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**Quantifying carbon budget, crop yields and their responses to
environmental variability using the *ecosys* model for U.S.
Midwestern agroecosystems**

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Abstract:

As one of the major agricultural production areas in the world, the United States (U.S.) Midwest plays a vital role in the global food supply and agricultural ecosystem services. Although significant efforts have been made in modeling the carbon cycle dynamics over this area, large uncertainty still exists in the previous simulations in terms of reproducing individual components of the carbon cycle and their responses to environmental variability. Here we evaluated the performance of an advanced agroecosystem model, *ecosys*, in simulating carbon budgets over the U.S. Midwest, considering both the magnitude of carbon flux/yield and its response to the environmental (climatic and soil) variability. We conducted model simulations and evaluations at 7 cropland eddy-covariance sites as well as over 293 counties of Illinois, Indiana, and Iowa in the U.S. Midwest. The site-level simulations showed that *ecosys* captured both the magnitude and seasonal patterns of carbon fluxes (i.e., net ecosystem carbon exchange, ecosystem gross primary production (GPP), and ecosystem respiration), leaf area index, and dynamic plant carbon allocation processes, with R^2 equal to 0.92, 0.87, 0.87, and 0.78 for GPP, NEE, Reco, and LAI, respectively across all the sites compared with the observations. For regional scale simulations, *ecosys* reproduced the spatial distribution and interannual variability of corn and soybean yields with the constraints of observed yields and a new remotely sensed GPP product, with R^2 of multi-year averaged simulated and observed yield equal 0.83 and 0.80 for corn and soybean, respectively. The simulated responses of carbon cycle dynamics to environmental variability were consistent with that from the empirical observations at both site and regional scales. Our results demonstrated the applicability of *ecosys* in simulating the carbon cycle and soil carbon sequestration of the U.S. Midwestern agroecosystems under different climate and soil conditions.

Keywords: carbon fluxes, crop yield, *ecosys*, agroecosystems, environmental variabilities, U.S. Midwest

1. Introduction

The terrestrial carbon balance of agroecosystems plays an important role in the global carbon cycle (Dold et al., 2017; Verma et al., 2005). Depending on the temporal and spatial scales used for accounting as well as the geographical regions, croplands can be either carbon sinks or sources for the atmospheric CO₂ (Blanco-Canqui and Lal, 2004; Kimble et al., 1998). In the U.S. Midwest, about 30–50% of soil organic carbon (SOC) has been lost when compared with that before cultivation for most croplands (Lal, 2002). Since SOC content is often positively related to soil fertility, SOC loss may enhance crop yield loss risk under future climate conditions (Lal, 2011, 2004, 2001). Fortunately, with recommended management practices (RMPs, i.e., conservation tillage, cover crops, and biosolids and manure, etc.), prior studies show that U.S. croplands have the potential to sequester about 75–208 Tg C/year, which may recover 50–70% of the depleted soil carbon (Jarecki and Lal, 2003; Lal, 2011, 2007, 2002; Meena et al., 2020; Hutchinson et al., 2007; Chambers et al., 2016). Hence, in order to help realize this carbon sequestration potential in U.S. croplands, and meanwhile ensure global food security, it is critical to accurately quantify the carbon balance of agroecosystems, including carbon fixation and emission.

The carbon inventory (West et al., 2013, 2010, 2008; West and Marland, 2002), derived from crop yield survey reports (Vogel, 2018), SOC measurements (van Wesemael et al., 2010), and observation-based gross primary production (GPP) estimations (Jiang et al., 2021), can provide several components of cropland carbon budget. Among these carbon inventory methods, soil-sampling-based SOC measurement is the most direct approach to investigate SOC change, but

it still has uncertainties associated with soil sampling strategies (i.e., sampling location, depth, and time), measurement methods, and duration of measurements (Jandl et al., 2014; Schrumpf et al., 2011; VandenBygaart and Angers, 2006). More importantly, it is difficult to scale up the soil sampling due to its high labor and financial costs. In the framework of carbon balance, SOC change can in principle be derived from the whole carbon mass balance, which requires different carbon cycling components (i.e., GPP, ecosystem respiration, and harvest etc.). However, most of the carbon inventories cover only part of the carbon cycling components, such as agroecosystem carbon input (i.e., GPP) or outputs (i.e., yield). Measurements of other key carbon cycling components (i.e., respiration and litterfall) of agroecosystems are still difficult and insufficient, especially at large scales (e.g. U.S. Midwest) (Osborne et al., 2010). All these factors limit wide and robust applications of carbon inventory to quantify agroecosystem carbon budgets and SOC change.

Alternatively, we can use process-based models to quantify the cropland carbon budget (Brilli et al., 2017; Huang et al., 2009; Wattenbach et al., 2010; Zhang et al., 2015). However, existing studies using process-based models for cropland carbon quantification have suffered from one or few of the following limitations. First, very few model-based quantifications of the cropland carbon budget have gone through rigorous model validation covering the whole agroecosystem carbon cycle (i.e., carbon fixation, carbon allocation, and respiration), especially at regional scales. Most process-based modeling studies for agroecosystems evaluated and constrained their models with a limited number of observational variables, such as crop yield (Gilhespy et al., 2014; Stehfest et al., 2007) and/or measured SOC (Li et al., 1997; Liu et al., 2006; Shirato, 2005). This lack of sufficient model constraint may cause simulations to be apparently right with wrong reasons (Peng et al., 2018). For example, models can generate the same crop yield with higher carbon fixation

but lower harvest index compared to the correct ones, because errors in plant carbon fixation can be reconciled by unconstrained fluxes of respiration and litterfall. Therefore, to ensure that the model simulates carbon emission and sequestration correctly in both short and long terms, we need to use more carbon-related observations with fine temporal resolution (i.e., daily GPP, NEE, Reco, LAI, plant carbon allocation, and phenology) to sufficiently constrain and validate the carbon cycling processes of the models.

Second, most existing model-based studies only calibrated and validated the models at a few specific sites due to limited availability of observations. In general, models involve both parameters that are site-specific (i.e., maturity group and climate zone) and parameters that are shared among sites at a regional scale (i.e., parameters controlling the temperature responses of activity of RuBP carboxylase-oxygenase) (Kuppel et al., 2012; Mäkelä et al., 2007). Thus optimizing a model at specific sites will tie the resultant model parameterization closely to the site information (e.g. climate, soil, groundwater depth, field microtopography, and land management practices etc.), so that the model may not be suitable for other sites and regions with different soil and climate conditions. To ensure the model parameterization can be robustly transferred to other sites or regions, systematic evaluations are needed. Specifically, we need to constrain and evaluate models under a wide range of soil and climate conditions, using diverse data such as large-scale carbon inventories (e.g. crop yield reports and crop progress reports) and satellite remote sensing carbon-related observations (e.g. GPP and LAI).

Finally, current model calibrations and validations have generally focused on matching the magnitude or time series of the target variables (e.g., GPP and yield) (Gurung et. al., 2020; Wang et al., 2020; Jin et al., 2017), which is achieved by minimizing a cost function (which measures model-data discrepancy) that does not take into account the relationship between these target

variables and environmental drivers. From the perspective of Bayesian inference (Tarantola, 2013), since only uncertainties in model parameters are constrained, such a practice leads to an underestimation of the prior information associated with the environmental drivers. To make a more comprehensive use of the information contained in observations and model driving variables, as well as to deliver more confident predictions of how agroecosystems will respond to environment changes, we thus further need to verify relationships between environment variables and model predicted variables to test whether the model can simulate emergent responses of those variables to environmental factors from empirical observations (Peng et al., 2020). The accurate representation of the response of the target variables to environmental factors (i.e., climate variability and soil conditions) will help expand the models to broader soil and climate conditions.

Based on the above rationale, to demonstrate a new standard to achieve a comprehensive constraint and evaluation of an agroecosystem simulator, in this study we used an advanced ecosystem model, *ecosys*, to simulate surface carbon fluxes and corn/soybean yield in the U.S. Midwest at both eddy-covariance sites and county scales for the three I states (Illinois, Iowa and Indiana). As one of the world's largest crop production areas, the U.S. Midwest produces about 85% of U.S. corn and soybean (USDA, 2020). The soil health and crop yield of the U.S. Midwest in the future is vital to the global food supply and agricultural ecosystem services. To improve the quantification of carbon cycle dynamics in the U.S. Midwest, both the absolute values of the simulated carbon fluxes and yield as well as the responses of those variables to the environmental variabilities were evaluated. Through the evaluations, we aim to evaluate the capability of *ecosys* in conducting spatiotemporal extrapolations of agroecosystem carbon cycle by addressing the following two questions: (1) To what extent can *ecosys* simulate agroecosystem carbon dynamics at different individual sites as well as across the broader regions in the U.S. Midwest? (2) How

well can *ecosys* capture responses of carbon fluxes and crop yield to environmental variabilities? Although we use *ecosys* as an example, the procedures for model evaluation described in this study are applicable to many other agroecosystem models.

2. Data and method

2.1 The process-based model *ecosys*

Ecosys is an advanced mechanistic ecosystem model developed to simulate water, energy, carbon, and nutrient cycles simultaneously for various ecosystems, including agroecosystems at the hourly step (Figure 1a) (Grant, 2001). It is one of the very few models that are formulated primarily based on biophysical and biochemical principles, with fully connected balances and interactions for water, energy, carbon and nutrient cycles in the soil-plant-atmosphere continuum, and has been extensively validated in various ecosystems ranging from agricultural (Grant et al., 2007, 2011; Mezbahuddin et al., 2020) to forest systems (Grant et al., 2010, 2006).

The *ecosys* model was built based on the strategy that pursues the mechanistic representations and model outputs as directly comparable to observations as possible to realistically inform agricultural practices, by combining reactive transport modeling and state of the art knowledge of biogeochemistry (Grant, 2001). For example, photosynthesis and plant hydraulics in *ecosys* are coupled through leaf osmotic pressure, and then turgor pressure and leaf water potential that is linked to stomatal conductance (Grant, 1995; Grant and Flanagan, 2007), rather than empirical stress functions (Van den Hoof et al., 2011; Liu et al., 2016; Yokohata et al., 2020), and all of which can be measured in the field (Salmon et al., 2020; Shekoofa et al., 2021; Xue et al., 2021). As it integrates the plant hydraulics closely with the plant photosynthesis (Grant et al., 1999), the plant stomata conductance in *ecosys* is directly controlled by the balance between

161 photosynthetic carbon assimilation and plant water hydraulics calculated for the soil-plant-
162 atmosphere continuum, which can properly resolve the plant response to drought (Mekonnen et
163 al., 2017). Due to the explicit simulation of plant hydraulic impacts on stomatal conductance, the
164 empirical crop response to atmospheric vapor pressure deficit does not need be prescribed as in
165 many other models (Van den Hoof et al., 2011; Liu et al., 2016; Yokohata et al., 2020). In response
166 to soil water and plant carbon stress, *ecosys* also dynamically adjusts the plants' carbon and
167 nutrient allocation strategies (Grant et al., 2001a), so that all plant organs will balance their
168 respective growth to help the plants survive the harsh growth conditions and flourish under
169 favorable conditions. In addition, the plant carbon and nutrient allocation is represented following
170 the source-storage-sink balance approach, rather than the fixed allometric relationship approach
171 adopted by most existing models (Grant, 1989b; Drewniak et al., 2013; Liu et al., 2016).

172 Moreover, *ecosys* employs much more complete physical and chemistry theories in
173 simulating soil related processes. Specifically, *ecosys* mechanistically resolves the oxygen stress
174 throughout the soil and plant roots (Grant, 1998), such that a flood condition will suppress plants'
175 growth and alter the soil carbon and nutrient cycling. In addition, *ecosys* explicitly includes
176 microbes' competitive and symbiotic nutrient interactions with plants (Grant and Pattey, 2003;
177 Grant et al., 2006; Grant and Pattey, 2008; Grant et al., 2016), enabling a nutrient-based analysis
178 of how various management practices could affect plant productivity. Meanwhile, soil organic
179 carbon dynamics in *ecosys* are driven explicitly by microbial community dynamics that emerge
180 from the interactions between bacteria and fungi, and another five functional groups carrying out
181 fermentation, methane and nitrogen cycling (Grant, 2013; Grant and Rochette, 1994). Emergent
182 microbial population structure, e.g. bacteria to fungi ratio, can be directly evaluated with respect
183 to field measurements (Anderson and Domsch, 1975; Bardgett and McAlister, 1999). Moreover,

the partitioning of soil carbon in *ecosys* is amenable to the density fractionation that is often used by empiricists to characterize soil organic matter. In addition, *ecosys* outputs profiles and fluxes of many easily measurable chemicals, including different phase existences of CO₂, CH₄, N₂O, NH₃, NO₃, HPO₄⁽²⁻⁾, etc. Finally, *ecosys* resolves many common agricultural practices, such as mixed cropping, depth dependent irrigation and tillage (Grant, 1997), banded vs broadcast fertilization (Grant et al., 2001b), soil liming, manure application (Grant et al., 2001c), denitrification inhibitor (Grant et al., 2020), and tile-drainage system (Mezbahuddin et al., 2017) etc. Finally, *ecosys* generally requires no calibration for the soil and hydrological processes due to its complete mechanistic thus provides scalability to regional scale applications (Grant et al., 2012). All these features make *ecosys* stand out as a unique simulator as compared to many other models that tend to lump processes into simplified representations. We here refer detailed information about the processes represented in *ecosys* to the supplement of Grant et al. (2019), and the code of *ecosys* can be obtained from the online repository (<https://github.com/jinyun1tang/ECOSYS>). Below we only describe major carbon cycling processes of agroecosystems simulated in *ecosys* (Eq. 1 and 2).

$$-NEE = GPP - Reco = GPP - (Ra + Rh) = GPP - ((Rm + Rg) + Rh) \quad (\text{Eq. 1})$$

$$NBP = -NEE - \text{Yield} - \epsilon \quad (\text{Eq. 2})$$

where NEE is net ecosystem exchange, GPP is gross primary production, Ra is ecosystem autotrophic respiration, Rh is ecosystem heterotrophic respiration, Reco is ecosystem respiration, Rm and Rg are plant maintenance and growth respiration, NBP is net biome productivity, Yield is harvested crop yield, and ϵ is the carbon losses caused by disturbances (e.g., fire) excluding harvest.

In *ecosys*, the change of SOC (ΔSOC) is equal to the difference between plant litter fall, Rh, and ecosystem carbon leakage, including CH₄ emission, dissolved organic (DOC) and

inorganic carbon (DIC) leaching, etc (Eq. 3a). For annual cropping system in most of the U.S. Corn Belt regions, we can use NBP to approximate ΔSOC at long term scales ($\geq$ annual scale). By using Eq. 3, most part of simulated cropland soil carbon balance can be directly backed up with the eddy covariance measurements or carbon inventory data, which provided another approach to evaluate and verify the model performance in carbon budget estimations (Baker and Griffis, 2005).

$$\Delta\text{SOC} = \text{Litter_Fall} - \text{Rh} - \varepsilon \quad (\text{Eq. 3a})$$

$$= (\text{GPP} - \text{Ra} - \text{Yield} + \text{Seed_C}) - \text{Rh} - \varepsilon \quad (\geq \text{annual scale for annual cropping systems})$$

$$= -\text{NEE} + \text{Seed_C} - \text{Yield} - \varepsilon = \text{NBP} + \text{Seed_C} - \varepsilon \quad (\geq \text{annual scale for annual cropping systems}) \quad (\text{Eq. 3c})$$

where Litter_Fall is the litter fall from plants, including leaf senescence, harvest residue, and root carbon exudation, Seed_C is the seed mass at planting, ε is the carbon leakage through CH_4 emission, and DOC and DIC are leaching terms.

Figure 1. (a) Major processes represented in the *ecosys* model (revised from (Grant, 2004)), and (b) locations of the seven flux towers and the three I states in the U.S. Midwest.

2.1.1 Photosynthesis (*GPP*)

The *ecosys* model uses a multiple-layer canopy module to simulate canopy light absorption and carbon assimilation (Grant et al., 1989). Photosynthesis of each individual leaf is calculated independently using the Farquhar model for C3 plants and explicitly considering the mesophyll-bundle sheath carbon exchange for C4 plants at hourly time step (Farquhar et al., 1980; Grant, 1989a) with specific azimuth, leaf inclination, exposure of light conditions (i.e., sunlit and shaded leaves), and canopy height. The canopy stomatal resistance (r_c) is controlled by canopy turgor potential ($\psi_t = \psi_c - \psi_{\pi p}$, where ψ_t , ψ_c , and $\psi_{\pi p}$ represent canopy turgor potential, total water potential and osmotic potential, respectively) and canopy photosynthesis (Eq. 4) (Grant, 1995; Grant et al., 1993). ψ_c is calculated through explicitly modeling the plant hydraulics, i.e., by balancing the root water uptake from different soil layers with that transferred from root to canopy, and transpired from the canopy to the atmosphere (Grant, 1995). Canopy photosynthesis is calculated by summing the photosynthesis of all individual leaves, and is coupled with the calculation of canopy stomatal resistance as:

$$r_{cmin} = 0.64 (C_b - C_i') / V_c' \quad r_c \text{ driven by rates of carboxylation vs. diffusion (Eq. 4a)}$$

$$r_c = r_{cmin} + (r_{cmax} - r_{cmin}) e^{(-\beta \psi_t)} \quad r_c \text{ constrained by water status (Eq. 4b)}$$

where r_c is canopy stomatal resistance to vapor flux, r_{cmin} is the minimum r_c at $\psi_c = 0$ MPa, C_b is the CO₂ concentration in canopy air, C_i' is the intercellular CO₂ concentration at $\psi_c = 0$ MPa, V_c' is the potential canopy CO₂ fixation rate at $\psi_c = 0$ MPa, r_{cmax} is canopy cuticular resistance to vapor flux, and β is the stomatal resistance shape parameter.

2.1.2 Carbon allocation, crop yield, and autotrophic respiration (R_a)

Ecosys simulates phenologically-driven plant carbon allocation to shoot and root (Grant, 1989b, 1989c). The dynamic ratio of shoot and root carbon allocation are functions of the number of phyllochron intervals and of the water and nutrient status of the plant (Grant, 1989b). The

allocated carbohydrate will be first used for maintenance respiration (R_m) in both shoot and root, which is calculated based on the canopy temperature (shoot)/soil temperature (root), shoot/root dry biomass, and nutrient stoichiometry. If the allocated carbohydrate can not meet the maintenance respiration, the unmet requirement is remobilized from the existing foliage carbohydrate pool, driving leaf senescence. Remaining carbohydrate after subtracting the maintenance respiration from total carbohydrate is used for growth respiration (R_g) and dry mass (DM) formation. For shoots, DM is partitioned to as many as seven organs, including leaf, sheath, stalk, soluble reserves, husk, cob, and grain, with dynamic partitioning coefficients varying with growth stages (Grant, 1989b). Before floral induction, the shoot DM only consists of leaf and sheath compartments. After floral induction and before anthesis, the shoot DM is allocated to all seven compartments except grain, which begins after anthesis, with partition coefficients calculated from organ growth curves (Grant, 1989b). The modelled yield upon harvest is determined by the seed number and kernel mass set during pre- and post-anthesis growth stages. The plant growth status during stem elongation and the length of post anthesis period together determine the seed number formulation. The kernel mass is determined by the seed growth during the early grain filling stage, limited by the predefined maximum kernel mass (Grant et al., 2011). The grain filling rate in *ecosys* is limited by canopy temperature, and soluble reserve carbon and reserve nutrients in the grain.

2.1.3 Heterotrophic respiration (R_h) and soil carbon dynamics

Ecosys computes R_h with explicit microbial dynamics that considers the stoichiometric interactions among carbon, nitrogen and phosphorus (Grant, 2013; Grant and Rochette, 1994). Specifically, organic matter and their transformation occur in five organic matter-microbial complexes, which are coarse woody litter, fine nonwoody litter (including root exudates), animal

manure (if applied), particulate organic matter (POM) and humus. Each complex has five organic states, including solid organic matter, sorbed organic matter, microbial residue, dissolved organic matter, and the decomposition agents (microbes), all of which are vertically resolved from the surface litter layer to the bottom of the soil column. The microbes include diverse functional groups, such as obligate aerobes (bacteria and fungi), aerobic and facultative nitrifiers, facultative anaerobes (denitrifiers), obligate anaerobes (fermenters), heterotrophic (acetotrophic) and autotrophic (hydrogenotrophic) methanogens, and aerobic and anaerobic heterotrophic diazotrophs (non-symbiotic N₂ fixers). In computing the organic matter transformation, solid organic matter is first decomposed by microbes as a function of active microbial biomass (as an approximation to the exoenzyme hydrolysis), the product (aka soluble organic matter) is then taken up by microbes in the presence of mineral soil sorption to support microbial catabolic activity (i.e., heterotrophic respiration), which drives microbial biomass growth and mortality. Mineralization associated with heterotrophic respiration produces ammonium, CO₂ and inorganic phosphorus to drive the metabolism of lithotrophic groups. To maintain the elemental stoichiometry, all microbial groups compete with plants for inorganic nutrients, such as ammonium, nitrate and dissolved inorganic phosphorus. Besides, aerobic microbes also compete with plant roots for oxygen. Therefore, the heterotrophic respiration simulated by *ecosys* comprehensively resolves important process constraints from microbial population dynamics, organic matter formation and destabilization, nutrient limitation and plant-microbial interaction as influenced by the soil physical conditions. Mechanistically, *ecosys* is well positioned to conduct a comprehensive assessment of SOC change and greenhouse gas budget of agroecosystems. More details on the soil biogeochemistry in *ecosys* can be found at Grant (2014).

2.2 Model setup

2.2.1 Site-scale simulation, calibration, and validation

We evaluated the performance of *ecosys* using seven agricultural sites from the AmeriFlux network (<https://ameriflux.lbl.gov/>) that span a wide range of climate and soil conditions (Figure 1b and Table 1) located in the U.S. Midwest. Among these sites, US-Ne1 planted corn during the study period, whereas other sites had corn-soybean rotations; US-Ne1 and US-Ne2 are irrigated sites, whereas other sites are rainfed. Ecosystem CO₂, water, and energy fluxes were measured using the eddy covariance technique at these sites (Baldocchi et al., 2001; Baldocchi, 2003).

The hourly gap-filled meteorological variables (i.e., air temperature, precipitation, downward shortwave radiation, humidity, and wind speed) from AmeriFlux and soil information (i.e., bulk density (BD), field capacity (FC), wilting point (WP), soil texture, saturated hydraulic conductivity (KSat), soil organic carbon (SOC), pH, and cation exchange capacity (CEC)) from the Gridded Soil Survey Geographic Database (gSSURGO) at these sites were used to drive *ecosys*. For US-Ne1, US-Ne2 and US-Ne3, detailed land management practices (including planting time and density, irrigation and fertilizer time and amount, tillage time and intensity) from the site records were also available as inputs for the model. For other sites, we used 7.5 plants/m² and 37.1 plants/m² for corn and soybean with the planting date from the Risk Management Agency (RMA) of United States Department of Agriculture (USDA) (Lobell et al., 2014), and applied 18g N/m²/year fertilizer before planting for corn years.

The time series of GPP, NEE, and Reco of US-Ne1-3 during 2001–2012 were obtained from the FLUXNET2015 Tier 1 dataset (<http://fluxnet.fluxdata.org/data/fluxnet2015-dataset/>), and the LAI and carbon allocation data at different growth stages for those three sites during 2003–2012 were obtained from Carbon Sequestration Program (CSP) at University of Nebraska-Lincoln's Agricultural Research and Development Center (<http://csp.unl.edu/Public/sites.htm>).

For other four sites, the gap-filled GPP, NEE, Reco, and LAI from the AmeriFlux website were used for the model evaluation. We fine tuned the rubisco carboxylation activity and plant maturity group parameters of corn and soybean to match the seasonal patterns and magnitude of GPP and LAI at US-Ne1, US-Ne2 and US-Ne3 sites. The tuned model was evaluated at US-Ne sites using NEE, Reco, and carbon allocation measurements, and at other sites using the observed GPP, NEE, Reco, and LAI data.

Table 1. Site information of selected flux towers in the U.S. Midwest for model evaluation.

Site	Latitude	Longitude	MAT (°C)	MAP (mm)	Simulation Period	Site Condition	Crop Types	References
US-Ne1	41.17	-96.48	10.4	710	2001-2012	Irrigated	Continuous corn	(Suyker et al., 2005; Suyker and Verma, 2012; Verma et al., 2005)
US-Ne2	41.16	-96.47	10.4	710	2001-2012	Irrigated	Soybean in even years before 2009, corn in other years	(Suyker et al., 2005; Suyker and Verma, 2012; Verma et al., 2005)
US-Ne3	41.17	-96.44	10.4	710	2001-2012	Rainfed	Corn in odd years and soybean in even years	(Suyker et al., 2005; Suyker and Verma, 2012; Verma et al., 2005)
US-Bo1	40.01	-88.29	11.5	821	2001-2008	Rainfed	Corn in odd years and soybean in even years	(Bernacchi et al., 2005; Meyers, 2004)
US-Br1	41.69	-93.69	9.1	938	2005-2011	Rainfed	Corn in odd years and soybean in	(Hernandez-Ramirez et al., 2011)

							even years	
							Corn in even	
US-	41.86	-88.22	9.5	972	2005-2011	Rainfed	years and soybean	(Allison et al., 2005)
Ib1							in odd years	
							Corn in odd years	
US-	44.71	-93.09	7.7	764	2004-2012	Rainfed	and soybean in	(Baker and Griffis, 2005;
Ro1							even years	Griffis et al., 2008)

2.2.2 Regional-scale crop yield and GPP simulation, calibration, and validation

For regional-scale simulations, we focused on the three I states (Illinois, Indiana, and Iowa), which is the major corn and soybean production area of the U.S. We conducted simulations at each county within three I states from 2001 to 2018 using corn-soybean rotation without irrigation (the major planting strategies within this area), using the North American Land Data Assimilation System (NLDAS-2) hourly meteorological data and gSSURGO soil data as inputs. NLDAS-2 meteorological data is from the integration of observation-based and model reanalysis data, with 0.125° spatial resolution covering central North America. The county scale meteorological variables were aggregated from the NLDAS-2 grids within that county. The National 2020 Cultivated Layer (based on 2016-2020 USDA Cropland Data Layer) (USDA, 2021) and gSSURGO datasets were used to obtain the county-scale soil properties (i.e., BD, soil texture, WP, FC, KSat, SOC, pH, and CEC) that correspond to the county-scale cropland majority soil type. For regional scale simulations, corn and soybean were also planted with 7.5 plants/m² and 37.1 plants/m² at the county scale based on the RMA planting date (2001–2012) (Lobell et al., 2014) and the state-scale/agricultural district-scale crop progress reports (2013–2018) depending on the

data availability (Figure S4), and all crops were harvested on October 31. The state-wise crop specific fertilizer information provided by USDA (USDA, 2019) was applied in the simulations.

For model calibration and evaluation, we used county-scale rainfed corn and soybean yield from USDA National Agricultural Statistics Service (NASS), and a new 250m resolution daily GPP estimation using MODIS-based soil-adjusted near-infrared reflectance of vegetation (SANIRv) and photosynthetically active radiation (PAR) (Jiang et al., 2021). The fixed linear yield trend was calculated from the NASS crop yield data for corn and soybean respectively for each county, and was used to adjust the simulated yield to year 2009 (the midpoint of 2001-2018). To constrain *ecosys* efficiently, we built the surrogate models for crop yield and GPP separately using the Long Short Term Memory networks (LSTM) to predict daily GPP and end-of-seasonal crop yield under different corn and soybean parameters. In these models, the daily climate meteorological data, three layers soil parameters (i.e., 0-5, 5-30, and 30-100cm), crop type, corn parameters, soybean parameters, fertilizer amount, planting and harvest date, and day of year (DOY) were used as input, and GPP or crop yield were used as output, respectively. The RMSE of the surrogate models are 13.5 gC/m² and 0.46 gC/m²/day for yield and GPP, respectively. The parameters for soybean include rubisco carboxylation activity, plant maturity group, maximum number of fruiting sites per reproductive node, and maximum rate of kernel filling, and for corn include fraction of leaf protein in bundle sheath chlorophyll, plant maturity group, maximum number of fruiting sites per reproductive node, and maximum rate of kernel filling. We conducted the parameter calibration for each county based on the surrogate models, and used data from even years during 2001 to 2018 for model constraint and those from odd years for model validation. In applying the constraint, the difference between simulated and observed yield, accumulated

growing season (i.e., May to September) GPP, and monthly growing season GPP were minimized using the cost function in Eq. 5c.

$$L(\text{soybean}) = NRMSE_{\text{yield}}(\text{soybean}) + NRMSE_{GPP}(\text{soybean}) + NRMSE_{GPP_monthly}(\text{soybean}) \quad (\text{Eq. 5a})$$

$$L(\text{corn}) = NRMSE_{\text{yield}}(\text{corn}) + NRMSE_{GPP}(\text{corn}) + NRMSE_{GPP_monthly}(\text{corn}) \quad (\text{Eq. 5b})$$

$$L = L(\text{soybean}) + L(\text{corn}) \quad (\text{Eq. 5c})$$

where $NRMSE_{\text{yield}}(\text{soybean})$ and $NRMSE_{\text{yield}}(\text{corn})$ are normalized RMSE of model simulated and measured crop yield for corn and soybean, respectively based on Eq. 6; $NRMSE_{GPP}(\text{soybean})$ and $NRMSE_{GPP}(\text{corn})$ are normalized RMSE of model simulated and measured growing season accumulated GPP for corn and soybean, respectively based on Eq. 7; $NRMSE_{GPP_monthly}(\text{soybean})$ and $NRMSE_{GPP_monthly}(\text{corn})$ are normalized RMSE of model simulated and measured growing season monthly GPP for corn and soybean, respectively based on Eq. 8.

$$NRMSE_{\text{yield}} = \frac{\sqrt{\frac{1}{9} \sum_{\text{year}=\text{even_years}} (\text{Yield}_{\text{sim}}(\text{year}) - \text{Yield}_{\text{obs}}(\text{year}))^2}}{\frac{1}{9} \sum_{\text{year}=\text{even_years}} \text{Yield}_{\text{obs}}(\text{year})} \quad (\text{Eq. 6})$$

$$NRMSE_{GPP} = \frac{\sqrt{\frac{1}{9} \sum_{\text{year}=\text{even_years}} (GPP_{\text{sim}}(\text{year}) - GPP_{\text{obs}}(\text{year}))^2}}{\frac{1}{9} \sum_{\text{year}=\text{even_years}} GPP_{\text{obs}}(\text{year})} \quad (\text{Eq. 7})$$

$$NRMSE_{GPP_monthly} = \frac{\sqrt{\frac{1}{5} \sum_{\text{month}=5} \left(\frac{\sum_{\text{year}=\text{even_years}} GPP_{\text{sim}}(\text{year}, \text{month})}{9} - \frac{\sum_{\text{year}=\text{even_years}} GPP_{\text{obs}}(\text{year}, \text{month})}{9} \right)^2}}{\frac{1}{45} \sum_{\text{year}=\text{even_years}} \sum_{\text{month}=5} GPP_{\text{obs}}(\text{year}, \text{month})}} \quad (\text{Eq. 8})$$

where $\text{Yield}_{\text{sim}}(\text{year})$ and $\text{Yield}_{\text{obs}}(\text{year})$ are the simulated and observed yield, $GPP_{\text{sim}}(\text{year})$ and $GPP_{\text{obs}}(\text{year})$ are the simulated and observed growing season accumulated GPP, $GPP_{\text{sim}}(\text{year}, \text{month})$ and $GPP_{\text{obs}}(\text{year}, \text{month})$ are the simulated and observed GPP at certain month, *even_years* is the years used for model constrain (i.e., 2002, 2004, ..., 2018).

3. Results

3.1 Site-scale validation of *ecosys* in simulating carbon dynamics

We compared observed and modelled GPP, NEE, Reco fluxes at 7 eddy-covariance sites in the U.S. Midwest (Figure 1b). The results indicate that *ecosys* can capture both the magnitude and seasonal patterns of these carbon fluxes with high accuracy at both daily and monthly scale (i.e., Figure 2, 3, S1, S2, and Table 2). The simulated GPP is consistent with the observations for both corn and soybean throughout the growing season, and can reflect the magnitude difference between corn and soybean during peak growing season. At the daily scale, R^2 and RMSE are 0.94 and 2.15 gC/m²/day for corn, and are 0.86 and 1.90 gC/m²/day for soybean at Ne3, respectively (Figure 2a). The seasonal pattern and magnitude of Reco, which is high during summer and low during winter in the U.S. Midwest, can be captured by *ecosys* for both corn and soybean with high modeling skills (i.e., $R^2=0.86$ and RMSE=2.04 gC/m²/day for corn, and $R^2=0.80$ and RMSE=1.37 gC/m²/day for soybean at Ne3, Figure 2c). As for NEE, the magnitude, peaking time, and zero-crossing points in observations are all captured by *ecosys* with $R^2=0.89$ and RMSE=1.73 gC/m²/day for corn, and $R^2=0.75$ and RMSE=1.27 gC/m²/day for soybean, respectively, at Ne3 (Figure 2b).

The comparison of observed and modelled above ground biomass (AGB) and its partition showed that the dynamics of AGB and its allocation to leaf, stem, and reproductive organs can be reproduced by *ecosys* for both corn and soybean, ensuring the application of *ecosys* for crop yield simulation (Figure 2, S1, and S2). The R^2 between the measured and simulated AGB and its leaf, stem, and reproductive percentages are 0.95, 0.92, 0.60, and 0.94 at Ne3. In both observations and simulations, during the early growing season, the AGB increase appears mostly as leaves to increase photosynthesis; during the peak growing season, the AGB increase is mostly found in

stem for plant structural support, and at the late stage, the AGB increase is mostly allocated to the reproductive organ for grain formulation.

We also compared the responses of the modelled and observed GPP, Reco, and NEE to the air temperature (T_a) and vapor pressure deficit (VPD) at those eddy-covariance sites (Figure 4 and S3). The results indicate that *ecosys* captured the responses of major carbon fluxes e.g. GPP, Reco and NEE to variations in air temperature and VPD at the eddy-covariance sites reasonably well. Taking corn as an example, when T_a is less than 30°C, GPP increases quickly, but stays stable when T_a becomes higher. Both observations and simulations show such a response, which is primarily controlled by the limitation of temperature on leaf rubisco activity. As for the response of GPP to VPD, GPP increases when VPD is small, but decreases when VPD gets higher in both observations and simulations, which reveals the emergent influence of VPD on crop stomatal conductance. The observed Reco showed a strong response to T_a (i.e., increases quickly with higher T_a when T_a is below the optimal value) and no significant response to VPD, which can also be captured by the *ecosys* simulations. As for NEE, the balance of carbon fixation and respiration shows similar responses to T_a and VPD as that of GPP in both observations and simulations. The reason that results in the similar response of NEE and GPP to environmental factors is that NEE is dominated by crop photosynthesis during peak growing season. Similar responses of GPP, NEE, and Reco to T_a and VPD are also captured by *ecosys* simulations for soybean at the eddy-covariance sites (i.e., Figure S3).

Figure 2. Comparing *ecosys* simulated GPP, NEE, Reco and carbon allocation with site observations at Mead Ne3 site in Nebraska for both corn (light yellow shaded) and soybean (light blue shaded).

Figure 3. Comparison of simulated and observed carbon fluxes (monthly) and LAI at the flux tower sites. Red dashed lines indicate the 1-to-1 line.

Corn

Figure 4. Responses of simulated and observed daily GPP, NEE, and Reco to air temperature and VPD for corn during peak growing season (June to August).

Table 2. Comparison statistics of *ecosys* simulated daily surface carbon fluxes with flux towers observations.

Sites	NEE			GPP			Reco		
	RMSE	Bias	R ²	RMSE	Bias	R ²	RMSE	Bias	R ²
	(gC/m ² /day)	(gC/m ² /day)		(gC/m ² /day)	(gC/m ² /day)		(gC/m ² /day)	(gC/m ² /day)	
Ne1	1.96	-0.60	0.86	2.44	0.67	0.93	1.92	-0.07	0.87
Ne2	1.67	-0.28	0.88	2.32	0.25	0.92	1.98	-0.03	0.83
Ne3	1.51	-0.04	0.86	2.02	-0.18	0.91	1.72	-0.11	0.79
Bo1	2.26	0.11	0.65	3.31	0.06	0.74	2.04	0.21	0.67
Br1	2.34	0.04	0.59	2.66	-0.06	0.80	1.46	-0.01	0.77
Ib1	1.90	-0.35	0.69	1.64	0.28	0.91	1.66	0.05	0.77
Ro1	1.77	-0.15	0.69	2.12	-0.69	0.89	1.30	-0.78	0.89

3.2 Regional-scale crop yield and gross primary productivity simulation

3.2.1 Regional-scale corn and soybean yield simulation

The comparison between modelled and NASS reported crop yield shows that *ecosys* can reproduce the spatial distribution and interannual variability of crop yield over three I States for both corn and soybean (Figure 5, Figure S5-7). Modeled long-term (2001–2018) averaged crop yield and NASS ground truth shows similar spatial patterns over three I States during both calibration years and validation years for both corn and soybean. The R², RMSE, and bias between the spatial patterns of modelled and measured yield are 0.83, 8.23 Bu/Acre, and 2.72 Bu/Acre for corn, and 0.80, 2.39 Bu/Acre, and 0.07 Bu/Acre for soybean, respectively. Long-term averaged

451 corn and soybean yield in the northern part of three I States are higher than that of the southern part
452 in both observations and simulations, which may be caused by the differences in soil (i.e., higher
453 SOC in the northern part) and climate conditions (i.e., more frequent heat stress and extreme
454 precipitation events in the southern part). The temporal variation of simulated average yield during
455 2001 to 2018 is also consistent with the observations with R^2 of 0.83 and 0.63 for corn and soybean,
456 respectively.

457

Figure 5. Comparison of *ecosys* simulated crop yield and NASS reported crop yield. (a) Spatial patterns of simulated and observed multi-year averaged corn yield in calibration and validation years. (b) Density scatter plot of simulated and observed multi-year averaged corn yield. Different colors mean different normalized density (means the ratio of points density to maximum points density, similar for the density scatter plot in other figures). (c) Spatial patterns of simulated and observed multi-year averaged soybean yield in calibration and validation years. (d) Density scatter

plot of simulated and observed multi-year averaged soybean yield. (e) and (f) is the time series of three I states averaged corn and soybean yield respectively. Light green shaded years in (e) and (f) are calibration years, and grey shaded years are validation years.

3.2.2 Regional-scale corn and soybean GPP simulation

We also compared the modeled long-term averaged GPP and a new satellite-based GPP estimation during the peak growing season (June to August). The spatial patterns of simulated and NIRv-based peak growing season accumulated GPP are similar during calibration years and validation years for both corn and soybean (Figure 6), which are consistent with the spatial patterns of yield (Figure 5). The R^2 , relative RMSE between the spatial patterns of modelled and NIRv-based GPP are 0.83 and 3.7% for corn, and 0.85 and 4.6% for soybean, respectively. The seasonal variation of GPP for both corn and soybean can also be captured by *ecosys* at regional scale when benchmarked with NIRv-based GPP (Figure 7, Figure S8, S10). For example, GPP of corn and soybean grows quickly from June to July, and peaks at July and August in both simulations and NIRv-based observations (Figure 7).

Figure 6. Comparison of *ecosys* simulated peak growing season accumulated (June to August) GPP and NIRv-based GPP. (a) Spatial patterns of simulated and NIRv-based multi-year averaged corn GPP in calibration and validation years. (b) Density scatter plot of simulated and NIRv-based multi-year averaged corn GPP. (c) Spatial patterns of simulated and NIRv-based multi-year averaged soybean GPP in calibration and validation years. (d) Density scatter plot of simulated and NIRv-based multi-year average soybean GPP.

Figure 7. Comparison of multi-year averaged *ecosys* simulated and NIRv-based monthly GPP for corn and soybean in validation years. (a) Simulated multi-year averaged monthly corn GPP during validation years. (b) NIRv-based multi-year averaged monthly corn GPP during validation years. (c) Comparison of simulated and NIRv-based multi-year averaged monthly corn GPP at Champaign, IL during validation years. (d) Simulated multi-year averaged monthly soybean GPP during validation years. (b) NIRv-based multi-year averaged monthly soybean GPP during validation years. (c) Comparison of simulated and NIRv-based multi-year averaged monthly soybean GPP at Champaign, IL during validation years.

3.3 Response of crop yield to environmental variability in the U.S. Midwest

Besides comparing the absolute value of modelled and observed yield and GPP, we also investigated the response of these variables to the environmental factors to evaluate whether the model could capture such response (Figure 8, Figure 9, and Figure S12). The LOWESS (LOcally

Weighted Scatterplot Smoothing) was used to fit the response of observed and modelled crop yield to key environmental factors, including Ta, precipitation, VPD, soil water content (SWC), bulk density, and SOC, in the U.S. Midwest for corn and soybean during the growing season (Figure 8).

We found that the trend and inflection points of the observation-based response curves can be simulated by *ecosys* at the regional scale for most of the months for climate variables and different depths for soil properties, demonstrating the ability of *ecosys* in capturing the response of crop yield to environmental variabilities in the U.S. Midwest. Both observations and simulations show that yield increases with increasing Ta until an optimal Ta value, and then decreases with higher Ta. The yield~Ta response is caused by the plant enzyme and growth activity with temperature, which is also reflected in the GPP~Ta response (Figure S12) during key growing months (i.e., July and August). For precipitation, the yield increases with increasing precipitation when precipitation is smaller and then decreases at higher precipitation, revealing the tradeoff between water limitation and excessive precipitation on crop growth (Li et al., 2019). For VPD, the yield increases with VPD when VPD is low, and decreases when VPD is higher, confirming the impacts of VPD on crop productivity in both photosynthesis (through the VPD control on stomatal conductance) (Ball, 1988; Grant et al., 1993) and crop yield (Kimm et al., 2020; Lobell et al., 2014; Zhou et al., 2020). The response of yield~SWC is similar to other environmental variables, revealing the tradeoff between water supply and oxygen stress at high soil moisture on crop growth. For both observations and simulations, the multi-year averaged crop yield decreases with larger bulk density and increases with larger SOC in the U.S. Midwest.

Figure 8. Fitted responses of *ecosys* simulated and observed crop yield to climate variables at three I States for corn and soybean using LOWESS. The shaded regions are the 95% confidence intervals of LOWESS.

Figure 9. Fitted responses of *ecosys* simulated and observed crop yield to soil conditions within different soil depths at three I States for corn and soybean using LOWESS. The shaded regions are the 95% confidence intervals of LOWESS.

4. Discussion

In this study, we used an advanced agroecosystem model, *ecosys*, to thoroughly simulate carbon budget for the U.S. Midwestern agroecosystems at both the site and regional scales. To address the gap that most previous model-based cropland carbon balance quantification studies with insufficient validations that only cover a small part of the carbon cycle components, we evaluated the model performance across a more comprehensive range of carbon cycle components, including carbon fixation, carbon allocation, and ecosystem respiration at site scale. In particular,

we tested *ecosys* performance at seven majority cropland eddy-covariance sites (with 55 site-years observations) across the U.S. Midwest regarding GPP, NEE, Reco, LAI, and carbon allocation. The model validation results reveal that *ecosys* can simulate the seasonal cycle and magnitude of agroecosystem carbon dynamics at different individual sites with high accuracy. Across all the sites, the R^2 of the simulated and observed value for GPP, NEE, Reco, and LAI were 0.92, 0.87, 0.87, and 0.78, respectively (Figure 3). In addition, the dynamics of above ground biomass and its allocation to leaf, stem, and reproductive can be reproduced by *ecosys* (Figure 3, S1, and S2). The overall model performance at Ne1, Ne2, and Ne3 sites are better than that at the other 4 sites in simulating GPP, NEE and Reco (Table 2), which may largely be attributed to the more accurate records of land management practice (i.e., planting date and planting density, tillage information, and irrigation information) at the Mead Ne sites.

Since the crop cultivar (e.g., maturity group) and management practices (e.g., fertilizer rate, planting date) may varies in spatial, and is hard to obtain the information at high resolution, we calibrated the *ecosys* model using the existing observations from both USDA survey for yield and satellite-based novel GPP estimations in even years to take the spatial variation of cultivars and management practices into account, and validated the model in odd years at the regional scale by simulating over 293 counties in the three I States. The model validation results show that *ecosys* can capture the spatial and temporal variability of crop yield as well as the magnitude and seasonal patterns of GPP for both corn and soybean across the broader regions in the U.S. Midwest. The R^2 of the multi-year averaged simulated and observed yield for corn and soybean is 0.83 and 0.80, respectively, showing the advanced ability of *ecosys* in capturing the crop yield spatial variance. Based on our best knowledge, such a high performance in simulating crop yield with a direct benchmark with county-level NASS data has not been achieved before (Zhang et al., 2015, 2020),

which is a strong demonstration of the ability of *ecosys* in simulating the carbon cycle for agroecosystems. The interannual variability of the observed crop yield can also be matched by *ecosys* simulations, but with some deviations at some years (i.e., 2003) between the observations and simulations for soybean, which may be caused by abiotic stress, such as pest, diseases, or uncaptured environmental impacts (e.g. hail, wind storm).

To fill the gap that previous studies only focused on matching the magnitude of the simulated target variables with observations, we also corroborated the simulated and observed responses of carbon-related variables to climate and soil variabilities in both the site scale and regional scale simulations. The response of the carbon flux and crop yield to the environmental variabilities obtained from the observations can be captured well by *ecosys*. For the site scale simulation, the responses of the modelled and observed GPP, Reco, and NEE to the ambient climate conditions (i.e., temperature and VPD) at the eddy-covariance sites are consistent; For the regional scale simulations, the responses of simulated crop yield/GPP to the environmental factors were similar to those of the observations during the growing season (i.e., Figure 8, Figure 9, and Figure S12). These results indicate the ability of *ecosys* in simulating carbon fluxes and crop yield across the border soil and weather conditions.

Through the comprehensive evaluation of the simulated carbon components with the observations (including GPP, Reco, and carbon allocation at eddy-covariance sites, and GPP and yield at regional scales), we are able to simulate the NBP at the regional scale (Figure 10). The simulated multi-year averaged NBP had higher negative correlation to the SOC and NEE in the same period with r of -0.88 and -0.52, respectively, which indicates that both SOC stock and NEE drives NBP dynamics across space. As indicated in Eq. 3, annually, the accumulated NBP is approximately equal to ΔSOC , assuming ϵ (carbon leakage through runoff and methane emission)

is sufficiently small. Our simulation results confirmed that using the carbon mass balance approach, we can regionally predict ΔSOC (Figure 10b). This means that our method has the potential to be applied for quantifying annual-scale soil carbon sequestration for agroecosystems. However, cautions are given that, in being able to ensure the carbon mass balance approach work or have a low uncertainty, rigorous tests of different carbon cycle components, i.e., GPP, Reco, and harvest carbon, all should be conducted - currently no existing modeling-based study has demonstrated such a capability except this current study.

Figure 10. Simulated multi-year averaged corn-soybean rotation cropland NBP during 2001-2018, and its correlation with Δ SOC, SOC content, NEE, and harvest over three I states. (a) Simulated multi-year averaged corn-soybean rotation cropland carbon budget over three I states during 2001

to 2018. (b) The scatter plot of simulated SOC change and the sum of seed mass at planting and NBP. (c) The scatter plot of averaged simulated SOC and NBP. (d) The scatter plot of simulated NEE and NBP. (e) The scatter plot of simulated harvest carbon and NBP. The black lines and shaded regions in (b)-(e) are the fitted linear regression models and the corresponding 95% confidence interval.

Although we had validated the ability of *ecosys* in simulating the carbon cycle processes for both crop yield and GPP, there are still some limitations in the regional scale carbon balance simulation that need to be further addressed. Specifically, in current simulations, we only focus on the case that with no tillage and no cover crop. In the U.S. Midwest, tillage and cover crop are the commonly adopted conservation practices (Deines et al., 2019; Seifert et al., 2019), and may change the soil carbon sequestration rate compared with the no till and no cover crop situation (Baker et al., 2007; Poeplau and Don, 2015). For the tillage practice, it may redistribute SOC content in the soil profile, affect the crop growth by influencing soil minimization and soil water content, and also affect ecosystem respiration especially R_h (Mehra et al., 2018). For cover crops, it may influence the SOC sequestration rate by increasing GPP in the winter period and competing the water and nutrients with the main crops in the summer (Abdalla et al., 2019). Studying the impacts of cover crop and tillage is beyond the scope of the current paper, but they are under active investigation in our other studies.

5. Conclusion

In conclusion, we evaluated an advanced agroecosystem model, *ecosys*, to thoroughly simulate carbon budget for the agroecosystems at 7 cropland eddy-covariance sites and 293 counties in the U.S. Midwest. Both the magnitude of simulated carbon flux/yield and their response

to the environmental variabilities had been compared with that from the observations. For site scale simulation, the R^2 of the simulated GPP, NEE, Reco, and LAI is 0.92, 0.87, 0.87, and 0.78, respectively. In addition, the dynamics of carbon allocation processes for both corn and soybean can also be reproduced by *ecosys*. For the regional scale simulation, the spatial pattern and interannual variance of crop yield are consistent with that from the USDA survey for both corn and soybean. Specifically, the R^2 of the multi-year averaged simulated and observed yield is 0.83 and 0.80 for corn and soybean, respectively; while the R^2 of spatial-averaged simulated and observed crop yield from 2001 to 2018 is 0.83 and 0.80 for corn and soybean, respectively. This study is a strong demonstration of the ability of *ecosys* in simulating the carbon cycle for agroecosystems. The response of carbon cycle processes/yield to the environmental variabilities obtained from the simulations is consistent with that from the observations at both site-scale and regional scale simulations, revealing the applicability of *ecosys* in simulating the impacts of future climate change on the carbon cycle of the U.S. Midwestern agroecosystems. In addition, by evaluating and constraining the majority carbon cycle process (i.e., GPP and yield at regional scale), we are able to simulate the net biome productivity, which can be applied to quantify the soil carbon sequestration of agroecosystems. The method and framework adopted in this study can also be applied to other land surface models and terrestrial biosphere models to improve the accounting of ecosystem carbon budget by integrating the mechanism models, observations, and advanced machine learning tools.

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6. Reference

- Abdalla, M., Hastings, A., Cheng, K., Yue, Q., Chadwick, D., Espenberg, M., Truu, J., Rees, R.M., Smith, P., 2019. A critical review of the impacts of cover crops on nitrogen leaching, net greenhouse gas balance and crop productivity. *Glob. Chang. Biol.* 25, 2530–2543. <https://doi.org/10.1111/gcb.14644>
- Allison, V.J., Michael Miller, R., Jastrow, J.D., Matamala, R., Zak, D.R., 2005. Changes in Soil Microbial Community Structure in a Tallgrass Prairie Chronosequence. *Soil Science Society of America Journal*. <https://doi.org/10.2136/sssaj2004.0252>
- Anderson, J.P.E., Domsch, K.H., 1975. Measurement of bacterial and fungal contributions to respiration of selected agricultural and forest soils. *Can. J. Microbiol.* 21, 314–322. <https://doi.org/10.1139/m75-045>

674 Baker, J.M., Griffis, T.J., 2005. Examining strategies to improve the carbon balance of
 675 corn/soybean agriculture using eddy covariance and mass balance techniques. *Agricultural*
 676 *and Forest Meteorology*. <https://doi.org/10.1016/j.agrformet.2004.11.005>

677 Baker, J.M., Ochsner, T.E., Venterea, R.T., Griffis, T.J., 2007. Tillage and soil carbon
 678 sequestration—What do we really know? *Agric. Ecosyst. Environ.* 118, 1–5.
 679 <https://doi.org/10.1016/j.agee.2006.05.014>

680 Baldocchi, D.D., 2003. Assessing the eddy covariance technique for evaluating carbon dioxide
 681 exchange rates of ecosystems: past, present and future. *Global Change Biology*.
 682 <https://doi.org/10.1046/j.1365-2486.2003.00629.x>

683 Baldocchi, D., Falge, E., Gu, L., Olson, R., Hollinger, D., Running, S., Anthoni, P., Bernhofer, C.,
 684 Davis, K., Evans, R., Fuentes, J., Goldstein, A., Katul, G., Law, B., Lee, X., Malhi, Y.,
 685 Meyers, T., Munger, W., Oechel, W., Paw, K.T., Pilegaard, K., Schmid, H.P., Valentini, R.,
 686 Verma, S., Vesala, T., Wilson, K., Wofsy, S., 2001. FLUXNET: A New Tool to Study the
 687 Temporal and Spatial Variability of Ecosystem–Scale Carbon Dioxide, Water Vapor, and
 688 Energy Flux Densities. *Bulletin of the American Meteorological Society*.
 689 [https://doi.org/10.1175/1520-0477\(2001\)082<2415:fantts>2.3.co;2](https://doi.org/10.1175/1520-0477(2001)082<2415:fantts>2.3.co;2)

690 Ball, J.T., 1988. An Analysis of Stomatal Conductance.

691 Bardgett, R.D., McAlister, E., 1999. The measurement of soil fungal:bacterial biomass ratios as
 692 an indicator of ecosystem self-regulation in temperate meadow grasslands. *Biol. Fertil. Soils*
 693 29, 282–290. <https://doi.org/10.1007/s003740050554>

694 Bernacchi, C.J., Hollinger, S.E., Meyers, T., 2005. The conversion of the corn/soybean ecosystem
 695 to no-till agriculture may result in a carbon sink. *Global Change Biology*.
 696 <https://doi.org/10.1111/j.1365-2486.2005.01050.x>

697 Blanco-Canqui, H., Lal, R., 2004. Mechanisms of Carbon Sequestration in Soil Aggregates.
 698 Critical Reviews in Plant Sciences. <https://doi.org/10.1080/07352680490886842>
 699 Brilli, L., Bechini, L., Bindi, M., Carozzi, M., Cavalli, D., Conant, R., Dorich, C.D., Doro, L.,
 700 Ehrhardt, F., Farina, R., Ferrise, R., Fitton, N., Francaviglia, R., Grace, P., Iocola, I., Klumpp,
 701 K., Léonard, J., Martin, R., Massad, R.S., Recous, S., Seddaiu, G., Sharp, J., Smith, P., Smith,
 702 W.N., Soussana, J.-F., Bellocchi, G., 2017. Review and analysis of strengths and weaknesses
 703 of agro-ecosystem models for simulating C and N fluxes. *Sci. Total Environ.* 598, 445–470.
 704 <https://doi.org/10.1016/j.scitotenv.2017.03.208>
 705 Chambers, A., Lal, R., Paustian, K., 2016. Soil carbon sequestration potential of US croplands and
 706 grasslands: Implementing the 4 per ThoUSAnd Initiative. *J. Soil Water Conserv.* 71, 68A-
 707 74A. <https://doi.org/10.2489/jswc.71.3.68A>
 708 Deines, J.M., Wang, S., Lobell, D.B., 2019. Satellites reveal a small positive yield effect from
 709 conservation tillage across the US Corn Belt. *Environ. Res. Lett.* 14, 124038.
 710 <https://doi.org/10.1088/1748-9326/ab503b>
 711 Dold, C., Büyükcangaz, H., Rondinelli, W., Prueger, J.H., Sauer, T.J., Hatfield, J.L., 2017. Long-
 712 term carbon uptake of agro-ecosystems in the Midwest. *Agric. For. Meteorol.* 232, 128–140.
 713 <https://doi.org/10.1016/j.agrformet.2016.07.012>
 714 Drewniak, B., Song, J., Prell, J., Kotamarthi, V.R., Jacob, R., 2013. Modeling agriculture in the
 715 Community Land Model. *Geosci. Model Dev.* 6, 495–515. [https://doi.org/10.5194/gmd-6-](https://doi.org/10.5194/gmd-6-495-2013)
 716 495-2013
 717 Farquhar, G.D., von Caemmerer, S., Berry, J.A., 1980. A biochemical model of photosynthetic
 718 CO₂ assimilation in leaves of C₃ species. *Planta*. <https://doi.org/10.1007/bf00386231>
 719 Gilhespy, S.L., Anthony, S., Cardenas, L., Chadwick, D., del Prado, A., Li, C., Misselbrook, T.,

Rees, R.M., Salas, W., Sanz-Cobena, A., Smith, P., Tilston, E.L., Topp, C.F.E., Vetter, S.,
Yeluripati, J.B., 2014. First 20 years of DNDC (DeNitrification DeComposition): Model
evolution. *Ecol. Modell.* 292, 51–62. <https://doi.org/10.1016/j.ecolmodel.2014.09.004>

Grant, R.F., Lin, S., Hernandez-Ramirez, G., 2020. Modelling nitrification inhibitor effects on
N₂O emissions after fall-and spring-Applied slurry by reducing nitrifier NH₄⁺ oxidation rate.
Biogeosciences 17, 2021–2039. <https://doi.org/10.5194/bg-17-2021-2020>

Grant, R., 1997. Changes in Soil Organic Matter under Different Tillage and Rotation:
Mathematical Modeling in ecosys. *Soil Science Society of America Journal.*
<https://doi.org/10.2136/sssaj1997.03615995006100040023x>

Grant, R., Arkebauer, T., Dobermann, A., Hubbard, K., Schimelfenig, T., Suyker, A., Verma, S.,
Walters, D., 2007. Net Biome Productivity of Irrigated and Rainfed Maize-Soybean
Rotations: Modeling vs. Measurements. *Agronomy Journal.*
<https://doi.org/10.2134/agronj2006.0308>

Grant, R., Barr, A., Black, T., Margolis, H., McCaughey, J., Trofymow, J., 2010. Net ecosystem
productivity of temperate and boreal forests after clearcutting—a Fluxnet-Canada
measurement and modelling synthesis. *Tellus B.* <https://doi.org/10.3402/tellusb.v62i5.16588>

Grant, R.F., 2014. Nitrogen mineralization drives the response of forest productivity to soil
warming: Modelling in ecosys vs. measurements from the Harvard soil heating experiment.
Ecol. Modell. 288, 38–46. <https://doi.org/10.1016/j.ecolmodel.2014.05.015>

Grant, R.F., 2013. Modelling changes in nitrogen cycling to sustain increases in forest productivity
under elevated atmospheric CO₂ and contrasting site conditions. *Biogeosciences.*
<https://doi.org/10.5194/bg-10-7703-2013>

Grant, R.F., 2004. Modeling topographic effects on net ecosystem productivity of boreal black

spruce forests. *Tree Physiol.* 24, 1–18. <https://doi.org/10.1093/treephys/24.1.1>

Grant, R.F., 2001. A Review of the Canadian Ecosystem Model — ecosys. Modeling Carbon and Nitrogen Dynamics for Soil Management. <https://doi.org/10.1201/9781420032635.ch6>

Grant, R.F., 1998. Simulation in ecosys of root growth response to contrasting soil water and nitrogen. *Ecol. Modell.* 107, 237–264. [https://doi.org/10.1016/S0304-3800\(97\)00221-4](https://doi.org/10.1016/S0304-3800(97)00221-4)

Grant, R.F., 1995. Salinity, water use and yield of maize: Testing of the mathematical model ecosys. *Plant and Soil.* <https://doi.org/10.1007/bf00011333>

Grant, R.F., 1989a. Test of a simple biochemical model for photosynthesis of maize and soybean leaves. *Agric. For. Meteorol.* 48, 59–74. [https://doi.org/10.1016/0168-1923\(89\)90007-5](https://doi.org/10.1016/0168-1923(89)90007-5)

Grant, R.F., 1989b. Simulation of Carbon Assimilation and Partitioning in Maize. *Agronomy Journal.* <https://doi.org/10.2134/agronj1989.00021962008100040004x>

Grant, R.F., 1989c. Simulation of Maize Phenology. *Agronomy Journal.* <https://doi.org/10.2134/agronj1989.00021962008100030011x>

Grant, R.F., Baldocchi, D.D., Ma, S., 2012. Ecological controls on net ecosystem productivity of a seasonally dry annual grassland under current and future climates: Modelling with ecosys. *Agric. For. Meteorol.* 152, 189–200. <https://doi.org/10.1016/j.agrformet.2011.09.012>

Grant, R.F., Flanagan, L.B., 2007. Modeling stomatal and nonstomatal effects of water deficits on CO₂ fixation in a semiarid grassland. *J. Geophys. Res. Biogeosciences* 112. <https://doi.org/10.1029/2006JG000302>

Grant, R.F., Goulden, M.L., Wofsy, S.C., Berry, J.A., 2001. Carbon and energy exchange by a black spruce-moss ecosystem under changing climate: Testing the mathematical model ecosys with data from the BOREAS experiment. *J. Geophys. Res. Atmos.* 106, 33605–33621. <https://doi.org/10.1029/2001JD900064>

766 Grant, R.F., Kimball, B.A., Conley, M.M., White, J.W., Wall, G.W., Ottman, M.J., 2011.
 767 Controlled Warming Effects on Wheat Growth and Yield: Field Measurements and
 768 Modeling. *Agronomy Journal*. <https://doi.org/10.2134/agronj2011.0158>
 769 Grant, R.F., Neftel, A., Calanca, P., 2016. Ecological controls on N₂O emission in surface litter
 770 and near-surface soil of a managed grassland: modelling and measurements. *Biogeosciences*
 771 13, 3549–3571. <https://doi.org/10.5194/bg-13-3549-2016>
 772 Grant, R.F., Pattey, E., 2008. Temperature sensitivity of N₂O emissions from fertilized
 773 agricultural soils: Mathematical modeling in ecosys. *Global Biogeochem. Cycles* 22, n/a-n/a.
 774 <https://doi.org/10.1029/2008GB003273>
 775 Grant, R.F., Pattey, E., 2003. Modelling variability in N₂O emissions from fertilized agricultural
 776 fields. *Soil Biol. Biochem.* 35, 225–243. [https://doi.org/10.1016/S0038-0717\(02\)00256-0](https://doi.org/10.1016/S0038-0717(02)00256-0)
 777 Grant, R.F., Pattey, E., Goddard, T.W., Kryzanowski, L.M., Puurveen, H., 2006. Modeling the
 778 Effects of Fertilizer Application Rate on Nitrous Oxide Emissions. *Soil Sci. Soc. Am. J.* 70,
 779 235–248. <https://doi.org/10.2136/sssaj2005.0104>
 780 Grant, R.F., Peters, D.B., Larson, E.M., Huck, M.G., 1989. Simulation of canopy photosynthesis
 781 in maize and soybean. *Agric. For. Meteorol.* 48, 75–92. [https://doi.org/10.1016/0168-](https://doi.org/10.1016/0168-1923(89)90008-7)
 782 [1923\(89\)90008-7](https://doi.org/10.1016/0168-1923(89)90008-7)
 783 Grant, R.F., Rochette, P., 1994. Soil Microbial Respiration at Different Water Potentials and
 784 Temperatures: Theory and Mathematical Modeling. *Soil Science Society of America Journal*.
 785 <https://doi.org/10.2136/sssaj1994.03615995005800060015x>
 786 Grant, R.F., Rochette, P., Desjardins, R.L., 1993. Energy Exchange and Water Use Efficiency of
 787 Field Crops: Validation of a Simulation Model. *Agronomy Journal*.
 788 <https://doi.org/10.2134/agronj1993.00021962008500040025x>

789 Grant, R.F., Goulden, M.L., Wofsy, S.C., Berry, J.A., 2001. Carbon and energy exchange by a
790 black spruce-moss ecosystem under changing climate: Testing the mathematical model
791 ecosys with data from the BOREAS experiment. *J. Geophys. Res. Atmos.* 106, 33605–33621.
792 <https://doi.org/10.1029/2001JD900064>

793 Grant, R., Juma, N.G., Robertson, J.A., Izaurrealde, R.C., McGill, W.B., 2001b. Long-Term
794 Changes in Soil Carbon under Different Fertilizer, Manure, and Rotation: Testing the
795 Mathematical Model ecosys with Data from the Breton Plots. *Soil Sci. Soc. Am. J.*, NATO
796 ASI Series Vol. I 38 65, 205–214. <https://doi.org/10.2136/sssaj2001.651205x>

797 Grant, R., Juma, N., Robertson, J., Izaurrealde, R., McGill, W., 2001c. Long-Term Changes in Soil
798 Carbon under Different Fertilizer, Manure, and Rotation. *Soil Science Society of America*
799 *Journal*. <https://doi.org/10.2136/sssaj2001.1872a>

800 Grant, R., Mekonnen, Z., Riley, W., 2019. Modeling Climate Change Impacts on an Arctic
801 Polygonal Tundra: 1. Rates of Permafrost Thaw Depend on Changes in Vegetation and
802 Drainage. *J. Geophys. Res. Biogeosci.* 124, 1308–1322.
803 <https://doi.org/10.1029/2018JG004644>

804 Grant, R.F., Wall, G.W., Kimball, B.A., Frumau, K.F.A., Pinter, P.J., Hunsaker, D.J., Lamorte,
805 R.L., 1999. Crop water relations under different CO₂ and irrigation: testing of ecosys with
806 the free air CO₂ enrichment (FACE) experiment. *Agric. For. Meteorol.* 95, 27–51.
807 [https://doi.org/10.1016/S0168-1923\(99\)00017-9](https://doi.org/10.1016/S0168-1923(99)00017-9)

808 Grant, R., Zhang, Y., Yuan, F., Wang, S., Hanson, P., Gaumont-Guay, D., Chen, J., Black, T.,
809 Barr, A., Baldocchi, D., Arain, A., 2006. Intercomparison of techniques to model water stress
810 effects on CO₂ and energy exchange in temperate and boreal deciduous forests. *Ecological*
811 *Modelling*. <https://doi.org/10.1016/j.ecolmodel.2006.02.035>

812 Griffis, T.J., Sargent, S.D., Baker, J.M., Lee, X., Tanner, B.D., Greene, J., Swiatek, E., Billmark,
 813 K., 2008. Direct measurement of biosphere-atmosphere isotopic CO₂ exchange using the
 814 eddy covariance technique. *Journal of Geophysical Research*.
 815 <https://doi.org/10.1029/2007jd009297>

816 Gurung, R.B., Ogle, S.M., Breidt, F.J., Williams, S.A., Parton, W.J., 2020. Bayesian calibration
 817 of the DayCent ecosystem model to simulate soil organic carbon dynamics and reduce model
 818 uncertainty. *Geoderma* 376, 114529. <https://doi.org/10.1016/j.geoderma.2020.114529>

819 Hernandez-Ramirez, G., Hatfield, J.L., Parkin, T.B., Sauer, T.J., Prueger, J.H., 2011. Carbon
 820 dioxide fluxes in corn–soybean rotation in the midwestern U.S.: Inter- and intra-annual
 821 variations, and biophysical controls. *Agricultural and Forest Meteorology*.
 822 <https://doi.org/10.1016/j.agrformet.2011.07.017>

823 Huang, Y., Yu, Y., Zhang, W., Sun, W., Liu, S., Jiang, J., Wu, J., Yu, W., Wang, Y., Yang, Z.,
 824 2009. Agro-C: A biogeophysical model for simulating the carbon budget of agroecosystems.
 825 *Agric. For. Meteorol.* 149, 106–129. <https://doi.org/10.1016/j.agrformet.2008.07.013>

826 Hutchinson, J.J., Campbell, C.A., Desjardins, R.L., 2007. Some perspectives on carbon
 827 sequestration in agriculture. *Agric. For. Meteorol.* 142, 288–302.
 828 <https://doi.org/10.1016/j.agrformet.2006.03.030>

829 Jandl, R., Rodeghiero, M., Martinez, C., Cotrufo, M.F., Bampa, F., van Wesemael, B., Harrison,
 830 R.B., Guerrini, I.A., Richter, D.D., Jr, Rustad, L., Lorenz, K., Chabbi, A., Miglietta, F., 2014.
 831 Current status, uncertainty and future needs in soil organic carbon monitoring. *Sci. Total*
 832 *Environ.* 468-469, 376–383. <https://doi.org/10.1016/j.scitotenv.2013.08.026>

833 Jarecki, M.K., Lal, R., 2003. Crop Management for Soil Carbon Sequestration. *Critical Reviews*
 834 *in Plant Sciences*. <https://doi.org/10.1080/713608318>

835

836 Jiang, C., Guan, K., Wu, G., Peng, B., Wang, S., 2021. A daily, 250 m and real-time gross primary
 837 productivity product (2000–present) covering the contiguous United States. *Earth Syst. Sci.*
 838 *Data* 13, 281–298. <https://doi.org/10.5194/essd-13-281-2021>

839 Jin, Z., Zhuang, Q., Wang, J., Archontoulis, S. V., Zobel, Z., Kotamarthi, V.R., 2017. The
 840 combined and separate impacts of climate extremes on the current and future US rainfed
 841 maize and soybean production under elevated CO₂. *Glob. Chang. Biol.* 23, 2687–2704.
 842 <https://doi.org/10.1111/gcb.13617>

843 Kimble, J.M., Follett, R.F., Vernon Cole, C., 1998. *The Potential of U.S. Cropland to Sequester*
 844 *Carbon and Mitigate the Greenhouse Effect*. CRC Press.

845 Kimm, H., Guan, K., Gentine, P., Wu, J., Bernacchi, C.J., Sulman, B.N., Griffis, T.J., Lin, C.,
 846 2020. Redefining droughts for the U.S. Corn Belt: The dominant role of atmospheric vapor
 847 pressure deficit over soil moisture in regulating stomatal behavior of Maize and Soybean.
 848 *Agric. For. Meteorol.* 287, 107930. <https://doi.org/10.1016/j.agrformet.2020.107930>

849 Kuppel, S., Peylin, P., Chevallier, F., Bacour, C., Maignan, F., Richardson, A.D., 2012.
 850 Constraining a global ecosystem model with multi-site eddy-covariance data.
 851 <https://doi.org/10.5194/bg-9-3757-2012>

852 Lal, R., 2011. Sequestering carbon in soils of agro-ecosystems. *Food Policy*.
 853 <https://doi.org/10.1016/j.foodpol.2010.12.001>

854 Lal, R., Follett, R.F., Stewart, B.A., Kimble, J.M., 2007. Soil carbon sequestration to mitigate
 855 climate change and advance food security. *Soil Sci.* 172.

856 Lal, R., 2004. Soil carbon sequestration impacts on global climate change and food security.
 857 *Science* 304, 1623–1627. <https://doi.org/10.1126/science.1097396>

858 Lal, R., 2002. Soil carbon dynamics in cropland and rangeland. *Environ. Pollut.* 116, 353–362.
859 [https://doi.org/10.1016/s0269-7491\(01\)00211-1](https://doi.org/10.1016/s0269-7491(01)00211-1)

860 Lal, R., 2001. World cropland soils as a source or sink for atmospheric carbon. *Advances in*
861 *Agronomy*. [https://doi.org/10.1016/s0065-2113\(01\)71014-0](https://doi.org/10.1016/s0065-2113(01)71014-0)

862 Li, C., Frolking, S., Crocker, G.J., Grace, P.R., Klír, J., Körchens, M., Poulton, P.R., 1997.
863 Simulating trends in soil organic carbon in long-term experiments using the DNDC model.
864 *Geoderma* 81, 45–60. [https://doi.org/10.1016/S0016-7061\(97\)00080-3](https://doi.org/10.1016/S0016-7061(97)00080-3)

865 Liu, X., Chen, F., Barlage, M., Zhou, G., Niyogi, D., 2016. Noah-MP-Crop: Introducing dynamic
866 crop growth in the Noah-MP land surface model. *J. Geophys. Res.* 121, 13,953–13,972.
867 <https://doi.org/10.1002/2016JD025597>

868 Liu, Y., Yu, Z., Chen, J., Zhang, F., Doluschitz, R., Axmacher, J.C., 2006. Changes of soil organic
869 carbon in an intensively cultivated agricultural region: a denitrification-decomposition
870 (DNDC) modelling approach. *Sci. Total Environ.* 372, 203–214.
871 <https://doi.org/10.1016/j.scitotenv.2006.09.022>

872 Li, Y., Guan, K., Schnitkey, G.D., DeLucia, E., Peng, B., 2019. Excessive rainfall leads to maize
873 yield loss of a comparable magnitude to extreme drought in the United States. *Glob. Chang.*
874 *Biol.* 25, 2325–2337. <https://doi.org/10.1111/gcb.14628>

875 Lobell, D.B., Roberts, M.J., Schlenker, W., Braun, N., Little, B.B., Rejesus, R.M., Hammer, G.L.,
876 2014. Greater sensitivity to drought accompanies maize yield increase in the U.S. Midwest.
877 *Science* 344, 516–519. <https://doi.org/10.1126/science.1251423>

878 Mäkelä, A., Pulkkinen, M., Kolari, P., Lagergren, F., Berbigier, P., Lindroth, A., Loustau, D.,
879 Nikinmaa, E., Vesala, T., Hari, P., 2007. Developing an empirical model of stand GPP with
880 the LUE approach: analysis of eddy covariance data at five contrasting conifer sites in Europe.

881 Glob. Chang. Biol. 0, 071124112207003–??? <https://doi.org/10.1111/j.1365->
882 2486.2007.01463.x

883 Meena, R.S., Kumar, S., Yadav, G.S., 2020. Soil Carbon Sequestration in Crop Production, in:
884 Meena, R.S. (Ed.), Nutrient Dynamics for Sustainable Crop Production. Springer Singapore,
885 Singapore, pp. 1–39. https://doi.org/10.1007/978-981-13-8660-2_1

886 Mehra, P., Baker, J., Sojka, R.E., Bolan, N., Desbiolles, J., Kirkham, M.B., Ross, C., Gupta, R.,
887 2018. A Review of Tillage Practices and Their Potential to Impact the Soil Carbon Dynamics,
888 in: Advances in Agronomy. Elsevier Inc., pp. 185–230.
889 <https://doi.org/10.1016/bs.agron.2018.03.002>

890 Mekonnen, Z.A., Grant, R.F., Schwalm, C., 2017. Carbon sources and sinks of North America as
891 affected by major drought events during the past 30 years. Agric. For. Meteorol. 244–245,
892 42–56. <https://doi.org/10.1016/j.agrformet.2017.05.006>

893 Meyers, T., 2004. An assessment of storage terms in the surface energy balance of maize and
894 soybean. Agricultural and Forest Meteorology.

895 Mezbahuddin, M., Grant, R.F., Flanagan, L.B., 2017. Coupled eco-hydrology and
896 biogeochemistry algorithms enable the simulation of water table depth effects on boreal
897 peatland net CO₂ exchange. Biogeosciences 14, 5507–5531. [https://doi.org/10.5194/bg-14-](https://doi.org/10.5194/bg-14-5507-2017)
898 5507-2017

899 Mezbahuddin, S., Spiess, D., Hildebrand, D., Kryzanowski, L., Itenfisu, D., Goddard, T., Iqbal, J.,
900 Grant, R., 2020. Assessing Effects of Agronomic Nitrogen Management on Crop Nitrogen
901 Use and Nitrogen Losses in the Western Canadian Prairies. Frontiers in Sustainable Food
902 Systems. <https://doi.org/10.3389/fsufs.2020.512292>

903 Osborne, B., Saunders, M., Walmsley, D., Jones, M., Smith, P., 2010. Key questions and

uncertainties associated with the assessment of the cropland greenhouse gas balance. *Agric. Ecosyst. Environ.* 139, 293–301. <https://doi.org/10.1016/j.agee.2010.05.009>

Peng, B., Guan, K., Chen, M., Lawrence, D.M., Pokhrel, Y., Suyker, A., Arkebauer, T., Lu, Y., 2018. Improving maize growth processes in the community land model: Implementation and evaluation. *Agric. For. Meteorol.* 250–251, 64–89. <https://doi.org/10.1016/j.agrformet.2017.11.012>

Peng, B., Guan, K., Tang, J., Ainsworth, E.A., Asseng, S., Bernacchi, C.J., Cooper, M., Delucia, E.H., Elliott, J.W., Ewert, F., Grant, R.F., Gustafson, D.I., Hammer, G.L., Jin, Z., Jones, J.W., Kimm, H., Lawrence, D.M., Li, Y., Lombardozzi, D.L., Marshall-Colon, A., Messina, C.D., Ort, D.R., Schnable, J.C., Vallejos, C.E., Wu, A., Yin, X., Zhou, W., 2020. Towards a multiscale crop modelling framework for climate change adaptation assessment. *Nat Plants* 6, 338–348. <https://doi.org/10.1038/s41477-020-0625-3>

Poeplau, C., Don, A., 2015. Carbon sequestration in agricultural soils via cultivation of cover crops – A meta-analysis. *Agric. Ecosyst. Environ.* 200, 33–41. <https://doi.org/10.1016/j.agee.2014.10.024>

Salmon, Y., Lintunen, A., Dayet, A., Chan, T., Dewar, R., Vesala, T., Hölttä, T., 2020. Leaf carbon and water status control stomatal and nonstomatal limitations of photosynthesis in trees. *New Phytol.* 226, 690–703. <https://doi.org/10.1111/nph.16436>

Seifert, C.A., Azzari, G., Lobell, D.B., 2019. Corrigendum: Satellite detection of cover crops and their effects on crop yield in the Midwestern United States (2018 *Environ. Res. Lett.* 13 064033). *Environ. Res. Lett.* 14, 039501. <https://doi.org/10.1088/1748-9326/aaf933>

Schrumpf, M., Schulze, E.D., Kaiser, K., Schumacher, J., 2011. How accurately can soil organic carbon stocks and stock changes be quantified by soil inventories? <https://doi.org/10.5194/bg->

8-1193-2011

Shekoofa, A., Safikhan, S., Snider, J.L., Raper, T.B., Bourland, F.M., 2021. Variation in stomatal conductance responses of cotton cultivars to high vapour pressure deficit under controlled and rainfed environments. *J. Agron. Crop Sci.* 207, 332–343. <https://doi.org/10.1111/jac.12440>

Shirato, Y., 2005. Testing the Suitability of the DNDC Model for Simulating Long-Term Soil Organic Carbon Dynamics in Japanese Paddy Soils. *Soil Sci. Plant Nutr.* 51, 183–192. <https://doi.org/10.1111/j.1747-0765.2005.tb00022.x>

Stehfest, E., Heistermann, M., Priess, J.A., Ojima, D.S., Alcamo, J., 2007. Simulation of global crop production with the ecosystem model DayCent. *Ecol. Modell.* 209, 203–219. <https://doi.org/10.1016/j.ecolmodel.2007.06.028>

Suyker, A.E., Verma, S.B., 2012. Gross primary production and ecosystem respiration of irrigated and rainfed maize–soybean cropping systems over 8 years. *Agricultural and Forest Meteorology*. <https://doi.org/10.1016/j.agrformet.2012.05.021>

Suyker, A.E., Verma, S.B., Burba, G.G., Arkebauer, T.J., 2005. Gross primary production and ecosystem respiration of irrigated maize and irrigated soybean during a growing season. *Agricultural and Forest Meteorology*. <https://doi.org/10.1016/j.agrformet.2005.05.007>

Tarantola, A., 2013. *Inverse Problem Theory: Methods for Data Fitting and Model Parameter Estimation*. Elsevier.

USDA, 2021. USDA National Agricultural Statistics Service National 2020 Cultivated Layer. Available at https://www.nass.usda.gov/Research_and_Science/Cropland/Release/index.php (accessed May 2021). USDA-NASS, Washington, DC.

USDA, 2020. Crop Production 2019 Summary.

950 https://www.nass.usda.gov/Publications/Todays_Reports/reports/cropan20.pdf

951 USDA, 2019. Fertilizer Use and Price. [https://www.ers.usda.gov/data-products/fertilizer-use-and-](https://www.ers.usda.gov/data-products/fertilizer-use-and-price/)

952 [price/](https://www.ers.usda.gov/data-products/fertilizer-use-and-price/)

953 VandenBygaart, A.J., Angers, D.A., 2006. Towards accurate measurements of soil organic carbon

954 stock change in agroecosystems. *Can. J. Soil Sci.* 86, 465–471. [https://doi.org/10.4141/S05-](https://doi.org/10.4141/S05-106)

955 106

956 van Wesemael, B., Paustian, K., Meersmans, J., Goidts, E., Barancikova, G., Easter, M., 2010.

957 Agricultural management explains historic changes in regional soil carbon stocks. *Proc. Natl.*

958 *Acad. Sci. U. S. A.* 107, 14926–14930. <https://doi.org/10.1073/pnas.1002592107>

959 Verma, S.B., Dobermann, A., Cassman, K.G., Walters, D.T., Knops, J.M., Arkebauer, T.J.,

960 Suyker, A.E., Burba, G.G., Amos, B., Yang, H., Ginting, D., Hubbard, K.G., Gitelson, A.A.,

961 Walter-Shea, E.A., 2005. Annual carbon dioxide exchange in irrigated and rainfed maize-

962 based agroecosystems. *Agric. For. Meteorol.* 131, 77–96.

963 <https://doi.org/10.1016/j.agrformet.2005.05.003>

964 Van den Hoof, C., Hanert, E., Vidale, P.L., 2011. Simulating dynamic crop growth with an adapted

965 land surface model – JULES-SUCROS: Model development and validation. *Agric. For.*

966 *Meteorol.* 151, 137–153. <https://doi.org/10.1016/j.agrformet.2010.09.011>

967 Vogel, F.A., 2018. Understanding USDA Crop Forecasts: March 1999 (Classic Reprint).

968 Forgotten Books.

969 Wang, S., Garcia, M., Ibrom, A., Bauer-Gottwein, P., 2020. Temporal interpolation of land surface

970 fluxes derived from remote sensing – results with an unmanned aerial system. *Hydrol. Earth*

971 *Syst. Sci.* 24, 3643–3661. <https://doi.org/10.5194/hess-24-3643-2020>

972 Wattenbach, M., Sus, O., Vuichard, N., Lehuger, S., Gottschalk, P., Li, L., Leip, A., Williams, M.,

973 Tomelleri, E., Kutsch, W.L., Buchmann, N., Eugster, W., Dietiker, D., Aubinet, M., Ceschia,
 974 E., Béziat, P., Grünwald, T., Hastings, A., Osborne, B., Ciais, P., Cellier, P., Smith, P., 2010.
 975 The carbon balance of European croplands: A cross-site comparison of simulation models.
 976 *Agric. Ecosyst. Environ.* 139, 419–453. <https://doi.org/10.1016/j.agee.2010.08.004>
 977 West, T.O., Bandaru, V., Brandt, C.C., Schuh, A.E., Ogle, S.M., 2011. Regional uptake and release
 978 of crop carbon in the United States. *Biogeosciences*. <https://doi.org/10.5194/bg-8-2037-2011>
 979 West, T.O., Brandt, C.C., Baskaran, L.M., Hellwinckel, C.M., Mueller, R., Bernacchi, C.J.,
 980 Bandaru, V., Yang, B., Wilson, B.S., Marland, G., Nelson, R.G., De la Torre Ugarte, D.G.,
 981 Post, W.M., 2010. Cropland carbon fluxes in the United States: increasing geospatial
 982 resolution of inventory-based carbon accounting. *Ecol. Appl.* 20, 1074–1086.
 983 <https://doi.org/10.1890/08-2352.1>
 984 West, T.O., Brandt, C.C., Wilson, B.S., Hellwinckel, C.M., Tyler, D.D., Marland, G., De La Torre
 985 Ugarte, D.G., Larson, J.A., Nelson, R.G., 2008. Estimating Regional Changes in Soil Carbon
 986 with High Spatial Resolution. *Soil Sci. Soc. Am. J.* 72, 285–294.
 987 <https://doi.org/10.2136/sssaj2007.0113>
 988 West, T.O., Brown, M.E., Duren, R.M., Ogle, S.M., Moss, R.H., 2013. Definition, capabilities and
 989 components of a terrestrial carbon monitoring system. *Carbon Management*.
 990 <https://doi.org/10.4155/cmt.13.36>
 991 West, T.O., Marland, G., 2002. A synthesis of carbon sequestration, carbon emissions, and net
 992 carbon flux in agriculture: comparing tillage practices in the United States. *Agric. Ecosyst.*
 993 *Environ.* 91, 217–232. [https://doi.org/10.1016/S0167-8809\(01\)00233-X](https://doi.org/10.1016/S0167-8809(01)00233-X)
 994 Xue, F., Tong, L., Liu, W., Cao, H., Song, L., Ji, S., Ding, R., 2021. Stomatal conductance of
 995 tomato leaves is regulated by both abscisic acid and leaf water potential under combined

996 water and salt stress 1–9. <https://doi.org/10.1111/ppl.13441>

997 Yokohata, T., Kinoshita, T., Sakurai, G., Pokhrel, Y., Ito, A., Okada, M., Satoh, Y., Kato, E., Nitta,
998 T., Fujimori, S., Felfelani, F., Masaki, Y., Iizumi, T., Nishimori, M., Hanasaki, N., Takahashi,
999 K., Yamagata, Y., Emori, S., 2020. MIROC-INTEG-LAND version 1: a global
1000 biogeochemical land surface model with human water management, crop growth, and land-
1001 use change. *Geosci. Model Dev.* 13, 4713–4747. <https://doi.org/10.5194/gmd-13-4713-2020>

1002 Zhang, X., Izaurrealde, R.C., Manowitz, D.H., Sahajpal, R., West, T.O., Thomson, A.M., Xu, M.,
1003 Zhao, K., LeDuc, S.D., Williams, J.R., 2015. Regional scale cropland carbon budgets:
1004 Evaluating a geospatial agricultural modeling system using inventory data. *Environmental*
1005 *Modelling & Software* 63, 199–216. <https://doi.org/10.1016/j.envsoft.2014.10.005>

1006 Zhang, Y., Gurung, R., Marx, E., Williams, S., Ogle, S.M., Paustian, K., 2020. DayCent Model
1007 Predictions of NPP and Grain Yields for Agricultural Lands in the Contiguous U.S. *J.*
1008 *Geophys. Res. Biogeosci.* <https://doi.org/10.1029/2020JG005750>

1009 Zhou, W., Guan, K., Peng, B., Shi, J., Jiang, C., Wardlow, B., Pan, M., Kimball, J.S., Franz, T.E.,
1010 Gentine, P., He, M., Zhang, J., 2020. Connections between the hydrological cycle and crop
1011 yield in the rainfed U.S. Corn Belt. *J. Hydrol.* 590, 125398.
1012 <https://doi.org/10.1016/j.jhydrol.2020.125398>