

UC Irvine

UC Irvine Previously Published Works

Title

TERRESTRIAL BIOSPHERE-ATMOSPHERE EXCHANGE IN HIGH-LATITUDES

Permalink

<https://escholarship.org/uc/item/62m9j8hw>

Journal

GLOBAL ATMOSPHERIC-BIOSPHERIC CHEMISTRY, 48

Authors

REEBURGH, WS
ROULET, NT
SVENSSON, BH

Publication Date

1994

Copyright Information

This work is made available under the terms of a Creative Commons Attribution License, available at <https://creativecommons.org/licenses/by/4.0/>

Peer reviewed

TERRESTRIAL BIOSPHERE-ATMOSPHERE EXCHANGE IN HIGH LATITUDES

William S. Reeburgh¹, Nigel T. Roulet² and Bo H. Svensson³

¹Department of Geosciences
University of California, Irvine
Irvine, CA 92717-3100

²Department of Geography
McGill University
Montreal, Quebec
Canada H3A 2K6

³Department of Microbiology
Swedish University of Agricultural Sciences
S-750 07 Uppsala, Sweden

ABSTRACT

This paper summarizes CH₄ flux measurements from high latitude tundra and wetland environments obtained during large field campaigns and a number of smaller local projects since 1986. Area-weighted time series CH₄ flux measurements at fixed sites in high latitude tundra and wetland environments suggest a CH₄ source strength of about 35 Tg yr⁻¹, which agrees reasonably with atmospheric model calculations. Individual flux measurements are subject to high spatial and temporal (daily to interannual) variability, and are strongly influenced by vegetation, soil temperature, thaw depth, and water table level, which influence the balance between CH₄ production and consumption. Field studies show that these subsurface parameters and CH₄ fluxes are often poorly correlated. Laboratory studies on varying water table level in tundra soil columns have shown long lag times and hysteresis in the fluxes under falling and rising water table conditions. These results indicate that the soil parameters are not independent variables and that a single parameter capable of predicting CH₄ flux is unlikely.

Boreal forest ecosystems appear to be a small net sink for CH₄. Microbially mediated oxidation of atmospheric CH₄ is the dominant process; wet bogs, fens, swamps, and beaver ponds are the only CH₄ sources. Microbial CH₄ oxidation is also an important modulator of tundra emission and has the potential to eliminate net CH₄ emission under warmer, drier conditions.

Studies to date have emphasized CH₄, but future integrated studies along boreal forest and tundra transects will include measurements of other gases and carbon dioxide fluxes for estimates of carbon fixation and storage. Field manipulation experiments aimed at determining the sensitivity of high latitude trace gas emissions to temperature and moisture variations are also planned. High latitude biosphere-atmosphere studies to date have a strong North American bias. Collaborative studies in northern European and Siberian wetlands are essential to understanding the importance of high latitude ecosystems in trace gas budgets.

INTRODUCTION

Two circumpolar components of Arctic and sub-Arctic ecosystems, tundra and boreal forest, have accumulated and stored soil carbon since the last glaciation. Tundra (area: 7.3×10^{12} m²; 7% of earth's land area) is the treeless Arctic and sub-Arctic plain populated with dwarf shrubs, graminoids and cryptogams, and is often underlain by permafrost. Tundra contains 13% of the earth's stored soil carbon. Boreal forest (11.1×10^{12} m²) is a general term describing the mixed deciduous-coniferous forest biome located between the tundra and temperate forest or prairie. It contains about 14% of the stored soil carbon (Post *et al.*, 1982). Boreal wetlands (area: 1.5×10^{12} m²) occupy regions between 45° and 60°N, and are counted as part of tundra under most vegetation classification schemes. Approximately 30% of all northern wetlands are contained in two continuous wetland complexes: the Hudson Bay Lowland in Canada (3.2×10^{10} m²) and the Vasuugan peatland (5.4×10^{10} m²) in Siberia. All of these systems are characterized by short (100 to 120 day) growing seasons, low temperatures, low decomposition rates, moist to waterlogged soils, and the presence of continuous or discontinuous permafrost.

Much of the stored soil carbon in high latitude systems is present as peat or frozen in permafrost, and is subject to limited biogeochemical degradation. The fate of this stored carbon under altered climate is a major question (Billings, 1987). Model projections (Mitchell, 1989) indicate that Arctic regions will be the first impacted by global warming. Permafrost temperature distributions (Lachenbruch and Marshall, 1986) are consistent with a decade-scale rise in temperature. We understand in broad terms that CH₄ emissions result from a balance between production (methanogenesis) and consumption (methylotrophy), and that this balance is governed by temperature, moisture/water table relations, and substrate quality, but which variables are the best predictors of methane flux, the importance of feedbacks, and generally, the role of tundra under altered climate remains unclear. Some projections (Khalil and Rasmussen, 1989; Lashoff, 1989; Oechel *et al.*, 1993) suggest that warmer climate will lead to warmer soil temperatures, thawing permafrost and increased CH₄ and CO₂ emission. Others (Whalen and Reeburgh, 1990a; Reeburgh *et al.*, 1993; Moore and Roulet, 1993; Roulet *et al.*, 1993) suggest that microbially mediated CH₄ oxidation, which is neglected in the above projections, has the potential to modulate CH₄ emission, and under appropriate conditions, consume atmospheric CH₄.

This paper is structured around progress on the tasks of the IGAC High Latitude Ecosystems as Sources and Sinks for Trace Gases (HESS) Activity, which include:

- Estimating the magnitude of high latitude ecosystem trace gas emissions.
- Understanding factors and processes controlling high latitude ecosystem gas exchange.
- Understanding sensitivity of high latitude trace gas sources and sinks to climate change.

Since these tasks are sequential, this paper will emphasize recent work on trace gas emissions from high latitude ecosystems and our current understanding of the role of high latitude environments in the CH₄ budget (Task 1), spell out our present understanding of the important controls on CH₄ fluxes (Task 2), and finally, outline manipulation experiments planned to determine sensitivity of high latitude systems to climate change (Task 3).

ROLE OF HIGH LATITUDE ENVIRONMENTS IN THE GLOBAL METHANE BUDGET (Task 1)

Early Studies

The possible importance of high latitude environments in radiatively active trace gas budgets was recognized early, but prior to 1986 only three studies provided information on CH₄ fluxes from high latitude systems (Clymo and Reddaway, 1971; Svensson and Rosswall, 1984a, b; Sebacher *et al.*, 1986). High latitude wetlands have been prominent as a potentially large term in the global CH₄ budget (Cicerone and Oremland, 1988; Matthews and Fung, 1987; Fung *et al.*, 1991), but their contribution has been uncertain because of three problem areas: lack of basic flux data, lack of understanding of the processes controlling fluxes, and uncertainties in scaling up from sites to regional and global scales. Large-scale coordinated field campaigns and individual projects have led to rapid growth in the high latitude CH₄ flux data base. The wetland CH₄ flux work has been summarized and reviewed recently by Bartlett and Harriss (1993). Because of the almost total absence of CH₄ flux data, net CH₄ fluxes have received major emphasis, but measurements of other gas fluxes are being incorporated in current work.

Large Field Campaigns

ABLE 3A. The Arctic Boundary Layer Expedition (ABLE 3A) was conducted in the Arctic and sub-Arctic of North America during July and August of 1988. Studies included a comprehensive tower, chamber and aircraft campaign at a site in the Yukon-Kuskokwim Delta near Bethel, Alaska, and flights around Barrow, Alaska. A summary of the results appears in a dedicated 1992 issue (number D15) of the *Journal of Geophysical Research*; the following discussion will emphasize the CH₄ studies.

Methane flux measurements were made with methods covering three scales; aircraft boundary layer measurements (1000 km), tower measurements (103 m) and chamber measurements (1 m²). Errors of approximately a factor of 2 between chamber and tower flux measurements resulted from poor resolution of CH₄ producing habitats in the tower footprint (Fan *et al.*, 1992). Mean aircraft fluxes were ~2 times higher than the tower fluxes. The aircraft data were obtained during afternoon in periods of good flying weather, and these conditions were associated with maximum tower fluxes. The tower and ground-based measurements were made in an area with small lakes, whose CH₄ flux varied with size and wind speed. A mean emission rate of 25 mg CH₄ m⁻² d⁻¹ for wet tundra, dry tundra and lakes, which is low compared to other sites, was made by areal weighting of vegetation types obtained from satellite imagery. Bartlett *et al.* (1992) extended this flux to a global emission rate for tundra of 11±3 Tg CH₄ yr⁻¹. Ritter *et al.* (1992) measured CH₄ source strengths of 25 to 85 mg CH₄ m⁻² d⁻¹ with aircraft boundary layer measurements, and established an averaged, seasonally adjusted source strength of 51 mg CH₄ m⁻² d⁻¹. A global estimate for tundra CH₄ emissions using this rate was 44 Tg yr⁻¹.

NOWES/ABLE 3B. The Northern Wetlands study (NOWES), a joint experiment between the Canadian Institute for Research in Atmospheric Chemistry (CIRAC) and the NASA Atmospheric Boundary Layer Experiment 3B, was conducted in 1990 to assess the importance of a large continuous wetland, the Hudson Bay Lowland in Northern Canada ($3.2 \times 10^{10} \text{ m}^2$), as a source/sink of biogenic gases under current climate conditions (Roulet *et al.*, 1992a). Most of the measurements of CH_4 flux in high latitude systems have been from relatively small isolated systems, i.e. wetlands with an area $<1000 \text{ km}^2$. A dedicated issue (number D1) similar to the ABLE 3A issue of the *Journal of Geophysical Research* appeared in 1994.

The experiment was conducted in two locations. The first location was a 120 km transect running from James Bay coast inland to the interior of the Hudson Bay lowland. Areal coverage of this transect by wetlands and lakes was 100%. The second location was near Schefferville, Quebec, in a region that is approximately 30% wetland. Fluxes of CO_2 , CH_4 , N_2O , NO_x , and NMHCs were measured at numerous sites with flux chambers. A micrometeorological flux tower at each study site was used to measure fluxes of CO_2 , CH_4 and O_3 . Periodic boundary layer flux measurements of CO_2 , CH_4 , CO and O_3 were made by aircraft during the experiment. In addition to flux measurements, the concentrations of other trace gases such as CO_2 , VOC, NO , N_2O , NO_2 , O_3 and some NMHC were made throughout the summer at Schefferville and an atmospheric observatory on the southern fringe of the lowland.

Low CH_4 fluxes for the Hudson Bay Lowland were the most surprising result of the campaign. Based on prior studies (Bartlett and Harriss, 1993) the Hudson Bay Lowland was expected to emit between 5 and 7 Tg $\text{CH}_4 \text{ yr}^{-1}$, but field measurements gave an estimate of $0.5 \pm 0.03 \text{ Tg CH}_4 \text{ yr}^{-1}$. (Roulet *et al.*, 1994). These estimates were obtained by weighting chamber flux measurements by the fractional component of each major ecosystem determined by remote sensing. Fluxes measured by aircraft averaged $16 \text{ mg CH}_4 \text{ m}^{-2} \text{ d}^{-1}$ (Ritter *et al.*, 1994), tower-derived fluxes averaged $16 \text{ mg CH}_4 \text{ m}^{-2} \text{ d}^{-1}$ (Edwards *et al.*, 1994), and the modeled chamber fluxes averaged $20 \text{ mg CH}_4 \text{ m}^{-2} \text{ d}^{-1}$. Methane fluxes measured by these three independent methods agreed to within a factor of 2 at all times, and to within a factor of 1.2 most of the time (Roulet *et al.*, 1993). However, Moore *et al.* (1993) have shown that comparisons between tower and aircraft fluxes to greater than a factor of 1.5 are not possible for sensible heat due to errors in measurements at different scales. These errors are larger in comparison to those for trace gases. The highest daily fluxes of CH_4 were from the shallow ponds on the surface of the wetlands: 40 to $150 \text{ mg CH}_4 \text{ m}^{-2} \text{ d}^{-1}$. (Hamilton *et al.*, 1994). The fluxes from the peat surfaces were much smaller: $<40 \text{ mg CH}_4 \text{ m}^{-2} \text{ d}^{-1}$ (Moore *et al.*, 1994). Approximately 10% of the fluxes measured from the peat surfaces indicated uptake of CH_4 . The NOWES CH_4 flux measurements were within 25% of those predicted by Fung *et al.* (1991).

The fluxes of N_2O were small, -3.8 to $7.9 \text{ mg m}^{-2} \text{ d}^{-1}$, yielding a total source of -1.2 GT yr^{-1} , an insignificant source in the global budget (Schiller and Hassie, 1994). Wetlands in general are NO_3^- depleted and limited, so their capacity to yield large amounts of N_2O is limited. The NMHC flux was measured in only a few locations, but was at least as large, and in many cases larger, than the CH_4 flux (Klinger *et al.*, 1994). The seasonal distribution of isoprene followed a pattern controlled by the source from vegetation and air temperature (Jobson *et al.*, 1994).

During NOWES there was little direct work on the controls of the fluxes of various trace gases. However, because the experiment was located along a transect in an area that has been isostatically rebounding since the last glaciation (Roulet *et al.*, 1992a), it presented an opportunity to study how 5000 years of peatland evolution influences the CH_4 flux. Moore *et al.* (1994) and Klinger *et al.* (1994) found clear relationships among various stages of ecological succession and CH_4 flux in individual wetland types, while Roulet *et*

al. (1994) showed there was a significant increase in CH₄ flux from the region of younger wetlands (1,000 to 2,000 years old) near the James Bay coast to older wetlands (~5000 years old) in the interior of the Hudson Bay lowland. This was in large part due to the increase in the area of shallow ponds on wetlands inland from the coast.

Individual Studies

While the large campaigns have emphasized integration over large areas, individual studies have emphasized surveys, time series flux measurements, and processes controlling trace gas fluxes in a variety of high latitude wetland (Roulet *et al.*, 1992b; Roulet *et al.*, 1992c) and boreal forest (Whalen *et al.*, 1991) environments. These studies provide long term information on spatial and temporal variability of CH₄ fluxes; one study (Whalen and Reeburgh, 1992) has continued for 6 years, providing information on interannual variability. At least two of these studies have involved winter measurements (Dise, 1992; Whalen and Reeburgh, 1988). Winter fluxes are less than 5% of the annual emission in high latitude tundra systems, but can contribute as much as 20% of the annual emission in lower latitude peatlands with less extensive freezing. Scaling up site scale measurements to regional and global scales has involved areal weighting of fixed site time series flux data (Whalen and Reeburgh, 1988, 1992) and transect studies (Whalen and Reeburgh, 1990b). Measurements in an Alaska boreal forest system (Whalen *et al.*, 1991) showed that large areas of moist to dry soils actively consume atmospheric CH₄. Local bogs and fens are the only CH₄-emitting sites in boreal forest systems.

Importance of High Latitude Systems in the Global Methane Budget

Because little was known about the CH₄ source strength of high latitude environments, a global emission estimate has been made using flux data from virtually every study. Table 1 summarizes these high latitude CH₄ emission estimates. Most of these studies assume an active period of 100 to 120 days and a distribution of environments related to the flux data available. Morrissey and Livingston (1992) point out that high natural spatial and temporal variability in fluxes make this practice questionable. The earliest estimate (Svensson and Rosswall, 1976) is based on a CH₄ flux/moisture relationship. The estimates of Sebacher *et al.* (1986) and Whalen and Reeburgh (1990b) are from transect sampling campaigns; the latter campaign involved sampling at fixed intervals to obtain an areally representative distribution of samples. Time series flux measurements from fixed sites were used in the estimates of Crill *et al.* (1988), Whalen and Reeburgh, (1988, 1992), and Moore *et al.* (1990). The estimate of Bartlett and Harriss (1993) results from a comprehensive review of published CH₄ flux data. Fung *et al.* (1991) used the GISS global 3-D tracer model, a compilation of geographic and seasonal CH₄ sources and sinks, and CH₄ carbon isotope ratios to provide constraints on global CH₄ budget terms. Several budget scenarios were tested for their ability to reproduce the meridional gradient and seasonal variations observed at the GMCC stations. Their preferred scenario calls for CH₄ emissions of 35 Tg yr⁻¹ from wetlands and tundra north of 50°. Reeburgh and Whalen (1992) pointed out that the high latitude CH₄ emission term seemed to be converging on about 35 Tg yr⁻¹. The additional global estimates since then, with the exception of the Bartlett *et al.* (1992) ABLE 3A estimate, lie between 20 and 45 Tg yr⁻¹ and have large (±50%) error limits. High latitude terrestrial environments contribute slightly less than 10% to the global CH₄ budget (Fung *et al.*, 1991).

Table 1. High latitude CH₄ emission estimates.

Reference	Region	Location	Study Type	CH ₄ Emission (Tg yr ⁻¹)
Svensson & Rosswall (1976)	Arctic tundra	Stordalen, Sweden	summer fluxes extended with moisture relationships	1 – 25
Sebacher <i>et al.</i> (1986)	tundra and taiga	Alaska	flux transect, Arctic and boreal regions	46 – 106
Matthews & Fung (1987)	forested bog 50° – 70°N	—	estimated area × literature emission rates	20 – 62
	non-forested bog 50° – 70°N	—	estimated area × literature emission rates	42
Crill <i>et al.</i> (1988)	all bogs > 40°N	Minnesota peatlands	summer flux time series at permanent sites	72
Whalen & Reeburgh (1990)	Arctic tundra	U. of Alaska Arboretum	annual flux time series at permanent sites	19 – 32
Aselmann & Crutzen (1989)	wetlands 50° – 70°N	—	estimated area × literature emission rates	23
Moore <i>et al.</i> (1990)	northern fens	Quebec, Canada	summer flux time series at permanent sites	14
Whalen & Reeburgh (1990)	Arctic tundra	Alaska	flux transect, Alaska Pipeline Haul Road	38 – 53
	taiga	Alaska	flux transect, Alaska Pipeline Haul Road	15
Fung <i>et al.</i> (1991)	wetlands > 50°N	—	3-D tracer model, estimates sources (scenario 7)	35
Whalen <i>et al.</i> (1991)	moist taiga	Bonanza Creek Exptl. Forest, Alaska	thaw season flux time series at permanent sites	-0.8
Whalen & Reeburgh (1992)	Arctic tundra	U. of Alaska Arboretum	annual flux time series (4 years of data)	42 ± 26
Bartlett <i>et al.</i> (1992)	ABLE 3A	Yukon- Kuskokwim Delta, Alaska	summer flux time series at permanent sites	11 ± 3
Ritter <i>et al.</i> (1992)	ABLE 3A	Yukon- Kuskokwim Delta, Alaska	aircraft boundary layer flux	44
Bartlett & Harriss (1993)	wetlands > 45°N	—	global summary	38

Present and Future Work

North America. *BOREAS*. The Boreal Ecosystem Atmosphere Study (BOREAS) is planned for 1994 at two locations in the boreal forest region of northern Canada. The objectives of the experiment are (1) to improve the understanding of the processes and states which govern the exchanges of energy, water, carbon and trace gases between boreal forest ecosystems and the atmosphere, with particular reference to those processes and states that may be sensitive to global climate change, and (2) to develop and validate remote sensing algorithms to extend this understanding of the processes from local to

regional scales. The focus of the research will be two 600 km² sites: one dry, warm boreal forest site located near Prince Albert in central Saskatchewan, and the other, a moist, cool boreal forest site located near Thompson, Manitoba. A unique aspect of BOREAS is the concurrent measurement of trace gases, tentatively CO, CH₄, some NMHCs, N₂O and NO_x, and all important physical, biogeochemical and ecological variables at all the important ecosystems that comprise what would be considered the boreal biome. Using this approach, sites of both net emissions and uptake of trace gases can be quantified and the total influence of this portion on the boreal biome can be assessed. The ecosystems include various forest types (aspen, spruce, pine), wetlands, and beaver ponds. Measurements will consist of a full set of chamber, micrometeorological, and aircraft derived fluxes. The experiment will run from snowmelt to freeze-up in 1994, with three intensive measurement periods where all three types of flux measurements will be conducted simultaneously. One tower at each site will measure fluxes from August, 1993 to September, 1994.

ARCSS/LAII Flux Study. The NSF-sponsored Arctic System Science Land-Air-Ice Interactions Flux Study (ARCSS/LAII Flux Study) will be conducted along a transect in Alaska from Toolik Lake in the Brooks Range foothills to the Arctic Coastal plain from 1993 to 1996. The goal of this study is to understand the processes controlling carbon storage and emission as CO₂ and CH₄. This study will involve measurements of emission and consumption of CO₂ and CH₄ using micrometeorological tower and chamber studies, coupled with aircraft flux measurements at intervals during one field season. The gas flux studies are integrated with hydrologic and soil studies, vegetation mapping and spectral reflectance measurements for interpretation of satellite data in this multi-year project.

Scandinavia. Several programs examining trace gas fluxes are underway in Scandinavia and are projected to continue until 1998. A Swedish effort to quantify CH₄ fluxes from peatlands (Mikkela *et al.*, 1992) also involves studies on the rates and controls on CH₄ production and oxidation (Svensson and Sundh, 1992). For example, Sundh *et al.* (1992) found that the oxidative capacity of a peatland is a direct function of the moisture content of the peatland. Other research is being conducted on the importance of peat quality on CH₄ flux (Nilsson, 1992) and the development of techniques such as FTIR for the measurement of the CH₄ flux.

Part of the Finnish Programme on Climate Change (SILMU) is examining the carbon balance of peatlands and climate change (Lainen and Päivänen, 1992). This experiment has been underway for three years and will continue for at least another three years. Chamber measurements of CO₂, CH₄ and N₂O flux are being made in areas subjected to various treatments and manipulations to determine the controls on trace gas emissions (Lien *et al.*, 1992; Martikainen *et al.*, 1992; Silvola *et al.*, 1992). The project includes experiments related to gas exchange from northern wetlands such as the emissions of CH₄ from virgin and drained peatlands, the effect of drainage intensity on carbon and nitrogen mineralization, rate of net peat accumulation, response of Sphagna to N deposition and increased CO₂ concentrations, and organic carbon in peatland waters. A particularly unique aspect of these experiments is the use of forest drainage sites as surrogates for climate change.

Russia. Several groups have initiated CH₄ flux studies in various parts of Russia. Panikov and Zelenev (1992) are working in Western Russia, and Samarkin *et al.* (1992) are involved in studies in the Kolyma lowland. A program of CO₂ flux measurements is underway near Chernski, Yakutia (S. Zimov, private communication), and a CH₄ flux measurement program is being established there. Established field stations and ongoing expeditions in these areas should facilitate collaborative flux time series studies.

CONTROLS ON HIGH LATITUDE TRACE GAS FLUXES (Task 2)

We understand in broad qualitative terms that CH₄ emissions result from a balance between production (methanogenesis) and consumption (methylotrophy). Physical factors that directly affect this balance are temperature, moisture, water table level, and substrate quality, but an understanding of the most important controls on CH₄ flux, the importance of feedbacks, and generally, the sensitivity of high latitude environments to altered climate remain unclear. Studies have concentrated on physical factors such as soil temperature because it is related to microbial activity, and the position of the water table, because it defines the zones of anaerobic (production) and aerobic (oxidation) processes. These are also the simplest variables to measure continuously.

Examples of relationships between CH₄ flux and soil temperature are summarized in Bartlett and Harriss (1993). A number of relationships have been reported for short term flux data sets using soil temperatures at fixed depths, but the number and variety of functions used (linear, polynomial, log-linear, and exponential), suggests that the relationship between temperature and CH₄ flux is not unique or straightforward (Whalen and Reeburgh, 1992).

Laboratory studies on the effect of water table level on CH₄ and carbon dioxide emission have been performed by Peterson *et al.* (1984) and Moore and Knowles (1987). Lowered wetland water tables are believed to control CH₄ fluxes by enhancing aerobic CH₄ oxidation. A lowered water table provides a more gas permeable surface zone that becomes aerobic through diffusive transport of atmospheric oxygen. This permits establishment of microbial populations capable of oxidation of CH₄ escaping from the deeper, waterlogged sediment layers. Whalen *et al.* (1993) showed that waterlogged tundra soils were anoxic only a few millimeters below the water table surface. Jar experiments with moist soils (Whalen and Reeburgh, 1990a; Whalen *et al.*, 1990) showed rapid equilibration with added tracers, suggesting gas phase diffusivities of $\sim 10^{-1}$ to 10^{-2} cm² sec⁻¹, 10^4 times more rapid than aqueous phase molecular diffusion. The CH₄ oxidizing populations responded rapidly to changes in CH₄ concentration and consumed CH₄ to threshold values of 0.1 ppm (Whalen *et al.*, 1992). Methane oxidation rates were very sensitive to water content, apparently a result of limiting transport by sealing inter-grain boundaries. The initial moist soil work suggested that oxidation could provide a rapid response and a negative feedback to increases in CH₄ flux.

The above rapid responses have not been observed in wetlands with waterlogged soils. Subsoil CH₄ distributions and fluxes appear to be decoupled, as high standing stocks of soil CH₄ and steep soil water CH₄ gradients show poor correlations with surface fluxes (Whalen and Reeburgh, 1992). Correlations between CH₄ flux and water table level in wetlands are generally poor, and appear to be masked in whole-season correlations (Whalen and Reeburgh, 1992). One possible explanation is that the CH₄ emission response occurs too rapidly to appear in whole-season correlations. The magnitude of CH₄ standing stocks in soil waters compared with the surface CH₄ fluxes suggests residence times of days. However, laboratory water table manipulation experiments on peat cores give results that suggest much slower responses. Moore and Roulet (1993) showed very dissimilar CH₄ flux responses under falling and rising water table conditions, and similar studies by Kizer and coworkers (private communication) showed that the CH₄ fluxes of cores subjected to water table adjustments stabilize only after periods of weeks. In contrast to the rapid response by moist soils, saturated or waterlogged soils show responses that may be related to flooding history or frequency. For example, in a falling water table situation, the newly exposed soils have only a history of anoxic conditions, and some time may be required to repopulate the zone with aerobic oxidizers. Reconciling the results of wetland field

observations with the laboratory core manipulations is an important problem that will require further study.

Whalen and Reeburgh (1992) noted that relationships between any single variable and CH₄ emission are site specific and of little value as predictors, emphasizing the fact that variables such as temperature and water table depth are not truly independent. They showed that parameters that integrate conditions over a season, such as thaw depth or centimeter-degrees, correlated best with CH₄ emission on a given date. Dise *et al.* (1993) showed that the position of the water table explained over 60% of the variance in CH₄ flux in Minnesota wetlands using multiple regression. Temperature explained 30% of the variance and a measure of substrate quality explained the remainder. Moore and Roulet (1993) showed that the response of CH₄ flux to a unit change in water table is similar among wetlands, but that the actual flux for a given water table varies considerably. Systems with small changes in water table level over the study period show no significant relationship between CH₄ flux and water table (Moore *et al.*, 1994). When this is the case, peat temperature dominates the system (Roulet *et al.*, 1992c).

The above attempts are directed toward a process-level understanding of the factors controlling CH₄ flux. Considering parameters like biomass and primary production rather than variables like temperature and water table level appears to offer a promising means of predicting CH₄ flux from some wetland systems. Recent work in diverse wetland systems shows strong linear correlations between CH₄ emission and biomass (Whiting *et al.*, 1991) and CH₄ emission and primary production (Aselmann and Crutzen, 1989; Whiting and Chanton, 1993). All sites considered were inundated and populated with plants capable of vascular transport. Concurrent measurements of net primary production and CH₄ flux showed that approximately 3% of daily net primary production was released as CH₄ (Whiting and Chanton, 1993). The correlations between biomass and primary production could result from either litter input or production of suitable methane-forming substrates below-ground by root exudation and root turnover, or they could be a result of the plants providing conduits for CH₄ emission. This relationship offers the potential of estimates of CH₄ emission through remote sensing of either photosynthetic activity or net primary production. Further work suggested by the net primary production/net CH₄ emission relationship includes determination of the origin and age of the methane-forming substrate, the role of below-ground CH₄ oxidation (Reeburgh *et al.*, 1993), and the applicability of the relationship to non-inundated wetlands free of vascular plants.

SENSITIVITY OF HIGH LATITUDE SYSTEMS TO CLIMATE CHANGE (Task 3)

Most climate models predict a large increase in winter temperature, and a moderate increase in summer temperature for northern regions. The predicted change in precipitation varies among models. Based on the output of the climate models, the summer climate of the northern wetland region is predicted to be warmer and slightly wetter, and it is not unreasonable to expect an increase in the annual emissions of CH₄ from northern wetlands. However, the biospheric feedbacks incorporating climate, hydrological and thermal regimes of wetlands, and the flux of CH₄ from wetlands are quite complicated and poorly understood. While an increase in summer precipitation would suggest an increase in the wetness of wetlands, the accompanying increase in temperature would also increase evaporative loss of water from wetlands. The direction of change in water storage in a wetland will be governed by the difference between precipitation input and evaporative and runoff outputs.

Models based solely on an increase in temperature suggest that the emissions of CH₄ from northern wetlands should increase in a 2 × CO₂ climate. Early models of Hameed and

Cess (1983) and Lashoff (1989) predicted an increase in CH_4 emissions with an increase in temperature. Harriss and Frolking (1992) developed a simple model based on the regressions of CH_4 flux and temperature developed by Crill *et al.* (1988) for peatlands in Minnesota. Using year-to-year temperature anomalies of $\pm 2^\circ\text{C}$ observed over the 20th century for five northern region wetlands, they estimate a $\pm 5 \text{ Tg yr}^{-1}$ variation in emissions. This deviation represents only 15% of the total emissions from northern wetlands for the expected range of temperature increase predicted by most GCMs for a $2 \times \text{CO}_2$ climate and suggests only a moderate feedback.

The CH_4 temperature feedback models assume that the moisture status of the wetland remains the same. However, Roulet *et al.* (1992b) using a simple hydrothermal model for northern fens, estimated a decrease in moisture storage of between 60 to 200 mm in a $2 \times \text{CO}_2$ climate. The same model predicted a 0.8 to 2.0°C increase in the peat temperature. The increase in peat temperature resulted in a 15% increase in CH_4 flux, a result very similar to that of Harriss and Frolking (1992). However, the decrease in water storage resulted in an 80% decrease in CH_4 flux, suggesting that the emissions of CH_4 from northern wetlands are more sensitive to change in moisture than temperature. A small decrease in water storage in a wetland can easily offset any temperature-induced increase in emissions. The models used to estimate the CH_4 biospheric feedback for wetlands are preliminary and crude. Several studies have used empirical approaches to examine the climate- CH_4 feedback in wetlands, either by attempting to correlate seasonal and inter-annual variability in the CH_4 with climate variables, or by measuring the CH_4 flux from wetlands that have undergone some drainage. The results of two studies conducted in Alaska suggest that the CH_4 flux from wetlands is extremely sensitive to change in moisture and temperature. A four-fold variation in annual CH_4 emissions from wetlands in the interior of Alaska was observed over a four year period (Whalen and Reeburgh, 1992). Livingston and Morrissey (1991) compared the difference in the CH_4 flux from Alaskan North Slope wetland with varying moisture and temperature regimes between two years with very different climates. They found a four-fold increase in flux with a 4°C increase in soil temperature and an increase in wetness. They also found a two-fold increase in flux with a 4°C increase in soil temperature under drier conditions. These results suggest a far stronger sensitivity of the CH_4 flux to the climate than was found in modeling studies.

A criticism of the use of normal climate variability as a surrogate for climate change is that ecosystems should be well-adjusted to this type of variability. However, with a persistent change in climate, ecosystems should be pushed beyond the range of normal variability. To find the effect of persistent change in the moisture, Roulet *et al.* (1993) measured the flux of CH_4 from several wetlands that have been drained. These wetlands experience the day-to-day natural variability in temperature and precipitation, but have a lower mean water table than natural wetlands. Their results show that even a slight drop in water table of 0.1 m effectively eliminates the flux. The flux decreased from 5 to $10 \text{ mg CH}_4 \text{ m}^{-2} \text{ d}^{-1}$ with an average water table of -0.2 m to between 0.1 and $-0.4 \text{ mg CH}_4 \text{ m}^{-2} \text{ d}^{-1}$ when the average water table was below -0.3 m . An additional drop in water table to -1.0 m did not increase CH_4 uptake. Whalen *et al.* (1993) used kinetic data on CH_4 oxidizers in soil cores to estimate that lowering the water table 10 to 40 cm could result in a zero net CH_4 flux.

All of the studies of biospheric feedback of CH_4 assume that there is no change in the structure and function of the wetland ecosystem. This is clearly a limiting assumption. Vegetation changes were obvious at the forest drainage sites used by Roulet *et al.* (1993) only after 10 years of drainage (J. Bubier, private communication). With changes in structure, changes in substrate quality and transport mechanism should be expected. Without an understanding of how the plant communities of wetlands respond to changes in

the climate it will be impossible to make any long-term predictions of future emissions from northern wetlands.

FUTURE STUDIES

A remarkable amount of data on trace gas emissions from high latitude systems has been obtained in a relatively short time, and the importance of these systems as global CH₄ sources is converging on a value between 20 and 45 Tg yr⁻¹. A first-order understanding of biogeochemical processes controlling the balance between CH₄ production and emission has resulted, and models dealing with sensitivity of emissions to various observed parameters are under development. Future work should emphasize the following.

Our current information on high latitude trace gas fluxes has a strong North American bias, and has placed almost total emphasis on CH₄ emissions. Studies in northern European and Siberian wetlands are beginning, and should be supported with collaborative studies and flux time-series measurements. Obtaining flux data from these systems, based on accepted site classification and nomenclature used in the other high latitude studies, is essential to understanding the global importance of high latitude systems. It is important that future work emphasize measurements of not only CH₄, but also of other gases, particularly CO₂.

Estimates of the sensitivity of high latitude systems to climate change are only beginning. Most use historical climate data (temperature, rainfall) and simple relationships to flux. We appreciate that none of the driving variables are independent and that a single process-level parameter capable of predicting CH₄ flux is unlikely. Feedbacks promise to be an important control and require more study. Realistic manipulation experiments on the 1 to 10 m scale are needed to confirm process-level controls and their role in feedbacks. Preliminary experiments involving manipulation of water table level and temperature are underway and should be repeated in a range of environments. The design and eventual results of these experiments are critical to scaling models of biogeochemical feedbacks to general circulation models and chemical tracer models.

Understanding microbially mediated CH₄ production and oxidation in wetlands and soils should be a high priority. Understanding the origin of the substrates important in CH₄ production are needed for a process-level understanding of the methane flux:primary production relationship. There are very few direct measurements of CH₄ oxidation, but the available information suggests that oxidation may be about half of the gross CH₄ budget (Reeburgh *et al.*, 1993)

REFERENCES

- Aselmann, I. and P.J. Crutzen, 1989, Global distribution of natural freshwater wetlands and rice paddies, and their net primary productivity, seasonality and possible methane emissions, *J. Atmos. Chem.*, 8:307-358.
- Bartlett, K.B., P.M. Crill, R.L. Sass, R.C. Harriss and N.B. Dise, 1992, Methane emissions from tundra environments in the Yukon-Kuskokwim Delta, Alaska, *J. Geophys. Res.*, 97:16,645-16,660.
- Bartlett, K.B., and R.C. Harriss, 1993, Review and assessment of methane emissions from wetlands, *Chemosphere*, 26:261-320.
- Billings, W.D., 1987, Carbon balance of Alaskan tundra and taiga ecosystems: Past, present and future, *Quat. Sci. Rev.*, 6:165-177.
- Cicerone, R.J. and R.S. Oremland, 1988, Biogeochemical aspects of atmospheric methane, *Global Biogeochem. Cycles*, 2: 299-327.
- Clymo, R.S. and E.J.F. Reddaway, 1971, Productivity of sphagnum (bog moss) and peat accumulation, *Hidrobiologia*, 12: 181-192.

- Crill, P.M. et al, 1988, Methane flux from Minnesota peatlands, *Global Biogeochem. Cycles*, 2: 371-384.
- Dise, N.B., 1992, Winter fluxes of methane from Minnesota peatlands, *Biogeochemistry*, 17:71-83.
- Dise, N., E. Gorham and E.S. Verry, 1993, Environmental factors controlling methane emissions from peatlands in northern Minnesota, *J. Geophys. Res.*, 98:10,583-10,594.
- Edwards, G.E., H.H. Neumann, G. den Hartog, G.W. Thurtell and G. Kidd, 1994, Eddy correlation measurements of methane fluxes using a tunable diode laser at the Kinosheo Lake tower during the Northern Wetlands Study (NOWES), *J. Geophys. Res.*, 99:1511-1518.
- Fan, S.M., S.C. Wofsy, P.S. Bakwin, D.J. Jacob, S.M. Anderson, P.L. Kebejian, J.B. McManus and C.E. Kolb, 1992, Micrometeorological measurements of CH₄ and CO₂ exchange between the atmosphere and subarctic tundra, *J. Geophys. Res.*, 97:16,627-16,643.
- Fung, I., J. John, J. Lerner, E. Matthews, M. Prather, L.P. Steele and P.J. Fraser, 1991, Three-dimensional model synthesis of the global methane cycle, *J. Geophys. Res.*, 96:13,033-13,065.
- Hameed, S. and R. Cess, 1983, Impact of global warming on biospheric sources of methane and its climate consequences, *Tellus*, 35B:1-7.
- Hamilton, J.D., C.A. Kelly, J.W.M. Rudd and R.H. Hesslein, 1994, Flux to the atmosphere of CH₄ and CO₂ from wetland ponds on the Hudson Bay Lowland (HBL), *J. Geophys. Res.*, 99:1495-1510.
- Harriss, R.C. and S. Frohling, 1993, The sensitivity of methane emissions from northern freshwater wetlands to global change, in: "Global Warming and Freshwater Ecosystems," P. Firth and S. Fisher (eds.), Springer-Verlag, New York, pg. 48, in press.
- Jobson, B.T., Z. Wu, H. Niki and L.A. Barrie, 1994, Seasonal trends of isoprene, C₂-C₅ alkanes, and acetylene at a remote boreal site in Canada, *J. Geophys. Res.*, 99:1589-1600.
- Khalil, M.A.K. and R.A. Rasmussen, 1989, Climate-induced feedbacks for the global cycles of methane and nitrous oxide, *Tellus*, 41B:554-559.
- Klinger, L.F., P.R. Zimmerman, J.P. Greenberg, L.E. Heidt, and A.B. Guenther, 1994, Carbon trace gases along a successional gradient in the Hudson Bay Lowland, *J. Geophys. Res.*, 1469-1494.
- Lachenbruch, A.H. and B.V. Marshall, 1986, Climate change: Geothermal evidence from permafrost in the Alaskan arctic, *Science*, 34:689-696.
- Lainen, J. and J. Päivänen, 1992, Carbon balance of peatlands and global climatic change: Summary, in: "Carbon Balance of Peatlands and Climate Change," Progress Report, SIMLU, Helsinki.
- Lashoff, D.A., 1989, The dynamic greenhouse: Feedback processes that may influence future concentrations of atmospheric trace gases and climate change, *Clim. Change*, 14:213-242.
- Lein, T., P. Martikainen, H. Nykänen and L. Bakken, 1992, Methane oxidation and fluxes in two drained peat soils, *SUO*, 43:231-236.
- Livingston, G.P. and L.A. Morrissey, 1991, Methane emissions from Alaskan arctic tundra in response to climatic change, in: "Role of Polar Regions in Global Change," G.L. Weller, C.L. Wilson and B.A.B. Severin (eds), University of Alaska, Fairbanks, pg. 372.
- Martikainen, P., H. Nykänen, P. Crill and J. Silvola, 1992, The effect of changing water table on methane fluxes at two Finnish mire sites, *SUO*, 43:237-240.
- Matthews, E. and I. Fung, 1987, Methane emission from natural wetlands: Global distribution, area, and environmental characteristics of sources, *Global Biogeochem. Cycles*, 1:61-86.
- Mitchell, J.F.B., 1989, The greenhouse effect and climate, *Rev. Geophys.*, 27:115-139.
- Mikkilä, C., I. Sundh, J. Eilertsson, B. Svensson and M. Nilsson, 1992, Methane emissions from a Swedish peatland area: Temporal and spatial variations, in: "Proceedings of the 9th International Peat Congress," D. Fredericksson (ed.), 3:152-165.
- Moore, K.E., D.A. Fitzjarrald and J.A. Ritter, 1993, How well can regional fluxes be derived from small-scale estimates?, *J. Geophys. Res.*, 98:7187-7198.
- Moore, T.R. and R. Knowles, 1989, The influence of water table levels on methane and carbon dioxide emissions from peatland soils, *Can. J. Soil Science*, 69:33-38.
- Moore, T.R., N. Roulet and R. Knowles, 1990, Spatial and temporal variations of methane flux from subarctic/northern fens, *Global Biogeochem. Cycles*, 4:29-46.
- Moore, T.R. and N.T. Roulet, 1993, Methane flux: water table relations in northern wetlands, *Geophys. Res. Lett.*, 20:587-590.
- Moore, T.R., A. Heyes and N.T. Roulet, 1994, Methane emission from wetlands of the southern Hudson Bay lowland, *J. Geophys. Res.*, 19:1455-1468.
- Morrissey, L.A. and G.P. Livingston, 1992, Methane emissions from Alaska arctic tundra: An assessment of local spatial variability, *J. Geophys. Res.*, 97:16,661-16,670.
- Nilsson, M., 1992, "Fungi and Bacteria in Peat and Peat-forming Plant Communities," Diss. Swedish Univ. Agric. Sciences, Dept. of Forest Site Research, Stencil #20.

- Oechel, W.C., S.J. Hastings, G. Vourlitis, M. Jenkins, G. Reichers and N. Grulke, 1993, Recent change of arctic tundra ecosystems from a net carbon dioxide sink to a source, *Nature*, 361:520-523.
- Panikov, N.S. and V.V. Zelenev, 1992, "Emission of CO₂ and CH₄ from Northern Wetlands to the Atmosphere: Dynamics, Controlling Factors, and Tentative Mechanisms," Russian Academy of Sciences, Puschino, Moscow Region.
- Peterson, K.M., W.D. Billings and D.N. Reynolds, 1984, Influence of water table and atmospheric CO₂ concentration on the carbon balance of arctic tundra, *Arctic and Alpine Res.*, 16:331-335.
- Post, W.M., W.R. Emanuel, P.J. Zinke and A.G. Stangenberger, 1982, Soil carbon pools and world life zones, *Nature*, 298:156-159.
- Reeburgh, W.S. and S.C. Whalen, 1992, High latitude ecosystems as CH₄ sources, *Ecol. Bull. (Copenhagen)*, 42:62-70.
- Reeburgh, W.S., S.C. Whalen and M.J. Alperin, 1993, The role of methylotrophy in the global methane budget, in: "Microbial Growth on C-1 Compounds," J.C. Murrell and D. Kelley (eds.), Intercept, Andover, pgs. 1-14.
- Ritter, J.A., J.D.W. Barrick, G.W. Sachse, G.L. Gregory, M.W. Woerner, C.E. Watson, G.F. Hill and J.E. Collins, Jr., 1992, Airborne flux measurements of trace species in the arctic boundary layer, *J. Geophys. Res.*, 97:16,601-16,625.
- Ritter, J.A., J.D.W. Barrick, C.E. Watson, G.W. Sachse, G.L. Gregory, B.E. Anderson, M.A. Worner and J.E. Collins, 1994, Airborne boundary layer flux measurements of trace gas species over Canadian boreal forest and northern wetland regions, *J. Geophys. Res.*, 99:1671-1686.
- Roulet, N.T., L. Barrie, W.A. Glooschenko, R.C. Harriss, J.I. MacPherson, T.R. Moore, H. Neuman and H.I. Schiff, 1992a, Northern Wetlands Study (NOWES): An assessment of the role of northern wetlands in the exchange of atmospheric trace gases, *Ecol. Bull. (Copenhagen)*, 42:77-85.
- Roulet, N.T., T.R. Moore, J. Bubier and P. LaFleur, 1992b, Northern fens: Methane flux and climatic change, *Tellus*, 42B:100-105.
- Roulet, N.T., R. Ash and T. Moore, 1992c, Low boreal wetlands as a source of atmospheric methane, *J. Geophys. Res.*, 97:3739-3749.
- Roulet, N.T., R. Ash, W. Quinton and T.R. Moore, 1993, Methane flux from drained northern peatlands: An analog for persistent water table lowering, *Global Biogeochem. Cycles*, 7:749-770.
- Roulet, N.T., A. Jano, C.A. Kelly, T.R. Moore, R. Protz, J.A. Ritter and W.R. Rouse, 1994, The Role of the Hudson Bay Lowland as a source of atmospheric methane, *J. Geophys. Res.*, 99:1423-1428.
- Samarkin, V.A., E.M. Rivkina and E.A. Bochamnikova, 1992, "Methane Biogeochemistry in Arctic Marsh Soils," Russian Academy of Sciences, Puschino, Moscow Region.
- Schiller, C.L. and D.R. Hassie, 1994, Emissions of nitrous oxide from the Hudson Bay Lowland, *J. Geophys. Res.*, 99:1573-1588.
- Sebacher, D.I., R.C. Harriss, K.B. Bartlett, S.M. Sebacher and S.S. Grice, 1986, Atmospheric methane sources: Alaskan tundra bogs, an alpine fen and a subarctic boreal marsh, *Tellus*, 38B:1-10.
- Silvola, J., P. Martikainen and H. Nykänen, 1992, A mobile automatic gas chromatograph to measure CO₂, CH₄ and N₂O fluxes from soil in the field, *SUO*, 43:263-266.
- Sundh, I., C. Mikkilä, M. Nilsson and B.H. Svensson, 1992, Potential methane oxidation in a Sphagnum peat bog: Relation to water table level and vegetation type, in: "Proceedings of the 9th International Peat Congress," D. Fredriksson (ed), 3:142-151.
- Svensson, B.H. and T. Rosswall, 1976, Methane production in tundra peat, in: "Microbial Production and Utilization of Gases (H₂, CH₄, CO)," H.G. Schelgel, G. Gottschalk, N. Pfennig, E. Goltze (eds.), Göttingen, pgs. 135-139.
- Svensson, B.H. and T. Rosswall, 1984, In situ methane production from acid peat communities with different moisture regimes in a subarctic mire, *Oikos*, 43:341-350.
- Svensson, B.H. and I. Sundh, 1992, Factors affecting methane production in peat soils, *SUO*, 43:183-190.
- Whalen, S.C. and W.S. Reeburgh, 1988, A methane flux time series for tundra environments, *Global Biogeochem. Cycles*, 2:399-409.
- Whalen, S.C. and W.S. Reeburgh, 1990a, Consumption of atmospheric methane by tundra soils, *Nature*, 346:160-162.
- Whalen, S.C. and W.S. Reeburgh, 1990b, A methane flux transect along the trans-Alaska pipeline haul road, *Tellus*, 42B:237-249.
- Whalen, S.C., W.S. Reeburgh and K.A. Sandbeck, 1990, Rapid methane oxidation in a landfill cover soil, *Appl. Environ. Microbiol.*, 56:3405-3411.
- Whalen, S.C. and W.S. Reeburgh, 1992, Interannual variations in tundra methane emission: A four-year time series at fixed sites, *Global Biogeochem. Cycles*, 6:139-159.

- Whalen, S.C., W.S. Reeburgh and V. A. Barber, 1992, Oxidation of methane in boreal forest soils: a comparison of seven measures, *Biogeochemistry*, 16:181-211.
- Whalen, S.C., W.S. Reeburgh and C.E. Reimers, 1993, Processes controlling methane fluxes from tundra environments, in: "Landscape Function: Implications for Ecosystem Response to Disturbance. A case Study in Arctic Tundra," J.F. Reynolds and J.D. Tenhunen (eds.), Springer-Verlag, New York, in press.
- Whiting, G.J., J.P. Chanton, D.S. Bartlett and J.D. Happell, 1991, Relationships between CH₄ emission, biomass, and CO₂ exchange in a subtropical grassland, *J. Geophys. Res.*, 96:13,067-13,071.
- Whiting, G.J. and J.P. Chanton, 1993, Primary production control of methane emission from wetlands, *Geophys. Res. Lett.*, submitted.