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Research

Arbuscular mycorrhiza convey significant plant carbon to a diverse hyphosphere microbial food web and mineral-associated organic matter

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Summary

• Arbuscular mycorrhizal fungi (AMF) transport substantial plant carbon (C) that serves as a substrate for soil organisms, a precursor of soil organic matter (SOM), and a driver of soil microbial dynamics. Using two-chamber microcosms where an air gap isolated AMF from roots, we ¹³CO₂-labeled *Avena barbata* for 6 wk and measured the C *Rhizophagus intrara-dices* transferred to SOM and hyphosphere microorganisms.

• NanoSIMS imaging revealed hyphae and roots had similar 13 C enrichment. SOM density fractionation, 13 C NMR, and IRMS showed AMF transferred 0.77 mg C g⁻¹ of soil (increasing total C by 2% relative to non-mycorrhizal controls); 33% was found in occluded or mineral-associated pools.

• In the AMF hyphosphere, there was no overall change in community diversity but 36 bacterial ASVs significantly changed in relative abundance. With stable isotope probing (SIP)-enabled shotgun sequencing, we found taxa from the Solibacterales, Sphingobacteriales, Myxococcales, and Nitrososphaerales (ammonium oxidizing archaea) were highly enriched in AMF-imported ¹³C (> 20 atom%). Mapping sequences from ¹³C-SIP metagenomes to total ASVs showed at least 92 bacteria and archaea were significantly ¹³C-enriched.

• Our results illustrate the quantitative and ecological impact of hyphal C transport on the formation of potentially protective SOM pools and microbial roles in the AMF hyphosphere soil food web.

Introduction

Arbuscular mycorrhizal fungi (AMF, phylum Mucoromycota, subphylum Glomeromycotina) form symbiotic associations with over 80% of vascular plant families (Schüßler et al., 2001; Spatafora et al., 2016) and facilitate plant nutrient uptake in exchange for photosynthetically derived carbon (C) (Smith & Read, 2008; Willis et al., 2013). Arbuscular mycorrhizal fungi consume 6.2% of plant photosynthate C on average (Hawkins et al., 2023) and grow extensive hyphal networks into soil (Jakobsen & Rosendahl, 1990; Tome et al., 2015), which can account for 15-30% of the soil microbial biomass (Leake et al., 2004; Parniske, 2008; Qin et al., 2017), and explore a soil volume substantially larger than fine roots alone (See et al., 2022). The composition, exudates, and interactions of AMF hyphae with plant roots and soil microbes play a critical and complex role in soil C processes (Wei et al., 2019; Domeignoz-Horta et al., 2021; Horsch et al., 2023). While soil C dynamics have been the focus of substantial research, gaps remain in our knowledge of the magnitude and fate of C fluxes within the mycorrhizal pathway (Sulman et al., 2018; Domeignoz-Horta *et al.*, 2021) and how surrounding soil biota are affected.

Plant-fixed C, distributed into soil as root biomass, mycorrhizal tissues, and their exudates, is the primary source of soil organic C and is transformed into soil organic matter (SOM) through diverse chemical and microbial processes (Torn et al., 1997; Trumbore, 2000; Schmidt et al., 2011). Since SOM holds not only organic C but also water and nutrients, its persistence is a major goal in climate change mitigation and sustainable land management. The mechanisms responsible for SOM persistence are complex and under active investigation (Jastrow et al., 2007; Treseder, 2016; Dynarski et al., 2020; Lehmann et al., 2020). Amidst the paths between plant photosynthate and SOM, AMF hyphae and soil microorganisms serve as primary intermediaries; their biomass and residues can contribute to slow-cycling and persistent forms of C that may become occluded within aggregates or associated with mineral surfaces (Miller & Jastrow, 2000; Dynarski et al., 2020; Angst et al., 2021; See et al., 2022). While occluded and mineral-associated soil C is not always more persistent due to potential desorption (Keiluweit et al., 2015; Li et al.,

2021), this C does typically have longer turnover times than other soil pools (Heckman *et al.*, 2022).

Several methods can distinguish C present as aggregateoccluded or mineral-associated C-forms. In the 'density fraction' approach, SOM is partitioned based on density into soil fractions that are operationally defined: particulate organic matter (free light fraction or light fraction), occluded within soil aggregates (occluded light fraction or occluded fraction), and mineralassociated (heavy fraction; Sollins *et al.*, 2006, 2009). Organic C in these soil fractions has distinct rates of biochemical and microbial degradation (Sollins *et al.*, 2006, 2009), and these fractions are widely thought to represent ecologically relevant soil components (Moni *et al.*, 2010; Hatton *et al.*, 2012) shaped by climate and ecosystem type (Sokol *et al.*, 2022).

Arbuscular mycorrhizal fungi may alter the relative distribution of organic C in different soil fractions (Orwin et al., 2011; Soudzilovskaia et al., 2015, 2019; Cheeke et al., 2016; Frey, 2019) and thereby influence the persistence of C in soil. Arbuscular mycorrhizal fungi have been shown to increase (Treseder, 2016; Wang et al., 2016; Zhang et al., 2020) or decrease (Hodge et al., 2001; Cheng et al., 2012; Herman et al., 2012; Paterson et al., 2016; Frey, 2019) soil C accumulation depending on the study system, environmental variables, and time frame. Arbuscular mycorrhizal fungi respire C and can lead to an increase in decomposition and soil respiration (Hodge et al., 2001; Cheng et al., 2012; Lang et al., 2021), reducing the amount of C available for SOM formation. Yet, AMF can promote soil C accumulation by increasing the formation of soil aggregates (Rillig et al., 2001b; Wilson et al., 2009; Leifheit et al., 2015) and SOM-mineral associations (Smits et al., 2009; See et al., 2022), and thereby protect organic C from decomposition. Arbuscular mycorrhizal fungi directly contribute C to SOM formation, as their cell materials become components of SOM when they senesce and are degraded (Kögel-Knabner, 2002; Langley & Hungate, 2003; Godbold et al., 2006). Arbuscular mycorrhizal fungi hyphae also release C compounds into surrounding soil (hyphosphere) as exudates (Hooker et al., 2007; Toljander et al., 2007) that can be transferred to other microbes (Herman et al., 2012; Kaiser et al., 2015; Bunn et al., 2019).

The presence of AMF may quantitatively and qualitatively alter soil bacterial communities, and AMF hyphae and spores provide important niches for bacterial interactions and growth. Bacteria can colonize living hyphae (Toljander et al., 2006, 2007) and form biofilms on hyphal surfaces (Lecomte et al., 2011), consume hyphal exudates (Kaiser et al., 2015), and help AMF mobilize nutrients in soil (Jiang et al., 2021). Bacteria can also attach to non-living hyphae and use them as a substrate (Toljander et al., 2006, 2007). Our prior work suggests AMF have diverse effects on nearby hyphosphere microbiomes: stimulating organic matter decomposition and nitrogen (N) transfer (Nuccio et al., 2013), supporting water transport (Kakouridis et al., 2022) and drought resilience (Hestrin et al., 2022), and stimulating cross-kingdom trophic interactions (Nuccio et al., 2022). However, these small-scale interactions can be difficult to directly measure, and under native soil conditions, it is particularly challenging to directly evaluate the hyphosphere effects without the influence of roots and rhizosphere organisms.

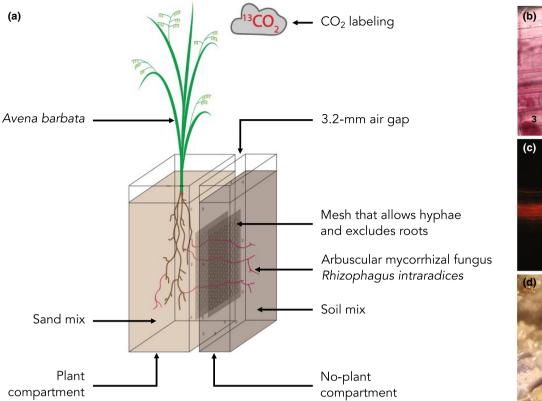
Arbuscular mycorrhizal fungi represent a globally important pathway for the flow of C from plants into soil (Hawkins et al., 2023), yet it remains unclear how AMF and their hyphosphere microbiome influence the initial incorporation of organic C into soil, separately from roots. In this study, we used ¹³C stable isotope tracing and molecular techniques to measure C transfer into soil, from the host plant Avena barbata, a widespread annual grass, via the AMF Rhizophagus intraradices. In a glasshouse experiment, we used a two-compartment microcosm design to isolate AMF transfer of C into soil. To assess how AMF affect the shortterm fate of plant-derived C, we used nanoscale secondary ion mass spectrometry (NanoSIMS) imaging and isotope ratio mass spectrometry (IRMS) to track ¹³C after it was fixed by host plants and transferred by AMF into soil during 6 wk of exponential phase plant growth. Using soil from the no-plant compartment, we characterized the form of AMF-contributed C by density fractionation into light fraction (likely still free hyphae), occluded fraction (contained within aggregate structures), and heavy fraction (mineralassociated) pools. Changes in C chemistry were assessed by ¹³C nuclear magnetic resonance (NMR) spectroscopy of soil aggregates. We also investigated the influence of AMF on hyphosphere soil microbial communities (independently from roots) by amplicon sequencing soil DNA and identifying the key microbial consumers of hyphal-derived C via stable isotope probing (SIP)-enabled metagenomic sequencing.

Materials and Methods

Experimental setup

Our experimental setup is described in detail in Kakouridis et al. (2022). In brief, three 2-wk-old A. barbata Pott ex Link seedlings were planted in the 'plant compartment' of twocompartment microcosms ($10 \times 2.5 \times 26.5$ cm; Fig. 1a), which was separated from the 'no-plant compartment' by a 3.2-mm air gap to prevent liquid water from traveling passively between compartments. Both sides of the air gap had nylon mesh, either 18 µm (allowing hyphae but excluding roots) or 0.45 µm (excluding both hyphae and roots). A total of 24 microcosms were used, 18 with 18 µm mesh, and six with 0.45 µm mesh. The plant compartment was packed with a 1 : 1 sand-clay mixture to a 1.21 g cm⁻³ density (referred to as the 'sand mix'). The no-plant compartment $(10 \times 1 \times 26.5 \text{ cm})$ was packed with a 1 : 1 soil-sand mixture to a 1.21 g cm⁻³ density (referred to as the 'soil mix') using fieldcollected soil containing live microbial communities. This soil (0-10 cm) was collected at the Hopland Research and Extension Center (38°59'35"N, 123°4'3"W) from a site where A. barbata was the dominant vegetation. Collected soil was air-dried, sieved (2 mm) to remove rocks and litter, and kept in a 4°C cold room before being mixed with sand and packed into the microcosms.

In the plant compartment, the sand mix was inoculated with 26 g of *R. intraradices* (N.C. Schenck & G.S. Sm.) C. Walker & A. Schuessler 2010 whole inoculum (accession no. AZ243, International Culture Collection of (Vesicular) Arbuscular Mycorrhizal Fungi (INVAM), The University of Kansas, Lawrence, KS, USA). *Rhizophagus intraradices* was selected because it naturally



Research



Fig. 1 Arbuscular mycorrhizal fungus (AMF) *Rhizophagus intraradices*, inoculated within the plant compartment, colonized roots of the host plant *Avena barbata* and grew across an air gap to reach the no-plant compartment. (a) Our experimental design tested for the movement of plant carbon (C) into root-free soil via AMF hyphae. This figure shows an AMF-permitted ¹³C microcosm ('+AMF') where plants were grown in a 99 atom% ¹³CO₂ atmosphere in Weeks 5–10 and AMF were able to access a no-plant compartment by crossing a 3.2-mm air gap (through a mesh with 18-µm pore size). Two other treatments included as follows: (1) AMF-excluded ¹³C controls ('-AMF') where plants were grown in a ¹³CO₂ atmosphere in Weeks 5–10, but AMF were not able to access the no-plant compartment because the mesh (0.45-µm pore size) between the compartments restricted both roots and hyphae from crossing from the plant compartment; and (2) AMF-permitted ¹²C controls ('¹²C'), where plants were grown in an ambient CO₂ atmosphere in Weeks 5–10 and AMF were able to access the no-plant compartment. All microcosms experienced an ambient CO₂ atmosphere in Weeks 1–4. Plants were harvested at the end of Week 10. In sum, the experiment included 12 +AMF microcosms, six –AMF microcosms, and six ¹²C microcosms. (b, c) *Avena barbata* roots stained with acid fuchsin showing AMF structures. (b) Bright field micrographs. (c) Fluorescence images at wavelengths λ_{ex} 596 nm and λ_{em} 615 nm, which target AMF. (d) Soil–sand mixture from the no-plant compartment of a +AMF microcosm with AMF hyphae visible under a dissecting microscope. In (b–d): 1, Root; 2, Hypha; 3, Arbuscule; 4, Vesicle.

colonizes *A. barbata* roots at the field site where seeds and soil were collected for this experiment (Kakouridis, 2021). Bone meal (78 mg) was mixed into the plant compartment to encourage AMF establishment, and into the no-plant compartment to act as a bait for AMF to cross the air gap (Supporting Information Table S1).

The microcosms were incubated in growth chambers in the Environmental Plant Isotope Chamber (EPIC) facility, located in the Oxford Tract Greenhouse at UC Berkeley, where environmental conditions were monitored and controlled. Three chambers were used, with eight microcosms in each, organized in a randomized fashion. Volumetric water content was monitored with electronic probes (EC-5; Decagon Services, Pullman, WA, USA), and maintained at *c*. 17% by watering three times weekly with autoclaved distilled water in both compartments. Ten milliliters of filter-sterilized Rorison's nutrient solution (Rorison & Rorison, 1987) was added to the plant compartment (low P) and no-plant compartment (high P) once per week (Table S1). The plant compartment of microcosms with 0.45 μ m mesh received twice as much nutrient solution as microcosms with 18 μ m mesh, to make up for the nutrients plants could obtain via AMF from the no-plant compartment (Table S1).

Twelve microcosms with 18 μ m mesh (AMF-permitted ¹³C microcosms, termed '+AMF') and six microcosms with 0.45 μ m mesh (AMF-excluded ¹³C microcosms, termed '-AMF') were placed in a ¹³C-labeled CO₂ atmosphere during Weeks 5–10. The remaining six microcosms with 18 μ m mesh (AMF-permitted ¹²C microcosms, termed '¹²C') remained in a natural abundance CO₂ atmosphere for the full 10 wk.

Harvest and sample processing

At the end of Week 10, all microcosms were destructively sampled. Shoots were cut at the base, dried at 60°C, and weighed for aboveground biomass (Table S1). Roots were gently harvested and divided into three aliquots so that each aliquot contained a randomized subsample of roots representing one-third of the root system: (1) To confirm AMF colonization under the microscope, roots were placed in distilled water for staining with acid fuchsin, (2) to confirm AMF presence by DNA sequencing, roots were placed in cell release buffer for molecular analysis (Brodie *et al.*, 2011), and (3) to assess the ¹³C enrichment via IRMS and to measure belowground biomass, roots were placed in paper envelopes and dried at 60°C (Table S1).

The sand mix and soil mix were collected and split into several aliquots: (1) 10 g for gravimetric water content was oven-dried at 105°C (Table S1) and (2) samples for molecular analysis were flash frozen in liquid nitrogen and stored at -80°C. In addition, the soil mix was preserved for hyphal extraction by storing at 4°C and for soil density fractionation by air drying.

Hyphae evident in the air gap were collected on the mesh facing the inside of the air gap using tweezers and placed into tubes for DNA extraction the same day. Visible hyphae were also collected from the plant and no-plant compartments of each microcosm using a dissecting scope and tweezers, then placed on silica wafers covered with carbon sticky tape for scanning electron microscopy (SEM) and NanoSIMS.

Microscopy

To observe and confirm AMF colonization under the microscope, roots were stained with acid fuchsin using a protocol modified from Habte & Osorio (2001) and described in Kakouridis *et al.* (2022). Stained roots were mounted on slides and observed under both bright field and fluorescence (λ_{ex} 596 nm/ λ_{em} 615 nm).

Molecular methods

To extract AMF spores and hyphae from the sand and soil mix, and then extract DNA from roots, soil mix, spores, and hyphae, we followed methods described in Kakouridis *et al.* (2022). In brief, we used a cell release buffer (Brodie *et al.*, 2011) to remove surface microbial cells from roots, hexametaphosphate to separate out AMF spores and hyphae, and conducted DNA extraction from roots, spores, hyphae, and soil mix with a DNeasy Power-Soil kit (Qiagen). To collect hyphosphere microbial communities, hyphae with soil attached were picked using tweezers under a dissecting microscope.

DNA extracted from roots, soil mix, spores and hyphae, and hyphosphere microbial communities was quantified with the Quant-iTTM PicoGreenTM dsDNA Assay Kit (Invitrogen), and concentrations were normalized to $5 \text{ ng }\mu\text{l}^{-1}$. To identify the AMF present, PCR was conducted using WANDA (Dumbrell *et al.*, 2011) and AML2 (Lee *et al.*, 2008) primers according to procedures described in Kakouridis *et al.* (2022). A sequence was considered a match for *R. intraradices* if query coverage and percent identity were both > 97%.

DNA extracts were used for both amplicon and SIPmetagenome sequencing. To identify the bacteria present on hyphae and in hyphosphere soil, a sequencing library was prepared using a phasing amplicon technique (Wu *et al.*, 2015) on the normalized DNA samples from +AMF and -AMF soil mix from the no-plant compartment with 515F/806R primers (Caporaso *et al.*, 2012) targeting the 16S rRNA gene V4 region. The library was prepared in the Zhou laboratory at the University of Oklahoma and sequenced on the Illumina MiSeq platform with 2 × 250 bp paired-end format.

16S rRNA sequence processing

Amplicon sequences were processed with QIIME2 (Bolyen *et al.*, 2019). After demultiplexing, a total of 637 832 raw sequences were obtained from the six +AMF and six –AMF soil samples. After primer trimming, sequences were denoised using DADA2 (Callahan *et al.*, 2016) and clustered into amplicon sequence variants (ASVs). The representative sequences of each ASV were then used to assign taxonomy based on a classifier trained with the SILVA database (v.132-99-515-806). Amplicon sequence variants unassigned at the domain level or identified as mitochondria or chloroplast sequences were discarded. This resulted in 3019 ASVs in 12 samples, with 23 502–41 551 sequences per sample. The SRA sequence records are accessible with the following link: https://www.ncbi.nlm.nih.gov/sra/PRJNA1051335.

¹³C DNA-SIP and metagenomic read processing

¹³C DNA-SIP and metagenomic sequencing of the fractionated DNA samples have been described previously by Nuccio *et al.* (2022). Briefly, 350 ng of ¹³C- and ¹²C-AMF DNA (n=3) was added to SIP density gradients, ultracentrifuged, fractionated, precipitated, and quantified with Lawrence Livermore National Laboratory's high-throughput HT-SIP pipeline (Nuccio *et al.*, 2022).

After quality control, we retrieved 16S sequences from the SIPmetagenomics dataset by mapping the raw SIP-metagenome reads (14 fractions per gradient) to our 3019 16S amplicon ASVs using BBSPLIT (Bushnell, 2014); this strategy allowed us to identify ¹³Cenriched organisms that may not have assembled in our previous work (Nuccio et al., 2022). Sequences were required to unambiguously map to a single ASV and mapped in 'semiperfect' mode, which requires a 100% sequence match (max index of 2) but allows N's and will allow a read to run off the end of the sequence. We generated an OTU table from the BBSPLIT ASV matches for each SIP fraction, and then calculated ASV atom percent excess (APE) using the QUANTITATIVE SIP (QSIP) pipeline (Hungate et al., 2015; Koch et al., 2018). We required an ASV be present in three replicates and at least three fractions. To determine whether an ASV was significantly enriched in ¹³C, we required the mean APE lower 90% confidence interval be significantly greater than zero, and the weighted average differences (WAD) of the ¹³C and ¹²C control density curves be significantly different (P-value < 0.05).

Soil density fractionation

We used a soil density fractionation method described in Fossum *et al.* (2022), which was modified from Hicks Pries *et al.* (2018) and Strickland & Sollins (1987). In brief, 20 g of air-dried soil

mix and 50 ml of sodium polytungstate (SPT; Geoliquids) were prepared to a density of 1.75 g cm^{-3} , centrifuged at 3700 g, and left to settle. Particles floating on top were defined as the light fraction. The light fraction was aspirated onto a 0.8-µm glass microfiber filter (Whatman, Cytiva, Marlborough, MA, USA), rinsed with Milli-Q water, and then dried at 55°C and weighed. To collect the occluded fraction, the remaining soil-SPT mixture was shaken with a benchtop mixer for 1 min, sonicated for 90 s, allowed to settle, and then centrifuged for 1 h. The occluded fraction was then removed by aspiration, dried, and weighed. The remaining sediment ($\rho > 1.75 \text{ g cm}^{-3}$) was defined as the heavy fraction. One hundred and fifty milliliters of Milli-Q water was added, vigorously shaken by hand, and then centrifuged for 20 min. The supernatant was aspirated and discarded. This process was repeated five times, or until the density of the supernatant reached 1 g cm $^{-3}$. The heavy fraction was transferred, dried, and weighed in the same manner as the light and occluded fractions. After drying, all fractions were ground to a fine powder.

¹³C isotope ratio mass spectrometry (IRMS)

Samples of shoots, roots, sand mix, soil mix, and soil fractions were finely ground, weighed, and analyzed for total C and ¹³C abundance by dry combustion on a PDZ Europa ANCA-GSL elemental analyzer interfaced to a PDZ Europa 20-20 isotope ratio mass spectrometer (Sercon Ltd, Cheshire, UK) with a precision for ¹³C measurements of 0.1 per mil.

Nanoscale secondary ion mass spectrometry (NanoSIMS)

To measure the atom% ¹³C of hyphae, mounted hyphae collected from the plant compartment, air gap, and no-plant compartment were coated with gold and mapped with an SEM (Inspect F; FEI, Hillsboro, OR, USA). The isotopic composition of the hyphae was measured by NanoSIMS (NanoSIMS 50; CAMECA, Gennevilliers, France). We collected a total of 150 analysis regions (30 µm by 30 μ m) to gain a broad survey of the distribution of ¹³C enrichment in hyphae. Each location was first sputtered with a c. 60 pA Cs^+ beam to a depth of *c*. 60 nm to enhance the yield of negative secondary ions and reach sputtering equilibrium (stable secondary ion counts; Ghosal et al., 2008). Then, a 2 pA Cs⁺ beam with a nominal spot size of 200 nm was used to raster the analysis area with 256×256 pixels with 20–30 scans per sample. Five masses were collected as follows: ${}^{16}O^-$, ${}^{12}C_2^-$, ${}^{12}C^{13}C^-$, ${}^{12}C^{14}N^-$, and ³¹P⁻ using electron multipliers and a mass resolving power c. 7000 (1.5× correction; Pett-Ridge & Weber, 2021). Pseudomonas stutzerii cells with known isotopic composition (previously analyzed by IRMS) were used as standards.

Images were processed using the L'IMAGE software (L.R. Nittler, www.limagesoftware.net). Images were corrected for deadtime and drift. Regions of interest (ROIs) were selected by first identifying hyphal structures from SEM and secondary electron imaging, and then based on areas of relatively uniform ¹³C enrichment from ratio images of ¹²C¹³C⁻: ¹²C₂⁻. Areas of high O⁻ counts (marker for minerals) were avoided. A total of 150 images were taken, from which a subset of 37 images with little charging and low O^- counts were selected for calculations of hyphae atom% ^{13}C .

Solid-state ¹³C nuclear magnetic resonance (NMR) spectroscopy

For +AMF microcosms and -AMF microcosms, hand-picked dried soil aggregate samples were finely ground. For logistical reasons, we were limited to running only six total samples for this method. To accommodate this limitation, the biological replicates were combined in pairs in order to reduce the number of samples to three for the +AMF microcosms and three for the -AMF microcosms. 1-100 mg of each of the six pooled samples were analyzed on a 500 MHz Bruker Avance-I Spectrometer (Bruker, Billerica, MA, USA) at the UC Davis NMR Facility as follows: MAS spinning speed 12 kHz, $^{13}C^{-1}H$ contact time 2 ms, relaxation delay time 1 s, number of scan 60 000–90 000 (signal-to-noise ratio: 15–20).

Statistical analyses

Statistical analyses were conducted using R v.3.6.1 (R Core Team, 2017). A one-way analysis of variance (ANOVA) coupled with Fisher's least-significant difference (LSD) test (package AGRICO-LAE, *P*-value adjustment using 'Holm') was used to differentiate means of atom% and the amount of ¹³C from different treatments.

A paired Wilcoxon test was used to compare the average atom% $^{13}\mathrm{C}$ of roots (IRMS) vs the average of hyphae NanoSIMS measurements from +AMF microcosms. One microcosm (#5) was excluded due to a technical issue with the hyphal sample preparation.

We calculated the alpha diversity (richness, Shannon index, and evenness, package VEGAN) and Faith's phylogenetic diversity (package PICANTE) of 16S rRNA detected prokaryotic communities (Kembel et al., 2010; Oksanen et al., 2019), and tested their differences between +AMF and -AMF treatments using a two-way ANOVA, treating 'chamber' as a random factor (and no significant chamber effect was observed). P-values were corrected using false discovery rate. The principal component analysis (package APE) and permutational multivariate analysis of variance (Adonis based on Bray-Curtis distance, package VEGAN) were performed to evaluate community composition difference among treatments. DESEQ analysis (Love et al., 2014) was used to identify ASVs responding to the presence of AMF. To reduce potential noise in modeling fitting, we removed ASVs that only occurred in one of the six biological replicate samples under either +AMF or -AMF conditions. After this prevalence filtering, 1534 ASVs were included in the DESEQ analysis for parameter estimation and model fitting.

Through the manuscript, unless otherwise stated, we present 'mean \pm standard error' where applicable.

Results

Confirmation of root colonization and hyphal networks in the no-plant compartment

Using WANDA-AML2 amplicon sequencing, we confirmed that roots of all microcosms were colonized by *R. intraradices*

and confirmed this AMF taxon in air gap hyphae and soil mix samples from +AMF and ¹²C microcosms. We also observed hyphae, spores, and arbuscules in roots stained with acid fuchsin using light and fluorescence microscopy (Fig. 1b,c), confirming that *R. intraradices* was actively growing in roots over the course of our experiment. We observed hyphae crossing the air gap, and extensive hyphal networks in the no-plant compartment of +AMF and ¹²C microcosms (Fig. 1d). In –AMF microcosms, we did not observe hyphae crossing the air gap nor hyphal networks in the soil mix.

Confirmation of ¹³C labeling and ¹³C presence in no-plant compartment of +AMF microcosms

The atom% ¹³C of shoots in +AMF and -AMF microcosms was $43.5 \pm 1.5\%$ and $43.7 \pm 1.7\%$, respectively, and $1.1 \pm 0.001\%$ in ¹²C microcosms (Fig. S1; Table S1). Root enrichment was 41.3 ± 1.9 atom% and 42.2 ± 2.0 atom% in +AMF and -AMF microcosms, respectively, and $1.1 \pm 0.002\%$ in ¹²C microcosms (Fig. S1; Table S1). The atom% ¹³C of the sand mix (plant compartment) was $5.3 \pm 0.4\%$ and $4.2 \pm 0.2\%$ in +AMF and -AMF microcosms, respectively, and $1.1 \pm 0.003\%$ in ^{12}C microcosms (Fig. S1; Table S1). The soil mix (no-plant compartment) of +AMF microcosms was significantly more ¹³C enriched than that of -AMF and ¹²C microcosms (atom% ¹³C $1.8 \pm 0.1\%$ vs $1.1 \pm 0.001\%$ and $1.1 \pm 0.0003\%$, respectively, P < 0.001; Fig. S1; Table S1). These results confirm there was significantly more ¹³C in the no-plant compartment of +AMF microcosms and that the no-plant compartments of -AMF microcosms were not contaminated by ¹³C. The presence of AMF did not change soil gravimetric water content, aboveground biomass, root : shoot ratio, shoot C : N ratio and %N, but significantly increased plant P content and decreased plant N:P ratio (Table S1; Kakouridis et al., 2022).

¹³C in light, occluded, and heavy fractions of SOM

Soil organic matter from the soil mix (no-plant compartment) was separated into light, occluded and heavy fractions using a sodium polytungstate density gradient approach. Based on ¹³C IRMS analysis of these density fractions, after 6 wk of labeling, 26.7 ± 3.0 mg of new ¹³C was present in the no-plant compartment soil mix of +AMF microcosms (Fig. 2a; Tables S1, S2). This represents 74.4 mg C, which is equivalent to 0.77 mg C g⁻¹ of the native soil material (not including the sand) or 2.05% of the total C in the soil mix in the no-plant compartment (Methods S1).

The 26.7 mg of newly added ¹³C found in the soil mix was distributed amongst the density fractions as follows: 17.8 mg ¹³C (representing 49.6 mg total C or 1.37% of the total soil mix C) in the light fraction, 2.2 mg ¹³C (representing 6.1 mg C or 0.17% of the total soil C) in the occluded fraction, and 6.7 mg ¹³C (representing 18.7 mg C or 0.52% of the total soil C) in the heavy fraction (Fig. 2b).

Hyphae highly enriched with ¹³C

To determine the average ¹³C enrichment of hyphae, we collected NanoSIMS images of hyphae from the plant compartment, air gap, and no-plant compartment of +AMF microcosms (Fig. 3a–e). Based on an average of 37 images, hyphae were $35.9 \pm 1.5\%$ ¹³C enriched (Fig. 3f; Table S3). A paired Wilcoxon test between five matched roots and hyphae samples indicated no statistical difference between the enrichment of roots and hyphae (P=0.063; Fig. 3f).

NMR spectra suggest carbohydrates in the presence of AMF

¹³C-NMR spectra were obtained for six soil mix samples (three from +AMF microcosms and three from -AMF microcosms;

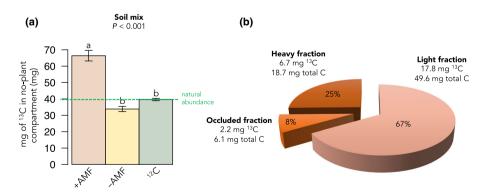


Fig. 2 Total carbon (C) and ¹³C transported by the arbuscular mycorrhizal fungus (AMF) *Rhizophagus intraradices* into a no-plant compartment and separated by soil organic matter fractions. (a) mg ¹³C in soil mix (no-plant compartment) in AMF-permitted ¹³C microcosms ('+AMF'), AMF-excluded ¹³C microcosms ('-AMF'), and AMF-permitted ¹²C microcosms ('¹²C') after 6 wk of ¹³CO₂ labeling of host plant *Avena barbata*. Different letters above bars represent statistically significant differences (one-way ANOVA and Fisher's LSD test); corresponding *P*-value is indicated above the plot. Error bars represent SE (*n* = 24). The dashed green line represents the natural abundance level of ¹³C. (b) Distribution of ¹³C contributed by AMF hyphae in light, occluded, and heavy soil fractions after 6 wk of labeling based on soil density fractionation followed by ¹³C isotope ratio mass spectrometry (IRMS) analysis.

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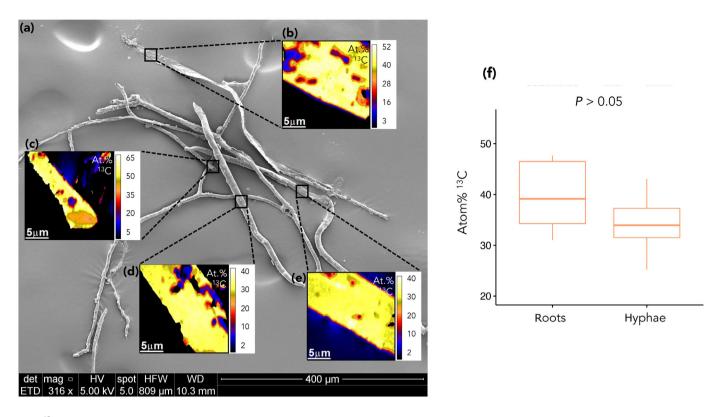


Fig. 3 ¹³C isotope enrichment of roots of the host plant *Avena barbata* and hyphae of the arbuscular mycorrhizal fungus (AMF) *Rhizophagus intraradices*. (a) Scanning electron microscopy image and (b–e) nanoscale secondary ion mass spectrometry (NanoSIMS) ¹³C abundance images of hyphae from the noplant compartment of an AMF-permitted ¹³C microcosm ('+AMF'). Areas with cooler colors (black, blue) have a lower atom% ¹³C while areas with warmer colors (orange, yellow) have a higher atom% ¹³C. (f) Atom% ¹³C of roots and hyphae from +AMF microcosms. Atom% ¹³C of roots was measured by isotope ratio mass spectrometry (IRMS). Atom% ¹³C of hyphae was calculated using the average of 37 NanoSIMS images. Whiskers represent the minimum and maximum values in the data; *P*-value is indicated above the plot (paired Wilcoxon test, *n* = 5).

Fig. 4). The NMR spectra for +AMF soil mix samples show greater variability and slightly higher peaks in the 45–110 ppm region, corresponding to the O-alkyl C functional group and indicative of carbohydrates (Kögel-Knabner, 1997), which is not observed anywhere else in the spectra (Fig. 4).

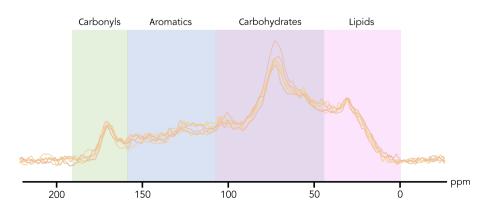
Hyphosphere community changes in response to AMF

There were no significant differences in the number of 16S sequences obtained from +AMF (32451 ± 4596) vs -AMF (34171 ± 6655) treatments after quality control. The 12 samples

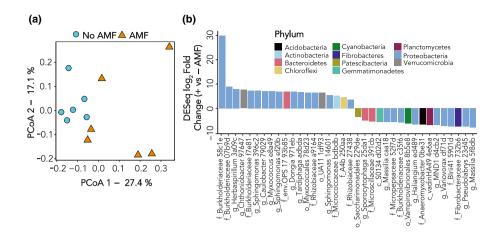
analyzed resulted in 3019 amplicon sequencing variants (ASVs), with 23 502–41 551 sequences per sample. There was no significant difference in the alpha diversity of prokaryotic communities in +AMF vs –AMF microcosms, as measured by richness, Shannon index, evenness, and Faith's phylogenetic diversity (ANOVA, P>0.05). However, community composition did differ between +AMF and –AMF treatments (Fig. 5a), measured via permutational multivariate analysis of variance (Adonis, $F_{1,8}$ = 2.23, P=0.004).

DESEQ analysis identified 19 ASVs that were significantly more abundant in the presence of AMF, and 17 ASVs that were

Fig. 4 Solid-state ¹³C nuclear magnetic resonance (NMR) spectra of the soil mix (no-plant compartment) of six microcosms. Three soil mix samples were from compartments accessible to hyphae of the arbuscular mycorrhizal fungus (AMF) *Rhizophagus intraradices* ('+AMF', in orange). Three soil mix samples were from compartments inaccessible to AMF hyphae ('-AMF', in yellow). Spectra intensities were normalized to the resonance at *ca.* 30 ppm.



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significantly less abundant in the presence of AMF (Fig. 5b; Table S4). Twenty-three of these 36 (65%) ASVs belong to the Proteobacteria Phylum. Fourteen increased in relative abundance in +AMF microcosms, and nine decreased. Other ASVs that significantly increased or decreased in relative abundance in +AMF microcosms are listed in Table S4.

¹³C enrichment of hyphosphere bacteria

Of the 3019 ASVs detected by 16S amplicon analysis, 2934 ASVs also had a semiperfect read match within the SIPmetagenomics dataset. We identified 92 ASVs significantly enriched in ¹³C (Fig. 6) among the total 682 ASVs that passed the QSIP criteria (e.g. detected in enough replicates and fractions). These ASVs span a diverse phylogeny, with Proteobacteria the dominant phylum. The most enriched taxa, with atom percent excess (APE) ${}^{13}C > 20\%$, include those from the delta-Proteobacteria order Myxococcales, Thaumarchaeota family Nitrososphaeraceae, the Acidobacteria genus Candidatus Solibacter, and the Bacteroidetes family AKYH767 (within the Sphingobacteriales). The most enriched ASV was from the Haliangium genus within the Myxococcales (APE 35.2%), an enrichment value similar to the AMF hyphae (35.9%) (note, Myxococcales has been proposed to be reorganized as the Myxococcota phylum in the GTDB taxonomy; Parks et al., 2018). There were 33 ASVs from diverse phylum and orders with APE of over 10% (Fig. 6).

Discussion

Arbuscular mycorrhizal fungi transport a significant amount of photosynthetic C into the soil matrix beyond the extent of roots alone, but the contribution of AMF to soil C retention lacks experimental exploration. We used ¹³CO₂ labeling of *A. barbata* in a two-compartment microcosm with an air gap designed to isolate *R. intraradices* hyphae from roots and quantify the amount and form of photosynthetic C transported via AMF hyphae into soil. After 6 wk of labeling, AMF hyphae crossing the air gap were ¹³C enriched at over 35 atom% and the total C transferred by hyphae and retained in soil represented over 2% of the total soil C, a third of which was found occluded within soil aggregates or stabilized in the mineral-associated fraction.

Fig. 5 Bacterial community composition in the soil mix (no-plant compartment) of 12 microcosms based on 16S rRNA gene amplicons. Six soil mix samples were from compartments accessible to hyphae of the arbuscular mycorrhizal fungus (AMF) *Rhizophagus intraradices* ('+AMF'). Six soil mix samples were from compartments inaccessible to AMF hyphae ('-AMF'). (a) Principal component analysis showing distinct community composition with or without AMF. (b) Amplicon sequence variants that significantly increased and decreased in relative abundance in +AMF microcosms compared to -AMF microcosms (DESEQ analysis of 16S rRNA reads, P < 0.05 adjusted for multiple comparisons).

AMF-transported C quickly enters protected pools

After 6 wk of labeling, the ¹³C that remained within the noplant compartment soil included ¹³C that was fixed by plants, transferred to AMF, and transported to the no-plant compartment, minus the ¹³C that left the system via soil respiration or dissolved organic carbon (DOC) in water that drained from the microcosms. Thirty-three percent of this remaining ¹³C was in either the occluded fraction or the heavy fraction. Organic C in both fractions is thought to be at least somewhat protected from decomposition by its location inside aggregates or its association with minerals (Kleber *et al.*, 2007; Keiluweit *et al.*, 2012; Throckmorton *et al.*, 2015; Kallenbach *et al.*, 2016).

The ¹³C enrichment of soil density fractions suggests that a large portion of the AMF-derived C was still present in the form of living or dead hyphae, with 67% of the ¹³C in the light fraction. The density of a fungal hypha has been estimated to be 1.1 g cm^{-3} (Bakken & Olsen, 1983), so hyphae should be concentrated in the light fraction unless they are in a protected form. Organic C in the light fraction is readily available to microbial and physical degradation and therefore tends to have a shorter residence time in soil compared with C in occluded and heavy fractions (Kleber *et al.*, 2007; Keiluweit *et al.*, 2012; Throckmorton *et al.*, 2015; Kallenbach *et al.*, 2016).

We found that a small fraction of the ¹³C, 8%, was in the occluded fraction inside aggregates that were sufficiently stable to retain structure and hold on to the ¹³C during the soil density fractionation process. As soil aggregation can physically protect C-rich litter from microbial degradation, an increase in aggregation may be an important mechanism for C sequestration (Rillig, 2004; Rillig & Mummey, 2006; Rillig et al., 2010). Arbuscular mycorrhizal fungi may increase the stability of soil aggregates by producing and releasing glomalin-related soil glycoproteins (GRSP) (Wright et al., 1996; Rillig et al., 2001a,b, 2002, 2010; Halvorson & Gonzalez, 2006). A recent review, however, argues that the evidence supporting the production of GRSP by AMF is primarily correlative and GRSP may have multiple non-AMF origins (Holátko et al., 2020). It is clear that AMF hyphae can increase soil aggregation by physically holding soil particles together in their intricate extraradical mycelium.

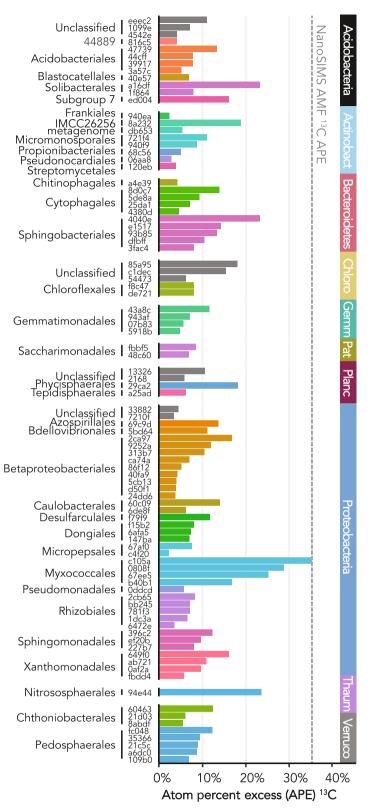


Fig. 6 Following DNA-Stable Isotope Probing (DNA-SIP), ¹³C-enriched hyphosphere taxa and their level of enrichment in ¹³C atom % excess (APE). The amplicon sequence variants (ASVs, first five alphanumeric characters shown) are SIP-metagenomic reads that were mapped to corresponding ASVs obtained from 16S amplicon sequencing of the same DNA samples. Bars are grouped by Phylum of the ASVs and colored by Order. The dashed line shows the APE of hyphae of the arbuscular mycorrhizal fungus (AMF) *Rhizophagus intraradices* estimated by nanoscale secondary ion mass spectrometry (NanoSIMS).

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The physical protection of C in aggregates can be visually observed in Fig. 1(d) where soil aggregates were mechanically broken open to reveal AMF hyphae.

Lastly, c. 25% of the transported ¹³C remained in soil within the heavy fraction. The origin of this mineral-associated organic C may have been hyphal exudates, hyphae that senesced and degraded to the point of being sorbed to mineral surfaces, or bacteria that consumed the exudates or hyphae (Pett-Ridge & Firestone, 2017). The transfer of root C into fungal and bacterial cell biomass appears to be a particularly important step that precedes C stabilization on mineral surfaces (Kleber et al., 2007, 2015; Keiluweit et al., 2012; Cotrufo et al., 2013: Throckmorton et al., 2015; Kallenbach et al., 2016; Pett-Ridge & Firestone, 2017). In an earlier glasshouse study, we used soil collected from the same field site and the same plant genus, but used only one compartment for both roots and hyphae. We found 43% of the C from root and hyphae in the heavy fraction after 12 wk of labeling (Pett-Ridge & Firestone, 2017). Carbon contributed by hyphae alone was more than half of that from both roots and hyphae, despite the shorter labeling time reported here (6 wk) compared to that of the previous study. Both sets of results demonstrate rapid interactions between root and hyphal exudates and soil minerals.

AMF contribute significant soil C

In our experiment, C transported by AMF hyphae made up over 2% of the total organic C in the no-plant compartment after a 6wk incubation. This amount of organic C is substantial relative to the estimate that up to 5% of the total organic C in soil is from living microbial biomass (Paul & Clark, 1989). This comparison is necessarily approximate because we do not know the effects that the no-plant compartment characteristics (1:1 mix of soil and sand, an air gap, and only 6 wk to develop) may have had relative to natural soil under equilibrium conditions. Nonetheless, this finding is consistent with estimates that AMF are 15-30% of the total soil microbial biomass in natural soil (Olsson & Wilhelmsson, 2000; Rillig et al., 2001b; Leake et al., 2004; Qin et al., 2017). Arbuscular mycorrhizal fungi have been observed to spread over 11 cm away from roots in a pot culture in 7 wk (Jakobsen et al., 1992), which is in line with observations in our study. It appears that AMF have the ability to grow hyphal networks quickly and thus rapidly distribute a substantial amount of C into the soil.

Notably, we did not measure DOC in the hyphal compartment. As such, our calculations do not take into account any ¹³C that desorbed into the supernatant during the fractionation process, or ¹³C that left the system via water that drained from the microcosms. Root and hyphal exudates in the form of DOC are key substrates for soil microbes and part of the pathway that generates mineral-associated forms of soil C, suggesting that our results may be an underestimation of the contribution of AMF to soil C pools.

In the NMR spectra, the presence of AMF seems to impact the 45–110 ppm region, typically associated with O-alkyl C, or carbohydrates, more than any other regions (Fig. 4). A higher peak

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in this region would be consistent with an increase in carbohydrate-rich biomass or hyphal deposition. While the -AMF soils all have a similar peak height and shape in this spectral region, some of the +AMF replicates have a broader or higher peak. It was not possible to standardize the amount of hyphal biomass or hyphal exudate in the soil samples from the noplant compartments analyzed by NMR, but it seems probable that at least one of the +AMF samples had a greater amount of a carbohydrate-like compound, leading to a higher peak. This would suggest that AMF in the soil away from the direct influence of roots produced hyphae and/or released metabolites that contained a large proportion of carbohydrates. Arbuscular mycorrhizal fungi hyphae contain structural chitin (an aminopolysaccharide) and are known to exude diverse compounds in the hyphosphere, including sugars (in particular hexoses) (Toljander et al., 2007; Bharadwaj et al., 2012; Sato et al., 2015; Luthfiana et al., 2021), so our observations of increased hyphosphere carbohydrates appear consistent with previous studies.

Enrichment of AMF C is comparable to that of roots

The NanoSIMS results show there was no significant difference between the ¹³C enrichment of roots and hyphae (Fig. 3f). Hyphae have been shown not only to bring plants nutrients but also to carry up to 34% of the water transpired by plants (Ruth *et al.*, 2011; Quiroga *et al.*, 2019; Kakouridis *et al.*, 2022). In fact, the evolution of land plants may have only been possible through the formation of mycorrhizal associations, because the algal ancestors to land plants were not adequately equipped to live on land independently (Jeffrey, 1962; Pirozynski & Malloch, 1975; van der Heijden *et al.*, 2015). NanoSIMS evidence from this study further supports the idea that AMF truly act as an extension of the root system.

Composition and activity of AMF hyphosphere microbial communities

Local-scale soil biogeochemistry is an important driver of microbial community composition (McGuire & Treseder, 2010; Wagg *et al.*, 2014; Graham *et al.*, 2016; Anthony *et al.*, 2020; Kivlin *et al.*, 2020). We were interested in how AMF affect the soil microbial community – separately from roots – which we assessed with both ¹³C-SIP analysis and amplicon 16S sequencing. These two methods ask fundamentally different questions: ¹³C-SIP asks what organisms partake in the AMF-C food web, either peripherally (low ¹³C enrichment) or centrally (high ¹³C enrichment), while amplicon sequencing asks how the presence vs absence of AMF enables dramatic (e.g. log₂fold) changes in the relative abundance of taxa. These two methods provide critical but different perspectives on the composition and activity of hyphosphere microbial communities.

Arbuscular mycorrhizal fungi frequently co-occur with bacteria in soil away from the direct influence of roots (Yuan *et al.*, 2021). Here, using 16S amplicon analysis, we found that the relative abundance of 19 ASVs significantly increased and the

relative abundance of 17 ASVs significantly decreased in the presence of AMF in the root-free hyphal compartment. Over half of these changed ASVs belonged to the Proteobacteria Phylum, suggesting a potential phylogenetic coherence of bacteria whose abundances strongly respond to the presence of AMF. These 36 ASVs only comprise 1.2% of all ASVs detected, but their abundance changes were sufficiently large to shift the overall community structure, as other studies have also discovered (Nuccio et al., 2013; Gui et al., 2017; Rodríguez-Caballero et al., 2017; Cao et al., 2020; Hao et al., 2020).

Some of the taxa that increased in abundance did not incorporate ¹³C. Two ASVs of the Burkholderiaceae family (58c1e and 07b9d) increased the most in relative abundance among all ASVs but were not ¹³C enriched. Arbuscular mycorrhizal fungi may have stimulated these organisms by mechanisms that did not involve ¹³C consumption, such as by changing local edaphic or nutrient conditions (e.g. increasing or decreasing N, P) or by non-assimilative C processes. In addition, we were unable to determine whether 10 of the 19 ASVs that increased in relative abundance incorporated ¹³C.

Compared with amplicon analysis, ¹³C-SIP identified many more organisms in the AMF food web. DNA-SIP isolates the populations actively assimilating a labeled substrate and does not detect DNA from inactive microbes, relic DNA fragments, or slow-growing communities. Even though only 23% of the ASVs detected via amplicon sequencing were captured by metagenomics and subsequent OSIP analysis, OSIP detected many organisms that took up ¹³C without detectable changes in relative abundance. The AMF hyphosphere is a microhabitat that can be difficult to isolate and sample without also collecting large amounts of background soil; in these instances, SIP may more sensitively identify populations responding to AMF than standard amplicon analyses. However, since shotgun metagenomic sequencing can hardly capture all 16S genes from a soil community, the true AMF food web likely contains more taxa than detected in this study.

This study goes beyond our previous work (Nuccio et al., 2022) by using 16S amplicon sequences from the same DNA samples as a reference database to identify more ¹³Cenriched hyphosphere taxa within our SIP-metagenomes. Previously, we used metagenome assembled genomes (MAGs), which limits the analysis to genomes amenable to assembly. In line with our previous findings, the three most enriched ASVs, including one with ¹³C enrichment close to that of AMF hyphae, all belong to the family Myxococcales in the genus Haliangium, likely facultative predators (Nuccio et al., 2022). An ammoniaoxidizing archaea (AOA), Nitrososphaera, was also highly enriched with 23% APE. We also identified multiple highly ¹³C enriched taxa that were not assembled into MAGs in our previous study. For example, ASVs belonging to Solibacterales, Subgroup 7, and Phycisphaerales were ¹³C enriched with over or nearly 20% APE. Solibacterales taxa are known to solubilize phosphorous (Bergkemper et al., 2016), which may synergistically interact with AMF by P-C exchange (Nacoon et al., 2020). Acidobacteria Subgroup 7 is reported to increase with pH (Jones et al., 2009) and has potential to use different carbohydrates (de

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Chaves et al., 2019). The high ¹³C enrichment of this Acidobacteria subgroup may be a combination of microhabitat change and substrate availability associated with AMF but needs further study. The Phycisphaerales ASV matched well to soil microbes from other studies (Zeglin et al., 2016; Addison et al., 2019), and our finding that it consumed AMF-derived C may be a first clue about its function in soil.

Some ASVs with moderate APE that are from Orders Rhizobiales, Caulobacterales, Sphingomonadales have been reported to associate with AMF (Hoseinzade et al., 2016; Agnolucci et al., 2019; Akvol et al., 2019; Hao et al., 2020; Yuan et al., 2021). Here, we confirmed their consumption of hyphae-derived C. Rhizobiales and Sphingomonadales are considered rhizosphere adaptors (Lei et al., 2019) and together consist of 4.8% of the total AMF hyphosphere ASVs in our study. The hyphosphere may have similar characteristics to the rhizosphere, in that low-molecular-weight C from hyphal exudates selectively enriches these taxa. Their moderate but not high levels of ¹³C incorporation may be limited by the relatively small portion of the cells that are spatially close to hyphae and have access to AMF C. The photosynthetic C transported by AMF thus facilitated a diverse hyphosphere microbial food web dependent on AMF hyphal C supply.

Taxa in the hyphosphere food web may contribute to incorporating AMF-derived C into soil, SOM formation, and stabilization. Resources derived from hyphae can increase microbial biomass, activity, frequency of interactions, and the rate of organic matter transformations (Sokol et al., 2022). Necromass from these microbial cells is also likely a significant source of SOM and may have formed part of the mineral-associated C that we recovered from the soil heavy fraction (Fossum et al., 2022; Sokol et al., 2022; Hu et al., 2023). Future studies that compare the AMF hyphosphere with and without its microbiome would be needed to quantify this contribution.

Conclusion

Our study shows that AMF move a substantial amount of C from plants beyond the rhizosphere. After only a few weeks, about a third of this mobilized C had made its way into aggregate-occluded and mineral-associated forms, which may have longer residence times in soil. In hyphosphere soil, we measured an increased presence of organic C compounds (such as chitin and hexoses) that are generally associated with AMF hyphae. Arbuscular mycorrhizal fungi C inputs modified the hyphosphere bacterial community, facilitating a diverse microbial food web that incorporated hyphae-derived C. This effect likely stimulated enhanced AMF (and thereby plant) access to N and P. Together, our findings indicate that AMF play a key role in the formation of SOM and thus could be employed in strategies for climate change mitigation and sustainable land management.

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Competing interests

None declared.

Author contributions

AK and MKF designed the experiment with the assistance of JAH, MY, KYE-M, EEN and JP-R. AK, JAH, MLM, CAF, and PKW performed the experiment with assistance from KYE-M. AK, JAH, MY, EEN, CAF, MLM, PSN, PKW and MKF analyzed the data with the assistance of KYE-M and JP-R. AK, MY, EEN, JP-R and MKF drafted the manuscript. AK, MY, EEN, JAH, CAF, KYE-M, PSN, PKW, JP-R and MKF all contributed to the final manuscript.

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Data availability

The data that support the findings of this study are openly available in the main text, supporting information, and stored in SRA under project ID PRJNA1051335.

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Supporting Information

Additional Supporting Information may be found online in the Supporting Information section at the end of the article.

Fig. S1 Atom% $^{13}\mathrm{C}$ in shoots, roots, sand mix, and soil mix in +AMF, -AMF, and $^{12}\mathrm{C}$ microcosms.

Methods S1 Detailed ¹³C calculations.

Table S1 Data used in statistical analyses.

Table S2 Data used to calculate how much carbon $(^{13}C \text{ and total } C)$ was transported by arbuscular mycorrhizal fungi into the noplant compartment in +AMF microcosms.

Table S3 Atom% ¹³C of hyphae from NanoSIMS measurements for the 37 samples discussed in the main text.

Table S4 Amplicon sequence variants that significantly increased or decreased in relative abundance in the soil mix (no-plant compartment) of +AMF microcosms.

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