Tool to Study the
- 687 Temporal and Spatial Variability of Ecosystem–Scale Carbon Dioxide, Water Vapor, and
- Energy Flux Densities. Bulletin of the American Meteorological Society.
- 689 https://doi.org/2.3.co;2">10.1175/1520-0477(2001)082<2415:fantts>2.3.co;2
- 690 Ball, J.T., 1988. An Analysis of Stomatal Conductance.
- Bardgett, R.D., McAlister, E., 1999. The measurement of soil fungal:bacterial biomass ratios as
 an indicator of ecosystem self-regulation in temperate meadow grasslands. Biol. Fertil. Soils
 29, 282–290. https://doi.org/10.1007/s003740050554
- Bernacchi, C.J., Hollinger, S.E., Meyers, T., 2005. The conversion of the corn/soybean ecosystem
 to no-till agriculture may result in a carbon sink. Global Change Biology.
 https://doi.org/10.1111/j.1365-2486.2005.01050.x

- Blanco-Canqui, H., Lal, R., 2004. Mechanisms of Carbon Sequestration in Soil Aggregates.
 Critical Reviews in Plant Sciences. https://doi.org/10.1080/07352680490886842
- Brilli, L., Bechini, L., Bindi, M., Carozzi, M., Cavalli, D., Conant, R., Dorich, C.D., Doro, L.,
- 700 Ehrhardt, F., Farina, R., Ferrise, R., Fitton, N., Francaviglia, R., Grace, P., Iocola, I., Klumpp,
- K., Léonard, J., Martin, R., Massad, R.S., Recous, S., Seddaiu, G., Sharp, J., Smith, P., Smith,
- W.N., Soussana, J.-F., Bellocchi, G., 2017. Review and analysis of strengths and weaknesses
- of agro-ecosystem models for simulating C and N fluxes. Sci. Total Environ. 598, 445–470.
- 704 https://doi.org/10.1016/j.scitotenv.2017.03.208
- 705 Chambers, A., Lal, R., Paustian, K., 2016. Soil carbon sequestration potential of US croplands and
- grasslands: Implementing the 4 per ThoUSAnd Initiative. J. Soil Water Conserv. 71, 68A707 74A. https://doi.org/10.2489/jswc.71.3.68A
- Deines, J.M., Wang, S., Lobell, D.B., 2019. Satellites reveal a small positive yield effect from
 conservation tillage across the US Corn Belt. Environ. Res. Lett. 14, 124038.
 https://doi.org/10.1088/1748-9326/ab503b
- 711 Dold, C., Büyükcangaz, H., Rondinelli, W., Prueger, J.H., Sauer, T.J., Hatfield, J.L., 2017. Long-
- term carbon uptake of agro-ecosystems in the Midwest. Agric. For. Meteorol. 232, 128–140.
 https://doi.org/10.1016/j.agrformet.2016.07.012
- Drewniak, B., Song, J., Prell, J., Kotamarthi, V.R., Jacob, R., 2013. Modeling agriculture in the
 Community Land Model. Geosci. Model Dev. 6, 495–515. https://doi.org/10.5194/gmd-6495-2013
- Farquhar, G.D., von Caemmerer, S., Berry, J.A., 1980. A biochemical model of photosynthetic
 CO2 assimilation in leaves of C3 species. Planta. https://doi.org/10.1007/bf00386231
- Gilhespy, S.L., Anthony, S., Cardenas, L., Chadwick, D., del Prado, A., Li, C., Misselbrook, T.,

720	Rees, R.M., Salas, W., Sanz-Cobena, A., Smith, P., Tilston, E.L., Topp, C.F.E., Vetter, S.,
721	Yeluripati, J.B., 2014. First 20 years of DNDC (DeNitrification DeComposition): Model
722	evolution. Ecol. Modell. 292, 51-62. https://doi.org/10.1016/j.ecolmodel.2014.09.004
723	Grant, R.F., Lin, S., Hernandez-Ramirez, G., 2020. Modelling nitrification inhibitor effects on
724	N2O emissions after fall-and spring-Applied slurry by reducing nitrifier NH4+ oxidation rate.
725	Biogeosciences 17, 2021–2039. https://doi.org/10.5194/bg-17-2021-2020
726	Grant, R., 1997. Changes in Soil Organic Matter under Different Tillage and Rotation:
727	Mathematical Modeling in ecosys. Soil Science Society of America Journal.
728	https://doi.org/10.2136/sssaj1997.03615995006100040023x
729	Grant, R., Arkebauer, T., Dobermann, A., Hubbard, K., Schimelfenig, T., Suyker, A., Verma, S.,
730	Walters, D., 2007. Net Biome Productivity of Irrigated and Rainfed Maize-Soybean
731	Rotations: Modeling vs. Measurements. Agronomy Journal.
732	https://doi.org/10.2134/agronj2006.0308
733	Grant, R., Barr, A., Black, T., Margolis, H., McCaughey, J., Trofymow, J., 2010. Net ecosystem
734	productivity of temperate and boreal forests after clearcutting-a Fluxnet-Canada
735	measurement and modelling synthesis. Tellus B. https://doi.org/10.3402/tellusb.v62i5.16588
736	Grant, R.F., 2014. Nitrogen mineralization drives the response of forest productivity to soil
737	warming: Modelling in ecosys vs. measurements from the Harvard soil heating experiment.
738	Ecol. Modell. 288, 38-46. https://doi.org/10.1016/j.ecolmodel.2014.05.015
739	Grant, R.F., 2013. Modelling changes in nitrogen cycling to sustain increases in forest productivity
740	under elevated atmospheric CO2 and contrasting site conditions. Biogeosciences.
741	https://doi.org/10.5194/bg-10-7703-2013
742	Grant, R.F., 2004. Modeling topographic effects on net ecosystem productivity of boreal black

743	spruce forests. Tree Physiol. 24, 1–18. https://doi.org/10.1093/treephys/24.1.1
744	Grant, R.F., 2001. A Review of the Canadian Ecosystem Model — ecosys. Modeling Carbon and
745	Nitrogen Dynamics for Soil Management. https://doi.org/10.1201/9781420032635.ch6
746	Grant, R.F., 1998. Simulation in ecosys of root growth response to contrasting soil water and
747	nitrogen. Ecol. Modell. 107, 237–264. https://doi.org/10.1016/S0304-3800(97)00221-4
748	Grant, R.F., 1995. Salinity, water use and yield of maize: Testing of the mathematical model
749	ecosys. Plant and Soil. https://doi.org/10.1007/bf00011333
750	Grant, R.F., 1989a. Test of a simple biochemical model for photosynthesis of maize and soybean
751	leaves. Agric. For. Meteorol. 48, 59-74. https://doi.org/10.1016/0168-1923(89)90007-5
752	Grant, R.F., 1989b. Simulation of Carbon Assimilation and Partitioning in Maize. Agronomy
753	Journal. https://doi.org/10.2134/agronj1989.00021962008100040004x
754	Grant, R.F., 1989c. Simulation of Maize Phenology. Agronomy Journal.
755	https://doi.org/10.2134/agronj1989.00021962008100030011x
756	Grant, R.F., Baldocchi, D.D., Ma, S., 2012. Ecological controls on net ecosystem productivity of
757	a seasonally dry annual grassland under current and future climates: Modelling with ecosys.
758	Agric. For. Meteorol. 152, 189–200. https://doi.org/10.1016/j.agrformet.2011.09.012
759	Grant, R.F., Flanagan, L.B., 2007. Modeling stomatal and nonstomatal effects of water deficits on
760	CO2 fixation in a semiarid grassland. J. Geophys. Res. Biogeosciences 112.
761	https://doi.org/10.1029/2006JG000302
762	Grant, R.F., Goulden, M.L., Wofsy, S.C., Berry, J.A., 2001. Carbon and energy exchange by a
763	black spruce-moss ecosystem under changing climate: Testing the mathematical model

- ecosys with data from the BOREAS experiment. J. Geophys. Res. Atmos. 106, 33605–33621.
- 765 https://doi.org/10.1029/2001JD900064

766	Grant, R.F., Kimball, B.A., Conley, M.M., White, J.W., Wall, G.W., Ottman, M.J., 2011.
767	Controlled Warming Effects on Wheat Growth and Yield: Field Measurements and
768	Modeling. Agronomy Journal. https://doi.org/10.2134/agronj2011.0158

- 769 Grant, R.F., Neftel, A., Calanca, P., 2016. Ecological controls on N2O emission in surface litter
- and near-surface soil of a managed grassland: modelling and measurements. Biogeosciences
- 771 13, 3549–3571. https://doi.org/10.5194/bg-13-3549-2016
- Grant, R.F., Pattey, E., 2008. Temperature sensitivity of N 2 O emissions from fertilized
 agricultural soils: Mathematical modeling in ecosys. Global Biogeochem. Cycles 22, n/a-n/a.
 https://doi.org/10.1029/2008GB003273
- Grant, R.F., Pattey, E., 2003. Modelling variability in N2O emissions from fertilized agricultural
 fields. Soil Biol. Biochem. 35, 225–243. https://doi.org/10.1016/S0038-0717(02)00256-0
- 777 Grant, R.F., Pattey, E., Goddard, T.W., Kryzanowski, L.M., Puurveen, H., 2006. Modeling the
- Effects of Fertilizer Application Rate on Nitrous Oxide Emissions. Soil Sci. Soc. Am. J. 70,
 235–248. https://doi.org/10.2136/sssaj2005.0104
- 780 Grant, R.F., Peters, D.B., Larson, E.M., Huck, M.G., 1989. Simulation of canopy photosynthesis
- in maize and soybean. Agric. For. Meteorol. 48, 75–92. https://doi.org/10.1016/01681923(89)90008-7
- Grant, R.F., Rochette, P., 1994. Soil Microbial Respiration at Different Water Potentials and
 Temperatures: Theory and Mathematical Modeling. Soil Science Society of America Journal.
 https://doi.org/10.2136/sssaj1994.03615995005800060015x
- Grant, R.F., Rochette, P., Desjardins, R.L., 1993. Energy Exchange and Water Use Efficiency of
 Field Crops: Validation of a Simulation Model. Agronomy Journal.
 https://doi.org/10.2134/agronj1993.00021962008500040025x

789	Grant, R.F., Goulden, M.L., Wofsy, S.C., Berry, J.A., 2001. Carbon and energy exchange by a
790	black spruce-moss ecosystem under changing climate: Testing the mathematical model
791	ecosys with data from the BOREAS experiment. J. Geophys. Res. Atmos. 106, 33605-33621.
792	https://doi.org/10.1029/2001JD900064

- Grant, R., Juma, N.G., Robertson, J.A., Izaurralde, R.C., McGill, W.B., 2001b. Long-Term
 Changes in Soil Carbon under Different Fertilizer, Manure, and Rotation: Testing the
 Mathematical Model ecosys with Data from the Breton Plots. Soil Sci. Soc. Am. J., NATO
 ASI Series Vol. I 38 65, 205–214. https://doi.org/10.2136/sssaj2001.651205x
- Grant, R., Juma, N., Robertson, J., Izaurralde, R., McGill, W., 2001c. Long-Term Changes in Soil
 Carbon under Different Fertilizer, Manure, and Rotation. Soil Science Society of America
 Journal. https://doi.org/10.2136/sssaj2001.1872a
- Grant, R., Mekonnen, Z., Riley, W., 2019. Modeling Climate Change Impacts on an Arctic
 Polygonal Tundra: 1. Rates of Permafrost Thaw Depend on Changes in Vegetation and
 Drainage. J. Geophys. Res. Biogeosci. 124, 1308–1322.
 https://doi.org/10.1029/2018JG004644
- Grant, R.F., Wall, G.W., Kimball, B.A., Frumau, K.F.A., Pinter, P.J., Hunsaker, D.J., Lamorte,

805 R.L., 1999. Crop water relations under different CO2 and irrigation: testing of ecosys with

- the free air CO2 enrichment (FACE) experiment. Agric. For. Meteorol. 95, 27–51.
 https://doi.org/10.1016/S0168-1923(99)00017-9
- Grant, R., Zhang, Y., Yuan, F., Wang, S., Hanson, P., Gaumont-Guay, D., Chen, J., Black, T.,
- 809 Barr, A., Baldocchi, D., Arain, A., 2006. Intercomparison of techniques to model water stress
- effects on CO2 and energy exchange in temperate and boreal deciduous forests. Ecological
- 811 Modelling. https://doi.org/10.1016/j.ecolmodel.2006.02.035

Griffis, T.J., Sargent, S.D., Baker, J.M., Lee, X., Tanner, B.D., Greene, J., Swiatek, E., Billmark,
K., 2008. Direct measurement of biosphere-atmosphere isotopic CO2exchange using the
eddy covariance technique. Journal of Geophysical Research.
https://doi.org/10.1029/2007jd009297

- Gurung, R.B., Ogle, S.M., Breidt, F.J., Williams, S.A., Parton, W.J., 2020. Bayesian calibration
- of the DayCent ecosystem model to simulate soil organic carbon dynamics and reduce model
 uncertainty. Geoderma 376, 114529. https://doi.org/10.1016/j.geoderma.2020.114529
- Hernandez-Ramirez, G., Hatfield, J.L., Parkin, T.B., Sauer, T.J., Prueger, J.H., 2011. Carbon
 dioxide fluxes in corn–soybean rotation in the midwestern U.S.: Inter- and intra-annual
 variations, and biophysical controls. Agricultural and Forest Meteorology.
 https://doi.org/10.1016/j.agrformet.2011.07.017
- 823 Huang, Y., Yu, Y., Zhang, W., Sun, W., Liu, S., Jiang, J., Wu, J., Yu, W., Wang, Y., Yang, Z.,

824 2009. Agro-C: A biogeophysical model for simulating the carbon budget of agroecosystems.

- Agric. For. Meteorol. 149, 106–129. https://doi.org/10.1016/j.agrformet.2008.07.013
- Hutchinson, J.J., Campbell, C.A., Desjardins, R.L., 2007. Some perspectives on carbon
 sequestration in agriculture. Agric. For. Meteorol. 142, 288–302.
 https://doi.org/10.1016/j.agrformet.2006.03.030
- Jandl, R., Rodeghiero, M., Martinez, C., Cotrufo, M.F., Bampa, F., van Wesemael, B., Harrison,
- 830 R.B., Guerrini, I.A., Richter, D.D., Jr, Rustad, L., Lorenz, K., Chabbi, A., Miglietta, F., 2014.
- 831 Current status, uncertainty and future needs in soil organic carbon monitoring. Sci. Total
- Environ. 468-469, 376–383. https://doi.org/10.1016/j.scitotenv.2013.08.026
- Jarecki, M.K., Lal, R., 2003. Crop Management for Soil Carbon Sequestration. Critical Reviews
- in Plant Sciences. https://doi.org/10.1080/713608318

836	Jiang, C., Guan, K., Wu, G., Peng, B., Wang, S., 2021. A daily, 250 m and real-time gross primary
837	productivity product (2000-present) covering the contiguous United States. Earth Syst. Sci.
838	Data 13, 281–298. https://doi.org/10.5194/essd-13-281-2021
839	Jin, Z., Zhuang, Q., Wang, J., Archontoulis, S. V., Zobel, Z., Kotamarthi, V.R., 2017. The
840	combined and separate impacts of climate extremes on the current and future US rainfed
841	maize and soybean production under elevated CO2. Glob. Chang. Biol. 23, 2687-2704.
842	https://doi.org/10.1111/gcb.13617
843	Kimble, J.M., Follett, R.F., Vernon Cole, C., 1998. The Potential of U.S. Cropland to Sequester
844	Carbon and Mitigate the Greenhouse Effect. CRC Press.
845	Kimm, H., Guan, K., Gentine, P., Wu, J., Bernacchi, C.J., Sulman, B.N., Griffis, T.J., Lin, C.,
846	2020. Redefining droughts for the U.S. Corn Belt: The dominant role of atmospheric vapor
847	pressure deficit over soil moisture in regulating stomatal behavior of Maize and Soybean.
848	Agric. For. Meteorol. 287, 107930. https://doi.org/10.1016/j.agrformet.2020.107930
849	Kuppel, S., Peylin, P., Chevallier, F., Bacour, C., Maignan, F., Richardson, A.D., 2012.
850	Constraining a global ecosystem model with multi-site eddy-covariance data.
851	https://doi.org/10.5194/bg-9-3757-2012
852	Lal, R., 2011. Sequestering carbon in soils of agro-ecosystems. Food Policy.
853	https://doi.org/10.1016/j.foodpol.2010.12.001
854	Lal, R., Follett, R.F., Stewart, B.A., Kimble, J.M., 2007. Soil carbon sequestration to mitigate
855	climate change and advance food security. Soil Sci. 172.
856	Lal, R., 2004. Soil carbon sequestration impacts on global climate change and food security.
857	Science 304, 1623–1627. https://doi.org/10.1126/science.1097396

- Lal, R., 2002. Soil carbon dynamics in cropland and rangeland. Environ. Pollut. 116, 353–362.
 https://doi.org/10.1016/s0269-7491(01)00211-1
- Lal, R., 2001. World cropland soils as a source or sink for atmospheric carbon. Advances in
 Agronomy. https://doi.org/10.1016/s0065-2113(01)71014-0
- Li, C., Frolking, S., Crocker, G.J., Grace, P.R., Klír, J., Körchens, M., Poulton, P.R., 1997.
- Simulating trends in soil organic carbon in long-term experiments using the DNDC model.
 Geoderma 81, 45–60. https://doi.org/10.1016/S0016-7061(97)00080-3
- Liu, X., Chen, F., Barlage, M., Zhou, G., Niyogi, D., 2016. Noah-MP-Crop: Introducing dynamic
- crop growth in the Noah-MP land surface model. J. Geophys. Res. 121, 13,953-13,972.
 https://doi.org/10.1002/2016JD025597
- Liu, Y., Yu, Z., Chen, J., Zhang, F., Doluschitz, R., Axmacher, J.C., 2006. Changes of soil organic
 carbon in an intensively cultivated agricultural region: a denitrification-decomposition
 (DNDC) modelling approach. Sci. Total Environ. 372, 203–214.
 https://doi.org/10.1016/j.scitotenv.2006.09.022
- Li, Y., Guan, K., Schnitkey, G.D., DeLucia, E., Peng, B., 2019. Excessive rainfall leads to maize
- yield loss of a comparable magnitude to extreme drought in the United States. Glob. Chang.

Biol. 25, 2325–2337. https://doi.org/10.1111/gcb.14628

- Lobell, D.B., Roberts, M.J., Schlenker, W., Braun, N., Little, B.B., Rejesus, R.M., Hammer, G.L.,
- 2014. Greater sensitivity to drought accompanies maize yield increase in the U.S. Midwest.
 Science 344, 516–519. https://doi.org/10.1126/science.1251423
- 878 Mäkelä, A., Pulkkinen, M., Kolari, P., Lagergren, F., Berbigier, P., Lindroth, A., Loustau, D.,
- Nikinmaa, E., Vesala, T., Hari, P., 2007. Developing an empirical model of stand GPP with
- the LUE approach: analysis of eddy covariance data at five contrasting conifer sites in Europe.

- 881 Glob. Chang. Biol. 0, 071124112207003-??? https://doi.org/10.1111/j.1365882 2486.2007.01463.x
- Meena, R.S., Kumar, S., Yadav, G.S., 2020. Soil Carbon Sequestration in Crop Production, in:
 Meena, R.S. (Ed.), Nutrient Dynamics for Sustainable Crop Production. Springer Singapore,
- Singapore, pp. 1–39. https://doi.org/10.1007/978-981-13-8660-2_1
- Mehra, P., Baker, J., Sojka, R.E., Bolan, N., Desbiolles, J., Kirkham, M.B., Ross, C., Gupta, R.,
- 2018. A Review of Tillage Practices and Their Potential to Impact the Soil Carbon Dynamics,
- in: Advances in Agronomy. Elsevier Inc., pp. 185–230.
 https://doi.org/10.1016/bs.agron.2018.03.002
- 890 Mekonnen, Z.A., Grant, R.F., Schwalm, C., 2017. Carbon sources and sinks of North America as
- affected by major drought events during the past 30 years. Agric. For. Meteorol. 244–245,
 42–56. https://doi.org/10.1016/j.agrformet.2017.05.006
- Meyers, T., 2004. An assessment of storage terms in the surface energy balance of maize andsoybean. Agricultural and Forest Meteorology.
- Mezbahuddin, M., Grant, R.F., Flanagan, L.B., 2017. Coupled eco-hydrology and
 biogeochemistry algorithms enable the simulation of water table depth effects on boreal
 peatland net CO2 exchange. Biogeosciences 14, 5507–5531. https://doi.org/10.5194/bg-14-
- 898 5507-2017
- 899 Mezbahuddin, S., Spiess, D., Hildebrand, D., Kryzanowski, L., Itenfisu, D., Goddard, T., Iqbal, J.,
- 900 Grant, R., 2020. Assessing Effects of Agronomic Nitrogen Management on Crop Nitrogen
- 901 Use and Nitrogen Losses in the Western Canadian Prairies. Frontiers in Sustainable Food
- 902 Systems. https://doi.org/10.3389/fsufs.2020.512292
- 903 Osborne, B., Saunders, M., Walmsley, D., Jones, M., Smith, P., 2010. Key questions and

- 904 uncertainties associated with the assessment of the cropland greenhouse gas balance. Agric.
 905 Ecosyst. Environ. 139, 293–301. https://doi.org/10.1016/j.agee.2010.05.009
- 906 Peng, B., Guan, K., Chen, M., Lawrence, D.M., Pokhrel, Y., Suyker, A., Arkebauer, T., Lu, Y.,
- 9072018. Improving maize growth processes in the community land model: Implementation and908evaluation.Agric.For.Meteorol.250-251,64-89.
- 909 https://doi.org/10.1016/j.agrformet.2017.11.012

- 910 Peng, B., Guan, K., Tang, J., Ainsworth, E.A., Asseng, S., Bernacchi, C.J., Cooper, M., Delucia,
- 911 E.H., Elliott, J.W., Ewert, F., Grant, R.F., Gustafson, D.I., Hammer, G.L., Jin, Z., Jones, J.W.,
- 912 Kimm, H., Lawrence, D.M., Li, Y., Lombardozzi, D.L., Marshall-Colon, A., Messina, C.D.,
- 913 Ort, D.R., Schnable, J.C., Vallejos, C.E., Wu, A., Yin, X., Zhou, W., 2020. Towards a
- 914 multiscale crop modelling framework for climate change adaptation assessment. Nat Plants
 915 6, 338–348. https://doi.org/10.1038/s41477-020-0625-3
- 916 Poeplau, C., Don, A., 2015. Carbon sequestration in agricultural soils via cultivation of cover crops
- 917
 A meta-analysis. Agric. Ecosyst. Environ. 200, 33–41.

 918
 https://doi.org/10.1016/j.agee.2014.10.024
- Salmon, Y., Lintunen, A., Dayet, A., Chan, T., Dewar, R., Vesala, T., Hölttä, T., 2020. Leaf carbon
 and water status control stomatal and nonstomatal limitations of photosynthesis in trees. New
 Phytol. 226, 690–703. https://doi.org/10.1111/nph.16436
- 922 Seifert, C.A., Azzari, G., Lobell, D.B., 2019. Corrigendum: Satellite detection of cover crops and
- their effects on crop yield in the Midwestern United States (2018 Environ. Res. Let. 13
- 924 064033). Environ. Res. Lett. 14, 039501. https://doi.org/10.1088/1748-9326/aaf933
- 926 carbon stocks and stock changes be quantified by soil inventories? https://doi.org/10.5194/bg-

Schrumpf, M., Schulze, E.D., Kaiser, K., Schumacher, J., 2011. How accurately can soil organic

927 8-1193-2011

- Shekoofa, A., Safikhan, S., Snider, J.L., Raper, T.B., Bourland, F.M., 2021. Variation in stomatal
 conductance responses of cotton cultivars to high vapour pressure deficit under controlled
 and rainfed environments. J. Agron. Crop Sci. 207, 332–343.
- 931 https://doi.org/10.1111/jac.12440
- Shirato, Y., 2005. Testing the Suitability of the DNDC Model for Simulating Long-Term Soil
 Organic Carbon Dynamics in Japanese Paddy Soils. Soil Sci. Plant Nutr. 51, 183–192.
 https://doi.org/10.1111/j.1747-0765.2005.tb00022.x
- Stehfest, E., Heistermann, M., Priess, J.A., Ojima, D.S., Alcamo, J., 2007. Simulation of global
 crop production with the ecosystem model DayCent. Ecol. Modell. 209, 203–219.

937 https://doi.org/10.1016/j.ecolmodel.2007.06.028

- Suyker, A.E., Verma, S.B., 2012. Gross primary production and ecosystem respiration of irrigated
 and rainfed maize–soybean cropping systems over 8 years. Agricultural and Forest
 Meteorology. https://doi.org/10.1016/j.agrformet.2012.05.021
- Suyker, A.E., Verma, S.B., Burba, G.G., Arkebauer, T.J., 2005. Gross primary production and
 ecosystem respiration of irrigated maize and irrigated soybean during a growing season.
 Agricultural and Forest Meteorology. https://doi.org/10.1016/j.agrformet.2005.05.007
- 944 Tarantola, A., 2013. Inverse Problem Theory: Methods for Data Fitting and Model Parameter
 945 Estimation. Elsevier.
- 946 USDA, 2021. USDA National Agricultural Statistics Service National 2020 Cultivated Layer.
- 947 Available at https://www.nass.usda.gov/Research_and_Science/Cropland/Release/index.php
- 948 (accessed May 2021). USDA-NASS, Washington, DC.
- 949 USDA, 2020. Crop Production 2019 Summary.

950	https://www.nass.usda.gov/Publications/Todays_Reports/reports/cropan20.pdf
951	USDA, 2019. Fertilizer Use and Price. https://www.ers.usda.gov/data-products/fertilizer-use-and-
952	price/
953	VandenBygaart, A.J., Angers, D.A., 2006. Towards accurate measurements of soil organic carbon
954	stock change in agroecosystems. Can. J. Soil Sci. 86, 465–471. https://doi.org/10.4141/S05-
955	106
956	van Wesemael, B., Paustian, K., Meersmans, J., Goidts, E., Barancikova, G., Easter, M., 2010.
957	Agricultural management explains historic changes in regional soil carbon stocks. Proc. Natl.
958	Acad. Sci. U. S. A. 107, 14926–14930. https://doi.org/10.1073/pnas.1002592107
959	Verma, S.B., Dobermann, A., Cassman, K.G., Walters, D.T., Knops, J.M., Arkebauer, T.J.,
960	Suyker, A.E., Burba, G.G., Amos, B., Yang, H., Ginting, D., Hubbard, K.G., Gitelson, A.A.,
961	Walter-Shea, E.A., 2005. Annual carbon dioxide exchange in irrigated and rainfed maize-
962	based agroecosystems. Agric. For. Meteorol. 131, 77–96.
963	https://doi.org/10.1016/j.agrformet.2005.05.003
964	Van den Hoof, C., Hanert, E., Vidale, P.L., 2011. Simulating dynamic crop growth with an adapted
965	land surface model - JULES-SUCROS: Model development and validation. Agric. For.
966	Meteorol. 151, 137–153. https://doi.org/10.1016/j.agrformet.2010.09.011
967	Vogel, F.A., 2018. Understanding USDA Crop Forecasts: March 1999 (Classic Reprint).
968	Forgotten Books.
969	Wang, S., Garcia, M., Ibrom, A., Bauer-Gottwein, P., 2020. Temporal interpolation of land surface
970	fluxes derived from remote sensing – results with an unmanned aerial system. Hydrol. Earth
971	Syst. Sci. 24, 3643–3661. https://doi.org/10.5194/hess-24-3643-2020
972	Wattenbach, M., Sus, O., Vuichard, N., Lehuger, S., Gottschalk, P., Li, L., Leip, A., Williams, M.,

973	Tomelleri, E., Kutsch, W.L., Buchmann, N., Eugster, W., Dietiker, D., Aubinet, M., Ceschia,
974	E., Béziat, P., Grünwald, T., Hastings, A., Osborne, B., Ciais, P., Cellier, P., Smith, P., 2010.
975	The carbon balance of European croplands: A cross-site comparison of simulation models.
976	Agric. Ecosyst. Environ. 139, 419-453. https://doi.org/10.1016/j.agee.2010.08.004
977	West, T.O., Bandaru, V., Brandt, C.C., Schuh, A.E., Ogle, S.M., 2011. Regional uptake and release
978	of crop carbon in the United States. Biogeosciences. https://doi.org/10.5194/bg-8-2037-2011
979	West, T.O., Brandt, C.C., Baskaran, L.M., Hellwinckel, C.M., Mueller, R., Bernacchi, C.J.,
980	Bandaru, V., Yang, B., Wilson, B.S., Marland, G., Nelson, R.G., De la Torre Ugarte, D.G.,
981	Post, W.M., 2010. Cropland carbon fluxes in the United States: increasing geospatial
982	resolution of inventory-based carbon accounting. Ecol. Appl. 20, 1074–1086
983	https://doi.org/10.1890/08-2352.1
984	West, T.O., Brandt, C.C., Wilson, B.S., Hellwinckel, C.M., Tyler, D.D., Marland, G., De La Torre

- Ugarte, D.G., Larson, J.A., Nelson, R.G., 2008. Estimating Regional Changes in Soil Carbon 985 with Sci. Soc. J. 72, 986 High Spatial Resolution. Soil Am. 285-294. https://doi.org/10.2136/sssaj2007.0113 987
- West, T.O., Brown, M.E., Duren, R.M., Ogle, S.M., Moss, R.H., 2013. Definition, capabilities and
 components of a terrestrial carbon monitoring system. Carbon Management.
 https://doi.org/10.4155/cmt.13.36
- 991 West, T.O., Marland, G., 2002. A synthesis of carbon sequestration, carbon emissions, and net
- 992 carbon flux in agriculture: comparing tillage practices in the United States. Agric. Ecosyst.
- 993 Environ. 91, 217–232. https://doi.org/10.1016/S0167-8809(01)00233-X
- Xue, F., Tong, L., Liu, W., Cao, H., Song, L., Ji, S., Ding, R., 2021. Stomatal conductance of
 tomato leaves is regulated by both abscisic acid and leaf water potential under combined

water and salt stress 1–9. https://doi.org/10.1111/ppl.13441

- 997 Yokohata, T., Kinoshita, T., Sakurai, G., Pokhrel, Y., Ito, A., Okada, M., Satoh, Y., Kato, E., Nitta,
- 998 T., Fujimori, S., Felfelani, F., Masaki, Y., Iizumi, T., Nishimori, M., Hanasaki, N., Takahashi,
- 999 K., Yamagata, Y., Emori, S., 2020. MIROC-INTEG-LAND version 1: a global
- biogeochemical land surface model with human water management, crop growth, and land-
- 1001 use change. Geosci. Model Dev. 13, 4713–4747. https://doi.org/10.5194/gmd-13-4713-2020
- 1002 Zhang, X., Izaurralde, R.C., Manowitz, D.H., Sahajpal, R., West, T.O., Thomson, A.M., Xu, M.,
- 1003 Zhao, K., LeDuc, S.D., Williams, J.R., 2015. Regional scale cropland carbon budgets:
- 1004 Evaluating a geospatial agricultural modeling system using inventory data. Environmental
- 1005 Modelling & Software 63, 199–216. https://doi.org/10.1016/j.envsoft.2014.10.005
- Zhang, Y., Gurung, R., Marx, E., Williams, S., Ogle, S.M., Paustian, K., 2020. DayCent Model
 Predictions of NPP and Grain Yields for Agricultural Lands in the Contiguous U.S. J.
 Geophys. Res. Biogeosci. https://doi.org/10.1029/2020JG005750
- 1009 Zhou, W., Guan, K., Peng, B., Shi, J., Jiang, C., Wardlow, B., Pan, M., Kimball, J.S., Franz, T.E.,
- 1010 Gentine, P., He, M., Zhang, J., 2020. Connections between the hydrological cycle and crop
- 1011 yield in the rainfed U.S. Corn Belt. J. Hydrol. 590, 125398.
- 1012 https://doi.org/10.1016/j.jhydrol.2020.125398