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Critical needs to close monitoring gaps in pantropical wetland CH $_4^{\,}$ emissions

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Critical needs to close monitoring gaps in pan-tropical wetland $CH₄$ emissions

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Abstract

Global wetlands are the largest and most uncertain natural source of atmospheric methane (CH4). The FLUXNET-CH⁴ synthesis initiative has established a global network of flux tower infrastructure, offering valuable data products and fostering a dedicated community for the measurement and analysis of methane flux data. Existing studies using the FLUXNET-CH $_4$ Community Product v1.0 have provided invaluable insights into the drivers of ecosystemto-regional spatial patterns and daily-to-decadal temporal dynamics in temperate, boreal, and Arctic climate regions. However, as the wetland CH_4 monitoring network grows, there is a critical knowledge gap about where new monitoring infrastructure ought to be located to improve understanding of the global wetland CH⁴ budget. Here we address this gap with a spatial representativeness analysis at existing and hypothetical observation sites, using 16 process-based wetland biogeochemistry models and machine learning. We find that, in addition to eddy covariance monitoring sites, existing chamber sites are important complements, especially over high latitudes and the tropics. Furthermore, expanding the current monitoring network for wetland CH⁴ emissions should prioritize, first, tropical and second, sub-tropical semi-arid wetland regions. Considering those new hypothetical wetland sites from tropical and semi-arid climate zones could significantly improve global estimates of wetland $CH₄$ emissions and reduce bias by 79% (from 76 to 16 TgCH⁴ y *−*1), compared with using solely existing monitoring networks. Our study thus demonstrates an approach for long-term strategic expansion of flux observations.

1. Introduction

Methane (CH_4) is a highly potent greenhouse gas with a global warming potential for sustained ecosystem emissions 45 times greater than that of carbon dioxide $(CO₂)$ over a 100 year time span (Neubauer and Megonigal [2015\)](#page-10-0). Among natural ecosystems, freshwater inland wetlands function as hotspots of sustained $CH₄$ emissions due to strong microbial activity operating under seasonal or permanent anaerobic soil conditions (Bridgham *et al* [2013](#page-10-1)). At a global scale, wetlands contribute approximately 20%–30% of total CH_4 emissions. However, uncertainties remain large within and between modeling approaches. Using bottom-up models, where parameterization is informed by process-level understanding, annual wetland CH_4 emissions to the atmosphere have been estimated at 102–182 TgCH₄ y^{−1}. Meanwhile, using top-down inversion models constrained by observed atmospheric CH₄ concentrations, yearly emissions have been estimated at 159– 200 TgCH⁴ y *−*1 (Saunois *et al* [2020](#page-11-0)). More precise quantification of wetland $CH₄$ emissions is much needed, yet it presents a formidable challenge to current state-of-the-art modeling approaches, evident in these wide model ensemble ranges.

Surface monitoring networks designed to directly measure wetland CH⁴ emissions can reduce modeled global CH_4 emission uncertainties by (1) informing bottom-up model parameterization (Yuan *et al* [2022b](#page-11-1), Chinta *et al* [2024](#page-10-2)), (2) providing ecosystem scale observational benchmarks (Yuan *et al* [2022a,](#page-11-2) Chang *et al* [2023,](#page-10-3) Ito *et al* [2023](#page-10-4)), and (3) providing data constraints for top-down inversions (Saunois et al [2020\)](#page-11-0). The FLUXNET-CH₄ Synthesis Activity, launched in 2018, is already facilitating global wetland CH⁴ model improvements through the recent compilation and publication of the FLUXNET-CH⁴ Community Product v1.0; a global, standardized dataset of eddy covariance (EC) flux measurements from 81 sites worldwide (Knox *et al* [2019,](#page-10-5) [2021,](#page-10-6)

Delwiche *et al* [2021\)](#page-10-7). The EC method allows for the continuous acquisition of high-frequency time series data of wetland CH_4 emissions at the ecosystem scale and can be paired with concurrent biometeorological and biogeochemical drivers of CH⁴ emission, e.g., air and soil temperature, vegetation activity, latent heat flux, and soil moisture (Keenan *et al* [2019\)](#page-10-8). Meanwhile, surface chamber measurements are also an important measurement strategy across wetland ecosystems (Turetsky *et al* [2014,](#page-11-3) Bao *et al* [2021\)](#page-10-9). Chamber-based observations provide CH⁴ emission estimates for specific wetland patches, including microtopographic features and plant functional types, and thus provide an important spatially explicit complement to EC towers. Combining both flux tower and chamber measurements will lead to better spatial coverage and will improve understanding of the spatiotemporal dynamics of wetland $CH₄$ emissions across ecosystem-to-global scales (Yuan *et al* [2024\)](#page-11-4).

The FLUXNET-CH₄ monitoring network is relatively new and still underdeveloped in contrast to the ecosystem $CO₂$ exchange monitoring systems (e.g. AmeriFlux, FLUXNET) (Baldocchi [2003](#page-10-10)), especially when considering spatial coverage (Delwiche *et al* [2021](#page-10-7)). Consequently, global upscaling efforts using the current network of *in situ* measurements of wetland CH⁴ emissions to reconstruct or forecast the global CH⁴ budget introduces large uncertainties because the spatiotemporal coverage of tower and chamber flux data is limited. In particular, there are only a handful of flux observation sites in the regions ranging from 60*◦* S to 30*◦* N, which account for*>*60% of global wetland CH⁴ emissions and *>*70% of emissions uncertainties (McNicol *et al* [2023](#page-10-11)). The ongoing expansion of the *in situ* tower and chamber CH⁴ flux network is therefore vital for a global $CH₄$ emission monitoring system (Griffis *et al* [2020,](#page-10-12) Soosaar *et al* [2022\)](#page-11-5). Yet, a comprehensive understanding of how this network will ultimately contribute to process-level understanding and predictability

of global wetland CH⁴ emissions remains largely unexplored.

Our study aims to outline potential advantages for wetland CH_4 emission monitoring at representative locations. We first evaluate the global representativeness of the existing observational network. Then we assess the relative benefit of adding new sites, particularly in currently under-sampled locations hypothesized to be important for the global $CH₄$ emission budget. It is important to note that we focus on the spatial representativeness of the wetland CH_4 monitoring network, defined by how well CH⁴ flux measurements obtained from a network depict the CH_4 flux conditions across a larger regional or global domain (Hargrove *et al* [2003](#page-10-13), Chu *et al* [2021](#page-10-14)). Specifically, we attempt to address two fundamentally important questions: (*1) How much uncertainty in global wetland CH⁴ emission estimates results from insufficient monitoring of diverse wetlands? (2) How should we prioritize future wetland CH⁴ monitoring sites to best improve global emission estimates?* To address these questions, we developed a machinelearning approach based on an expanded observational dataset and the latest Global Carbon Project CH⁴ (GCP-CH4) bottom-up model ensembles (Ito *et al* [2023](#page-10-4), Zhang *et al* [2023\)](#page-11-6).

2. Methodology and data

Overall, our approach combines multi-model ensemble estimates of global wetland CH₄ emissions (section [2.1\)](#page-4-0), site locations from current EC and chamber measurements (section [2.2](#page-4-1)), and spatial representative analysis with machine learning approach (section [2.3\)](#page-4-2) to evaluate the representativeness of the existing monitoring network. In section [2.4](#page-5-0), we further quantify the potential benefit of hypothetical monitoring sites from different climate zones. Lastly, we examine the uncertainties associated with randomness of new site locations within each climate zone and the selection of GCP model ensembles (section [2.5](#page-5-1)).

2.1. GCP-CH⁴ global wetland CH⁴ emission

The global monthly estimates of wetland $CH₄$ emissions came from the latest GCP-CH₄ products based on a cohort of biogeochemical models. This product extended the previous 2000–2017 estimate (Saunois *et al* [2020\)](#page-11-0) to 2000–2020 (Ito *et al* [2023](#page-10-4)) and all models applied the same Wetland Area and Dynamics for Methane Modeling (WAD2M) (dynamic) global wetland extent estimates and climate forcing (Zhang *et al* 2021). Wetland CH₄ emissions were all gridded to 0.5*◦ ×* 0.5*◦* spatial resolution.

The GCP-CH₄ multi-model product provided 16 bottom-up model estimates of global wetland CH⁴ emissions. Although these models calculate CH_4 production, oxidation, and transport based on broadly similar wetland hydrological and biogeochemical

processes, the global estimates differed by up to a factor of two. Therefore, we calculated the multimodel ensemble mean for all 16 models and used it as our benchmark (figure $1(a)$ $1(a)$) to minimize potential bias due to individual models.

2.2. Wetland CH⁴ monitoring networks

The existing network for monitoring wetland CH⁴ emissions has two types of measurements: highfrequency measurements using the EC technique (Knox *et al* [2019\)](#page-10-5) and chamber measurements (Turetsky *et al* [2014](#page-11-3)). In this study, we considered 84 EC sites (Deshmukh *et al* [2020](#page-10-15), Delwiche *et al* [2021](#page-10-7), Roman *et al* [2021](#page-11-8)) and 96 chamber sites (table S1), for a total of 180 globally-distributed sites. The existing monitoring network (figure [1](#page-5-2)(b)) provides valuable insights into the diverse locations that possess the capacity for monitoring, although chamber measurements are not continuous in a high-frequency manner (e.g. 30 min data). We used the location information of the monitoring network (section [2.3](#page-4-2)) to sample corresponding gridcells (2000–2020 monthly time series) from the GCP-CH₄ models (figure [1](#page-5-2) blue arrow).

This work represents an idealized experiment aimed at forward-looking assessment to determine the optimal locations for EC tower sites. We acknowledge that the existing EC and chamber datasets lack continuous temporal coverage from 2000 to 2020. We focus on the spatial representativeness of the monitoring network rather than the temporal variability (seasonal, inter-annual) of CH_4 emissions. We assessed the accuracy of a global wetland $CH₄$ emission reconstruction, utilizing the machine learning model described in section [2.3.](#page-4-2) This evaluation was conducted with samples from corresponding gridcells at CH_4 monitoring sites (figure $1(b)$ $1(b)$).

2.3. Experiment design for representativeness analysis of existing sites

Here, we define the 'representativeness' of the wetland CH_4 monitoring sites by how well CH_4 emission measurements at those sites could represent the spatial variability of global wetland $CH₄$ emissions (Hargrove *et al* [2003,](#page-10-13) Chu *et al* [2021](#page-10-14)). Based on this definition, we designed a *sample-and-reconstruct* approach to evaluate the representativeness of the sampled sites.

First, we assembled a dataset from the GCP-CH⁴ model ensemble mean at existing monitoring sites, including air temperature, rainfall, pressure, wind speed, solar radiation, relative humidity, inundation fraction, longitude, latitude, and $CH₄$ emissions. Then, we trained (using 80% randomly sampled data) and tested (with the remaining 20%) an XGBoost (eXtreme Gradient Boosting) model (Chen *et al* [2015](#page-10-16)) to capture the relationship between environmental predictors and wetland $CH₄$ emissions. The XGBoost model is a widely used ensemble learning

algorithm. This algorithm employs a strategic optimization process to iteratively refine decision trees, thereby achieving good predictive precision and generalization across a diverse range of complex datasets (Osman *et al* [2021](#page-11-9), Guo *et al* [2023\)](#page-10-17).

Then, we extrapolated the trained model to all wetland gridcells in the GCP-CH₄ benchmark and calculated the global pattern of wetland $CH₄$ emissions. The reconstructed global $CH₄$ emission was compared with the GCP-CH₄ model benchmark (figure [1](#page-5-2) green arrow). With this approach, site representativeness could be confirmed if the sampled sites were sufficient to reconstruct the GCP-CH⁴ model benchmark (Shirley *et al* [2023](#page-11-10)).

2.4. Experiment design for strategic network expansion

Besides the existing monitoring network sites (section [2.2](#page-4-1)), we are also interested in seeking underrepresented locations for future network expansion. We evaluated the benefit of new monitoring sites by quantifying how adding $CH₄$ emission measurements would improve the reconstruction global CH⁴ emission estimate (figure $1(c)$ $1(c)$). Twenty randomly selected gridcells with a mean annual inundation fraction larger than 1% at each of 5 Köppen climate zones (tropical, semi-arid, temperate, continental, and polar zones (Kalvová *et al* [2003\)](#page-10-18)) were selected, and additional samples were drawn from GCP-CH⁴ models at those hypothetical new sites (figure [1](#page-5-2) purple arrow).

We conducted three sets of experiments with this dataset: (1) only at FLUXNET-CH₄ EC network sites (hereafter referred to as 'network1'); (2) at FLUXNET- CH⁴ EC and chamber sites (hereafter referred to as 'network2'); and (3) at FLUXNET-CH⁴ EC sites, chamber sites, and 20 hypothetical new sites. The value of adding a set of new sites could be quantified by the relative reduction in $CH₄$ emission bias, compared to the GCP-CH₄ model ensemble benchmark.

2.5. Uncertainty associated with hypothetical sites locations and model selection

Each climate zone covers many wetland gridcells, which may significantly differ in $CH₄$ emission magnitude and seasonality. To quantify the sampling uncertainty for the 20 hypothetical new sites within each climate zone, we randomly repeated the site selection 10 times and reported the associated uncertainty using the standard deviation of the reconstruction errors.

The GCP-CH₄ includes 16 bottom-up models, each of which exhibits different spatial and temporal patterns of wetland CH_4 emissions. In the main analysis (section [2.1\)](#page-4-0), we used the model ensemble mean to avoid the uncertainty associated with any individual model. However, each GCP models' performance vary dramatically when compared with FLUXNET-CH⁴ site observations (Chang *et al* [2023\)](#page-10-3). To determine whether our main conclusions depend on the choice of the multi-model ensemble mean, we

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also conducted the same analysis using the top three models selected by the FLUXNET-CH₄ dataset and the International Land Model Benchmarking package (Chang *et al* [2023](#page-10-3)).

3. Results and discussion

3.1. Is the current monitoring network sufficient for global wetland CH⁴ emission estimates?

Our model showed high performance when trained and tested at current EC sites (figure $S1(a)$) and chamber sites (figure $S1(b)$), with R^2 values for training and testing both *>*0.95, and mean absolute percentage errors of only 1%–6%. These results suggest that the trained XGBoost ML models robustly captured the relationships between environmental drivers (e.g. temperature, inundation) and wetland CH⁴ emissions, and were thus reliable tools for interpolation across regions that experience similar environmental conditions.

Extrapolating the XGBoost ML model to global wetlands, reconstructed $CH₄$ emissions with network1's sites generally overestimated wetland CH⁴ emissions across all climate zones (figure [2](#page-7-0) blue line), with the highest biases at tropical and high-latitude wetland ecosystems. The reconstruction with network1's sites estimated annual wetland $CH₄$ emissions to be 282 TgCH⁴ year*−*¹ , which is 70% higher than the GCP-CH₄ model benchmark of 165 TgCH⁴ year*−*¹ . Although the reconstruction using network2*′* sites still overestimated global wetland CH⁴ emissions (242 TgCH⁴ year*−*¹), significant improvements were achieved over tropical and highlatitude regions (figure [2](#page-7-0) orange line). For example, the regional reconstruction error was reduced from 16.9 TgCH⁴ year*−*¹ to 4.2 TgCH⁴ year*−*¹ over regions north of 60 N by considering chamber observational sites.

Although both models (trained with network1 vs. network2 sites) achieved similar high accuracy during training, the global extrapolations were substantially different between the two trained models (figure [2\)](#page-7-0). This result indicates that, on the one hand, the model can accurately capture relationships between environmental predictors and CH_4 emissions, but on the other hand, extrapolation was sensitive to the number and locations of monitoring sites. After considering current chamber sites as EC tower sites, however, relatively high errors still exist over tropical and tropical regions. In summary, even if long-term continuous data were available across the existing EC and chamber monitoring network, large errors would persist in global (e.g. 70% over-estimation of global emissions) and regional (e.g. 51% over-estimation in the tropics) estimates. Therefore, more sites are needed to reduce regional and global biases.

3.2. Where to prioritize new monitoring sites of wetland CH⁴ emissions?

The existing EC and chamber monitoring network has most of its sites in continental and polar Köppen climate zones (figure $1(b)$ $1(b)$). This incomplete spatial coverage is consistent with our results that a large portion of reconstruction errors occurred over tropical regions (30*◦* S–30*◦* N) (figure [2\)](#page-7-0). Specifically, the reconstruction with EC and chamber sites (network2) estimated 184 TgCH⁴ year*−*¹ emissions from tropical wetland, which 51% higher than the GCP-CH⁴ model benchmark of 122 TgCH⁴ year*−*¹ over this region). To quantify the benefit of adding new tropical sites, we randomly selected 10 sets of 20 monitoring sites over the tropical zone and trained and tested 10 new XGBoost models. The new models consistently had high accuracy over training and testing datasets with R^2 both $>$ 0.95 (figure S2). The newly reconstructed global wetland $CH₄$ emissions were substantially improved over the tropics (figure [3](#page-7-1) green line), and overestimated emissions only by 18% (figure [4](#page-8-0) green bar). The reconstruction was not sensitive to the random selection of tropical sites (figure [3](#page-7-1) green shaded area), partly because the hypothetical site locations were limited to GCP-CH₄ gridcells with *>*1% surface inundation. The reconstruction errors were most prominent between 30*◦* S and 10*◦* S. In summary, sampling new sites from tropical wetlands could reduce the reconstruction error from 76 to 30 TgCH⁴ year*−*¹ (absolute error).

Similarly, global reconstructions with 10 sets of 20 new sites from semi-arid, temperate, continental, and polar Köppen climate zones indicate that new sites over semi-arid regions reduced the reconstruction error from 10 to 5 TgCH⁴ year*−*¹ over southern hemisphere sub-tropical semi-arid regions (figure [3](#page-7-1) magenta line), while new sites in temperate, continental, and polar Köppen climate zones provided much smaller benefits. Overall, by comparing global reconstruction accuracy, we conclude that expanding the monitoring network for wetland $CH₄$ emissions should prioritize, first, tropical and second, subtropical semi-arid regions

Our sample-and-reconstruct approach might be sensitive to the spatial variation of wetland CH_4 emissions in our GCP-CH⁴ model benchmark (i.e. the multi-model ensemble mean of the 16 GCP-CH⁴ models). We note that $GCP-CH₄$ estimates are not real observations, and individually show deviations from observed emissions (Chang *et al* [2023](#page-10-3)). To investigate the robustness of our conclusions, we conducted an additional analysis that used only the 3 highest scoring GCP-CH₄ models based on an independent global benchmark analysis (Chang *et al* [2023\)](#page-10-3). The top 3 models (TRIPLEX, LPJ-MPI, and DLEM) were used to derive the global mean spatial pattern of CH_4 emissions by sampling the

Figure 2. Reconstructed global wetland CH₄ emissions using samples from network1 (eddy covariance sites only, blue line) or from network2 sites (eddy covariance and chamber sites, orange line). The shaded area around the line is the uncertainty range. The inserted panel shows the number of EC and chamber sites along latitudes.

Köppen climate zones (see distribution in figure [1\(](#page-5-2)c)).

models at locations of existing and hypothetical monitoring networks (same as the previous analyses). Although the benchmark has changed, our sampleand-reconstruct approach suggested similar results: (1) observations from the chamber site locations are an important supplement to the EC monitoring network; (2) tropical wetlands are not well represented in the existing monitoring network, thus future development of tropical sites is critical to reducing biases in global wetland CH_4 emissions budgets; and (3) more observational sites are needed across sub-tropical

semi-arid Köppen climate zones (30*◦* S–10*◦* S) where wetlands are sparsely distributed (figures S3 and S4).

3.3. Sensitivity to the number new monitoring sites Given our conclusion that prioritizing monitoring sites over tropical regions can substantially reduce CH⁴ emission uncertainties, we next evaluated how much the global reconstruction bias could be reduced by increasing the number of hypothetical sites. We therefore conducted additional model experiments that randomly sampled 50, 100, 150, and 200

sites across the tropics. Each random sampling was repeated 10 times to evaluate the impacts of random selection. As the number of monitoring sites increased from 20 to 200, the global reconstruction error declined from 30 to 23 TgCH⁴ year*−*¹ (figure [4](#page-8-0)).

Furthermore, sampling sites from tropical Africa, America, and Asia individually did not significantly change the reconstruction. The results suggested that sampling sites across the entire tropical region is to be preferred (figure [5](#page-8-1) green bar). Sampling sites from tropical Africa are better than sampling sites from tropical America or tropical Asia, because of the relatively lower reconstruction error (figure [5](#page-8-1) magenta bar versus purple and cyan bars).

3.4. Towards a representative CH⁴ monitoring network

We acknowledge that the establishment of new monitoring sites is a complex problem, *e.g.* should a network preferentially monitor high emission (hotspots) sites (Gloor *et al* [2021,](#page-10-19) France *et al* [2022,](#page-10-20) Rößger *et al* [2022,](#page-11-11) Shaw *et al* [2022,](#page-11-12) Soosaar *et al* [2022\)](#page-11-5) or

better cover wetlands that span the emissions range. For large-scale analysis, spatial representativeness is an important metric (Yang *et al* [2008\)](#page-11-13). For analyses of $EC CO₂$ emissions, it has been reported that the historical placement of EC towers in highly productive landscapes may bias the global upscaling effort (Ran *et al* [2016\)](#page-11-14). Previous analysis using the FLUXNET-CH⁴ dataset has highlighted the lack of spatial coverage over tropical wetlands (Delwiche *et al* [2021\)](#page-10-7). Such gaps in representativeness have been identified in both geographic space (based on longitude and latitude) and bioclimatic space (based on temperature, latent heat, and vegetation index). In addition to the existing tropical sites in the FLUXNET-CH₄ monitoring network, we considered chamber site locations, thus greatly enhancing the spatial coverage over tropical wetlands. However, even with these additional sites, our analysis still showed significant biases (figure [2](#page-7-0)), indicating the need for more EC sites at tropical and semi-arid wetland sites. Our best reconstruction used existing sites plus 20 hypothetical tropical and semi-arid wetland sites. The resultant global reconstruction of the GCP-CH⁴ benchmark, has an absolute bias of only 16 TgCH⁴ y *−*1 , while the global

reconstruction using solely existing sites has an absolute bias of 76 TgCH⁴ y *−*1 .

Although a previous study (Delwiche *et al* [2021](#page-10-7)) using dissimilarity metrics at FLUXNET-CH $_4$ sites suggested relatively high representativeness over arctic wetlands, our analysis quantitatively showed the high value of monitoring the wetland-dense regions (i.e. Siberian lowland and Hudson Bay lowland), that are currently represented in the chamber network, but not the FLUXNET-CH₄ network.

Our sample-and-reconstruct approach with hypothetical new sample sites was able to address the question of where to prioritize new sites, from the perspective of global CH₄ emissions estimates, modeling, and analysis. This work represents an idealized experiment aimed at determining the optimal locations for placing EC tower sites. We found that future investments in monitoring wetlands over tropical and sub-tropical semi-arid Köppen climate zones are recommended (figure [3](#page-7-1)). We also recommend enhanced monitoring of the fine-scale wetland hotspots across temperate, continental, and polar zones. We acknowledge that our analysis is subject to the spatial resolution and accuracy of the GCP-CH⁴ model benchmark (0.5*◦ ×* 0.5*◦* , *∼*50 Km), and thus insensitive to fine-scale heterogeneity of real-world wetlands. Previous efforts with a similar sample-andreconstruct approach for $CO₂$ emissions have shown the importance of finer-scale heterogeneity in the representativeness analysis and choice of new monitoring sites (Shirley *et al* [2023](#page-11-10)). Further, our analysis did not consider seasonal and inter-annual variability in $CH₄$ emissions and its drivers, but the methods we describe here are well-suited to explore those issues in future work.

This study used 0.5*◦ ×* 0.5*◦* resolution gridded products for wetland $CH₄$ emissions, which means that fine-scale processes and interactions influencing CH⁴ emissions were not considered. We acknowledge that sub-grid heterogeneity is an important factor in representativeness analysis, though accounting for such heterogeneity is challenging at a global scale. This challenge arises because: (1) high-resolution gridded products for wetland $CH₄$ emissions do not exist; and (2) within the current monitoring network, only a limited number of 0.5*◦ ×* 0.5*◦* gridcells are covered by multiple monitoring sites, making it difficult to robustly estimate sub-grid heterogeneity.

We conservatively selected two gridcells with the largest number of monitoring sites for the EC and chamber networks, respectively (figure S5). The gridcell at Sacramento–San Joaquin Delta, USA, is covered by 9 EC sites, while the gridcell at Stordalen, Sweden, is covered by 17 chamber sites. The derived sub-grid heterogeneity in these gridcells is relatively similar, with coefficients of variation (CV) of 72% and 68%, respectively $(CV = standard deviation of$ annual CH₄ emission divided by mean annual CH₄ emission). However, all other 0.5*◦ ×* 0.5*◦* gridcells contain fewer than 5 EC sites or 10 chamber sites, making it difficult to generalize this sub-grid heterogeneity to global wetlands.

Although applying sub-grid heterogeneity to global estimates is challenging, we argue that our representativeness analysis framework (figure [1](#page-5-2)) could be seamlessly adapted to a finer resolution when highresolution gridded products become available in the future. Ideally, using a product at the hundred-meter scale for wetland CH_4 emissions, combined with EC site data, would significantly improve the analysis. Therefore, a high-resolution global product of wetland CH⁴ emissions is urgently needed to improve representativeness analysis and guide the strategic deployment of new wetland monitoring sites.

4. Conclusions

In this study, we assessed the representativeness of the global wetland CH_4 emissions monitoring network using a machine learning-based sample-andreconstruct approach. We utilized a combination of the GCP-CH⁴ model simulations, which provide estimates of wetland $CH₄$ emissions from 2000 to 2020, and existing wetland $CH₄$ monitoring locations including both EC and chamber sites. Our results revealed that, while our machine learning model effectively captured predictive relationships at existing sites, extrapolation to global wetlands often led to a large overestimation of $CH₄$ emissions, particularly in tropical and arctic regions. We conclude that the existing monitoring network lacks the necessary spatial representativeness to reliably construct global wetland $CH₄$ emission estimates. Considering new monitoring sites, particularly over tropical and semi-arid climate zones, could substantially improve the network's representativeness and reduce global wetland CH_4 emission by 79% (from 76 to 16 TgCH₄ y⁻¹)

Data availability statement

All data that support the findings of this study are included within the article and the supplementary file: GCP-CH₄ bottom-up estimates of wetland $CH₄$ emission are archived in the International Carbon Observation System: [https://doi.org/10.18160/gcp](https://doi.org/10.18160/gcp-ch4-2019)[ch4-2019.](https://doi.org/10.18160/gcp-ch4-2019)

FLUXNET-CH4 data are available at [https://](https://fluxnet.org/data/fluxnet-ch4-community-product/) [fluxnet.org/data/fluxnet-ch4-community-product/,](https://fluxnet.org/data/fluxnet-ch4-community-product/) <https://doi.org/10.5281/zenodo.4672601>. Chamber site locations are documented in supplementary file.

All data that support the findings of this study are included within the article (and any supplementary files).

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