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1 Fugitive Emissions and Health Implications of Plancha-Type Stoves

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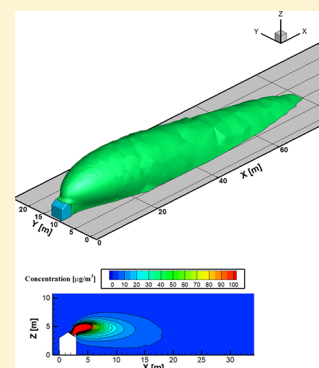
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11 **S** Supporting Information

12 **ABSTRACT:** Plancha-type stoves have been widely disseminated in Mexico and Central
13 America, but the contribution of fugitive emissions from these stoves to indoor air
14 concentrations has been poorly quantified. In this study, fugitive emissions were measured for
15 four plancha-type cookstoves most disseminated in Mexico (Patsari, ONIL, Ecostufa, and
16 Mera-Mera). In controlled testing, fugitive emissions from plancha-type chimney stoves ($n =$
17 15 for each stove) were on average $5 \pm 3\%$ for $PM_{2.5}$ and $1 \pm 1\%$ for CO, much lower than
18 defaults in WHO Guidelines ($25 \pm 10\%$). Using a Monte Carlo single zone model with locally
19 measured parameters, average kitchen concentrations resulting from fugitive emissions were 15
20 $\pm 9 \mu\text{g}/\text{m}^3$ for $PM_{2.5}$ and $0.06 \pm 0.04 \text{ mg}/\text{m}^3$ for CO. On the basis of these models, plancha-
21 type stoves meet benchmarks for WHO Air Quality Guidelines (AQG) Interim Target I for
22 $PM_{2.5}$ and the 24 h AQG for CO, respectively, with on average 97% of homes meeting the
23 guideline for $PM_{2.5}$. Similarly, all four plancha-type stoves were ISO IWA Tier 4 for indoor
24 emissions of CO and Tier 3 for indoor emissions of $PM_{2.5}$. Three-dimensional computational
25 fluid dynamics (CFD) analysis was used to estimate neighborhood pollution impacts of
26 upstream chimney emissions. When chimney emissions were included as background concentrations combined with indoor
27 contributions from fugitive emissions, plancha-type stoves would still meet the WHO AQG Annual Interim Target I for $PM_{2.5}$
28 and the 24 h AQG for CO for the scenario modeled in this study.



1. INTRODUCTION

29 Field studies have shown that well-functioning plancha-type
30 chimney stoves that vent emissions outside the home result in
31 significant exposure benefit compared to open fire stoves,¹ as
32 only a fraction of the emissions enter the kitchen via fugitive
33 emissions and reinfiltration, especially if the plancha stove is
34 used exclusively and other combustion sources are removed
35 from the kitchen.² Plancha-type cookstoves have been widely
36 disseminated in Mexico, Guatemala, El Salvador, Honduras,
37 and Nicaragua, where they have become widely accepted in
38 local communities, and somewhat less in Panama and Costa
39 Rica.³ In Mexico, a total of more than 600 000 plancha-type
40 stoves were disseminated between 2007 and 2012, mostly
41 through the *Programa Nacional de Estufas de Leña*.⁴ While
42 plancha-type cookstoves are very popular and widely
43 disseminated in Latin America, they do not easily fit into
44 performance frameworks currently being developed by the
45 cookstove community.

46 The International Workshop Agreement (IWA) 11:2012
47 established tiers of performance for cookstove efficiency,
48 emissions, and safety.⁵ The IWA emission benchmarks were

49 established with tiers for indoor emission rates and for total
50 emission factors (based on useful energy delivered) spanning
51 performance of traditional open fires (Tier 0) to aspirational
52 targets (Tier 4). Although the fraction of emissions that does
53 not go up the chimney (fugitive emissions) may be used to
54 estimate the tiers for indoor emissions, the protocols/testing
55 methods for simultaneous assessment of fugitive and chimney
56 emissions have not yet been established. Further, the indoor
57 emission tiers do not account for the reinfiltration of emissions
58 from the chimney back into the indoor environment.

The World Health Organization (WHO) Indoor Air Quality
59 Guidelines recommend emission rates for both vented and
60 unvented stoves,⁶ where the emission rate for vented stoves
61 used a normal distribution for the fraction of emissions
62 entering the kitchen (fugitive and reinfiltration), ranging from
63 1% to 50% with a mean of 25% and standard deviation of 10%
64

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65 of the emissions from an unvented stove. The estimate,
66 however, was roughly based on the ratio of concentrations in
67 homes with chimney stoves to those without chimney stoves
68 rather than direct studies of fugitive emissions. The
69 contributions from reinfiltration were also not included in
70 the WHO guidance. Cognizant of the gaps in information for
71 specific stove types, the IWA workshop attendees recom-
72 mended that new protocols be developed or current protocols
73 be updated to more adequately address a larger number of
74 stove and fuel types, including plancha-type stoves.

75 Indoor pollutant concentrations in a kitchen with a plancha
76 stove are a function of fugitive emissions from the stove,
77 reinfiltration of emissions from the chimney of the same stove
78 back into the home, and infiltration of ambient pollutants from
79 other external sources and upstream village stoves, in addition
80 to ventilation rate, room volume, and mixing. Determination of
81 whether plancha-type stoves should be promoted in an area to
82 actually achieve substantive health benefits therefore depends
83 on the amount of fugitive emissions from the stoves, the
84 density of homes that contribute to downwind neighborhood
85 pollution levels, and relative costs and accessibility of cleaner
86 fuels, in addition to considerations relating to adoption, user
87 preferences, and cooking tasks. For plancha-type stoves to be
88 better incorporated into ISO:IWA 11:2012 tiers of perform-
89 ance for emissions and for better estimation of relative health
90 benefits that may accrue from promotion of plancha-type
91 stoves, fugitive emissions and the impact of the chimney
92 emissions on other village houses downstream need to be
93 evaluated.

94 In this paper, fugitive emissions for four Mexican plancha-
95 type stoves are presented that show plancha stoves meet ISO
96 IWA benchmarks for indoor emissions of PM_{2.5} Tier 3 and CO
97 Tier 4 with mean kitchen concentrations below the WHO
98 Interim Target 1 (WHO IT1) when combined with local
99 parameters on building ventilation rates, room volumes, and
100 cooking times. Further, the impact of the chimney emissions
101 on surrounding houses downstream is demonstrated through
102 the application of a three-dimensional computational fluid
103 dynamics (CFD) analysis to establish criteria for housing
104 densities where plancha-type stoves could be promoted
105 without major neighborhood pollution effects in surrounding
106 houses.

2. MATERIALS AND METHODS

107 One of the key parameters that need to be measured for
108 adequately estimating the impact of chimney stoves on indoor
109 air pollution is the fraction of pollutants that do not exit
110 through the chimney but leak through the combustion
111 chamber entry or from other parts of the stove into the
112 kitchen, known as fugitive emissions “*f*”. For example, as stated
113 above, the WHO roughly assumed that, for chimney stoves, *f*
114 ranges from 1% to 50% with a mean of 25% and standard
115 deviation of 10%.⁶

116 **2.1. Measurements of Fugitive Emissions.** Fugitive
117 emissions were measured as a fraction of the overall emission
118 by using two nested hoods to capture separately the emissions
119 through the chimney and fugitive emissions to the room at the
120 same time (Figure S1), consistent with the system specified in
121 ISO Standard 19867-1. Flue emissions and fugitive emissions
122 were measured for four plancha stove models with a chimney
123 (Patsari, ONIL, Mera-Mera, and Ecostufa; see Medina et al.⁷),
124 which are the models most disseminated in Mexico, using the
125 Water Boiling Test (WBT) protocol version 4.2.3.⁸ Fuels used

were White Oak (*Quercus spp.*) in all tests with average
moisture content on a wet basis of $7.0 \pm 0.6\%$ with a range of
6.1–8.2% measured using a Protimeter Timbermaster Wood
Moisture Meter (GE, Billerica, MA). Further details are
described in Ruiz and Masera.⁹

Measurement of CO₂ and CO emissions was performed
using a Laboratory Emissions Measurement System (LEMS)
(Aprovecho Research Center, Oregon USA), and a Q-Trak
(Model 7575, TSI Inc., Shoreview, MN) was used for fugitive
emissions. Real-time concentrations are measured in both
equipment using a nondispersive infrared (NDIR) sensor for
CO₂ and an electrochemical cell to measure CO. CO₂ and CO
sensors were calibrated using zero air and a mixture of 500
ppm of CO and 5000 ppm of CO₂.¹⁰ Factory calibrated Q-trak
have been shown to perform well in quantifying lower
concentrations of CO₂ and CO in comparison with stainless
steel canisters collected during the same emissions tests and
analyzed by GC FID equipped with preconcentrator and
methanizer.¹¹ PM_{2.5} mass from both flue emissions and fugitive
emissions was collected isokinetically with a PM_{2.5} cyclone on
fiberglass filters for gravimetric determination.¹² The sample
flow for fugitive emissions was 4 L/min with a 47 mm filter
and 16.7 L/min with a 102 mm filter for chimney emissions,
and all filters were weighed on a microbalance with a
resolution of 1 μg .

**2.2. Impact of Fugitive Emissions on Indoor Air
Quality and Neighborhood Air Pollution.** A Monte Carlo
single zone model was used to estimate the impact of fugitive
emissions from plancha-type stoves on indoor air quality⁶ and
is described mathematically in the Supporting Information.
The impact of the chimney emissions on surrounding houses
downstream is demonstrated through the application of a
three-dimensional computational fluid dynamics (CFD)
analysis described in further detail with methodology
validation in the Supporting Information.

Indoor air concentrations estimated using the single-zone
model and impacts of emissions on neighborhood pollution
were based on locally measured housing characteristics and
ventilation parameters (Table S2). Briefly, kitchens in
Michoacán typically incorporate the kitchen and dining room
with a mean volume of $41 \pm 20 \text{ m}^3$ ($n = 627$ measurements).
In regard to time, women usually start cooking in the early
morning. For Mexican customs in rural areas, there are no
cooking events (breakfast, lunch, and dinner); instead, there is
a period of operation in which the cooks cook foods that
include tortillas (tortillas are a garnish of any dish at any time
of day), breakfast, and dinner. The mean operating period per
day of the stoves is extensive $259 \pm 123 \text{ min}$ ($n = 30$
measurements) and similar to that reported by Armendáriz-
Arnez et al.¹³ When one considers nominal air exchange rate,
kitchen designs in rural Michoacán are diverse: without walls,
three walls, four walls, brick walls, and wood plank walls with
and without gaps, which affects the air exchange between the
kitchen and the exterior. Armendáriz-Arnez et al.¹³ reported
43% with wooden planks and gaps <1 cm, 33% with gaps <1
cm, and 19% without gaps. Typically, however, in these
regions, kitchens have four walls made of wooden planks with
gaps of approximately 1 cm,^{13,14} a door, and one window. Air
exchange rates were estimated after Traynor et al. using CO
decay rates in a test kitchen (Figure S2) with similar volume
and construction as a typical kitchen ($n = 31$ measurements).¹⁵
CO decay rates were measured with Q-Trak (Model 7575, TSI

Table 1. Fugitive Emissions from Plancha-Type Stoves in Mexico^a

parameter	n	PM _{2.5} emissions rate (mg/min)			CO emissions rate (mg/min)		
		chimney	fugitive	fraction of overall	chimney	fugitive	fraction of overall
ONIL	15	52 (32)	2.1 (1.3)	0.05 (0.03)	594 (332)	12 (12)	0.02 (0.02)
Ecostufa	15	75 (52)	3.5 (1.9)	0.06 (0.04)	931 (588)	5 (3)	0.01 (0.01)
Mera-Mera	15	76 (47)	2.4 (1.6)	0.03 (0.02)	1244 (543)	20 (16)	0.01 (0.01)
Patsari	15	50 (21)	3.9 (2.1)	0.07 (0.03)	1645 (965)	11 (11)	0.01 (0.00)
all stoves	60	63 (41)	3.0 (1.8)	0.05 (0.03)	1104 (743)	12 (12)	0.01 (0.01)

^aMean (standard deviation).

Table 2. Estimates of 24 h Mean Indoor Concentrations and Percent of Simulations Meeting WHO Air Quality Guidelines and Interim Target 1

	particulate matter PM _{2.5} (μg m ⁻³)				percent meeting annual WHO guideline		ISO IWA indoor emission PM _{2.5} mg min ⁻¹	carbon monoxide (mg m ⁻³)				percent meeting 24 h WHO guideline AQG (7 mg m ⁻³)	ISO IWA indoor emission CO g min ⁻¹
	mean (SD)	median	10%	90%	interim target-1 (35 μg m ⁻³)	AQG (10 μg m ⁻³)		mean (SD)	median	10%	90%		
ONIL	10 (6)	9	5	17	99%	58%	3	0.06 (0.04)	0.05	0.02	0.1	100%	4
Ecostufa	17 (10)	15	9	28	96%	17%	3	0.02 (0.01)	0.02	0.01	0.04	100%	4
Mera-Mera	12 (7)	10	5	20	99%	48%	3	0.10 (0.06)	0.09	0.04	0.16	100%	4
Patsari	19 (11)	17	9	31	94%	13%	3	0.05 (0.03)	0.05	0.02	0.09	100%	4
all plancha stoves	15 (9)	13	7	25	97%	30%	3	0.06 (0.04)	0.05	0.02	0.11	100%	4
LPG	4 (2)	3	2	7	100%	98%	4	0.15 (0.09)	0.14	0.07	0.26	100%	4

188 Inc., Shoreview, MN) every 2 s, located at a height of 1.5 m in
 189 a room with a mixing fan to ensure well mixed concentration.
 190 **2.2.1. Impacts of Emissions on Neighborhood Pollution.**
 191 Smoke temperatures were assumed to be 318 °C based on
 192 measured flue gas temperatures at the top of the flue of
 193 plancha stoves during performance testing, and exit velocities
 194 were 0.5 m/s (Table S2). Other sources of air pollutants can
 195 (and often do) increase background ambient concentrations.
 196 Similar to the single zone model used to derive emission rates
 197 for stoves to meet indoor air quality guidelines and interim
 198 targets,⁶ background ambient concentrations were not
 199 incorporated into the model as the objective was to determine
 200 emission rates that would result in an incremental increase in
 201 exposures. These concentrations can be readily incorporated
 202 on the basis of the local context, but since background ambient
 203 concentrations vary widely, the results are of limited
 204 application. Wind speeds were based on local wind speeds in
 205 the Purepecha region. On the basis of a logarithmic profile,
 206 wind speed near the ground would be expected to be lower
 207 than wind speed typically measured at a height of 10 m for
 208 meteorological data. Low wind speeds are conservative as
 209 higher wind speeds would increase particulate dispersion and
 210 result in lower estimates of exposure. The criteria for dispersed
 211 settlements based on minimum distances between homes were
 212 also conservatively estimated to be protective of health by
 213 minimizing the dispersion of the plumes by wind and
 214 estimating the distance directly downwind based on the center
 215 of the plume.

3. RESULTS

216 **3.1. Pollutant Emissions Rate (CO and PM_{2.5}).** Table 1
 217 shows fugitive emissions rates of PM_{2.5} and carbon monoxide
 218 (CO) from four plancha stove types and the relative fraction of
 219 fugitive emissions to the overall emission rate from the stove.

Fugitive PM_{2.5} emissions from the Patsari stove were 1.6
 220 times those from the Mera-Mera, while chimney emissions
 221 from the Patsari were ~34% lower. As a result, fugitive
 222 emissions from the Patsari are a larger fraction of overall
 223 emissions from the stove compared to the stoves with greater
 224 chimney emissions. The fractions of fugitive emissions for
 225 PM_{2.5} were also substantially higher than those for CO, which
 226 may be the result of a greater fraction of PM_{2.5} emissions
 227 escaping during lighting or from larger fuel pieces that stick out
 228 of the entrance to the combustion chamber. 229

In controlled testing, the fractions of fugitive emissions from
 230 plancha-type stoves were substantially lower than the mean of
 231 25% of the overall emission (standard deviation of 10%; range
 232 1–50) guidance by the World Health Organization (WHO)
 233 Indoor Air Quality Guidelines for the fraction of emissions
 234 entering the kitchen. The average fugitive emission for the 4
 235 stove types tested was 5 ± 3% for PM_{2.5} emissions and 1 ± 1%
 236 for CO emissions. As a result, ISO tier performance should be
 237 assessed through direct measurement of fugitive emissions
 238 combined with local ventilation parameters for input into the
 239 single zone model. 240

241 Table 2 and Figure 1 show estimates of the impact of 241
 242 fugitive emissions on indoor air quality using the Monte Carlo
 243 single zone model, the percentage meeting WHO air quality
 244 guidelines and interim targets, and the ISO IWA 11:2012
 245 indoor emissions tiers of performance. The impact of LPG
 246 emissions, assuming all emissions enter the room, have been
 247 added for comparison as LPG is often cited as a clean-burning
 248 cooking fuel. 248

The means for each distribution of PM_{2.5} concentrations
 249 ranged from 10 to 19 μg/m³ depending on the stove; all are
 250 below the WHO Interim Target-1 of 35 μg/m³. The means of
 251 the CO (24 h) distributions ranged from 0.02 to 0.10 mg/m³
 252 depending on the stove, which are well below the WHO Air 253

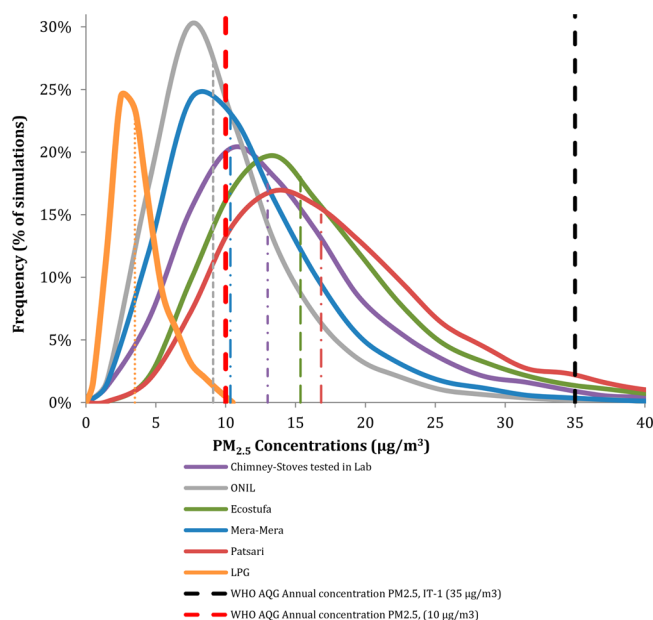


Figure 1. Modeled 24 h mean indoor air concentration distributions of $\text{PM}_{2.5}$ from plancha-type stoves and LPG in reference to WHO IAQ guidelines. Vertical bars show geometric means.

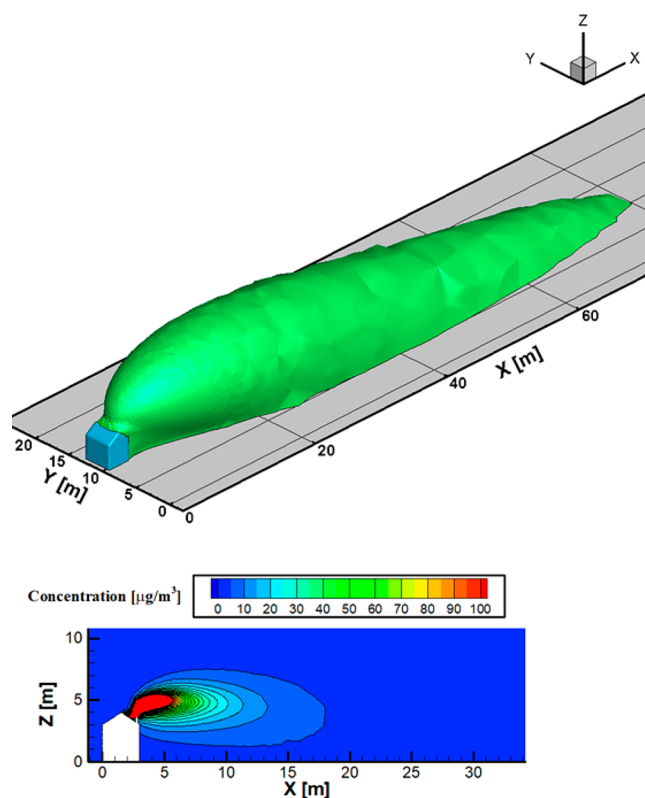


Figure 2. Pollutant plume using iso-surface of concentration (C) of $1 \mu\text{g}/\text{m}^3$ for emission rate of $40 \text{ mg}/\text{min}$ and contour of pollutant concentration with distance X and height Z in the center of the plume.

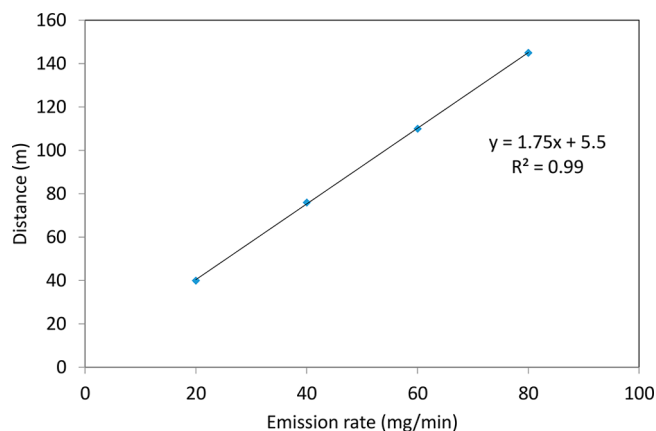


Figure 3. Distance between buildings for $\text{PM}_{2.5}$ emission concentration to decrease to $1 \mu\text{g}/\text{m}^3$ for different chimney emission rates.

254 Quality Guideline (AQG) of $7 \text{ mg}/\text{m}^3$ recommended by
 255 WHO.¹⁶ Fugitive emissions of $\text{PM}_{2.5}$ from plancha-type
 256 chimney stoves resulted in simulated incremental increases in
 257 indoor concentrations of $\text{PM}_{2.5}$ that would meet WHO Interim
 258 Target 1 for greater than 97% of homes. In addition, for the
 259 best performance of these stoves in terms of fugitive emissions,
 260 58% of homes would have an incremental increase in indoor
 261 concentrations of $\text{PM}_{2.5}$ that was less than the WHO AQG.
 262 Overall indoor air concentrations were a factor 2.5–4.7 higher
 263 than similar homes simulated with LPG stoves, which is
 264 reflected in the percentages of homes meeting the AQG for
 265 $\text{PM}_{2.5}$. ISO IWA indoor emission tiers for CO were all Tier 4
 266 for plancha-type stoves. Using emissions rates from laboratory
 267 tests of LPG reported by Shen et al.¹⁷ and the same single zone
 268 model parameters as above, mean indoor $\text{PM}_{2.5}$ concentrations
 269 were $1.0 \pm 0.6 \mu\text{g}/\text{m}^3$.

270 In order to estimate the impact of plancha-type stove
 271 emissions on surrounding houses downstream, a three-
 272 dimensional computational fluid dynamics (CFD) analysis
 273 was performed to determine housing densities where plancha-
 274 type stoves could be promoted without major neighborhood
 275 pollution effects in surrounding houses. Using parameters
 276 estimated for the Purepecha region of Mexico on wind speed,
 277 housing volume, ground roughness, gas temperature, etc.
 278 (Table S2), plume concentrations were estimated for emission
 279 rates of 20, 40, 60, and $80 \text{ mg}/\text{min}$. Although average values
 280 for wind speed were used, plume concentrations will depend
 281 on local wind speeds and other meteorological parameters.
 282 Figure 2 shows the iso-surface of concentration is equal to 1
 283 $\mu\text{g}/\text{m}^3$ for the emission rate of $40 \text{ mg}/\text{min}$ and contour of
 284 pollutant concentration with distance and height in the center
 285 of the plume.

286 Figure 3 shows the distance between buildings for emission
 287 concentration to decrease to $1 \mu\text{g}/\text{m}^3$ at the breathing height
 288 of the individual for different chimney emissions rates. As
 289 emissions rates for plancha-type stoves range between 50 and
 290 $76 \text{ mg}/\text{min}$ (Table 1), stoves plumes would decrease to $1 \mu\text{g}/$

m^3 at an average of approximately 93 m for the Patsari, 96 m 291
 for the ONIL, 136 m for the Ecostufa, and 136 m for the Mera- 292
 Mera, with an average of 115 m for all chimney stoves. 293

4. DISCUSSION

4.1. Fugitive Emissions. Chimney and fugitive emissions 294
 from four widely disseminated Mexican wood burning plancha- 295
 type chimney stoves were measured to allow better estimation 296
 of indoor air concentrations and neighborhood air pollution 297
 impacts on other village houses downstream and relative health 298
 benefits that may accrue from promotion of plancha-type 299
 stoves. Measurements show that fugitive emissions from well- 300
 functioning chimney stoves are a smaller fraction of total 301

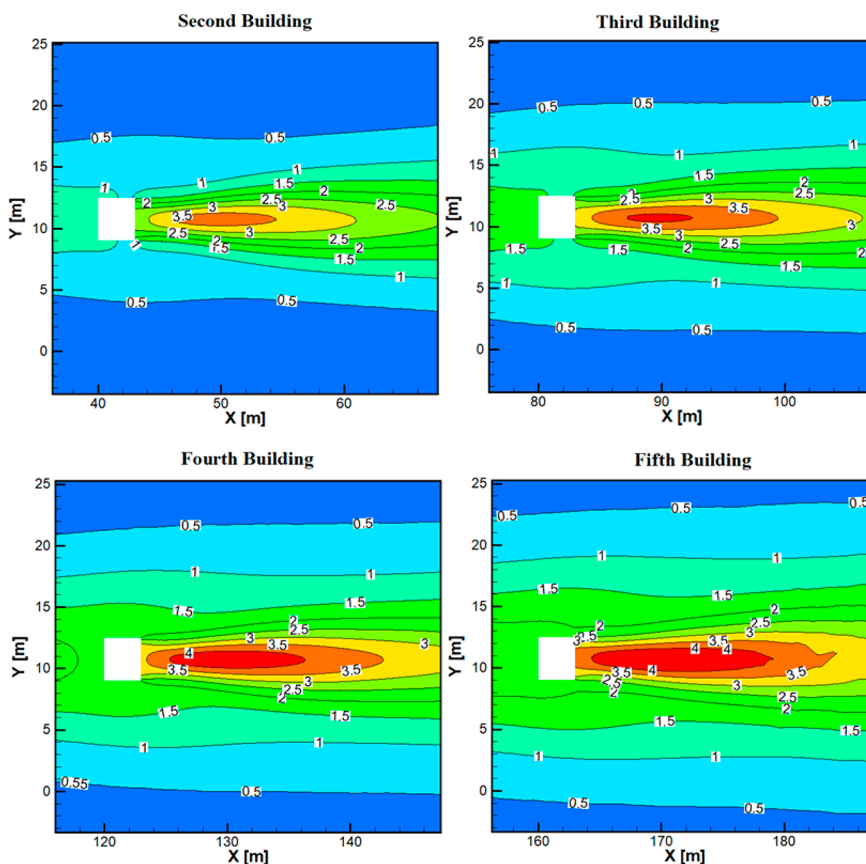


Figure 4. Contour of pollutant concentration at 1.5 m for 5 homes downwind in the center of the plume with each home emitting at 20 mg/min and a distance of 40 m between homes.

302 emissions than estimated for the WHO defaults of fraction of
 303 emissions entering the kitchen. Plancha-type chimney stoves
 304 vent on average $95 \pm 3\%$ and $99 \pm 1\%$ of total $PM_{2.5}$ and CO
 305 emissions out of the kitchen, respectively. Models that link
 306 emissions from cookstoves with indoor concentrations of
 307 $PM_{2.5}$ ¹⁸ have been useful in evaluating which stove types would
 308 meet WHO air quality guidelines in indoor environments.⁶
 309 Using a Monte Carlo single zone model parametrized for local
 310 Mexican building ventilation rates, building volume, and
 311 cooking time, 24 h IAP concentration levels resulting from
 312 these stoves are estimated to be on average $15 \pm 9 \mu\text{g}/\text{m}^3$ for
 313 $PM_{2.5}$ and $0.06 \pm 0.04 \mu\text{g}/\text{m}^3$ for CO. Modeled concentrations
 314 showed good agreement with estimated 24 h mean indoor
 315 concentrations from Blanco et al.¹⁹ of $19.9 \pm 42.6 \mu\text{g}/\text{m}^3$ and
 316 $1.6 \pm 3.5 \text{ mg}/\text{m}^3$ for $PM_{2.5}$ and CO, respectively, based on
 317 indoor concentrations during water boiling tests using Patsari
 318 and ONIL stoves in a simulated kitchen with dimensions
 319 similar to that modeled in the current study. On the basis of
 320 these models, plancha stoves meet benchmarks for WHO Air
 321 Quality Guidelines (AQG) Interim Target I for $PM_{2.5}$ and the
 322 24 h AQG for CO, with on average 92% of homes meeting the
 323 guideline for $PM_{2.5}$. Similarly, all four plancha-type stoves were
 324 ISO Tier 4 for indoor emissions of CO and Tier 3 for indoor
 325 emissions of $PM_{2.5}$, respectively.
 326 The Global Alliance for Clean Cookstoves considers stoves
 327 that meet ISO IWA Tier 3 for indoor emissions as “clean” for
 328 health.²⁰ Leaving aside issues related to the basis for this
 329 classification, modeled indoor air concentrations using the
 330 Monte Carlo single zone model based on fugitive emissions
 331 measured in the current study implies that the most

disseminated Mexican stoves would meet this classification. 332
 Thus, these measurements suggest that significant benefits to 333
 indoor air quality can result from venting emissions outside the 334
 home through a chimney. While it is clear that the best/ 335
 cleanest options which can feasibly displace traditional stove 336
 use should be pursued in different regional contexts, this may 337
 include chimney stoves in areas where the use of clean fuels is 338
 limited by distribution networks, local infrastructure, fuel 339
 pricing, political climate, affordability, and other factors. The 340
 benefits of venting outdoors have largely been overlooked by 341
 the most recent generations of forced draft stoves. Recent 342
 measurements of some of these stoves however have shown 343
 that exposure reductions have failed to materialize in 344
 uncontrolled field trials.^{21,22} When used in indoor kitchens, 345
 ventilation of these stoves with flues or chimneys, however, 346
 likely would have achieved substantial exposure reductions 347
 provided that traditional stoves were not used alongside stoves 348
 with chimneys. Clearly, there are a range of other barriers and 349
 complications in installing chimneys; however, the benefits 350
 may merit the extra effort, and many stoves globally 351
 incorporate chimneys including heating stoves in China^{21,23} 352
 and Mongolia,²⁴ among others, and plancha-type stoves across 353
 Central America. When used outdoors, emission rates can be 354
 substantially higher and still maintain exposure contributions 355
 below WHO air quality guidelines.²⁵ 356
4.2. Contribution to Neighborhood Emissions. 357
 Although indoor fugitive emissions from Mexican plancha- 358
 type stoves meet GACC criteria for being “clean” in indoor 359
 kitchens, there are clearly scenarios where promotion of typical 360
 plancha-type stoves would not be advisable due to the density 361

362 of surrounding homes and the neighborhood pollution impacts
 363 of the emissions vented through the chimney (although more
 364 advanced plancha-type stoves with cleaner combustion or
 365 alternative fuels may resolve some of these concerns). Indoor
 366 pollutant concentrations in a kitchen with a plancha stove are a
 367 function of fugitive emissions from the stove, re-infiltration of
 368 emissions from the chimney of the same stove back into the
 369 home, and infiltration of ambient pollutants from other
 370 external sources and upstream village stoves. Aside from
 371 important considerations related to user preferences, adoption,
 372 use intensity, and purchasing power, from a health perspective,
 373 determination of whether plancha stoves should be promoted
 374 in an area depends on the density of homes that contribute to
 375 downwind neighborhood pollution levels and relative costs and
 376 accessibility of cleaner fuels.

377 Determination of whether the indoor air concentrations are
 378 lower than WHO AQG and Interim Targets depends on the
 379 ambient concentrations upwind, which are a function of the
 380 density of homes in the area where dissemination of plancha-
 381 type stoves are being considered. To provide a benchmark for
 382 criteria for where dissemination of stoves could be considered,
 383 three-dimensional computational fluid dynamics (CFD)
 384 analysis was used to estimate the distance required between
 385 homes for chimney emissions outdoors to drop to $1 \mu\text{g}/\text{m}^3$
 386 between homes. As a first step, these criteria can be useful in
 387 spatial analyses to evaluate potential areas where installation of
 388 plancha-type stoves could be promoted where the emissions do
 389 not substantially increase neighborhood pollution levels, an
 390 approach similar to that used for estimating impacts of outdoor
 391 cooking on neighboring houses.²⁵ Consequently, a mechanistic
 392 model was used to evaluate plausible contributions to the
 393 neighborhood pollution impacts from houses with similar
 394 emissions spaced equally downwind in the center of the plume
 395 and resultant health implications. Figure 4 shows plume
 396 concentrations at the breathing height of an individual (1.5 m)
 397 for 5 houses in the downwind direction with each home
 398 emitting at 20 mg/min and a distance of 40 m between homes
 399 such that the emissions from the first house reach $1 \mu\text{g}/\text{m}^3$ at
 400 the second home.

401 Figure 4 demonstrates the increase in neighborhood
 402 pollution as we move downwind, as a result of pollutant
 403 emissions from houses upwind, which is countered by plume
 404 dispersal. Since the distance between homes is linearly
 405 proportional to the emissions rate, for the Patsari stove, Figure
 406 5 shows the incremental increase in concentration in
 407 neighborhood pollution for homes spaced 100 m apart such
 408 that emissions from the first house reach $1 \mu\text{g}/\text{m}^3$ at the
 409 second home. The figure demonstrates that neighborhood
 410 pollution impacts reach $4 \mu\text{g}/\text{m}^3$ by 8 km or 80 homes

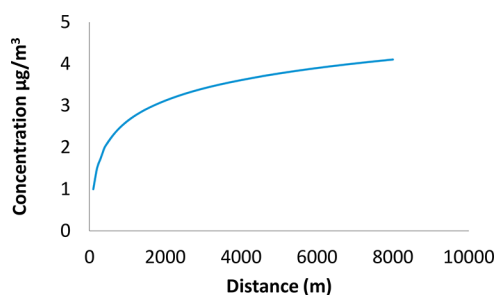


Figure 5. Increase in concentration at each successive building with distance for homes with Patsari stoves spaced 100 m apart.

downstream under this scenario. When added as background
 concentrations to the Monte Carlo single zone model
 simulation, plancha-type stoves would still meet WHO IT1
 of $35 \mu\text{g}/\text{m}^3$.

5. LIMITATIONS

Use of Monte Carlo simulations and three-dimensional
 computational fluid dynamics (CFD) analysis clearly repre-
 sents an idealized set of assumptions and scenarios that are not
 intended to represent actual communities, rather they are
 intended to facilitate the establishment of benchmarks that can
 be further evaluated and tested with field measurements. In
 particular, simulated indoor air concentrations are sensitive to
 assumptions on kitchen ventilation (airflow and mixing
 volume). Although ventilation rates were measured in a
 kitchen with similar construction, dimensions, and door/
 window openings as the average kitchen in these regions of
 Michoacán, in practice, kitchens vary considerably particularly
 where self-built structures are not standardized. These
 simulations do not capture the range of variability in kitchen
 design or the actual indoor air concentrations in individual
 kitchens; rather, they are intended to give an idea of indoor air
 concentrations in a typical kitchen for the region. Similarly,
 there are many differences in kitchen construction and design
 in different climatic conditions in Mexico and other regions of
 the world, which would impact kitchen ventilation and indoor
 air concentrations and, thus, the number of simulations and
 homes meeting guidelines. For example, Park and Lee report
 lower average air exchange rates of 12.2 air changes per hour
 (ACH) in 18 homes based on CO decay for a single event of 7
 min or longer. Actual air exchange rates were likely higher
 however as they assumed zero background CO concentration,
 well mixed air, and no CO source strength during the decay
 although the source was not removed from the room.²⁶ CO
 emissions can be very significant in the smoldering phase of the
 fire, and emissions are typically quite stratified and not well
 mixed in village kitchens. Smith et al. estimated that air
 exchanges greater than 35 ACH would be required for their
 modeled concentration to agree with measured concentrations
 in Indian village homes.²⁷ Bhanger reported mean air exchange
 rates of 15.8 ± 4.04 ACH for 10 homes with mud walls using
 traditional stoves in Kaldari, India.²⁸ Cowlin measured 6
 kitchens in San Lorenzo Guatemala and reported air exchange
 rates of 15.8 ± 5 ACH for closed kitchens, 18.3 ± 7.8 ACH for
 partially closed, and 31.4 ± 18.8 ACH for open homes.²⁹ In
 addition, they showed that height of measurement significantly
 impacts air exchange estimates in open homes when a mixing
 fan is not present with lower-height measurements producing
 lower air exchange estimates.²⁹ In the current study, the mean
 ventilation rates were substantially higher than many of those
 reported in the literature but similar to Cowlin estimates for
 open homes at higher measurement heights. A mixing fan was
 present in the current study which may account for the higher
 estimates, and kitchen construction in this region is typically in
 a separate room from the main house with large eave spaces
 and ~ 1 cm gaps between wall planks. In addition, simulated
 indoor concentrations agree well with measured concentra-
 tions in homes.¹⁹ To evaluate the impact of air exchange
 rates, therefore, and to illustrate how these stoves might
 perform in other environments, Figure 6 shows the percent of
 simulations meeting WHO guidelines under different air
 exchange profiles with other parameters kept as above. With
 air exchange rates above 40 ACH, there is not a large

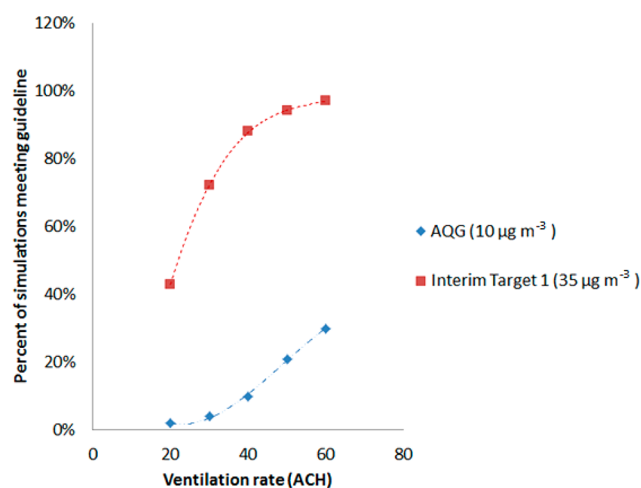


Figure 6. Percent of simulations meeting WHO guidelines under different air exchange profiles.

472 difference in the number of simulations meeting Interim
473 Target 1 but a significantly greater fraction meeting the air
474 quality guideline. Conversely, the fraction meeting Interim
475 Target 1 decreases rapidly below 30 ACH. Where direct
476 measurements of fugitive emissions are not possible, these
477 results give a baseline for the estimated fraction of simulations
478 meeting the guidelines when combined with local data on
479 kitchen ventilation.

480 Modeling homes spaced equally apart directly downwind
481 and emitting all at the same time represents an idealized
482 conservative scenario to be more protective of health. While
483 the modeled indoor air concentrations based on the Monte
484 Carlo single zone model agree well with controlled measure-
485 ments, they differ substantially from actual communities due to
486 the upwind contributions of a variety of sources including
487 homes and small scale industries, variable wind speed in actual
488 communities, temporal (daily and seasonal) differences in
489 emissions, behavior, meteorological parameters, etc. The
490 impact of upwind sources on the effective indoor air reductions
491 seen from plancha-type stove dissemination in communities
492 has been demonstrated by Armendáriz-Arnez et al.¹³ While the
493 neighborhood pollution impacts are modeled from stove
494 emissions, the contributions of ambient background emissions
495 from further upstream are not included, as they are spatially
496 and temporally variable in different countries around the world
497 and limit application of these analyses. Background ambient
498 concentrations can be added for different regions, however, to
499 make these analyses more relevant to the local context.

500 Stacking of cooking stoves (the practice of using more than
501 one stove) has been widely reported in Mexico and other
502 places where households do not transition fully to cleaner
503 burning cookstoves but retain the traditional stove, often
504 outdoors for some cooking tasks.³⁰ The current analyses do
505 not include stacking in homes as the purpose was to evaluate
506 the potential benefits of plancha-type stoves and to determine
507 spatial criteria where these stoves could be promoted. Similar
508 to LPG, where stacking occurs and the traditional fire is used
509 indoors for some tasks, air quality guidelines may not be met.
510 However, in this region, use of the traditional fire in
511 combination with the plancha-type stove tends to be in
512 outdoor environments for heating water.

■ ASSOCIATED CONTENT

513

📄 Supporting Information

514

The Supporting Information is available free of charge on the
ACS Publications website at DOI: 10.1021/acs.est.8b01704. 515

Emissions hood configuration; test kitchen; Monte 517
Carlo and CFD model parametrizations; locally 518
measured housing characteristics and ventilation param- 519
eters; CFD methodology validation (PDF) 520

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Notes

526

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