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### Authors

Fang, Ting  
Huang, Yu-Kai  
Wei, Jinlai  
[et al.](#)

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# Superoxide Release by Macrophages through NADPH Oxidase Activation Dominating Chemistry by Isoprene Secondary Organic Aerosols and Quinones to Cause Oxidative Damage on Membranes

Ting Fang, Yu-Kai Huang, Jinlai Wei, Jessica E. Monterrosa Mena, Pascale S. J. Lakey, Michael T. Kleinman, Michelle A. Digman, and Manabu Shiraiwa\*



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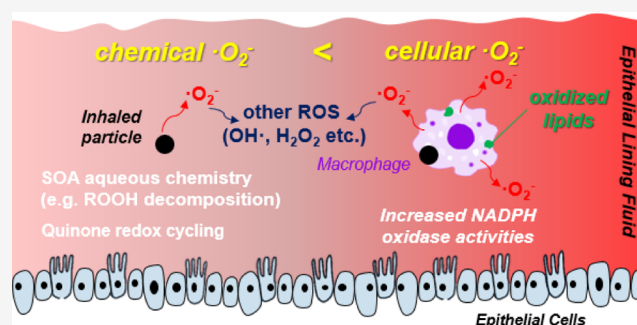
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**ABSTRACT:** Oxidative stress mediated by reactive oxygen species (ROS) is a key process for adverse aerosol health effects. Secondary organic aerosols (SOA) account for a major fraction of fine particulate matter, and their inhalation and deposition into the respiratory tract causes the formation of ROS by chemical and cellular processes, but their relative contributions are hardly quantified and their link to oxidative stress remains uncertain. Here, we quantified cellular and chemical superoxide generation by 9,10-phenanthrenequinone (PQN) and isoprene SOA using a chemiluminescence assay combined with electron paramagnetic resonance spectroscopy as well as kinetic modeling. We also applied cellular imaging techniques to study the cellular mechanism of superoxide release and oxidative damage on cell membranes. We show that PQN and isoprene SOA activate NADPH oxidase in macrophages to release massive amounts of superoxide, overwhelming the superoxide formation by aqueous chemical reactions in the epithelial lining fluid. The activation dose for PQN is 2 orders of magnitude lower than that of isoprene SOA, suggesting that quinones are more toxic. While higher exposures trigger cellular antioxidant response elements, the released ROS induce oxidative damage to the cell membrane through lipid peroxidation. Such mechanistic and quantitative understandings provide a basis for further elucidation of adverse health effects and oxidative stress by fine particulate matter.

**KEYWORDS:** lipid peroxidation, cell membrane fluidity, cellular mechanism, antioxidant response elements, reactive oxygen species, aqueous chemistry, epithelial lining fluid



## INTRODUCTION

Air pollution with high concentrations of fine particulate matter (PM) causes several millions of premature deaths per year globally.<sup>1,2</sup> PM exposure is linked to oxidative stress and inflammation through the generation of reactive oxygen species (ROS) including superoxide ( $\text{O}_2^{\bullet-}$ ), the hydroxyl radical ( $\text{OH}^{\bullet}$ ), and hydrogen peroxide ( $\text{H}_2\text{O}_2$ ).<sup>3</sup> ROS play a central role in physiological processes as signaling molecules,<sup>4</sup> but excess generation of ROS can overwhelm the antioxidant defense capacity to cause oxidative stress and respiratory diseases.<sup>5–7</sup> Among all ROS,  $\text{O}_2^{\bullet-}$  is particularly of great interest as it can be converted by chemical and enzymatic processes into  $\text{H}_2\text{O}_2$ , which is a precursor of highly reactive  $\text{OH}^{\bullet}$  radicals.<sup>8–10</sup> A major fraction of fine PM is secondary organic aerosols (SOA),<sup>11</sup> which are generated *via* atmospheric oxidation of volatile organic compounds (VOCs), forming a myriad of oxygenated compounds such as hydroperoxides and alcohols.<sup>12,13</sup> In the aqueous phase, these compounds can undergo a cascade of chemical reactions to form  $\text{O}_2^{\bullet-}$ .<sup>14</sup> Fine PM also contains redox-active compounds with high oxidative

potential as introduced into the atmosphere from various sources including traffic-related emissions and biomass burning.<sup>15</sup> Upon inhalation and respiratory deposition of PM in the epithelial lining fluid covering human airways, these compounds can trigger redox reactions to convert  $\text{O}_2$  into  $\text{O}_2^{\bullet-}$ .<sup>16,17</sup> Hence, there are a growing number of oxidative potential measurements using acellular assays such as dithiothreitol and ascorbic acid assays emerging as new plausible metrics for PM toxicity.<sup>15,18</sup>

Macrophages are the first cellular responders of the innate immune system that protect the lung from infection through bacteria, microbes, and pathogens by releasing  $\text{O}_2^{\bullet-}$  after phagocytosis through a process called the “respiratory burst”

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due to transient consumption of oxygen.<sup>19</sup> This process also occurs upon exposure to atmospheric PM.<sup>20–22</sup> Thus,  $\bullet\text{O}_2^-$  can be generated both chemically and cellularly upon PM respiratory deposition; however, the relative importance of these pathways is poorly quantified. A previous study has quantified chemical and cellular  $\text{H}_2\text{O}_2$  production from macrophages upon exposure to naphthalene SOA,<sup>23</sup> but very limited research has been conducted in quantifying cellular  $\bullet\text{O}_2^-$ . Cellular release of  $\bullet\text{O}_2^-$  can be triggered by a number of different enzymatic systems such as mitochondrial oxidative phosphorylation, NAD(P)H [reduced nicotinamide adenine dinucleotide (phosphate)] oxidase, and xanthine oxidase.<sup>24</sup> However, very little is known about which  $\bullet\text{O}_2^-$  generation mechanism is activated by inhaled PM. While  $\bullet\text{O}_2^-$  can be converted into a less reactive form (e.g.,  $\text{H}_2\text{O}_2$ ) by SOD, excess  $\bullet\text{O}_2^-$  is known to be cytotoxic as it can interfere with lipids, proteins, and DNA,<sup>6,7,19,25</sup> but a specific consequence of PM-triggered  $\bullet\text{O}_2^-$  is yet to be identified. Several methods have been developed to detect cellular  $\bullet\text{O}_2^-$  using spectroscopic,<sup>26</sup> fluorescence,<sup>27,28</sup> or luminescence<sup>29,30</sup> assays. The Diogenes chemiluminescence assay is suitable for monitoring cellular  $\bullet\text{O}_2^-$  production as Diogenes is a very sensitive  $\bullet\text{O}_2^-$  chemiluminescence enhancer that is non-denaturing to living cells; however, the chemiluminescence readouts represent light intensity without providing a quantitative measure for the production rate or concentration.<sup>30</sup>

Here, we develop a method to quantify superoxide generation by combining the Diogenes chemiluminescence assay with electron paramagnetic resonance (EPR) spectroscopy. RAW 264.7 macrophages, a widely-applied and established macrophage cell line for studying the oxidative stress responses of macrophages<sup>19,31–34</sup> (see also Table S1), are exposed to 9,10-phenanthrenequinone (PQN) and isoprene SOA. PQN is one of the most abundant quinones in atmospheric PM with cytotoxic effects both *in vitro* and *in vivo*.<sup>35</sup> Quinones are important components in anthropogenic SOA as they are generated by the oxidation of polycyclic aromatic hydrocarbons and they can also be directly emitted *via* diesel exhaust, tire wear, and biomass burning associated with soot and humic-like substances.<sup>36,37</sup> Quinones are redox active and accept electrons from antioxidants to form semiquinones,<sup>38,39</sup> which can react with  $\text{O}_2$  to form  $\bullet\text{O}_2^-$ , which can further be converted into  $\text{H}_2\text{O}_2$ .<sup>8</sup> This pathway is a major source of  $\text{H}_2\text{O}_2$  for naphthalene SOA.<sup>40</sup> Isoprene is the most abundant biogenic VOC emitted from plants. The isoprene SOA represents a major component of biogenic PM,<sup>12,13</sup> and it has been found to induce oxidative stress in human lung cells.<sup>41</sup> Quinones and isoprene SOA can also generate superoxide by chemical processes, which we quantify using EPR with a spin-trapping technique and kinetic modeling. The phasor approach to fluorescence lifetime imaging (Phasor-FLIM) is a state-of-the-art cellular imaging technique that has been previously used to study cellular metabolism in detail.<sup>42–45</sup> Here, we apply Phasor-FLIM for the first time to study NADPH oxidase activities as a potential mechanism of cellular  $\bullet\text{O}_2^-$  release after PM exposure. In addition, we apply Laurdan FLIM and the third harmonic generation (THG) microscopy<sup>46</sup> to investigate the impacts of PQN and isoprene SOA on cell membrane fluidity and lipids, respectively.

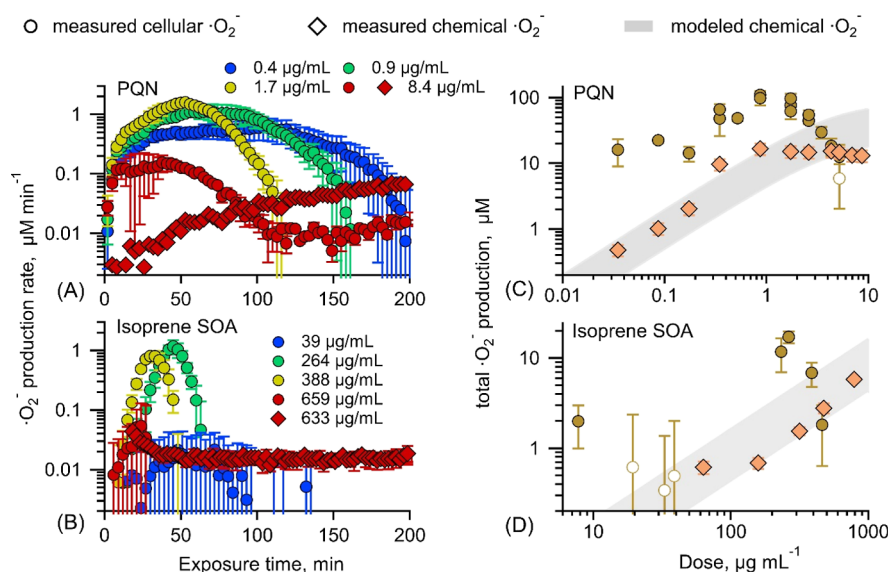
## METHODS

**Sample Collection and Preparation.** Isoprene SOA particles were generated with  $\bullet\text{OH}$  photo-oxidation in a 19 L potential aerosol mass (PAM) chamber<sup>47</sup> (see Figure S1 for the schematic of the system). 100–500  $\mu\text{L}$  of isoprene was placed in an open 1.5 mL amber glass vial, which was kept inside a glass bottle. VOC gases were mixed with a carrier flow of 0.5 L  $\text{min}^{-1}$  of purified air from a zero-air generator (model 7000, Environics) and combined with humidified zero air (Perma Pure humidifier, MH-110-12P-4) prior to the PAM chamber inlet.  $\bullet\text{OH}$  was generated by UV radiation (185 nm) through the photolysis of water molecules with a relative humidity of  $\sim 40\%$ . The flow rate in the chamber was 6.5 L  $\text{min}^{-1}$ , resulting in a mean residence time of approximately 3 min. While  $\bullet\text{OH}$  concentrations are higher ( $\sim 10^{10}$   $\text{cm}^{-3}$ ) than ambient levels ( $\sim 10^6$   $\text{cm}^{-3}$ ), the PAM-generated SOA are found to be similar to ambient and chamber-generated SOA in terms of the yield, oxidation state, hygroscopicity, and chemical composition.<sup>48–50</sup> Multiple PAM experiments were conducted, and the isoprene SOA had a mean particle diameter of 400 nm with a mass concentration of particles of  $\sim 100$ – $1000$   $\mu\text{g m}^{-3}$ , as measured using a scanning mobility particle sizer (SMPS, Grimm Aerosol Technik). The SOA particles were collected on pre-weighed 47 mm polytetrafluoroethylene (PTFE) filters (Millipore FGLP04700, 0.2  $\mu\text{m}$  pore size) at a flow rate of 5 L  $\text{min}^{-1}$  for 6 h. The filter samples were weighed and stored at  $-18$   $^\circ\text{C}$  for less than a week before analysis. The SOA samples were extracted (with an extraction efficiency of  $86 \pm 3\%$ ) and diluted in different volumes of the incomplete medium (Dulbecco's modified Eagle medium, DMEM, GIBCO) to achieve various doses for cell exposure. A 5 mM PQN stock solution was prepared in dimethyl sulfoxide and kept in a freezer ( $-18$   $^\circ\text{C}$ ). The PQN solutions and SOA extracts were diluted in incomplete medium and kept at room temperature (roughly 25  $^\circ\text{C}$ ) for no longer than 2 h before cell exposure.

**Cell Culture and Cytotoxicity.** Macrophage cells (ATCC TIB-71) were obtained and passaged in complete medium (DMEM supplemented with 10% FBS and 1% penicillin streptomycin) until  $>80\%$  confluent. Cells were then seeded at a density of  $4 \times 10^4$  cells/mL with 200  $\mu\text{L}$  per well into 96-well plates (Corning) and incubated at 37  $^\circ\text{C}$  and 5%  $\text{CO}_2$  in an incubator for about 2 h for cells to fully adhere to the bottom of the culture plate. Cell density was kept the same for superoxide measurements and cell imaging and is within the typical ranges used for PM exposure studies in other studies (details in Table S1).

Cell cytotoxicity was measured using the CellTox Green Cytotoxicity Assay, which measures the changes in membrane integrity that may occur as a result of cell death. In brief, after cells were cultured in 96-well plates, the complete medium was replaced by 100  $\mu\text{L}$  of incomplete medium and 50  $\mu\text{L}$  of the CellTox Green reagent for each well. The CellTox Green reagent was produced according to the package directions. To facilitate dye/DNA binding, cells were incubated in the dark for at least 15 min before 50  $\mu\text{L}$  of samples was added to initiate exposure. Fluorescence signals were then measured using a microplate reader (Promega, GloMax) for 4 h with excitation and emission wavelengths of 475 and 525 nm, respectively. The fluorescence signals were given as the relative light unit (RLU).

**Superoxide Measurements.** A Diogenes chemiluminescence assay combined with EPR spectroscopy was applied to



**Figure 1.** Time profiles (A, B) and dose–response curves (C, D) of cellular (circles) and chemical (diamonds)  $\bullet\text{O}_2^-$  production upon exposure to PQN and the isoprene SOA. Markers in panels (A) and (B) are color-coded and labeled with the doses (in  $\mu\text{g mL}^{-1}$ ). Data points with error bars represent the average and uncertainties calculated from error propagation based on variabilities from samples and controls (see the Statistical Analyses section for details). Exposure groups statistically insignificant ( $p$ -value  $> 0.05$ , unpaired  $t$ -test) compared to vehicle controls are plotted as open circles in C and D. Chemical production of  $\bullet\text{O}_2^-$  was also simulated using kinetic models with shaded areas representing model uncertainties.

quantify cellular and chemical  $\bullet\text{O}_2^-$  production with and without superoxide dismutase (SOD) pretreatments. The Diogenes probe reacts with  $\bullet\text{O}_2^-$  to emit flash a chemiluminescence signal that is proportional to the  $\bullet\text{O}_2^-$  production rate. The output of the Diogenes chemiluminescence is in the RLU. To convert the RLU to  $\bullet\text{O}_2^-$  production rates, the Diogenes chemiluminescence assay was calibrated with EPR coupled with a spin-probe technique using the standardized cell-free  $\bullet\text{O}_2^-$  generation system—the hypoxanthine (HX) and xanthine oxidase (XO) system (Figure S2).  $\bullet\text{O}_2^-$  production rates calculated from the EPR spin probe method show a linear relationship with the RLU from chemiluminescence ( $R^2 = 0.993$ , Figure S3). Therefore, the slope from the linear regression was used to convert the Diogenes chemiluminescence data to the  $\bullet\text{O}_2^-$  production rate in the unit of micromoles per minute (see further details in the Supporting Information).

For cellular  $\bullet\text{O}_2^-$  measurements, after the cells were cultured in 96-well plates, the medium was gently removed, and 50  $\mu\text{L}$  of the Diogenes probe made in sterile water was added to each well to incubate for 2 min. Then, 100  $\mu\text{L}$  of incomplete medium (for SOD-pretreated wells, 85  $\mu\text{L}$  of incomplete medium, and 15  $\mu\text{L}$  of 1250 U  $\text{mL}^{-1}$  SOD in PBS buffer were added instead), and 50  $\mu\text{L}$  of the sample was added to start the exposure. For inhibitor experiments that used Apocynin (Apo) as a specific NADPH oxidase inhibitor,<sup>51</sup> 10  $\mu\text{L}$  of Apo (roughly 100  $\mu\text{M}$ ) was added before sample exposure. In order to prevent cross-well interferences, black sealing films (AbsorbMax, EXCEL Scientific, Inc.) were used to seal the bottom of the plates. Continuous luminescence measurements were conducted using the plate reader every 3 min for 4 h to obtain the  $\bullet\text{O}_2^-$  production rate (Figure 1A,B). The total  $\bullet\text{O}_2^-$  production (Figure 1C,D) was obtained by integrating the area under the curve from the kinetics of  $\bullet\text{O}_2^-$  release as a function of exposure time. Incomplete medium and filter blanks were used as vehicle controls for PQN and SOA, respectively. The  $\bullet\text{O}_2^-$  production rate and total production from exposure to

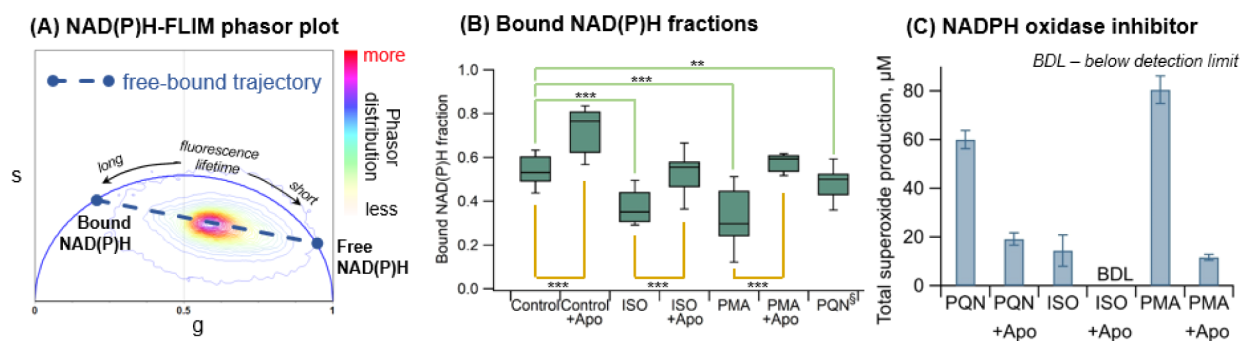
samples were corrected by subtracting those from controls so that the control-corrected data represent cellular  $\bullet\text{O}_2^-$  induced by exposure to samples. Unexposed (*i.e.*, vehicle controls) macrophages release  $\bullet\text{O}_2^-$  at  $\sim 0.06 \mu\text{M min}^{-1}$  with a total production of  $\sim 14 \mu\text{M}$  (Figure S4), which represents  $\bullet\text{O}_2^-$  production from a normal metabolism regulated by mitochondrial respiration.<sup>6</sup> Cell-free wells were also used as “acellular controls” to capture any background signals, which were subtracted from the data with cells when necessary.

The measurements of chemical  $\bullet\text{O}_2^-$  production followed the same protocol as that of cellular  $\bullet\text{O}_2^-$  measurements except cells were absent and the reaction time started when PQN or isoprene SOA particles were dissolved or extracted in incomplete medium (pH 7.4). Accounting for the time it took to prepare samples, chemical  $\bullet\text{O}_2^-$  measurements usually started at 15 min.

**Kinetic Modeling.** A kinetic multi-layer model of the surface and bulk chemistry in the epithelial lining fluid (KM-SUB-ELF)<sup>16</sup> was used to estimate the chemical  $\bullet\text{O}_2^-$  production from PQN. KM-SUB-ELF treats mass transport and chemical reactions involving ROS, antioxidants, surfactants, and redox-active compounds in the epithelial lining fluid. The model considers the redox cycling reactions between ascorbate and quinones to generate  $\bullet\text{O}_2^-$ .<sup>4,25</sup> The model simulations were conducted with PQN doses ranging from 0 to 10  $\mu\text{g mL}^{-1}$ . Initial ascorbate concentrations of 10 and 40  $\mu\text{M}$ <sup>52</sup> were used to obtain total  $\bullet\text{O}_2^-$  production with a reaction time of 4 h. The model results from 10 and 40  $\mu\text{M}$  ascorbate correspond to the upper and lower bounds of the KM-SUB-ELF model in Figure 1C.

Chemical  $\bullet\text{O}_2^-$  production from the isoprene SOA was estimated using a SOA aqueous kinetic model from our previous work.<sup>14</sup> The kinetic model includes chemical reactions of SOA components, ROS coupling reactions, and radical–radical reactions. SOA chemistry includes decomposition of organic hydroperoxides (ROOH), generating  $\bullet\text{OH}$ ,  $\bullet\text{OH}$  oxidation of primary and secondary alcohols (ROH), and





**Figure 2.** NADPH oxidase activities. (A) Phasor plot of the FLIM images from RAW 264.7 macrophages treated with the control (vehicle) and all samples. (B) NAD(P)H bound fractions based on phasor locations in (A) for macrophages treated with the control and samples with and without NADPH oxidase inhibitor apocynin (Apo). (C) Effect of apocynin on total  $\cdot\text{O}_2^-$  production. Bars with error bars represent the average from triplicates and the standard deviation. Unpaired *t*-test, \*\*\**p* < 0.0001 and \*\**p* < 0.001. ISO and PMA denote isoprene SOA and phorbol 12-myristate 13-acetate, respectively. <sup>§</sup>FLIM images were taken from cells after 10 min exposure of PQN, washed with PBS buffer, and replaced with fresh incomplete medium.

the subsequent reaction with  $\text{O}_2$  to form  $\alpha$ -hydroxyperoxy radicals, which can decompose to generate the hydroperoxy radical  $\text{HO}_2\cdot$ , and  $\cdot\text{OH}$  oxidation of SOA components. The model simulations were conducted with initial SOA concentrations ranging from 0 to  $1000 \mu\text{g mL}^{-1}$ , 50% of ROH, and 1–4% ROOH of total SOA mass. Model uncertainties are associated with uncertainties in the relative abundance of ROOH and ROH groups in the SOA.

#### Phasor Approach to Fluorescence Lifetime Imaging.

The NADPH oxidase complex is an important source of  $\cdot\text{O}_2^-$  in phagocytosis as activated by bacterial products and cytokines.<sup>53</sup> The activation of NADPH oxidase generates  $\cdot\text{O}_2^-$  through the reaction:  $\text{NADPH} + 2\text{O}_2 \rightarrow \text{NADP}^+ + \text{H}^+ + 2\cdot\text{O}_2^-$ .<sup>54,55</sup> Here, we utilize the auto-fluorescence of NADPH measured using FLIM combined with experiments using NADPH oxidase activator PMA<sup>56</sup> and inhibitor Apocynin to study the NADPH oxidase activities. It should be noted that NADH (nicotinamide adenine dinucleotide) has identical fluorescence properties to NADPH,<sup>57</sup> and thus, FLIM cannot differentiate between NADH and NADPH, and we denote the FLIM signals as NAD(P)H. Note that NADH reacts with oxygen to produce either water or  $\text{H}_2\text{O}_2$ <sup>58</sup> or produces significantly lower  $\cdot\text{O}_2^-$  than NADPH.<sup>59</sup> NAD(P)H expresses in two forms inside cells, bound and free states. Each state has its own fluorescence lifetime, and in the Phasor-FLIM method, the NAD(P)H fluorescence lifetime is visualized in polar coordinates by calculating the Fourier sine and cosine transformations of the FLIM photon histogram curve.<sup>43</sup> Since oxidized  $\text{NAD(P)}^+$  does not emit auto-fluorescence, the conversion of NAD(P)H to  $\text{NAD(P)}^+$  reduces the fraction of bound NAD(P)H and causes a shift from the bound end to free end of the NAD(P)H free-bound trajectory.<sup>60,61</sup> Therefore, the relative locations on the trajectory can be used to obtain the bound NAD(P)H fractions, and the decrease of bound NAD(P)H fractions indicates the activation of NADPH oxidase activities (see further details in the Supporting Information).

For FLIM imaging, cells were seeded at a density of  $4 \times 10^4$  cells  $\text{mL}^{-1}$  and incubated overnight before exposure to PQN and isoprene SOA for imaging. PQN and SOA in media in the absence of cells produced low fluorescence background signals. Cells were imaged in a  $37^\circ\text{C}$  and 5%  $\text{CO}_2$  environment before exposure and at different time points after the addition of samples (see Table S2). Since NAD(P)H mainly resides in cell

membranes and cytoplasm<sup>62</sup> and since the fluorescence lifetimes of nucleus do not change significantly before and after sample exposure (Figure S5), the nucleus was cropped out, that is, only the fluorescence signals from the cell membranes and cytoplasm were included in calculating the bound NAD(P)H fractions. The bound NAD(P)H fractions from all data points for each sample were averaged from multiple cells ( $N = 12\text{--}23$ ) and are shown in Figure 2B. PQN exhibited an unusually long fluorescence lifetime distribution, and the phasors from PQN fall outside of the NAD(P)H free-bound trajectory. We conducted hyperspectral imaging to confirm that the long fluorescence lifetime was caused by phosphorescence from triplets of PQN (see the Supporting Information and Figure S6).<sup>63</sup> To avoid the lifetime of phosphorescence interfering with the NAD(P)H fluorescence lifetime, we used PBS buffer to wash cells after 10 min of incubation. After replacing with fresh incomplete medium, cells were loaded back to the FLIM system for imaging. Using this method, the lifetime phasors of PQN fall onto the NAD(P)H free-bound trajectory, as shown in Figure S5.

The FLIM Laurdan imaging technique was used to measure cell membrane fluidity as a result of sample exposure.<sup>64</sup> Laurdan (6-dodecanoyl-2-dimethylaminonaphthalene) is a fluorescent membrane marker used to investigate membrane fluidity. Shorter Laurdan fluorescence lifetimes correspond to increases in membrane fluidity due to lipid peroxidation. After seeding cells in an eight-well plate, complete medium was replaced with incomplete medium, and Laurdan dye was added to each well to achieve a final concentration of  $5 \mu\text{M}$  and incubated under the cellular condition for 30 min before exposure to the control, PQN, PMA, and isoprene SOA. After some period of exposure, the cells ( $4 \times 10^4$  cells  $\text{mL}^{-1}$ ) were loaded to the FLIM system for imaging at 800 nm excitation wavelength (see further details in the Supporting Information and Table S2).

**Third Harmonic Generation Imaging.** The THG imaging technique was applied to detect the accumulation of lipids inside the cells after exposure to isoprene SOA and PQN.<sup>46</sup> The excitation of the THG signal requires laser wavelengths of up to  $\sim 1000$  nm, which conventional tunable Titanium Sapphire lasers could not achieve. The THG imaging was carried out in the deep imaging via emission recovery (DIVER) system from the LFD, which uses Spectra Physics Insight DS + femtosecond laser tunable in the range of 68–

1300 nm. The actual focal depth difference was found to be  $\sim 0.75$  mm for different wavelength excitations, which for NAD(P)H is 740 nm and for THG is 1050 nm. The THG signal is generated at the interface between media with the difference in the third-order nonlinear susceptibility, refractive index, and dispersion and can be used to detect lipids. The THG images were taken before and after 10 min exposure to isoprene SOA and PQN. Note that THG microscopy is not affected by long-lifetime phosphorescence from PQN.

**Statistical Analyses.** For Diogenes chemiluminescence data, for each dose, triplicate measurements were performed on the same experimental day along with vehicle and acellular controls, which were used to correct the experimental data. Experiments with some different doses were performed on different days. One-way ANOVA followed by Tukey's post-hoc test on the total superoxide production from the same dose of PQN ( $1.74 \mu\text{g mL}^{-1}$ ) performed on different days shows that no significant difference was determined between the mean of data ( $p > 0.1$ ). The time profile of superoxide production and total production are presented as the mean of the triplicate measurements with error bars representing the uncertainties propagated from the standard deviations of triplicate measurements of samples and controls. For imaging, all data points for each sample were averaged from multiple cells ( $N > 10$ ). Results were analyzed by using unpaired student *t*-tests. The results were considered statistically significant at a  $p$ -value  $\leq 0.05$  for all exposure groups compared to control groups.

## RESULTS AND DISCUSSION

**Dose–Response Relationship of Superoxide Production.** Figure 1A,B shows the measured temporal evolution of control-corrected chemical and cellular  $\bullet\text{O}_2^-$  production rates upon exposure to selected doses of PQN and isoprene SOA, respectively. Note that similar temporal trends are observed in other doses. The cellular  $\bullet\text{O}_2^-$  production rate increases to  $\sim 0.4 \mu\text{M min}^{-1}$  when exposed to  $0.4 \mu\text{g mL}^{-1}$  PQN. Upon an increase of the dose to  $0.9$  and  $1.7 \mu\text{g mL}^{-1}$ , cellular  $\bullet\text{O}_2^-$  production rates increase and reach higher maxima at  $\sim 1$  and  $\sim 1.6 \mu\text{M min}^{-1}$  after  $\sim 50$  min, respectively, followed by decreases over time. At a higher dose of  $8.4 \mu\text{g mL}^{-1}$ , macrophages respond faster but reach a lower peak at  $\sim 0.2 \mu\text{M min}^{-1}$  at 40 min and decrease more swiftly afterward. At all doses, cellular  $\bullet\text{O}_2^-$  production rates eventually decrease to the basal level, indicating the completion of cellular response after exposure to samples. In the absence of cells, PQN with a dose of  $8.4 \mu\text{g mL}^{-1}$  triggers redox reactions of semiquinone that can react with  $\text{O}_2$  to produce  $\bullet\text{O}_2^-$ . The reduction of PQN to its semiquinone form is the limiting step, leading to a gradual increase of  $\text{O}_2^-$  production rate to reach  $\sim 0.1 \mu\text{M min}^{-1}$  over a few hours.

Similar temporal trends are observed for cellular  $\bullet\text{O}_2^-$  release from the isoprene SOA (Figure 1B), while much higher doses are required to trigger  $\bullet\text{O}_2^-$  release from macrophages. Upon increasing doses from  $39$  to  $659 \mu\text{g mL}^{-1}$  of the isoprene SOA, cellular  $\bullet\text{O}_2^-$  production rates reach their peaks above  $1 \mu\text{M min}^{-1}$  for  $264 \mu\text{g mL}^{-1}$  and  $\sim 0.05 \mu\text{M min}^{-1}$  for  $659 \mu\text{g mL}^{-1}$ . While PQN stimulates macrophages for  $\bullet\text{O}_2^-$  release for a few hours, the isoprene SOA induces it for less than 2 h. In the absence of cells,  $633 \mu\text{g mL}^{-1}$  of the isoprene SOA produces  $\bullet\text{O}_2^-$  via aqueous reactions of SOA components with a comparable rate to the peak cellular release rate. Compared to the cellular release, the chemical  $\bullet\text{O}_2^-$  production rate from the isoprene SOA stays

relatively constant (Figures 1B and S7) due to the relatively fast onset of aqueous reactions. Note that a previous study by Tong *et al.*<sup>40</sup> did not observe the formation of chemical  $\bullet\text{O}_2^-$  by the isoprene SOA in the surrogate lung fluid (pH of 7.4) using the EPR spin trapping technique. We have very recently found that the EPR spin trap method is not ideal to measure  $\bullet\text{O}_2^-$  in neutral conditions due to the low trapping rate of  $\bullet\text{O}_2^-$ ,<sup>65</sup> which likely explains the difference between Tong *et al.* and this work.

The dose–response relationships are shown in Figure 1C,D, where responses are given as total  $\bullet\text{O}_2^-$  production as obtained from the integration of control-corrected  $\bullet\text{O}_2^-$  production rates over exposure time. The measured chemical total  $\bullet\text{O}_2^-$  production increases with an increase of dose, and it plateaus at  $\sim 10 \mu\text{M}$ . PQN generates  $\bullet\text{O}_2^-$  via redox-cycling reactions with antioxidants,<sup>4,16</sup> while the isoprene SOA yields  $\bullet\text{O}_2^-$  via a series of aqueous reactions including decomposition of  $\alpha$ -hydroperoxyl radicals and organic hydroperoxides as well as OH oxidation of primary or secondary alcohols.<sup>14</sup> We applied kinetic models that include these chemical mechanisms to simulate chemical  $\bullet\text{O}_2^-$  production,<sup>14,16</sup> producing consistent results as the measurements within model uncertainties (shaded areas). While the model does not perfectly reproduce the observed temporal evolution of the  $\bullet\text{O}_2^-$  production rate (Figure S7), this agreement is still remarkable, given that the reaction system is very complicated with numerous reactions involved;<sup>14</sup> the discrepancies may be due to incomplete chemical mechanisms, untreated species contained in media, and the inexplicit treatment of pH that can modulate ROS formation.<sup>66</sup>

Interestingly, cellular  $\bullet\text{O}_2^-$  release induced by PQN and the isoprene SOA shows inverted U shape dose–response distributions. Under low doses, macrophages release  $\bullet\text{O}_2^-$  under regular cellular metabolisms with a control level of  $\sim 14 \mu\text{M}$  (Figure S4B), which is higher than the chemical  $\bullet\text{O}_2^-$  production ( $< 0.05 \mu\text{M}$ ). Once the dose reaches a threshold, macrophages are activated, and cellular total  $\bullet\text{O}_2^-$  production increases sharply, further dominating over chemical  $\bullet\text{O}_2^-$  production. The onset dose of cell activation to release an excess amount of  $\bullet\text{O}_2^-$  for PQN ( $\sim 0.2 \mu\text{g mL}^{-1}$ ) is much lower than for the isoprene SOA ( $> 40 \mu\text{g mL}^{-1}$ ), indicating that macrophages are more sensitive to PQN than to the isoprene SOA in producing ROS. This suggests that quinones are intrinsically more toxic than the isoprene SOA, consistent with previous studies that show that anthropogenic aromatics have higher toxicity compared to biogenic aerosols.<sup>67–69</sup>

Cellular  $\bullet\text{O}_2^-$  production induced by PQN and isoprene SOA then decreases at higher doses, surpassed by chemical  $\bullet\text{O}_2^-$  production (PQN at  $\sim 5 \mu\text{g mL}^{-1}$  and SOA at  $\sim 450 \mu\text{g mL}^{-1}$ ). The decreases of cellular  $\bullet\text{O}_2^-$  release at higher doses (Figure 1C,D) and longer exposure time (Figure 1A,B) are not due to cell death. Cell cytotoxicity measurements show that cells exposed to the doses of up to  $5 \mu\text{g mL}^{-1}$  for PQN and  $923 \mu\text{g mL}^{-1}$  for the isoprene SOA have similar fluorescence signals (unpaired *t*-test,  $p$ -value  $> 0.05$ ) to that of filter blanks (fresh filter without particle collection), which were much lower than those from the positive control lysed cells ( $p$ -value  $< 0.05$ ) (Figure S8), confirming that cells are alive during the whole course of exposure. Instead, the decreased cellular release of  $\bullet\text{O}_2^-$  is most likely due to redox homeostasis by cells upregulating the antioxidant response elements to scavenge ROS for protection against proinflammatory effects.<sup>70</sup> Previous

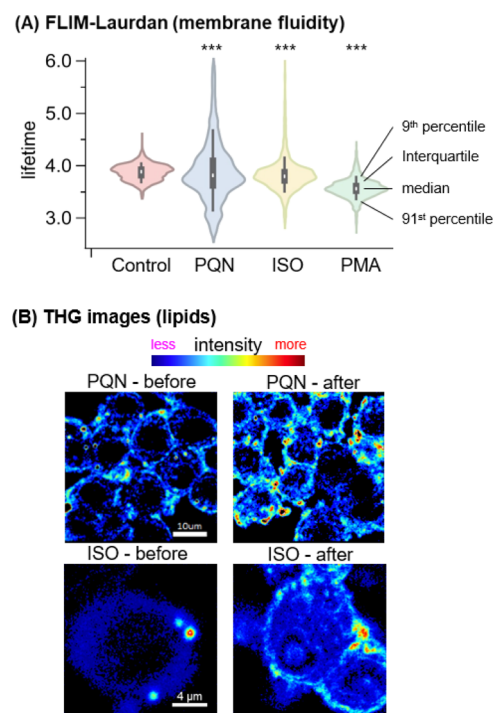
studies have found that antioxidant enzymes can be upregulated by the transcription factor Nrf2 when oxidative stress is induced upon exposure to ambient and diesel exhaust PM containing polycyclic aromatic hydrocarbons and quinones.<sup>20,71,72</sup> Based on the kinetics of cellular  $\bullet\text{O}_2^-$  releases (Figure 1A,B), the differences in the rates of decrease and time to reach peak production rates from different doses suggest that the activation of antioxidant response elements is also dose-dependent. The inverted U shape time profiles and the dose–response relationship also imply that the cellular antioxidant defense mechanisms occur after the activation of macrophage to release  $\bullet\text{O}_2^-$ .

The conversion of doses to ambient PM concentrations is highly complex as respiratory deposition is largely affected by the nature of breathing (nasal vs oral, tidal volume, and frequency), individual differences in the lung anatomy, the airflow patterns in the lung airways, and the presence of deposition hot spots in lung.<sup>73</sup> Phalen *et al.* (2006) demonstrated by dosimetry model predictions that the surface PM deposition can reach up to  $85.5 \mu\text{g cm}^{-2}$  but vary by up to 3 orders of magnitude in the tracheobronchial region.<sup>74</sup> With a bottom area of  $0.32 \text{ cm}^2$  and a working volume of  $0.2 \text{ mL}$  of a 96-well plate, the surface doses for the SOA and PQN in the current study are  $5\text{--}412$  and  $0.02\text{--}5.3 \mu\text{g cm}^{-2}$ , respectively. Combining all the uncertainties and variabilities (up to 4 orders of magnitude) and also considering that the surface area of the alveolar region is about 200 times larger and the deposition in the alveolar region is up to 3 folds higher than in the tracheobronchial region,<sup>75</sup> the applied surface doses are within relevant ranges for real-life exposure scenarios and consistent with doses applied in previous submerged cell exposure studies (details in Table S1).

**Cellular Superoxide Release by NADPH Oxidase Activation.** We then applied cellular imaging techniques to study the mechanism of cellular  $\bullet\text{O}_2^-$  release. Selected doses that activate massive cellular  $\bullet\text{O}_2^-$  release (with minor contribution from chemical  $\bullet\text{O}_2^-$  formation) were used for exposure (Table S2). The phasors for cell membranes and cytoplasm mainly fall along a metabolic trajectory represented by a line joining positions of free NAD(P)H and protein-bound NAD(P)H on a phasor plot (Figure 2A).<sup>43,45</sup> The relative locations on the trajectory were used to obtain the bound NAD(P)H fractions (Figure 2B). Macrophages exposed to PQN and the isoprene SOA show lower bound NAD(P)H fractions than the controls, indicating that the bound-state NADPH is oxidized to non-fluorescent  $\text{NADP}^+$ , releasing superoxide:  $\text{NADPH} + 2\text{O}_2 \rightarrow \text{NADP}^+ + \text{H}^+ + 2\bullet\text{O}_2^-$ .<sup>54,55</sup> A bound-to-free shift of NAD(P)H is observed upon exposure to phorbol 12-myristate 13-acetate (PMA), commonly used as an inducer to activate NADPH oxidase and increase endogenous  $\bullet\text{O}_2^-$  production (Figure S9). Additionally, apocynin (or Apo), a specific NADPH oxidase inhibitor, results in higher bound NAD(P)H fractions (Figure 2B) in cells than those without inhibitors. Note that inhibitor experiments for PQN are not available due to the interference by phosphorescence of PQN. Taking into account the difference in the doses and exposure times for FLIM (Table S2), the bound fractions in Figure 2B for PQN and the isoprene SOA are not directly comparable. Consequently, apocynin reduces cellular  $\bullet\text{O}_2^-$  production substantially upon exposure to PQN, the isoprene SOA, and PMA (Figure 2C). These observations strongly indicate that PQN and the isoprene SOA trigger cellular  $\bullet\text{O}_2^-$  production mainly through activating the NADPH oxidase. We observe

that the bound NAD(P)H fraction for controls with Apo is higher than isoprene SOA and PMA with Apo. The higher fraction in controls may be explained by the increase of NADH bound fractions from the increase of mitochondrial oxidative phosphorylation (OXPHOS), as a recent study showed that Apo can enhance ATP production and mitochondrial membrane potential, which are determined by the regulation of OXPHOS.<sup>76</sup> This may further suggest that the exposure to PQN and the isoprene SOA induced a certain level of mitochondrial dysfunction, although future dedicated studies are necessary to further investigate this aspect.

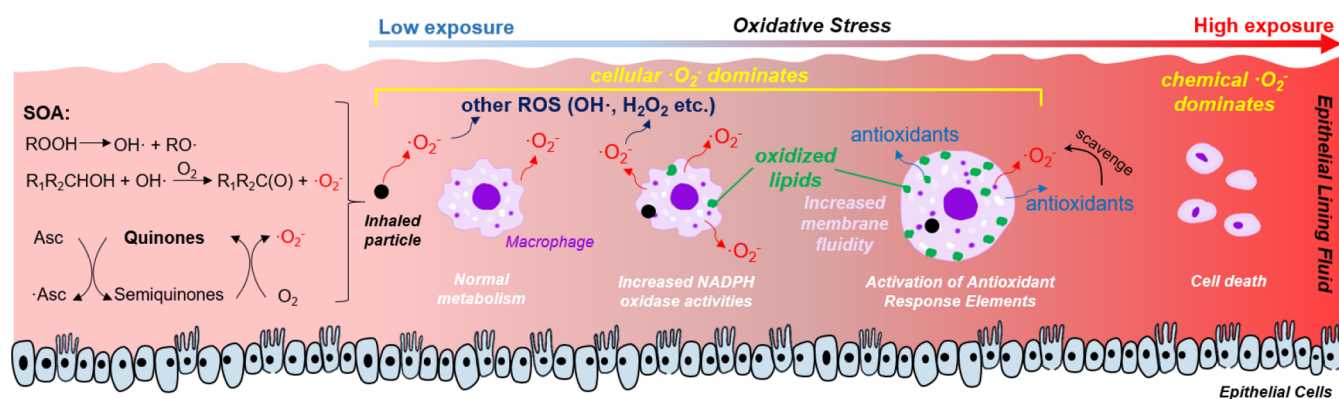
**Oxidative Stress on Cell Membranes.** To investigate the impact of PQN and the isoprene SOA on cell membranes, we quantified the lifetime changes of a solvatochromic probe (Laurdan) and the THG images of cells after exposure. The images of FLIM-Laurdan are shown in Figure S10. Since Laurdan dye is incorporated into the hydrophobic phase in the membrane,<sup>64</sup> only the fluorescence signals from cell membranes were selected for fluorescence lifetime calculation. A total of 15 cells from each sample were averaged and are presented as violin plots in Figure 3A. We observe significant



**Figure 3.** Oxidative stress on cell membranes. (A) Membrane fluidity measured by Laurdan FLIM. (B) Locations and intensity of lipid accumulation from THG imaging on cells treated with PQN ( $1.74 \mu\text{g mL}^{-1}$ ) and isoprene SOA (ISO,  $305 \mu\text{g mL}^{-1}$ ) before and after 10 min exposure. Unpaired *t*-test, \*\*\**p* < 0.0001.

decreases in the fluorescence lifetimes of the Laurdan probe on cell membranes from macrophages exposed to PQN, isoprene SOA, and PMA, indicating that these samples cause an increase in membrane fluidity, an important parameter in membrane integrity and cell health. In addition, the THG microscopy imaging shows that exposure to PQN and the isoprene SOA causes an increase of the THG signal around cell membranes, suggesting the accumulation of lipids (Figure 3B). Bright-field cell images also suggest that macrophages may have taken up oxidized low-density lipoprotein (LDL), forming foam cells





**Figure 4.** Multi-tier chemical and cellular response mechanisms upon PM deposition in the epithelial lining fluid. Cellular  $\cdot\text{O}_2^-$  release via activation of NADPH oxidase mostly dominates over chemical  $\cdot\text{O}_2^-$  production, causing lipid peroxidation and activation of antioxidant response elements.

(Figure S11), a key event suggested to involve the activation of NADPH oxidase observed in previous studies.<sup>77,78</sup> The doses for PQN and the isoprene SOA are 1.74 and 305  $\mu\text{g mL}^{-1}$ , respectively, and cellular  $\cdot\text{O}_2^-$  dominates chemical production. Therefore, with the revelation of increased membrane fluidity and laden lipids around the cell membranes, we suggest that macrophages are undergoing lipid peroxidation caused by  $\cdot\text{O}_2^-$  or other ROS produced thereafter. Note that the Diogenes probe can diffuse through the cell membrane, but SOD is membrane-impermeant. The addition of SOD largely eliminated the chemiluminescence intensity, suggesting that the observed cellular  $\cdot\text{O}_2^-$  signals were primarily from extracellular  $\cdot\text{O}_2^-$ . This is consistent with NADPH oxidase being a transmembrane protein that regulates redox signaling by reducing extracellular  $\text{O}_2$  to  $\cdot\text{O}_2^-$ .<sup>53</sup> This is also consistent with the observed oxidative stress induced by PQN and the isoprene SOA, which occurred mainly on the cell membrane by lipid peroxidation. Liu *et al.* (2022) suggested that the diffusion of chemically formed  $\text{H}_2\text{O}_2$  into cells is a main contributor to intracellular  $\text{H}_2\text{O}_2$  upon exposure to the naphthalene SOA.<sup>23</sup> As the chemical lifetime of  $\cdot\text{O}_2^-$  is much shorter than that of  $\text{H}_2\text{O}_2$ ,  $\cdot\text{O}_2^-$  is not expected to readily pass through the cell membranes. Although the penetration of  $\cdot\text{O}_2^-$  through the membrane could still occur with the assistance of anion channels,<sup>79</sup> our study indicates that this pathway is not significant. Note that the dismutation of extracellular  $\cdot\text{O}_2^-$  can occur and the product  $\text{H}_2\text{O}_2$  can diffuse across the membrane into the cytoplasm to cause oxidative damage intracellularly.

**Multi-Tier Response Mechanism.** Figure 4 summarizes multi-tier chemical and cellular response mechanisms upon PQN and SOA exposure in the epithelial lining fluid. At low doses,  $\cdot\text{O}_2^-$  is mainly produced from normal cellular metabolism *via* mitochondrial respiration with minor contributions from chemical reactions. After a threshold dose to macrophages, NADPH oxidase activities are upregulated for respiratory burst, releasing massive amounts of  $\cdot\text{O}_2^-$ , which can cause oxidative stress by increasing cell membrane fluidity through lipid peroxidation. Further increases of doses lead to the activation of antioxidant response elements, reducing the net cellular  $\cdot\text{O}_2^-$  production. At very high doses and long exposure times, chemical  $\cdot\text{O}_2^-$  production becomes comparably important or dominant if the escalation of oxidative stress leads to cell death. Given that cellular  $\cdot\text{O}_2^-$  release mostly dominates over its chemical production by PM reactive and redox-active components, widely-applied acellular assays that

measure oxidative potential and ROS activity may need to be interpreted with caution.

The mechanistic understandings obtained in this study provide a basis for further elucidation of adverse health effects and oxidative stress upon respiratory deposition of PM. Future studies should be extended to include other chemical compounds that have high oxidative potential and redox activity such as transition metals; the presence of metals in ambient PM might enhance the importance of chemical ROS.<sup>17,80–82</sup> While the current work is based on the standard submerged cell culture method, this classical condition should be extended to represent more realistic conditions with multiple types of cells including macrophages and epithelial and endothelial cells to simulate synthetic interactions between cell populations<sup>83</sup> with an application of the air–liquid interface for simulating exposure and respiratory deposition of aerosol particles.<sup>67</sup>

## ■ ASSOCIATED CONTENT

### Supporting Information

The Supporting Information is available free of charge at <https://pubs.acs.org/doi/10.1021/acs.est.2c03987>.

Additional experimental details, materials, and methods, including chemicals and materials, SOA generation by the PAM chamber, cell culture, cytotoxicity, calibration of the Diogenes chemiluminescence assay and EPR spin probe, Phasor-FLIM, phosphorescence from PQN, and FLIM-Laurdan (PDF)

## ■ AUTHOR INFORMATION

### Corresponding Author

Manabu Shiraiwa – Department of Chemistry, University of California, Irvine 92697 California, United States;  
[orcid.org/0000-0003-2532-5373](https://orcid.org/0000-0003-2532-5373); Email: [m.shiraiwa@uci.edu](mailto:m.shiraiwa@uci.edu)

### Authors

Ting Fang – Department of Chemistry, University of California, Irvine 92697 California, United States;  
[orcid.org/0000-0002-4845-2749](https://orcid.org/0000-0002-4845-2749)

Yu-Kai Huang – Department of Biomedical Engineering, University of California, Irvine 92697 California, United States

Jinlai Wei – Department of Chemistry, University of California, Irvine 92697 California, United States;  
[orcid.org/0000-0002-4741-9015](https://orcid.org/0000-0002-4741-9015)



Jessica E. Monterrosa Mena – Division of Occupational and Environmental Medicine, University of California, Irvine 92697 California, United States; [orcid.org/0000-0002-3315-4431](https://orcid.org/0000-0002-3315-4431)

Pascale S. J. Lakey – Department of Chemistry, University of California, Irvine 92697 California, United States; [orcid.org/0000-0003-2923-4073](https://orcid.org/0000-0003-2923-4073)

Michael T. Kleinman – Division of Occupational and Environmental Medicine, University of California, Irvine 92697 California, United States

Michelle A. Digman – Department of Biomedical Engineering, University of California, Irvine 92697 California, United States

Complete contact information is available at:  
<https://pubs.acs.org/10.1021/acs.est.2c03987>

## Notes

The authors declare no competing financial interest.

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## REFERENCES

- (1) Lelieveld, J.; Evans, J. S.; Fnais, M.; Giannadaki, D.; Pozzer, A. The contribution of outdoor air pollution sources to premature mortality on a global scale. *Nature* **2015**, *525*, 367–371.
- (2) Cohen, A. J.; Brauer, M.; Burnett, R.; Anderson, H. R.; Frostad, J.; Estep, K.; Balakrishnan, K.; Brunekreef, B.; Dandona, L.; Dandona, R.; Feigin, V.; Freedman, G.; Hubbell, B.; Jobling, A.; Kan, H.; Knibbs, L.; Liu, Y.; Martin, R.; Morawska, L.; Pope, C. A., III; Shin, H.; Straif, K.; Shaddick, G.; Thomas, M.; van Dingenen, R.; van Donkelaar, A.; Vos, T.; Murray, C. J. L.; Forouzanfar, M. H. Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: an analysis of data from the Global Burden of Diseases Study 2015. *Lancet* **2017**, *389*, 1907–1918.
- (3) Shiraiwa, M.; Ueda, K.; Pozzer, A.; Lammel, G.; Kampf, C. J.; Fushimi, A.; Enami, S.; Arangio, A. M.; Fröhlich-Nowoisky, J.; Fujitani, Y.; Furuyama, A.; Lakey, P. S. J.; Lelieveld, J.; Lucas, K.; Morino, Y.; Pöschl, U.; Takahama, S.; Takami, A.; Tong, H.; Weber, B.; Yoshino, A.; Sato, K. Aerosol Health Effects from Molecular to Global Scales. *Environ. Sci. Technol.* **2017**, *51*, 13545–13567.
- (4) D'Autréaux, B.; Toledano, M. B. ROS as signalling molecules: mechanisms that generate specificity in ROS homeostasis. *Nat. Rev. Mol. Cell Biol.* **2007**, *8*, 813–824.
- (5) Nel, A. Air pollution-related illness: Effects of particles. *Science* **2005**, *308*, 804–806.

(6) Winterbourn, C. C. Reconciling the chemistry and biology of reactive oxygen species. *Nat. Chem. Biol.* **2008**, *4*, 278–286.

(7) Sies, H.; Berndt, C.; Jones, D. P. Oxidative Stress. *Annu. Rev. Biochem.* **2017**, *86*, 715–748.

(8) Lelieveld, S.; Wilson, J.; Dovrou, E.; Mishra, A.; Lakey, P. S. J.; Shiraiwa, M.; Pöschl, U.; Berkemeier, T. Hydroxyl Radical Production by Air Pollutants in Epithelial Lining Fluid Governed by Interconversion and Scavenging of Reactive Oxygen Species. *Environ. Sci. Technol.* **2021**, *55*, 14069–14079.

(9) Hayyan, M.; Hashim, M. A.; AlNashef, I. M. Superoxide Ion: Generation and Chemical Implications. *Chem. Rev.* **2016**, *116*, 3029–3085.

(10) Babior, B. M. Superoxide: a two-edged sword. *Braz. J. Med. Biol. Res.* **1997**, *30*, 141–155.

(11) Jimenez, J. L.; Canagaratna, M. R.; Donahue, N. M.; Prevot, A. S. H.; Zhang, Q.; Kroll, J. H.; DeCarlo, P. F.; Allan, J. D.; Coe, H.; Ng, N. L.; Aiken, A. C.; Docherty, K. S.; Ulbrich, I. M.; Grieshop, A. P.; Robinson, A. L.; Duplissy, J.; Smith, J. D.; Wilson, K. R.; Lanz, V. A.; Hueglin, C.; Sun, Y. L.; Tian, J.; Laaksonen, A.; Raatikainen, T.; Rautiainen, J.; Vaattovaara, P.; Ehn, M.; Kulmala, M.; Tomlinson, J. M.; Collins, D. R.; Cubison, M. J.; Dunlea, E. J.; Huffman, J. A.; Onasch, T. B.; Alfarra, M. R.; Williams, P. I.; Bower, K.; Kondo, Y.; Schneider, J.; Drewnick, F.; Borrmann, S.; Weimer, S.; Demerjian, K.; Salcedo, D.; Cottrell, L.; Griffin, R.; Takami, A.; Miyoshi, T.; Hatakeyama, S.; Shimono, A.; Sun, J. Y.; Zhang, Y. M.; Dzepina, K.; Kimmel, J. R.; Sueper, D.; Jayne, J. T.; Herndon, S. C.; Trimborn, A. M.; Williams, L. R.; Wood, E. C.; Middlebrook, A. M.; Kolb, C. E.; Baltensperger, U.; Worsnop, D. R. Evolution of organic aerosols in the atmosphere. *Science* **2009**, *326*, 1525–1529.

(12) Claeys, M.; Graham, B.; Vas, G.; Wang, W.; Vermeylen, R.; Pashynska, V.; Cafmeyer, J.; Guyon, P.; Andreae, M. O.; Artaxo, P.; Maenhaut, W. Formation of secondary organic aerosols through photooxidation of isoprene. *Science* **2004**, *303*, 1173–1176.

(13) Wennberg, P. O.; Bates, K. H.; Crounse, J. D.; Dodson, L. G.; McVay, R. C.; Mertens, L. A.; Nguyen, T. B.; Praske, E.; Schwantes, R. H.; Smarte, M. D.; St Clair, J. M.; Teng, A. P.; Zhang, X.; Seinfeld, J. H. Gas-Phase Reactions of Isoprene and Its Major Oxidation Products. *Chem. Rev.* **2018**, *118*, 3337–3390.

(14) Wei, J.; Fang, T.; Wong, C.; Lakey, P. S. J.; Nizkorodov, S. A.; Shiraiwa, M. Superoxide Formation from Aqueous Reactions of Biogenic Secondary Organic Aerosols. *Environ. Sci. Technol.* **2021**, *55*, 260–270.

(15) Daellenbach, K. R.; Uzu, G.; Jiang, J.; Cassagnes, L.-E.; Leni, Z.; Vlachou, A.; Stefanelli, G.; Canonaco, F.; Weber, S.; Segers, A.; Kuenen, J. J. P.; Schaap, M.; Favez, O.; Albinet, A.; Aksoyoglu, S.; Dommen, J.; Baltensperger, U.; Geiser, M.; El Haddad, I.; Jaffrezo, J.-L.; Prévôt, A. S. H. Sources of particulate-matter air pollution and its oxidative potential in Europe. *Nature* **2020**, *587*, 414–419.

(16) Lakey, P. S.; Berkemeier, T.; Tong, H.; Arangio, A. M.; Lucas, K.; Pöschl, U.; Shiraiwa, M. Chemical Exposure-response Relationship between Air Pollutants and Reactive Oxygen Species in the Human Respiratory Tract. *Sci. Rep.* **2016**, *6*, 32916.

(17) Fang, T.; Lakey, P. S. J.; Weber, R. J.; Shiraiwa, M. Oxidative Potential of Particulate Matter and Generation of Reactive Oxygen Species in Epithelial Lining Fluid. *Environ. Sci. Technol.* **2019**, *53*, 12784–12792.

(18) Gao, D.; Ripley, S.; Weichenthal, S.; Godri Pollitt, K. J. G. Ambient particulate matter oxidative potential: Chemical determinants, associated health effects, and strategies for risk management. *Free Radical Biol. Med.* **2020**, *151*, 7–25.

(19) Forman, H. J.; Torres, M. Reactive Oxygen Species and Cell Signaling. *Am. J. Respir. Crit. Care Med.* **2002**, *166*, S4–S8.

(20) Gurgueira, S. A.; Lawrence, J.; Coull, B.; Murthy, G. G.; González-Flecha, B. Rapid increases in the steady-state concentration of reactive oxygen species in the lungs and heart after particulate air pollution inhalation. *Environ. Health Perspect.* **2002**, *110*, 749–755.

(21) Kleinman, M. T.; Bufalino, C.; Rasmussen, R.; Hyde, D.; Bhalla, D. K.; Mautz, W. J. Toxicity of chemical components of

- ambient fine particulate matter (PM 2.5) inhaled by aged rats. *J. Appl. Toxicol.* **2000**, *20*, 357–364.
- (22) Beck-Speier, I.; Dayal, N.; Karg, E.; Maier, K. L.; Schumann, G.; Schulz, H.; Semmler, M.; Takenaka, S.; Stettmaier, K.; Bors, W.; Ghio, A.; Samet, J. M.; Heyder, J. Oxidative stress and lipid mediators induced in alveolar macrophages by ultrafine particles. *Free Radic. Biol. Med.* **2005**, *38*, 1080–1092.
- (23) Liu, F.; Saavedra, M. G.; Champion, J. A.; Griendling, K. K.; Ng, N. L. Prominent Contribution of Hydrogen Peroxide to Intracellular Reactive Oxygen Species Generated upon Exposure to Naphthalene Secondary Organic Aerosols. *Environ. Sci. Technol. Lett.* **2020**, *7*, 171–177.
- (24) Sachse, A.; Wolf, G. Angiotensin II–Induced Reactive Oxygen Species and the Kidney. *J. Am. Soc. Nephrol.* **2007**, *18*, 2439.
- (25) Kumagai, Y.; Arimoto, T.; Shinyashiki, M.; Shimojo, N.; Nakai, Y.; Yoshikawa, T.; Sagai, M. Generation of reactive oxygen species during interaction of diesel exhaust particle components with NADPH-cytochrome P450 reductase and involvement of the bioactivation in the DNA damage. *Free Radical Biol. Med.* **1997**, *22*, 479–487.
- (26) Teufelhofer, O.; Weiss, R.-M.; Parzefall, W.; Schulte-Hermann, R.; Micksche, M.; Berger, W.; Elbling, L. Promyelocytic HL60 Cells Express NADPH Oxidase and Are Excellent Targets in a Rapid Spectrophotometric Microplate Assay for Extracellular Superoxide. *Toxicol. Sci.* **2003**, *76*, 376–383.
- (27) Holevinsky, K. O.; Nelson, D. J. Simultaneous Detection of Free Radical Release and Membrane Current during Phagocytosis. *J. Biol. Chem.* **1995**, *270*, 8328–8336.
- (28) Gomes, A.; Fernandes, E.; Lima, J. L. F. C. Fluorescence probes used for detection of reactive oxygen species. *J. Biochem. Biophys. Methods* **2005**, *65*, 45–80.
- (29) Messner, K. R.; Imlay, J. A.; In vitro quantitation of biological superoxide and hydrogen peroxide generation. *Methods in Enzymology*; Academic Press, 2002; Vol. 349, pp 354–361.
- (30) Yamazaki, T.; Kawai, C.; Yamauchi, A.; Kuribayashi, F. A highly sensitive chemiluminescence assay for superoxide detection and chronic granulomatous disease diagnosis. *Trop. Med. Health* **2011**, *39*, 41–45.
- (31) Li, N.; Sioutas, C.; Cho, A. K.; Schmitz, D.; Misra, C.; Sempf, J.; Wang, M.; Oberley, T.; Froines, J.; Nel, A. Ultrafine particulate pollutants induce oxidative stress and mitochondrial damage. *Environ. Health Perspect.* **2003**, *111*, 455–460.
- (32) Hiura, T. S.; Kaszubowski, M. P.; Li, N.; Nel, A. E. Chemicals in diesel exhaust particles generate reactive oxygen radicals and induce apoptosis in macrophages. *J. Immunol.* **1999**, *163*, 5582–5591.
- (33) He, M.; Ichinose, T.; Yoshida, S.; Ito, T.; He, C.; Yoshida, Y.; Arashidani, K.; Takano, H.; Sun, G.; Shibamoto, T. PM2.5-induced lung inflammation in mice: Differences of inflammatory response in macrophages and type II alveolar cells. *J. Appl. Toxicol.* **2017**, *37*, 1203–1218.
- (34) Franzl, L. M.; Bratt, J. M.; Williams, K. M.; Last, J. A. Why is particulate matter produced by wildfires toxic to lung macrophages? *Toxicol. Appl. Pharmacol.* **2011**, *257*, 182–188.
- (35) Yang, M.; Ahmed, H.; Wu, W.; Jiang, B.; Jia, Z. Cytotoxicity of Air Pollutant 9,10-Phenanthrenequinone: Role of Reactive Oxygen Species and Redox Signaling. *BioMed Res. Int.* **2018**, *2018*, 9523968.
- (36) Antiñolo, M.; Willis, M. D.; Zhou, S.; Abbatt, J. P. D. Connecting the oxidation of soot to its redox cycling abilities. *Nat. Commun.* **2015**, *6*, 6812.
- (37) Tian, Z.; Zhao, H.; Peter, K. T.; Gonzalez, M.; Wetzel, J.; Wu, C.; Hu, X.; Prat, J.; Mudrock, E.; Hettlinger, R.; Cortina, A. E.; Biswas, R. G.; Kock, F. V. C.; Soong, R.; Jenne, A.; Du, B.; Hou, F.; He, H.; Lundeen, R.; Gilbreath, A.; Sutton, R.; Scholz, N. L.; Davis, J. W.; Dodd, M. C.; Simpson, A.; McIntyre, J. K.; Kolodziej, E. P. A ubiquitous tire rubber–derived chemical induces acute mortality in coho salmon. *Science* **2021**, *371*, 185.
- (38) Kumagai, Y.; Shinkai, Y.; Miura, T.; Cho, A. K. The Chemical Biology of Naphthoquinones and Its Environmental Implications. *Annu. Rev. Pharmacol. Toxicol.* **2012**, *52*, 221–247.
- (39) Charrier, J. G.; McFall, A. S.; Richards-Henderson, N. K.; Anastasio, C. Hydrogen Peroxide Formation in a Surrogate Lung Fluid by Transition Metals and Quinones Present in Particulate Matter. *Environ. Sci. Technol.* **2014**, *48*, 7010–7017.
- (40) Tong, H.; Lakey, P. S. J.; Arangio, A. M.; Socorro, J.; Shen, F.; Lucas, K.; Brune, W. H.; Pöschl, U.; Shiraiwa, M. Reactive Oxygen Species Formed by Secondary Organic Aerosols in Water and Surrogate Lung Fluid. *Environ. Sci. Technol.* **2018**, *52*, 11642–11651.
- (41) Lin, Y.-H.; Arashiro, M.; Clapp, P. W.; Cui, T.; Sexton, K. G.; Vizuete, W.; Gold, A.; Jaspers, L.; Fry, R. C.; Surratt, J. D. Gene Expression Profiling in Human Lung Cells Exposed to Isoprene-Derived Secondary Organic Aerosol. *Environ. Sci. Technol.* **2017**, *51*, 8166–8175.
- (42) Ranjit, S.; Malacrida, L.; Stacic, M.; Gratton, E. Determination of the metabolic index using the fluorescence lifetime of free and bound nicotinamide adenine dinucleotide using the phasor approach. *J. Biophot.* **2019**, *12*, No. e201900156.
- (43) Digman, M. A.; Caiola, V. R.; Zamai, M.; Gratton, E. The Phasor Approach to Fluorescence Lifetime Imaging Analysis. *Biophys. J.* **2008**, *94*, L14–L16.
- (44) Weber, G. Resolution of the fluorescence lifetimes in a heterogeneous system by phase and modulation measurements. *J. Phys. Chem.* **1981**, *85*, 949–953.
- (45) Stringari, C.; Edwards, R. A.; Pate, K. T.; Waterman, M. L.; Donovan, P. J.; Gratton, E. Metabolic trajectory of cellular differentiation in small intestine by Phasor Fluorescence Lifetime Microscopy of NADH. *Sci. Rep.* **2012**, *2*, 568.
- (46) Dvornikov, A.; Malacrida, L.; Gratton, E. The DIVER Microscope for Imaging in Scattering Media. *Methods Protoc.* **2019**, *2*, 53.
- (47) Kang, E.; Root, M. J.; Toohey, D. W.; Brune, W. H. Introducing the concept of Potential Aerosol Mass (PAM). *Atmos. Chem. Phys.* **2007**, *7*, 5727–5744.
- (48) Lambe, A. T.; Chhabra, P. S.; Onasch, T. B.; Brune, W. H.; Hunter, J. F.; Kroll, J. H.; Cummings, M. J.; Brogan, J. F.; Parmar, Y.; Worsnop, D. R.; Kolb, C. E.; Davidovits, P. Effect of oxidant concentration, exposure time, and seed particles on secondary organic aerosol chemical composition and yield. *Atmos. Chem. Phys.* **2015**, *15*, 3063–3075.
- (49) Lambe, A. T.; Ahern, A. T.; Williams, L. R.; Slowik, J. G.; Wong, J. P. S.; Abbatt, J. P. D.; Brune, W. H.; Ng, N. L.; Wright, J. P.; Croasdale, D. R.; Worsnop, D. R.; Davidovits, P.; Onasch, T. B. Characterization of aerosol photooxidation flow reactors: heterogeneous oxidation, secondary organic aerosol formation and cloud condensation nuclei activity measurements. *Atmos. Meas. Tech.* **2011**, *4*, 445–461.
- (50) Peng, Z.; Jimenez, J. L. Radical chemistry in oxidation flow reactors for atmospheric chemistry research. *Chem. Soc. Rev.* **2020**, *49*, 2570–2616.
- (51) El-Sawalhi, M. M.; Ahmed, L. A. Exploring the protective role of apocynin, a specific NADPH oxidase inhibitor, in cisplatin-induced cardiotoxicity in rats. *Chem.-Biol. Interact.* **2014**, *207*, 58–66.
- (52) Mudway, I. S.; Kelly, F. J. Ozone and the lung: a sensitive issue. *Mol. Aspect. Med.* **2000**, *21*, 1–48.
- (53) Bedard, K.; Krause, K.-H. The NOX Family of ROS-Generating NADPH Oxidases: Physiology and Pathophysiology. *Physiol. Rev.* **2007**, *87*, 245–313.
- (54) Cross, A. R.; Segal, A. W. The NADPH oxidase of professional phagocytes—prototype of the NOX electron transport chain systems. *Biochim. Biophys. Acta* **2004**, *1657*, 1–22.
- (55) Wientjes, F. B.; Segal, A. W. NADPH oxidase and the respiratory burst. *Semin. Cell Biol.* **1995**, *6*, 357–365.
- (56) Nguyen, G. T.; Green, E. R.; Meccas, J. Neutrophils to the ROScues: Mechanisms of NADPH Oxidase Activation and Bacterial Resistance. *Front. Cell. Infect. Microbiol.* **2017**, *7*, 373.
- (57) Rehman, A. U.; Anwer, A. G.; Gosnell, M. E.; Mahub, S. B.; Liu, G.; Goldys, E. M. Fluorescence quenching of free and bound NADH in HeLa cells determined by hyperspectral imaging and

- unmixing of cell autofluorescence. *Biomed. Opt. Express* **2017**, *8*, 1488–1498.
- (58) Yang, X.; Ma, K. Characterization of an Exceedingly Active NADH Oxidase from the Anaerobic Hyperthermophilic Bacterium *Thermotoga maritima*. *J. Bacteriol.* **2007**, *189*, 3312–3317.
- (59) Li, J.-M.; Shah, A. M. Differential NADPH- versus NADH-dependent superoxide production by phagocyte-type endothelial cell NADPH oxidase. *Cardiovasc. Res.* **2001**, *52*, 477–486.
- (60) Pate, K. T.; Stringari, C.; Sprowl-Tanio, S.; Wang, K.; TeSlaa, T.; Hoverter, N. P.; McQuade, M. M.; Garner, C.; Digman, M. A.; Teitell, M. A.; Edwards, R. A.; Gratton, E.; Waterman, M. L. Wnt signaling directs a metabolic program of glycolysis and angiogenesis in colon cancer. *EMBO J.* **2014**, *33*, 1454–1473.
- (61) Stringari, C.; Cinquin, A.; Cinquin, O.; Digman, M. A.; Donovan, P. J.; Gratton, E. Phasor approach to fluorescence lifetime microscopy distinguishes different metabolic states of germ cells in a live tissue. *Proc. Natl. Acad. Sci. U. S. A.* **2011**, *108*, 13582.
- (62) Panday, A.; Sahoo, M. K.; Osorio, D.; Batra, S. NADPH oxidases: an overview from structure to innate immunity-associated pathologies. *Cell. Mol. Immunol.* **2015**, *12*, 5–23.
- (63) Togashi, D. M.; Nicodem, D. E. Photophysical studies of 9,10-phenanthrenequinones. *Spectrochim. Acta, Part A* **2004**, *60*, 3205–3212.
- (64) Golfetto, O.; Hinde, E.; Gratton, E. Laurdan Fluorescence Lifetime Discriminates Cholesterol Content from Changes in Fluidity in Living Cell Membranes. *Biophys. J.* **2013**, *104*, 1238–1247.
- (65) Wei, J.; Fang, T.; Shiraiwa, M. Effects of Acidity on Reactive Oxygen Species Formation from Secondary Organic Aerosols. *ACS Environ. Au* **2022**, *2*, 336–345.
- (66) Wei, J.; Fang, T.; Shiraiwa, M. Effects of Acidity on Reactive Oxygen Species Formation from Secondary Organic Aerosols. *ACS Environ. Au* **2022**, *2*, 336–345.
- (67) Offer, S.; Hartner, E.; Di Bucchianico, S.; Bisig, C.; Bauer, S.; Pantzke, J.; Zimmermann Elias, J.; Cao, X.; Binder, S.; Kuhn, E.; Huber, A.; Jeong, S.; Käfer, U.; Martens, P.; Mescerikovas, A.; Bendl, J.; Brejcha, R.; Buchholz, A.; Gat, D.; Hohaus, T.; Rastak, N.; Jakobi, G.; Kalberer, M.; Kanashova, T.; Hu, Y.; Ogris, C.; Marsico, A.; Theis, F.; Pardo, M.; Gröger, T.; Oeder, S.; Orasche, J.; Paul, A.; Ziehm, T.; Zhang, Z.-H.; Adam, T.; Sippula, O.; Sklorz, M.; Schnelle-Kreis, J.; Czech, H.; Kiendler-Scharr, A.; Rudich, Y.; Zimmermann, R. Effect of Atmospheric Aging on Soot Particle Toxicity in Lung Cell Models at the Air–Liquid Interface: Differential Toxicological Impacts of Biogenic and Anthropogenic Secondary Organic Aerosols (SOAs). *Environ. Health Perspect.* **2022**, *130*, 027003.
- (68) Tuet, W. Y.; Chen, Y.; Xu, L.; Fok, S.; Gao, D.; Weber, R. J.; Ng, N. L. Chemical oxidative potential of secondary organic aerosol (SOA) generated from the photooxidation of biogenic and anthropogenic volatile organic compounds. *Atmos. Chem. Phys.* **2017**, *17*, 839–853.
- (69) Tuet, W. Y.; Chen, Y.; Fok, S.; Champion, J. A.; Ng, N. L. Inflammatory responses to secondary organic aerosols (SOA) generated from biogenic and anthropogenic precursors. *Atmos. Chem. Phys.* **2017**, *17*, 11423–11440.
- (70) Dröge, W. Free radicals in the physiological control of cell function. *Physiol. Rev.* **2002**, *82*, 47–95.
- (71) Pardo, M.; Qiu, X.; Zimmermann, R.; Rudich, Y. Particulate Matter Toxicity Is Nrf2 and Mitochondria Dependent: The Roles of Metals and Polycyclic Aromatic Hydrocarbons. *Chem. Res. Toxicol.* **2020**, *33*, 1110–1120.
- (72) Li, N.; Alam, J.; Venkatesan, M. I.; Eiguren-Fernandez, A.; Schmitz, D.; Di Stefano, E.; Slaughter, N.; Killeen, E.; Wang, X.; Huang, A.; Wang, M.; Miguel, A. H.; Cho, A.; Sioutas, C.; Nel, A. E. Nrf2 Is a Key Transcription Factor That Regulates Antioxidant Defense in Macrophages and Epithelial Cells: Protecting against the Proinflammatory and Oxidizing Effects of Diesel Exhaust Chemicals. *J. Immunol.* **2004**, *173*, 3467.
- (73) Porra, L.; Dégrugilliers, L.; Broche, L.; Albu, G.; Strengell, S.; Suhonen, H.; Fodor, G. H.; Peták, F.; Suortti, P.; Habre, W.; Sovijärvi, A. R. A.; Bayat, S. Quantitative Imaging of Regional Aerosol Deposition, Lung Ventilation and Morphology by Synchrotron Radiation CT. *Sci. Rep.* **2018**, *8*, 3519.
- (74) Phalen, R. F.; Oldham, M. J.; Nel, A. E. Tracheobronchial Particle Dose Considerations for In Vitro Toxicology Studies. *Toxicol. Sci.* **2006**, *92*, 126–132.
- (75) Vicente, E. D.; Alves, C. A.; Martins, V.; Almeida, S. M.; Lazaridis, M. Lung-deposited dose of particulate matter from residential exposure to smoke from wood burning. *Environ. Sci. Pollut. Res.* **2021**, *28*, 65385–65398.
- (76) Du, Z.-D.; Yu, S.; Qi, Y.; Qu, T.-F.; He, L.; Wei, W.; Liu, K.; Gong, S.-S. NADPH oxidase inhibitor apocynin decreases mitochondrial dysfunction and apoptosis in the ventral cochlear nucleus of D-galactose-induced aging model in rats. *Neurochem. Int.* **2019**, *124*, 31–40.
- (77) Aviram, M.; Rosenblat, M.; Etzioni, A.; Levy, R. Activation of NADPH oxidase is required for macrophage-mediated oxidation of low-density lipoprotein. *Metabolism* **1996**, *45*, 1069–1079.
- (78) Jay Forman, H.; Torres, M. Redox signaling in macrophages. *Mol. Aspects. Med.* **2001**, *22*, 189–216.
- (79) Mumbengegwi, R.; Li, Q.; Li, C.; Bear, E.; Engelhardt, F. Evidence for a Superoxide Permeability Pathway in Endosomal Membranes. *Mol. Cell. Biol.* **2008**, *28*, 3700–3712.
- (80) Charrier, J. G.; Anastasio, C. On dithiothreitol (DTT) as a measure of oxidative potential for ambient particles: evidence for the importance of soluble transition metals. *Atmos. Chem. Phys.* **2012**, *12*, 9321–9333.
- (81) Guo, H.-b.; Li, M.; Lyu, Y.; Cheng, T.-t.; Xv, J. j.; Li, X. Size-resolved particle oxidative potential in the office, laboratory, and home: Evidence for the importance of water-soluble transition metals. *Environ. Pollut.* **2019**, *246*, 704–709.
- (82) Shirmohammadi, F.; Hasheminassab, S.; Wang, D.; Saffari, A.; Schauer, J. J.; Shafer, M. M.; Delfino, R. J.; Sioutas, C. Oxidative potential of coarse particulate matter (PM<sub>10-2.5</sub>) and its relation to water solubility and sources of trace elements and metals in the Los Angeles Basin. *Environ. Sci.: Processes Impacts* **2015**, *17*, 2110–2121.
- (83) Li, N.; Wang, M.; Oberley, T. D.; Sempf, J. M.; Nel, A. E. Comparison of the Pro-Oxidative and Proinflammatory Effects of Organic Diesel Exhaust Particle Chemicals in Bronchial Epithelial Cells and Macrophages. *J. Immunol.* **2002**, *169*, 4531.