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Long-term Exposure to Air Pollution and Markers of Inflammation, Coagulation, and Endothelial Activation:

A Repeat-measures Analysis in the Multi-Ethnic Study of Atherosclerosis (MESA)

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Abstract

Background—Air pollution is associated with cardiovascular disease, and systemic inflammation may mediate this effect. We assessed associations between long- and short-term concentrations of air pollution and markers of inflammation, coagulation, and endothelial activation.

Methods—We studied participants from the Multi-Ethnic Study of Atherosclerosis from 2000 to 2012 with repeat measures of serum C-reactive protein (CRP), interleukin-6 (IL-6), fibrinogen, D-dimer, soluble E-selectin, and soluble Intercellular Adhesion Molecule-1. Annual average concentrations of ambient fine particulate matter (PM_{2.5}), individual-level ambient PM_{2.5} (integrating indoor concentrations and time–location data), oxides of nitrogen (NO_x), nitrogen dioxide (NO₂), and black carbon were evaluated. Short-term concentrations of PM_{2.5} reflected the day of blood draw, day prior, and averages of prior 2-, 3-, 4-, and 5-day periods. Random-effects models were used for long-term exposures and fixed effects for short-term exposures. The sample size was between 9,000 and 10,000 observations for CRP, IL-6, fibrinogen, and D-dimer; approximately 2,100 for E-selectin; and 3,300 for soluble Intercellular Adhesion Molecule-1.

Results—After controlling for confounders, $5 \mu g/m^3$ increase in long-term ambient PM_{2.5} was associated with 6% higher IL-6 (95% confidence interval = 2%, 9%), and 40 parts per billion increase in long-term NO_x was associated with 7% (95% confidence interval = 2%, 13%) higher

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level of D-dimer. $PM_{2.5}$ measured at day of blood draw was associated with CRP, fibrinogen, and E-selectin. There were no other positive associations between blood markers and short- or long-term air pollution.

Conclusions—These data are consistent with the hypothesis that long-term exposure to air pollution is related to some markers of inflammation and fibrinolysis.

Inflammation plays an important role in the initiation and development of atherosclerosis and in precipitation of cardiovascular events. Animal, epidemiologic, and controlled exposure studies have provided evidence that air pollution causes an inflammatory response in the vasculature, which stimulates the process of atherosclerosis. Air pollution-induced inflammation can occur through autonomic nervous system imbalance (ie, sympathetic nervous system activation and/or parasympathetic nervous system withdrawal) or through localized inflammation in the lungs that spills over into the bloodstream. However, the exact biological mechanisms are still unclear.

The role of coagulation in cardiovascular disease has also been well established. 3,4 Moreover, the relation between inflammatory markers and the coagulation cascade in cardiovascular disease 5 makes these markers relevant to the study of air pollution health effects. However, the evidence supporting the hypothesis that air pollution increases concentrations of coagulation-related blood markers is mixed. $^{6-10}$

The endothelial cell layer of blood vessels is dynamic, changing with factors such as age, altitude, exercise and diet, 11 smoking, hypertension, 12,13 and various disease states. 14,15 The endothelium is also actively involved in regulating blood coagulation and inflammatory response. $^{16-18}$ A negative association between long-term, but not short-term, concentrations of particulate matter <2.5 μ m in aerodynamic diameter (PM_{2.5}) and flow-mediated dilation—a marker of endothelial function—has been documented. 19

Much of the existing literature has focused on associations between recent air pollution exposure and markers of inflammation and coagulation. 1,2 A review of the association between C-reactive protein (CRP) and particulate matter concluded that epidemiologic evidence is inconsistent at best. 2 The potential for air pollution to contribute to long-term inflammatory responses may be more relevant to the development of cardiovascular disease. The studies that have explored the association between long-term air pollution and blood markers have focused on markers of inflammation and coagulation and not on markers of endothelial activation. 20–23

This study examined the association between long-term pollutant concentrations and several blood markers that may be biologically relevant to the mechanism by which air pollution exposure results in cardiovascular disease. The blood markers of interest are CRP, interleukin-6 (IL-6), fibrinogen and D-dimer, and markers of endothelial activation soluble E-selectin and soluble intercellular adhesion molecule-1 (sICAM-1). As a secondary aim, we examined the association between short-term $PM_{2.5}$ concentrations and blood markers.

METHODS

The Multi-Ethnic Study of Atherosclerosis (MESA) is a longitudinal epidemiologic study designed to examine the progression of subclinical and clinical cardiovascular disease among adults free from such disease at baseline. Prom July 2000 to August 2002 (baseline examination), the study recruited 6,814 white, African-American, Hispanic, and Chinese men and women age 45 to 84 years from six US communities (Baltimore, MD; Chicago, IL; Winston-Salem, NC; Los Angeles, CA; New York, NY; and St. Paul, MN). Four follow-up exams were conducted between 2002 and 2012. Exam 2 was held between Fall 2002 and Winter 2004, the third examination between Spring 2004 and Fall 2005, the fourth between Fall 2005 and Spring 2007, and the fifth exam was from Spring 2010 to Winter 2012. All examinations included a blood draw, anthropometric measurements, and the collection of questionnaire data. Institutional review board approval was granted at each study site, and written informed consent was obtained from participants.

Blood Markers

The blood markers of interest were measured at baseline for most MESA participants, and in subsets of individuals at follow-up through three ancillary studies. The MESA abdominal body composition study, conducted over exams 2 and 3, assessed CRP, fibringen, and IL-6 in 1970 existing MESA participants. The MESA and Air Pollution (MESA Air) assessed CRP, D-dimer, fibrinogen, E-selectin, and sICAM-1 in approximately 715 participants at exams 4 and 5. Two-hundred fifty-seven of these participants were newly recruited into the MESA cohort during follow-up exam 4 (September 2005 to May 2007) and followed through exam 5. These new participants were recruited from the Los Angeles and New York areas to increase the range of air-pollution exposures among the study population.²⁵ Finally, the MESA Stress examined D-dimer and IL-6 in 1,002 participants over exams 3 and 4. A follow-up study to the MESA Stress was conducted during exam 5 that measured CRP, Ddimer, fibrinogen, and IL-6 in about 1,300 participants. Because these ancillary studies did not collect each blood marker at each visit, our final sample of participants are missing a Ddimer measurement from exam 2 and E-selectin and sICAM-1 measurements from exams 2 and 3. In addition, only a subset of baseline participants had measurements of E-selectin and sICAM-1. Table 1 provides sample sizes by examination for each blood marker. Participants in our study had up to four measures of biomarkers collected between 2000 and 2012 (mean 1.6 observations).

At all exams, blood was drawn after 12 hours of fasting. All plasma assays were performed at the Laboratory for Clinical Biochemistry Research (University of Vermont, Burlington, VT) with test characteristics as previously described. High-sensitivity CRP and fibrinogen antigen assays were performed using BNII nephelometers (N High Sensitivity CRP and N Antiserum to Human Fibrinogen, respectively; Dade Behring Inc., Deerfield, IL). Both IL-6 and sICAM-1 were measured by ultrasensitive enzyme-linked immunosorbent assay (Quantikine HS Human IL-6 Immunoassay and Parameter Human sICAM-1 Immunoassay, respectively; R&D Systems, Minneapolis, MN, respectively). Fibrin fragment D-dimer was measured using an immuno-turbidimetric assay (Liatest D-DI; Diagnostica Stago, Parsippany, NJ). E-selectin was measured using a high-sensitivity quantitative sandwich

enzyme (Parameter Human sE-selectin Immunoassay; R&D Systems, Minneapolis, MN). With the exception of E-selectin, all analytes have been previously evaluated for within- and between-individual variation.^{27–29}

Air Pollution Concentrations

Several ambient pollutant concentrations were analyzed, including: $PM_{2.5}$, ambient-origin individual-level $PM_{2.5}$ (integrating infiltration into the home and time-location data), oxides of nitrogen (NO_x), nitrogen dioxide (NO_2), and light absorption coefficient, which represents black carbon (BC). NO_x , NO_2 , and BC are considered markers of traffic-related air pollution (TRAP). All pollution concentrations were predicted at the participant's residential address. Participants who moved during the year had data incorporated for their new address when possible.

Concentrations of pollutants were calculated from a spatiotemporal prediction model which used a unified modeling approach for all pollutants.^{30,31} To derive these predictions, we used data from several sources: regulatory monitoring stations from the US Environmental Protection Agency's Air Quality System (AQS), monitors deployed by MESA Air at fixed sites throughout the study area, outdoor monitors at participant's homes, and monitors placed to better capture roadway concentration gradients, 32 with the exception of the BC model, which did not utilize AQS monitoring data. These different sources of monitoring data provided detailed spatial and temporal data to better characterize within-site variability. Further investigation revealed no systematic differences between the different monitors used to derive our predictions. ^{32,33} Models used geographic covariates such as roadway density, land use, and outputs from dispersion models. Over 150 geographic covariates were available for use in these models. To select the most relevant ones, we employed a partial least squares approach, which is similar to principal-components analysis. Smoothed time trends were extracted from AQS and fixed site monitors to assess temporal trends. Models were built separately for each pollutant and each city in the study. Cross validation was used to select the best final model. In eTable 1 (http://links.lww.com/EDE/A886), we report the site-specific leave-one-out cross-validated R^2 based on fit to the 1-1 line and R^2 based on fit to the regression line. The cross-validated R^2 based on the 1-1 line are the most relevant measure of prediction accuracy, but they are generally lower than those based on the regression line, which we also report to facilitate comparison with research that uses R^2 based regression line.³⁰

 $PM_{2.5}$, NO_x , and NO_2 pollutant concentrations were time varying and estimated the annual average concentration for the year before the participant's date of blood draw. The individual level $PM_{2.5}$ predictions were of ambient origin (ie, they do not include indoor sources). They integrate data from time–activity questionnaires kept by the participant to assess the amount of time spent indoors versus outdoors by season during the course of a typical weekday and weekend. The season-specific residential infiltration fraction is also incorporated into this individual-level metric and is derived from questionnaires about characteristics of the participant's home as well as indoor–outdoor residential pollution sampling in a subset of homes (5% from each study site). 34 BC was not time varying because predictions were only available from 2006 to 2008.

Short-term concentrations of PM_{2.5} reflected several averaging periods before the date of blood draw, including: day of blood draw, prior day and moving average of prior 2, 3, 4, and 5 day periods. Because the dominant variability in short-term PM_{2.5} concentrations is temporal rather than spatial, short-term PM_{2.5} concentrations are based on daily observations from a single central site monitor in each region. We did not assess short-term exposures for NO_x, NO₂, and BC because daily central site measurements were not available or did not adequately capture the spatio-temporal variation. To control for temporal confounding, the short-term PM_{2.5} concentrations were pre-adjusted using splines for calendar time (12 degrees of freedom [df]/year), temperature (6 df/year), and relative humidity (6 df/year), and an indicator variable for day of the week. This approach has been demonstrated to be an effective alternative to standard semi-parametric adjustment that can increase precision for epidemiologic models assessing health effects of short-term pollutant exposures.³¹

Covariates

The following covariates were included in the analysis: age, race/ethnicity (white, black, Hispanic, or Chinese), gender, study exam (one through five), site, individual socioeconomic status (SES) defined as education (less than or equal to high school, some college, greater than or equal to college graduate), income (specified continuously as permanent income, which is the average of all reported income over all exams), and employment status (working outside the home or not), neighborhood SES (summary index derived from factor analysis), 35 smoking status (current, former, never smoker), secondhand smoke exposure (yes or no), current alcohol consumption (yes or no), body mass index (weight in kilograms/height in meters squared), waist-hip ratio (waist circumference/hip circumference), diabetes (normal, borderline, or treated/untreated as defined by the 2003 American Diabetes Association fasting blood glucose criteria algorithm), ³⁶ hypertension (yes/no as defined by the criteria in the sixth report of the Joint National Committee on Prevention, Detection, Evaluation and Treatment of High Blood Pressure),³⁷ and use of any one of the following anti-inflammatory medications: aspirin, NSAIDs or oral antiinflammatory medications, hormone replacement therapy for women, and lipid-lowering medications. For the sICAM-1 models, we included the single nucleotide polymorphism RS4591 (specified as dichotomous for those with and without the T-allele), which may affect the assays ability to accurately detect circulating levels of sICAM-1.³⁸

In addition, we included city-specific trends in calendar time, modeled with splines that use 4 df for each year the blood marker was collected; therefore CRP, IL-6, and fibrinogen models included 48 df (4 df \times 12 years), D-dimer 40 df (4 df \times 10 years), and E-selectin and sICAM-1 20 df (4 df \times 5 years). In MESA, there is clustering of participants within neighborhood and participants from the same neighborhoods came in for examinations at about the same time, resulting in an association between time and long-term air pollution concentrations. Furthermore, there is a documented seasonality in several of the blood markers under study (CRP, IL-6, fibrinogen, and sICAM-1), $^{39-41}$ thus necessitating the adjustment for calendar time.

Analysis

The outcomes, main exposure, and most covariates in the analytic models were time varying. A few covariates were fixed or treated as such, including: race/ethnicity, gender, education, neighborhood SES, and income. CRP, IL-6, and D-dimer were log-transformed for analysis. We excluded some extreme outlying values for each blood marker: CRP 30 mg/liter (n = 76), IL-6 10 pg/ml (n = 64), D-dimer 18 µg/ml (n = 12), fibrinogen 1,000 mg/dl (n = 6), E-selectin > 132 ng/ml (n = 27), and sICAM-1 > 537 ng/ml (n = 44). Outliers were identified by plotting the distribution of the outcomes. Both log-transformation and exclusion of outliers helped reduce excessive skewness and kurtosis of the outcomes.

For long-term air pollution exposures, we used hierarchical models with a person-level random intercept to account for within-person clustering of blood markers. All other variables were treated as fixed. For short-term exposures, we used fixed effects models specifying person as the fixed effect to handle within-person variability of the outcome. For both long- and short-term exposures, we evaluated three models, each with incrementally more covariates. Model 1 was adjusted for age, race/ethnicity, gender, site, and exam. Model 2 additionally included income, education, employment, neighborhood SES, health behaviors (smoking, second-hand smoke exposure and alcohol consumption), cardiovascular disease risk factors (BMI, waist-hip ratio, and physical activity) and splines for calendar time and was considered our primary model. Model 3 adjusted for variables from model 2 as well as diabetes, hypertension, and anti-inflammatory medications, which may lie on the causal pathway between air pollution and the blood markers of interest. In addition, included in the online supplement (http://links.lww.com/EDE/A886) is model 2a, which includes all the variables in model 1 (described above) as well as the SES variables (income, education, employment, and neighborhood SES). This model staging approach allows readers to ascertain which set of confounders influence parameter estimates the most. Parameter estimates were reported as 5 µg/m³ increase in both ambient and individual PM_{2.5}, 40 parts per billion (ppb) increase in NO_x , 17 ppb increase in NO_2 , and 0.7 10^{-6} m⁻¹ increase in BC. These values are not interquartile ranges of the pollutants but are close to the interquartile range. Because interquartile ranges depend on site and time frame, these values were chosen to simplify interpretation of results. Short-term models were similar to long-term but did not include splines for calendar time because the short-term exposures were pre-adjusted to control for temporal confounding.³¹ When the associations between long-term concentrations of air pollution and blood markers were statistically significant, we further included short-term concentrations of PM_{2.5} in the analysis and conducted stratified analysis to assess effect modification by key covariates. In sensitivity analysis, we employed timevarying nearest monitor predictions and time-invariant predictions calculated for each participant's baseline home address from Jan 1 to Dec 30, 2000 for all pollutants using the same model staging approach described above.

RESULTS

The maximum number of observations and participants for long-term models is provided in Table 1. For short-term models, each participant was required to have at least two observations, resulting in fewer observations for each pollutant-blood marker combination

(eTable 2; http://links.lww.com/EDE/A886). Observations were excluded because of missing data on exposures, outcomes, or covariates and because of outliers for blood markers. The number of participants missing air pollution concentrations varied by pollutant resulting in different final sample sizes for each pollutant—blood marker combination.

The study population was on average 62 years old, 53% female and 27% African-American, 22% Hispanic, 11% Chinese and 39% white (Tables 2 and 3). Correlation coefficients between long-term pollutants were highest for NO_2 and NO_x ($R^2 = 0.93$) and lowest between ambient $PM_{2.5}$ and BC ($R^2 = 0.50$; eTable 3; http://links.lww.com/EDE/A886). The distribution of site-specific long-term annual average pollutant concentrations are shown in eFigure1 (http://links.lww.com/EDE/A886). The 4- and 5-day average short-term air pollution concentrations were highly correlated ($R^2 = 0.97$), whereas the day of blood draw and 5-day average were only moderately correlated ($R^2 = 0.47$; eTable 4; http://links.lww.com/EDE/A886).

For most blood markers, means across quartiles of air pollution concentrations did not vary (Table 4). Means for several of the blood markers for quartiles of BC appeared to increase monotonically, ie, the highest quartile of BC had the highest values of the blood markers. The pattern was similar for IL-6 and sICAM-1 for quartiles of NO_x and sICAM-1 and NO_2 . The means of D-dimer and quartiles of $PM_{2.5}$, however, were higher for the lowest quartile (0.36 µg/ml) but lower for the highest quartile of $PM_{2.5}$ (0.29 µg/ml).

After controlling for confounders in model 2, a 5 μ g/m³ increase in ambient and individual PM_{2.5} was associated with higher levels of IL-6 (6% [95% confidence interval {CI} = 2, 9] and 4% [95% CI = 1, 8], respectively). The parameter estimates for BC, NO_x, and NO₂ were substantially attenuated with the addition of confounders (Figure 1). The SES covariates were responsible for most of the confounding in these models (eTable 5; http://links.lww.com/EDE/A886). A 40 ppb increase in NO_x was associated with 7% higher D-dimer level (95% CI = 2, 13; model 2, Figure 2). The association with NO₂ was similar but attenuated. There was no association between D-dimer and ambient PM_{2.5}, individual PM_{2.5} or BC. The inclusion of additional potential mediators (model 3) had little impact on the point estimates for the associations shown in Figures 1 and 2. We evaluated effect modification for IL-6-ambient PM_{2.5} and D-dimer-NO_x associations (Table 5). In both cases, CIs overlapped between groups (eg, men and women), but the magnitude of parameter estimates was larger for older persons, those not working outside the home, smokers, and hypertensives. In the D-dimer-NO_x association, estimates were larger for women, obese, and diabetic participants.

CRP, E-selectin, and sICAM-1 showed little association with any of the long-term air pollutants (Table 6). The data suggested a negative association between fibrinogen and several air pollutant concentrations, whereas higher pollutant levels were associated with lower levels of fibrinogen (eg, 17 ppb increase in NO_2 was associated with 13.1 mg/dl decrease in fibrinogen (95% CI = -19.82, -6.37). We performed sensitivity analysis using nearest monitor $PM_{2.5}$ data and year 2000 estimates and found largely similar results with improved precision for the time-varying exposures (eTable 6; http://links.lww.com/EDE/A886).

A 5 μ g/ml increase in PM_{2.5} on the day of blood draw was associated with 0.60 ng/ml increase in E-selectin (95% CI = 0.06, 1.14) and suggestive of a positive association with CRP (1% difference, 95% CI = 0, 3) and fibrinogen (1.16 mg/dl, 95% CI = -0.28, 2.61; model 2, Table 7). However, the association between D-dimer and sICAM-1 and the 3-, 4-, and 5-day lagged PM_{2.5} were suggestive of an inverse relation. Associations between other blood markers and short-term PM_{2.5} were null. Finally, including short-term PM_{2.5} exposures in the long-term associations between PM_{2.5} and IL-6, and between NO_x and D-dimer, did not substantially alter point estimates (eFigure 2; http://links.lww.com/EDE/A886).

DISCUSSION

We found evidence of a positive association between long-term $PM_{2.5}$ and a marker of inflammation (IL-6), and between NO_x (a TRAP) and a marker of fibrinolysis (D-dimer). Based on previous research, the magnitude of effects is likely in the range of what is to be expected for a typical long-term ambient exposure.²² Furthermore, the findings were not consistent across all the available markers of inflammation and coagulation or across all TRAPs, thus we approach our findings with some caution. The research on short-term air pollution exposure and blood markers has commonly found inconsistencies across categories of blood markers (eg, findings were robust for sICAM-1 but null for e-Selectin, two markers of endothelial activation).¹⁰ There was no association among CRP, fibrinogen, e-Selectin, and sICAM-1 and long-term concentrations of air pollution.

In assessing effect modification, we found suggestive evidence that older individuals, smokers, and participants with hypertension experienced larger increases in IL-6 and D-dimer compared with younger, non-smoking, and normotensive participants. This finding is consistent with some studies on short-term air pollution and inflammation and coagulation ^{9,42,43} and with our general understanding of which populations are more susceptible to the health effects of air pollution. ¹

To put our results in context, it is useful to compare the magnitude of effects we observed for air pollutants with those observed for other common risk factors, such as smoking and gender. A 5 μ g/m³ increase in PM_{2.5} corresponds to about 0.081 pg/ml higher IL-6 and a 40 ppb increase in NO_x to about 0.018 μ g/ml higher D-dimer. In our study, current smokers had about 0.32 pg/ml increase in IL-6 (22%) compared with those who never smoked and women had about 0.044 μ g /ml increase in D-dimer (17%) compared with men. These are about four and two times higher, respectively, than the estimated effects of long-term concentrations of air pollution.

Two previous studies found a positive association between long-term air pollution and IL-6 and NO_2 (a marker of TRAP). 20,23 In our study, however, IL-6 was positively associated with $PM_{2.5}$. IL-6 is an acute phase reactant elaborated by T-cells and macrophages, so our result suggests an effect of air pollution early in the inflammatory process. Similar to Forbes et al, 21 we did not find an association with the other inflammatory marker, CRP; a German study did find a positive association between CRP and $PM_{2.5}$, but only among men. 22

Fibrinogen, an acute phase reactant and coagulation factor, is to our knowledge the only marker involved in the coagulation cascade that has been previously studied with respect to long-term air pollution. A positive association between fibrinogen and long-term $PM_{2.5}$ was found in German men but not in an English or a Swedish cohort^{21–23}; our study replicates the null findings of these two studies. We also found a null association between long-term air pollutants and E-selectin and sICAM-1, neither of which has been evaluated in existing research.

The association between long-term air pollution and D-dimer has not been previously evaluated, thus our findings of a positive association between NO_x and D-dimer are, we believe, novel. D-dimer is a fibrin degradation product and is produced as a result of the breakdown of fibrin clots. Epidemiologic and controlled exposure studies of short-term D-dimer and air pollutants have been largely null, 6.7,10,44 thus our results suggesting long-term air pollution has a role to play in fibrinolysis require additional confirmation.

We found some evidence of positive associations between day of blood draw and CRP, fibrinogen, and E-selectin. The evidence in this area is mixed: some studies have found similar positive associations for acute air pollution and CRP, ^{10,45} fibrinogen, ^{44,46} and E-selectin, ⁴⁷ whereas several others have found null associations. ² A study using the MESA cohort at baseline examined PM_{2.5} 1 day, 2 days, and 7 days before blood draw and found null effects for both CRP and IL-6. ⁴⁸ It is unclear why associations between 3-, 4-, and 5-day averages and D-dimer and sICAM-1 were negative. The use of fixed effects models for the short-term exposures leveraged the within-person variability in blood markers and controlled for unmeasured confounders better than a random-effects model. The random effects model (eTable 7; http://links.lww.com/EDE/A886) gave similar results, except there was no association between day of blood draw and fibrinogen and there was a suggestion of a positive association between E-selectin and day before blood draw. The associations between 3-, 4-, and 5-day averages and D-dimer and sICAM-1 were negative in both models.

Consistent with Hoffmann et al, 22 our study that found positive long-term associations between PM_{2.5} and IL-6 and between NO_x and D-dimer were not greatly impacted by inclusion of acute PM_{2.5}. We cannot, however, be certain that this association is free from confounding by short-term air pollution because we did not assess acute NO_x exposures in the NO_x–D-dimer association and our acute exposures are not spatially resolved.

Our approach to air pollution exposure assessment advances the field in two ways. First, individual PM_{2.5} prediction incorporates participant- and season-specific information about time participants spent indoors versus outdoors and residential infiltration efficiencies.³⁴ To the best of our knowledge, this is the first study to use predictions that more realistically represent concentrations of ambient-origin PM_{2.5}. Comparing the results of ambient to individual PM_{2.5}, we see point estimates closer to the null and slight improvements in precision that may represent a reduction in one dimension of measurement error. Second, our use of pre-adjusted acute exposure is a novel approach which can improve efficiency while effectively minimizing concerns over statistical power.³¹ It should be noted that black carbon predictions relied solely on study-specific monitors (central monitoring data is

unavailable), and in turn parameter estimates for BC from health effects models were less precise than those for other air pollutants (Figures 1 and 2, Table 6).

Sensitivity analysis using nearest-monitor PM_{2.5} and year 2000 estimates did not change the conclusions of our study (eTable 6; http://links.lww.com/EDE/A886). A few point estimates changed signs, but given the wide confidence intervals containing the null, the study conclusions remain the same.

The strengths of our study included improvements in air pollution exposure assessment, the large number of measurements, the use of repeat blood measures performed at a central laboratory, and a well-characterized cohort that allowed good control for confounding. The study of some biological markers can be challenging because of inadequate reproducibility of assays. Each one of the biomarkers used in this study is highly reproducible and if large effects had existed, we likely would have detected them.

Our study had a few limitations. Given the relative inconsistency of our results and the multiple comparisons made, statistically significant results may be due to chance alone. However, our findings were in line with previous research. In addition, because NO_2 is not measured directly (it is the difference between concentrations of NO and NO_x) and because there are fewer AQS sites for NO compared with NO_x , precision of our NO_2 estimates is slightly worse than $PM_{2.5}$ and NO_x . Also, the MESA population is healthier than the general public: all participants were free of cardiovascular disease at baseline and those who remained in the cohort over time reflect a healthy participant bias. Several studies have shown stronger effects of air pollution on blood markers in populations with diabetes, hypertension, or heart disease. 9,10,49 Finally, generalizability of our study is limited to older adults residing in mostly urban areas.

Overall, we found evidence that long-term exposure to air pollution was associated with some markers of inflammation and fibrinolysis, and short-term exposure was associated with some markers of inflammation, coagulation, and endothelial activation.

Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

Acknowledgments

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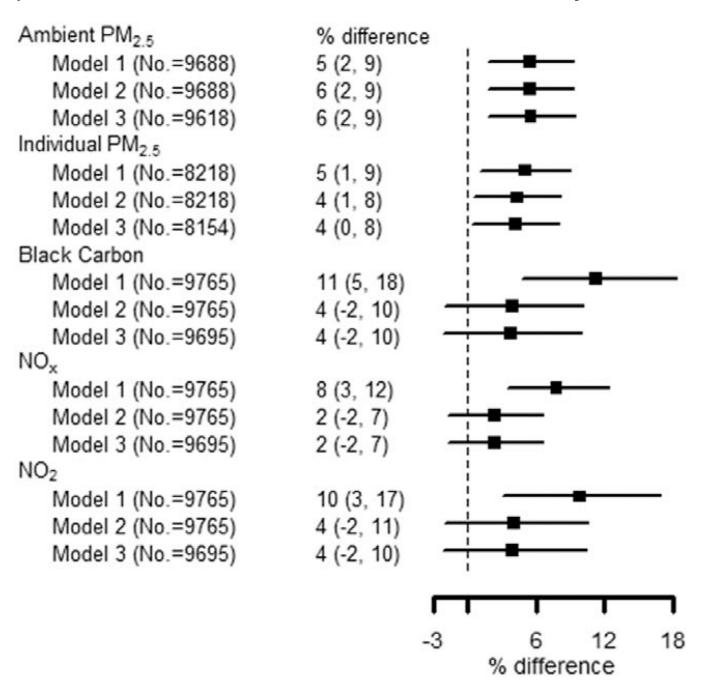


FIGURE 1.

Associations between IL-6 and long-term pollutant concentrations. Air pollution metrics are annual average concentrations. Percent differences in IL-6 are reported as $5 \mu g/m^3$ increase in PM_{2.5}, $0.7 \cdot 10^{-6} m^{-1}$ increases in black carbon, 40 ppb increase in NO_x, and 17 ppb increase in NO₂. Model 1 is adjusted for age, race/ethnicity, gender, exam, and site. Model 2 is adjusted for the covariates in Model 1 as well as for education, employment, income, neighborhood SES, recent infection, second-hand smoke exposure, smoking status, alcohol consumption, physical activity, BMI, waist–hip ratio, and splines for calendar time. Model 3

is adjusted for the covariates in Model 2 as well as for hypertension, diabetes, and anti-inflammatory medications.

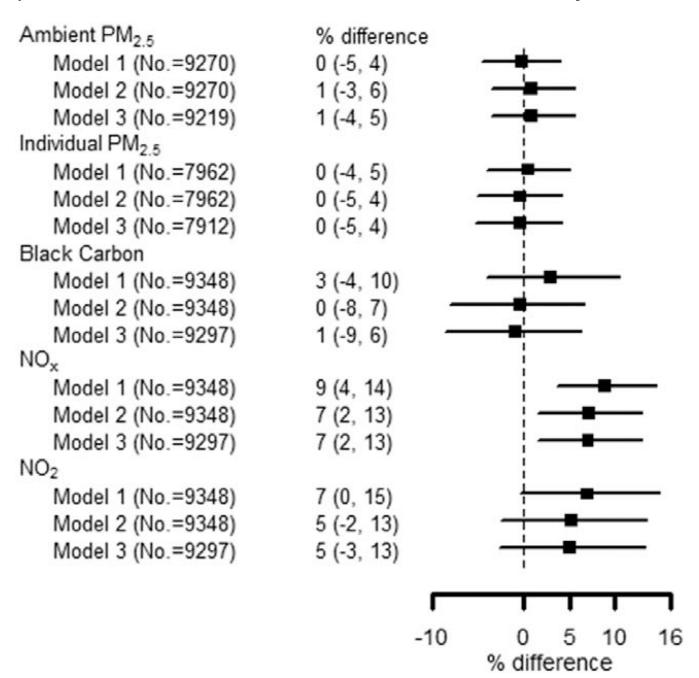


FIGURE 2.

Associations between D-dimer and long-term pollutant concentrations. Air pollution metrics are annual average concentrations. Percent differences are reported as $5 \mu g/m^3$ increase in $PM_{2.5}$, $0.7 \cdot 10^{-6} m^{-1}$ increases in black carbon, 40 ppb increase in NO_x , and 17 ppb increase in NO_2 . Model 1 is adjusted for age, race/ethnicity, gender, exam, and site. Model 2 is adjusted for the covariates in Model 1 as well as education, employment, income, neighborhood SES, recent infection, second-hand smoke exposure, smoking status, alcohol consumption, physical activity, BMI, waist–hip ratio, and splines for calendar time. Model 3

is adjusted for the covariates in Model 2 as well as for hypertension, diabetes, and anti-inflammatory medications. BMI indicates body mass index.

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TABLE 1

Sample Size for Each Blood Marker at Each Exam, After Exclusion of Missing and Outlier Outcome Data

	Exam 1 (2000–2002), No.	Exam 2 (2002–2004), No.	Exam 3 (2004–2005), No.	Exam 4 (2005–2007), No.	Exam 5 (2010–2012), No.	Total No. of Observations	% of Participants with 2 Observations	No. of Unique Participants ^a
CRP	6,713	092	1,160	189	1,870	11,190	47	6,889
IL-6	6,617	754	1,683	335	1,265	10,654	42	6,663
D-dimer	6,761	,	651	1,017	1,879	10,308	30	7,029
Fibrinogen	6,765	611	1,187	7111	1,876	11,318	48	7,037
E-selectin	686	,	,	704	711	2,404	11	1,251
sICAM-1	2,594	1		701	712	4,007	11	2,865
Total no.	6,814	780	1,753	1,019	1,882	12,248		

Empty cells indicate that the blood marker was not measured at this exam.

 $^{\rm d}$ Reflects participants from MESA parent study (n = 6,814) and MESA Air new recruits (n = 257).

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TABLE 2

Number (%) of Demographic Characteristics and Cardiovascular Disease Risk Factors at Baseline Exam for All Study Participants

Participant Characteristic	No. (%)
Female	3,740 (53)
Black^a	1,933 (27)
$\mathrm{Hispanic}^a$	1,547 (22)
Chinese a	805 (11)
Current smokers ^b	903 (13)
Former smokers ^b	2,607 (37)
Exposed to second-hand smoke	3,170 (46)
Current alcohol	3,903 (56)
Hypertensive	3,162 (45)
Treated and untreated diabetic $^{\mathcal{C}}$	905 (13)
Borderline diabetic $^{\mathcal{C}}$	990 (14)
Anti-inflammatory medications	3,814 (54)
1 acute infection	2,277 (32)

 $^{^{}a}$ Category not shown is white.

 $^{{}^{}b}\mathrm{Category}$ not shown is non-smoker.

 $^{^{\}it c}$ Category not shown is normal blood sugar.

 TABLE 3

 lard Deviation) of Blood Markers, Short- and Long-term Air Pollution, and Cardiovascula

Mean (Standard Deviation) of Blood Markers, Short- and Long-term Air Pollution, and Cardiovascular Disease Risk Factors at Baseline Exam for All Study Participants

Participant Characteristic	Mean (Standard Deviation)
Cardiovascular disease risk factors	
Age (years)	62 (10)
Intentional exercise (MET-min/wk)	1,576 (2,344)
BMI (kg/m²)	28.4 (5.5)
Waist-hip ratio	0.9 (0.1)
Annual average air pollution concentrations a	
Ambient $PM_{2.5}$ (µg/m ³)	16.5 (3.4)
Individual $PM_{2.5}$ (µg/m ³)	10.9 (3.6)
NO_x (ppb)	49.8 (27.0)
NO ₂ (ppb)	21.5 (9.1)
Black carbon (10^{-6} m^{-1})	0.8 (0.4)
Short-term $PM_{2.5}$ concentrations ($\mu g/m^3$) a	
Day of blood draw	17.2 (10.2)
Day prior	16.7 (10.1)
2-day average	16.6 (9.1)
3-day average	16.7 (8.4)
4-day average	16.8 (7.9)
5-day average	16.9 (7.5)
Blood markers	
CRP (mg/liter)	3.5 (4.3)
CRP (geometric mean)	1.9 (3.1)
IL-6 (pg/ml)	1.6 (1.2)
IL-6 (geometric mean)	1.2 (1.9)
Fibrinogen (mg/dl)	348.8 (74.9)
D-Dimer (µg/ml)	0.4 (0.5)
D-Dimer (geometric mean)	0.2 (2.5)
E-selectin (ng/ml)	50.8 (23.2)
sICAM-1 (ng/ml)	267.2 (71.3)

 $^{^{}a}$ Means calculated for 6,814 exam 1 participants recruited from 2000 to 2002. Excludes participants recruited at exam 4 (2005–2007) for MESA Air ancillary study (n = 257) because of declining trends in air pollution. Means (standard deviation) for 257 MESA Air participants at recruitment are as follows: ambient PM2.5 = 14.4 (3.6), individual PM2.5 = 10.3 (3.2), NO_{χ} = 37.3 (11.7), NO_{χ} = 16.5 (4.3), BC = 0.9 (0.3), day of blood draw = 14.4 (8.8), day prior = 13.9 (7.9), 2-day average = 14.0 (7.0), 3-day average = 13.9 (6.3), 4-day average = 13.8 (5.8), and 5-day average = 14.1 (5.6).

BMI = body mass index; CRP = C-reactive protein; MET = metabolic equivalents.

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TABLE 4

Adjusted Mean Concentrations (Standard Deviation) of Blood Markers by Quartile of Long-term Air Pollution Concentrations for Study Participants at Baseline Exama

Fibrinogen (mg/dl) D-Dimer (µg/ml) E-Selectin (ng/ml) sICAM-1 (ng/ml) 271.82 (5.51) 265.86 (3.81) 258.08 (3.07) 267.42 (4.85) 261.03 (7.42) 260.87 (3.65) 261.63 (3.2) 252.11 (4.03) 266.69 (3.39) 251.74 (4.29) 274.16 (4.33) 264.19 (3.36) 259.49 (3.3) 274.01 (4.6) 267.92 (3.3) 253.41 (4.6) 0.30 0.003 0.001 0.11 54.75 (1.84) 54.85 (1.69) 54.15 (2.11) 56.58 (1.91) 56.19 (2.07) 58.07 (2.58) 57.03 (3.42) 54.23 (1.88) 56.95 (2.78) 56.95 (1.85) 55.07 (2.52) 53.42 (2.24) 55.86 (1.71) 55.61 (1.82) 57.53 (2.33) 53.71 (2.35) 0.98 0.86 0.34 0.31 0.37 (0.02) 0.35 (0.02) 0.34 (0.02) 0.38 (0.02) 0.35 (0.02) 0.29 (0.03) 0.34 (0.02) 0.38 (0.02) 0.34 (0.02) 0.36 (0.02) 0.34 (0.02) 0.35 (0.02) 0.36 (0.02) 0.34(0.02)0.34(0.02)0.33(0.02)0.50 0.03 0.75 0.89 340.68 (2.44) 346.75 (2.19) 344.42 (2.06) 351.86 (2.07) 346.88 (3.37) 341.31 (2.68) 342.89 (2.11) 348.52 (2.03) 350.86 (2.96) 346.48 (2.05) 344.33 (2.28) 348.35 (3.13) 345.65 (2.1) 347.98 (2.98) 343.26 (2.88) 343.2 (2.79) 0.08 0.03 0.40 0.46 IL-6 (mg/liter) 1.58 (0.03) 1.61 (0.05) 1.47 (0.04) .46 (0.04) ..51 (0.05) .49 (0.04) .49 (0.03) .57 (0.03) .65 (0.05) .47 (0.03) .55 (0.03) .57 (0.04) .51 (0.05) .55 (0.06) .38 (0.04) .44 (0.05) 0.0006 0.21 0.02 0.75 CRP (mg/liter) 3.11 (0.19) 3.22 (0.17) 3.09 (0.15) 3.32 (0.12) 3.43 (0.12) 3.07 (0.2) 2.87 (0.16) 3.1 (0.13) 3.41 (0.18) 3.01 (0.17) 3.31 (0.12) 3.36 (0.13) 3.28 (0.18) 3.25 (0.13) 3.41 (0.14) 3.6 (0.12) 0.95 0.03 0.38 0.87 Black carbon (10⁻⁶ m⁻¹) Ambient PM_{2.5} (µg/m³) 15.8 to < 17.7 42.4 to <71.9 14.5 to <15.8 0.51 to <0.65 0.65 to <1.17 26.0 to <42.4 13.8 to <19.9 19.9 to <30.4 NO_x (ppb) NO₂ (ppb) $P_{
m trend}^{} p$ $P_{\mathrm{trend}}^{}$ $P_{
m trend}^{} p$ $P_{
m trend}^{} p$ <0.51 <26.0 <13.8 <14.5 17.7 71.9 1.17 30.4

ame and an are adjusted for age, race/ethnicity, gender, and site using linear regression models. Air pollution metrics are annual average concentrations before date of blood draw. Outliers for markers of inflammation, coagulation and endothelial activation have been excluded. Includes 6,814 exam 1 participants only, recruited from 2000 to 2002.

TABLE 5

Percent Differences in Concentrations of Blood Markers (IL-6 and D-Dimer) Associated with Incremental Increase in Long-term Pollutant Concentration (Ambient $PM_{2.5}$ and NO_x), in Models Stratified by Individual Demographic and Health Characteristics^a

	% Difference in IL-6 Associated with Ambient PM _{2.5} (95% CI)	% Difference in D-Dimer Associated with NO _x (95% CI)
Women	4 (-1, 9)	10 (3, 19)
Men	8 (3, 14)	3 (-4, 11)
<65 years old	3 (-2, 9)	5 (-2, 12)
65 years old	9 (3, 14)	11 (3, 21)
White	5 (-3, 13)	11 (1, 23)
Non-white	4 (0, 9)	4 (-3, 11)
Some college	8 (3, 14)	8 (1, 16)
High school	3 (-2, 9)	8 (-1, 17)
Working outside home	4 (-1, 10)	5 (-2, 12)
Not working outside home	7 (1, 13)	12 (3, 21)
Not current smoker	5 (1, 9)	6 (0, 12)
Current smoker	9 (-2, 22)	16 (0, 35)
Not obese	6 (1, 11)	6 (-1, 13)
Obese	5 (-1, 11)	11 (2, 21)
Not diabetic	5 (1, 10)	6 (0, 12)
Diabetic	6 (-3, 17)	23 (5, 43)
Not hypertensive	4 (-1, 9)	3 (-4, 10)
Hypertensive	9 (3, 15)	12 (4, 22)

 $[^]a$ Air pollution metrics are annual average concentrations. Percent differences in IL-6 and D-dimer are relative to 5 µg/m 3 increase in ambient PM2.5 and 40 ppb increase in NO $_x$ and are derived from random effects models. Models are adjusted for age, race, gender, exam, education, employment, income, neighborhood SES, second-hand smoke exposure, smoking status, alcohol consumption, physical activity, BMI, waist–hip ratio, recent infection and for calendar time splines, except when model is stratified by one of the aforementioned variables. IL-6 models include 48 df calendar time splines (4 df * 12 years) and D-dimer models include 40 df splines (4df * 10 years).

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TABLE 6

Percent Differences and Difference from Mean Blood Markers per Increment of Long-term Pollutant Concentrations (5 µg/m³ Increase in PM2.5, 0.7 10⁻⁶ m⁻¹ Increase in Black Carbon, 40 ppb Increase in NO_x, and 17 ppb Increase in NO₂) from Random-effects Models^a

	Z	CRP, % Difference	Z	Fibrinogen, mg/dl	Z	E-Selectin, ng/ml	Z	sICAM-1, ng/ml
Ambient PM _{2.5}								
Model 1^b	10,120	3 (-2, 9)	10,155	-0.42 (-4.33, 3.49)	2,182	2.3 (0.54, 4.05)	3,374	0.97 (-5.11, 7.06)
Model 2^c	10,120	1 (-4, 6)	10,232	-3.45 (-7.43, 0.52)	2,182	1.08 (-0.66, 2.82)	3,374	-2.07 (-7.69, 3.56)
Model 3d	10,045	0 (-5, 6)	10,232	-3.29 (-7.28, 0.7)	2,160	0.87 (-0.88, 2.63)	3,349	-2.07 (-7.72, 3.59)
Individual PM _{2.5}								
Model 1	8,651	0 (-6, 6)	8,741	-0.34 (-4.54, 3.86)	2,026	1.64 (-0.28, 3.57)	2,985	-0.58 (-6.58, 5.42)
Model 2	8,651	-2 (-8, 4)	8,741	-2.31 (-6.46, 1.84)	2,026	0.71 (-1.2, 2.62)	2,985	-3.02 (-8.47, 2.43)
Model 3	8,581	-2 (-8, 4)	8,669	-2.25 (-6.41, 1.9)	2,004	0.53 (-1.4, 2.46)	2,960	-2.84 (-8.32, 2.64)
Black carbon								
Model 1	10,197	5 (-4, 15)	10,310	-3.51 (-10, 2.98)	2,202	-0.89 (-4.26, 2.49)	3,461	0.94 (-8.2, 10.08)
Model 2	10,197	2 (-7, 12)	10,310	-6.03 (-12.62, 0.56)	2,202	-2.69 (-6.16, 0.78)	3,461	-1.02 (-9.32, 7.28)
Model 3	10,122	3 (-6, 12)	10,233	-5.7 (-12.3, 0.91)	2,180	-2.67 (-6.15, 0.8)	3,436	-0.58 (-8.9, 7.74)
NO_x								
Model 1	10,197	0 (-7, 6)	10,310	2.37 (-2.19, 6.93)	2,202	1.85 (-0.21, 3.92)	3,461	3.88 (-1.97, 9.73)
Model 2	10,197	-4 (-11, 2)	10,310	-0.54 (-5.24, 4.15)	2,202	0.49 (-1.61, 2.58)	3,461	-0.13 (-5.67, 5.41)
Model 3	10,122	-4 (-11, 2)	10,233	-0.54 (-5.24, 4.16)	2,180	0.3 (-1.8, 2.41)	3,436	0.07 (-5.49, 5.62)
NO_2								
Model 1	10,197	-5 (-15, 4)	10,310	-12.55 (-19.06, -6.03)	2,202	1.07 (-1.94, 4.08)	3,461	-0.93 (-9.54, 7.68)
Model 2	10,197	-7 (-17, 2)	10,310	-13.1 (-19.82, -6.37)	2,202	-0.57 (-3.72, 2.59)	3,461	-0.54 (-8.78, 7.7)
Model 3	10,122	-7 (-17, 3)	10,233	$-13.01 \; (-19.75, -6.28)$	2,180	-0.53 (-3.7, 2.64)	3,436	-0.09 (-8.37, 8.18)

aAir pollution metrics are annual average concentrations.

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 $b \hspace{-0.5em} \mbox{Model 1}$ is adjusted for age, race/ethnicity, gender, exam, and site.

BMI, waist-hip ratio, recent infection, and for calendar time splines. CRP, IL-6, and fibrinogen models include 48 df calendar time splines (4 df * 12 years), D-dimer models include 40 df splines (4df * 10 CModel 2 is adjusted for the covariates in Model 1 as well as education, employment, income, neighborhood SES, second-hand smoke exposure, smoking status, alcohol consumption, physical activity, years) and E-selectin and sICAM-1 models includes 20 df splines (4df * 5 years). sICAM-1 models were also adjusted for the gene RS4591 (no T versus at least one T).

 $d_{\rm M}$ Model 3 is adjusted for the covariates in Model 2 as well as for hypertension, diabetes, and anti-inflammatory medications.

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TABLE 7

Percent Differences and Difference from Mean Blood Markers per 5 µg/m³ Increase in Short-term PM_{2.5} Concentrations from Fixed Effects Models

	CRP (% Difference)	IL-6 (% Difference)	D-Dimer (% Difference)	Fibrinogen (mg/dl)	Fibrinogen (mg/dl)	sICAM-1 (ng/ml)
Day of blood draw						
Model 1 ^a	2 (0, 3)	0 (-1, 1)	1 (-1, 3)	1.09 (-0.36, 2.54)	0.61 (0.05, 1.16)	0.43 (-1.73, 2.59)
Model 2^b	1 (0, 3)	0 (-1, 1)	1 (-1, 3)	1.16 (-0.28, 2.61)	0.6 (0.06, 1.14)	0.43 (-1.7, 2.55)
Model $3^{\mathcal{C}}$	1 (0, 3)	0 (-1, 1)	1 (-1, 3)	1.19 (-0.28, 2.66)	0.6 (0.05, 1.14)	0.42 (-1.74, 2.58)
Day prior						
Model 1	0 (-1, 2)	0 (-1, 1)	0 (-1, 2)	-0.37 (-1.75, 1)	0.27 (-0.26, 0.81)	-1.03 (-3.01, 0.96)
Model 2	0 (-1, 2)	0 (-1, 1)	0 (-1, 2)	-0.33 (-1.7, 1.04)	0.39 (-0.13, 0.91)	-0.6 (-2.57, 1.36)
Model 3	1 (-1, 2)	0 (-1, 1)	0 (-1, 2)	-0.27 (-1.66, 1.12)	0.35 (-0.18, 0.89)	-0.88 (-2.89, 1.13)
Average 2 day						
Model 1	-1 (-3, 1)	0 (-1, 1)	0 (-2, 2)	$-0.08 \; (-1.73, 1.57)$	0.3 (-0.36, 0.96)	-1.43 (-3.86, 1)
Model 2	-1 (-3, 1)	0 (-1, 1)	0 (-2, 2)	$-0.04 \; (-1.69, 1.6)$	0.44 (-0.2, 1.08)	-0.93 (-3.33, 1.47)
Model 3	-1 (-3, 2)	0 (-1, 1)	0 (-2, 2)	0.05 (-1.63, 1.73)	0.27 (-0.39, 0.93)	-1.29 (-3.75, 1.17)
Average 3 day						
Model 1	-2 (-4, 1)	0 (-1, 2)	-2 (-4, 0)	0.16 (-1.73, 2.05)	-0.23 (-0.96, 0.51)	-3.99 (-6.75, -1.23)
Model 2	-1 (-4, 1)	0 (-1, 2)	-2 (-4, 0)	0.12 (-1.76, 2)	-0.02 (-0.74, 0.69)	-3.21 (-5.93, -0.48)
Model 3	-1 (-4, 1)	0 (-1, 2)	-2 (-4, 0)	0.26 (-1.66, 2.18)	$-0.22\ (-0.95,0.51)$	-3.58 (-6.36, -0.79)
Average 4 day						
Model 1	-2 (-4, 1)	1 (-1, 2)	-2 (-5, 0)	-0.56 (-2.64, 1.52)	-0.61 (-1.42, 0.2)	-5.19 (-8.24, -2.13)
Model 2	-1 (-4, 1)	1 (-1, 2)	-2 (-5, 0)	-0.51 (-2.58, 1.56)	-0.37 (-1.16, 0.42)	-4.22 (-7.24, -1.21)
Model 3	-1 (-4, 1)	1 (-1, 2)	-2 (-5, 0)	-0.32 (-2.43, 1.79)	-0.59 (-1.39, 0.22)	-4.5 (-7.57, -1.42)
Average 5 day						
Model 1	-2 (-5, 1)	1 (0, 3)	-3 (-5, 0)	-1.59 (-3.89, 0.71)	-0.57 (-1.42, 0.28)	-4.17 (-7.47, -0.88)
Model 2	-2 (-4, 1)	1 (0, 3)	-3 (-5, 0)	-1.4 (-3.68, 0.89)	-0.33 (-1.16, 0.5)	-3.13 (-6.38, 0.13)
Model 3	-2 (-4, 1)	1 (0, 3)	-3 (-5, 0)	-1.22 (-3.55, 1.1)	-0.5(-1.35, 0.34)	-3.23 (-6.53, 0.08)

 $^{^{\}it a}$ Model 1 is adjusted for site, age, race/ethnicity, gender, and exam.

bModel 2 is adjusted for all variables in Model 1 as well as education, employment, income, neighborhood SES, recent infection, second-hand smoke exposure, smoking status, alcohol consumption, BMI, exercise, waist-hip ratio. sICAM-1 models also adjusted for the gene RS4591 (no T versus at least one T).