1	Eco-evoluti	onary dynamics of interference competition	
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37 Abstract

Theorists have identified several mechanisms through which species that compete exploitatively 38 39 for resources could coexist. By contrast, under current theory, interference competitors could 40 coexist only in rare circumstances. Yet, some types of interference competition, such as 41 interspecific territoriality, are common. This mismatch between theory and nature inspired us to model interference competition in an eco-evolutionary framework. We based the model on the 42 life cycle of territorial birds, and ran simulations to examine whether natural selection could 43 rescue a superior interference competitor from extinction without driving a superior exploitative 44 45 competitor extinct. We found that coexistence between interference competitors can occur over a 46 wide range of ecological plausible scenarios, and up to the highest levels of resource overlap. An important caveat is that coexistence requires the species to coevolve. Reductions in population 47 48 size and levels of genetic variation could destabilize coexistence between interference 49 competitors, and thereby increase extinction rates over current estimates.

50 Introduction

51 The term "competition" encompasses two rather distinct categories of ecological interactions. Exploitative competition occurs indirectly through depletion of a shared, limiting resource 52 (Krebs 2001). Interference competition is any costly interaction between individuals over access 53 54 to a resource, aside from resource depletion itself, regardless of whether the resource is shared or 55 limiting (Krebs 2001). Aggression and allelopathy are probably the most common types of interference competition in animals and plants, respectively (Inderjit et al. 2011; Meiners et al. 56 57 2012; Grether *et al.* 2013). Interference competition occurs in microbes as well, through stabbing 58 or poisoning of neighboring cells (Cordero & Datta 2016; Ghoul & Mitri 2016; García-Bayona & Comstock 2018). Competition of any kind could prevent species from coexisting, but the 59 60 ecological literature is replete with models showing, collectively, that species that compete 61 exploitatively can coexist under a variety of circumstances (Chesson 2000; HilleRisLambers et al. 2012). A consistent theme emerging from these models is the central role of trade-offs 62 63 facilitating coexistence. By contrast, there have been very few attempts to incorporate interference competition into coexistence theory. Case & Gilpin (1974) added interference 64 competition parameters to Lotka-Volterra competition equations and showed that coexistence 65 66 between interference competitors is theoretically possible, but they did not clarify the 67 circumstances under which this might occur. Using a mechanistic model, Vance (1984) showed 68 that a superior interference competitor can drive a superior exploitative competitor extinct, and 69 that the starting population densities can determine which species prevails. For coexistence to 70 occur, interference competition has to be strong enough to override the superior exploitative 71 competitor's advantage, and yet each species has to reduce its own resource encounter rate much 72 more than the other species' (Vance 1984), which seems rather improbable (there appear to be no documented empirical examples). Based on a two-consumer model with explicit resource
dynamics, Amarasekare (2002) concluded that coexistence cannot occur under interference
competition, even when one species is a superior exploitative competitor and the other is a
superior interference competitor, unless the superior interference competitor also consumes or
parasitizes the other species. Thus, species engaging only in costly forms of interference
competition, such as interspecific territoriality and allelopathy, were predicted to not be able to
coexist (Amarasekare 2002).

Yet, costly forms of interference competition are common in nature, as ecologists have long 80 81 been aware (Darwin 1859; Case & Gilpin 1974; Diamond 1978; Schoener 1983; Dobson 1985). 82 For example, nearly a third of all 322 species of passerine birds that breed in North America 83 defend territories against one or more other species (Drury et al. 2020), and roughly half of the sympatric sister species in this clade are interspecifically territorial (Cowen et al. 2020; Drury et 84 al. 2020). Ordinarily, mismatches between theory and nature would inspire theorists to question 85 86 the assumptions of their models and seek ways to bring them more closely in line with reality, but that has yet to occur for interference competition. Empirical research on this subject has 87 88 continued advancing, largely in a vacuum of formal theory (Grether et al. 2009, 2013, 2017; 89 Peiman & Robinson 2010; Freshwater et al. 2014; Martin & Ghalambor 2014; Tobias et al. 90 2014). For example, interspecific aggression appears to cause some species of rodents to 91 partition resources in ways that enable them to coexist (Ziv & Kotler 2003; Levy et al. 2011; 92 Pasch et al. 2013). As another example, interspecifically territorial birds and damselflies show 93 clear evidence of having evolved in response to each other (Tobias et al. 2014; Drury et al. 2020; 94 Grether et al. 2020; McEachin et al. 2021), which indicates that they have been in contact for 95 many generations.

96 Different types of interference competition probably have different effects on coexistence. 97 Recent theoretical and empirical advances have established that allelopathy might actually 98 promote coexistence in plants (Hierro & Callaway 2021). Here we focus on interspecific 99 territoriality between phenotypically similar species. This effectively narrows the scope to 100 animals, but within animals, territoriality is common in many taxonomic groups, including 101 insects, crustaceans, gastropods, fishes, amphibians, reptiles, birds, and mammals. Territoriality is a form of social dominance in which the resident individual has priority of access to resources 102 103 or mates at a particular location. One important but often overlooked feature of interspecific 104 territoriality is that it reduces interspecific exploitative competition by reducing spatial overlap. It is a mutually costly interaction only in the sense that the population mean fitness of both 105 106 species would be higher in the other species' absence (Grether et al. 2017). At the individual level, territoriality has winners and losers, with the interaction rarely being symmetrical between 107 108 species (Martin et al. 2017).

109 Asymmetries in interference competition should be able to mediate trade-offs that promote coexistence between resource competitors. For example, if species 1 can profitably exclude 110 species 2 from high-suitability habitat patches, but species 2 is a superior exploitative competitor 111 112 that can persist in habitat patches too low in suitability to support species 1, it follows that there 113 must be a region of parameter space within which the species can coexist indefinitely. This is 114 essentially the situation hypothesized to enable pied flycatchers and collared flycatchers to 115 coexist in a mosaic of deciduous and coniferous forest (Qvarnström et al. 2009; Veen et al. 116 2010). However, our goal was not to evaluate whether interspecifically territorial species could 117 coexist in some static region of parameter space, but instead to determine whether species could 118 evolve into that parameter space, under realistic conditions, and remain there indefinitely. More

specifically, we sought to determine whether a superior interference competitor could be rescued 119 120 from extinction by agonistic character displacement, without driving a superior exploitative 121 competitor extinct. Agonistic character displacement (ACD) can be defined as phenotypic evolution caused by interspecific interference competition (Grether et al. 2009). ACD can cause 122 species to diverge or converge in competitor recognition and the traits used to recognize 123 124 competitors, depending on whether recognizing individuals of other species as competitors is adaptive (i.e., increases an individual's fitness) (Grether et al. 2009). If our hypothetical species 125 126 came into secondary contact and were similar enough phenotypically to recognize each other as 127 competitors but were not actually in competition for limiting resources, they would be expected to diverge in competitor recognition until interference competition was eliminated. If instead the 128 129 species overlapped substantially in resource use, interspecific territoriality might be adaptive for species 1 (the superior interference competitor) but not for species 2 (the superior exploitative 130 competitor), in which case species 1 would be under selection to converge while species 2 was 131 132 under selection to diverge, potentially resulting in species 1 chasing species 2 through evolutionary time. Could species 1 converge rapidly enough to avoid extinction, and if so, would 133 134 species 2 go extinct, or would the species' population sizes stabilize in a type of coevolutionary 135 stalemate or zero-sum game?

To determine what is theoretically possible while staying within the bounds of biological realism, we constructed an individual-based eco-evolutionary model based on the life cycle of territorial passerine birds, with survivorship and reproductive parameters from the published literature on pied and collared flycatchers (Qvarnström *et al.* 2009; Veen *et al.* 2010). In secondary contact simulations, we varied the level of dietary overlap, the proportions of highand low-suitability habitat, the mean difference between the species in territorial fighting ability,

and the initial level of phenotypic divergence, to examine how these factors influence
community composition over evolutionary timescales. The results show that ACD can indeed
rescue a superior interference competitor from extinction without driving the other species
extinct, but also that other outcomes are possible, depending on the parameter settings and
chance events (e.g., mutation and genetic drift).

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148 Material and methods

149 The model. Here we provide a conceptual overview; a detailed mathematical description of the 150 model is provided in the Supporting Information. Interference competition is mechanistically explicit in the model, but exploitative competition is modeled implicitly. The exploited resources 151 152 could represent any entity, the depletion of which reduces recruitment to the next generation. For *Ficedula* flycatchers, there is evidence that nest cavities are the key limiting resource (von 153 Haartman 1956; Vallin et al. 2012), but for generality we consider the limiting resource to be 154 155 food. Territory competition and reproduction occur during an annual breeding season. The breeding habitat is divided into discrete territories. If more than one breeding pair of the same 156 157 species occupies a territory, the pairs compete exploitatively for food to feed their nestlings, 158 which reduces nestling survival. If breeding pairs of different species occupy the same territory, 159 they compete exploitatively in proportion to the level of dietary overlap between the species. To 160 represent ecologically relevant trade-offs, baseline nestling survival is higher in species 2 than in 161 species 1. We assume a spatially heterogeneous habitat in which territories vary in suitability. 162 Due to selection on habitat preferences in the evolutionary past (not modeled), individuals only 163 accept territories that exceed a suitability threshold, with species 1's suitability threshold higher

than species 2's. Thus, some territories serve as refuge for species 2, free from interspecificcompetition.

Territoriality is governed by four polygenic male traits: a phenotypic cue (z) that is visible or audible from a distance (e.g., a prominent color patch), the central location (μ) and width (σ) of a Gaussian recognition function (Okamoto & Grether 2013), and a fighting ability state variable. As the sender's value of z approaches the receiver's value of μ , the probability that the receiver recognizes the sender as a competitor increases. The width of the function (σ) determines how close the sender's value of z must be to the receiver's value of μ for a given probability of recognition.

173 Territory settlement occurs at the start of each annual breeding season. Males of both 174 species preferentially settle in unoccupied, high-suitability territories. Once the high-suitability 175 territories are occupied, males of species 2 settle in unoccupied refuge territories. Any males still without territories enter occupied territories and encounter residents. If neither male recognizes 176 177 the other as a competitor, they both stay. If one male recognizes the other as a competitor, they 178 fight, one male wins, the other male leaves, and the fighting ability of both males is reduced by 179 the same absolute amount. Residents have no inherent advantage over intruders. The probability 180 of the male with higher fighting ability winning increases asymptotically with the difference between them in fighting ability. Males without a territory keep trying to acquire one until the 181 182 territory settlement period ends. Then females arrive and pair up (monogamously) with 183 conspecific territory holders. Individuals that fail to secure breeding territories remain unmated 184 and do not compete with breeders. The fighting ability of surviving males is replenished between 185 breeding seasons.

By setting the baseline nestling survival of species 2 higher than that of species 1, we made species 2 the superior exploitative competitor. In the absence of interference territoriality, species 2 would out-reproduce species 1. If the species overlapped much in diet, species 2's population would grow at the expense of species 1's population, and thus species 1 would be driven to extinction.

191 To make species 1 the superior interference competitor, we set its baseline fighting ability 192 higher than that of species 2. With large differences between the species in fighting ability, 193 species 1 is always favored to win interspecific fights, but with smaller differences, a male of 194 species 1 that declined in fighting ability due to prior fights could be favored to lose.

The model captures key trade-offs that have been hypothesized to enable pied and collared flycatchers to coexist (Qvarnström *et al.* 2009; Veen *et al.* 2010). Species 1 represents collared flycatchers, species 2 represents pied flycatchers, and the habitat suitability threshold represents the ratio of deciduous to coniferous trees (which is related to food availability). However, our model is not meant to match the flycatcher system in every detail. Also, we varied parameters that are not known to vary in that system, for the purposes of reaching general conclusions about systems in which similar trade-offs might occur.

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Model implementation. We built the model on the sPEGG (simulating Phenotypic Evolution on
General Purpose Graphics Processing Units) modeling framework (Okamoto & Amarasekare
2018). sPEGG is a library of open-source code for explicitly tracking the survival, reproduction
and genotypes of individuals, with customizable modules for mutation, recombination,
migration, gene flow, etc. Individual-level processes (birth, death, behaviour, inheritance,

208 development) were characterized as realizations of stochastic processes. We customized sPEGG

for the life cycle and reproductive biology of passerine birds and added modules simulatinghabitat selection and territoriality.

211

Allopatric period and secondary contact. For the first 1000 generations (years) of each 212 213 simulation, populations of the two species breed in different communities with identical habitats. 214 Initially, both species are genetically homogeneous. As genetic variation builds gradually through mutation and recombination, the species evolve to their respective optima for σ and 215 216 degrees of separation between z and μ , and population sizes stabilize. The first generation is 217 highly territorial because z equals μ within species, but if territoriality proved maladaptive, traits z and μ would diverge from each other until the probability of males recognizing conspecific 218 219 males as competitors was negligible. Thus, intraspecific territoriality is not assumed by the 220 model, and would be lost if selection opposed it. It is even possible to have interspecific territoriality without intraspecific territoriality, if one species' mean value of μ diverged from its 221 222 mean value of z and in doing so converged on the other species' mean value of z, although it is 223 difficult to imagine a scenario in which selection would favor that outcome.

With the model parameterized with estimates of clutch size, nestling survival, and adult survival from the published literature on pied and collared flycatchers (Table S1), intraspecific territoriality is adaptive for both species, albeit more strongly so for species 1 than species 2, because species 1's baseline nestling survival is lower, and thus the fitness consequences of not being territorial are more severe for species 1. Starting in the 1001st year, individuals move from one community to the other between breeding seasons, with a probability of 0.5, which brings the species into contact. Simulations ran for 10001 years in total.

232	Parameter sets, model output, and outcome classifications. We varied dietary overlap (27		
233	levels), the fraction of refuge territories (14 levels), the species difference in fighting ability (6		
234	levels), and the initial phenotypic difference between the species in traits z and μ (2 levels), in a		
235	fully factorial design, for a total of 4536 parameter sets. Each parameter set was run multiple		
236	times (mean \pm s.d. = 13.5 \pm 1.6) with different random seeds (61386 simulations in total).		
237	Parameter settings remained constant for the duration of a simulation. The model recorded the		
238	species, sex, age, phenotypes, and nestling survival of all individuals in selected years.		
239	We considered the ecological outcome to be "coexistence" if both species persisted until the		
240	end of the simulation (9000 years). R scripts were used to classify the evolutionary outcome		
241	based on changes in the population means of z and μ (see Figs. S1-S7). We restricted the		
242	evolutionary outcome classifications to simulations in which both species persisted > 999 years		
243	because distinguishing between some alternative outcomes (e.g., unilateral divergence versus		
244	chasing) can be difficult over shorter time periods, especially in regions of parameter space		
245	where both outcomes are possible. Further information on the evolutionary outcome		
246	classifications is provided in the Supporting Information.		
247	Forking, defined as one species' mean values of z and μ being on opposite sides of the other		
248	species' mean values of z and μ , occurred in 1584 (2.6%) of the simulations. The usual cause		
249	was that z_2 and μ_2 separated somewhat during the allopatric period and were on opposite sides of		
250	z_1 and μ_1 when contact occurred (Fig. S8). We do not consider forking to be a plausible outcome		
251	in nature and therefore re-ran simulations in which forking occurred, using different random		
252	seeds.		
253			

254 Results

255 When the species were evenly matched in fighting ability, coexistence occurred only at relatively 256 low levels of dietary overlap, and if one species prevailed, it was always the superior exploitative 257 competitor (Fig. 1). Coexistence became progressively less likely as the fraction of the habitat 258 where the superior exploitative competitor could breed without interference from the other 259 species (henceforth, refuge habitat) increased (Fig. 1, upper left). In the region of parameter 260 space where coexistence occurred, one or both species diverged from the other in competitor recognition until interspecific interference competition was eliminated (Fig. 1, lower left; Figs. 261 262 S1-S3).

263 The eco-evolutionary picture changed rather dramatically when the inferior exploitative competitor was a superior interference competitor. The zone in parameter space where the 264 265 species coexisted was substantially larger, but also, at the lowest fractions of refuge habitat, the superior exploitative competitor was likely to go extinct, especially when dietary overlap was 266 high (Fig. 1, upper right). This was true even when the species difference in fighting ability was 267 268 quite small (Fig. S9, upper right). As the species difference in fighting ability increased, 269 coexistence became possible even with 100% dietary overlap and up to a refuge habitat fraction 270 of 0.85 (Fig. 1, upper right). When dietary overlap was high, the superior interference competitor 271 evaded competitive exclusion by chasing the superior exploitative competitor through 272 evolutionary time (Fig. 1, lower right; Figs. S4, S11, S12). As the difference between the species 273 in fighting ability increased, the level of dietary overlap at which the evolutionary dynamics 274 switched from divergence to chasing progressively decreased (Figs. S11 and S12). Unexpectedly, the relationship between the probability of coexistence and dietary overlap 275 276 was U-shaped when the species differed in fighting ability (Figs. 2, Fig. S15). The rarity of

277 coexistence at intermediate levels of dietary overlap arose from the high rate of extinction of the

278 superior interference competitor (Fig. S17), except at the lowest fractions of refuge habitat, 279 where the superior exploitative competitor was at greater risk of extinction (Fig. S19). To clarify 280 why coexistence occurred more often at high than intermediate levels of dietary overlap, we 281 examined all cases in which both species persisted for at least 1000 years after contact. This 282 revealed that extinction at intermediate levels of dietary overlap usually occurred while the 283 superior interference competitor was chasing the superior exploitative competitor (Fig. S13; for examples, see Figs. S5 and S6). Evidently, selection on the superior interference competitor to 284 285 recognize the superior exploitative competitor was not strong enough at intermediate levels of 286 dietary overlap to maintain interspecific territoriality and forestall competitive exclusion. The phenotypic similarity of the species at the time of secondary contact influenced the 287 outcome in predictable ways. In the simulations summarized in Figs. 1 and 2, the species were 288 initially 1 s.d. apart in the trait used for competitor recognition (z), which corresponds to a 289 heterospecific recognition probability of 0.85. Comparable results with an initial phenotypic 290 291 difference of 2 s.d. and heterospecific recognition probability of 0.25 are shown in Figs. S10, 292 S12 and S16. The lower probability of heterospecific recognition resulted in a broader zone in 293 which the species always coexisted when they were evenly matched in fighting ability (Figs. S10 294 and S16, upper left), but coexistence was less likely to occur at high levels of dietary overlap 295 when the species differed in fighting ability (Figs. S10 and S16). Intuitively, it makes sense that 296 the superior interference competitor would be at a disadvantage if it has to close a larger 297 phenotypic gap to recognize heterospecifics consistently.

Indeed, when dietary overlap was high, the probability of coexistence was strongly,
positively affected by whether heterospecifics recognized each other as competitors in the last
years of contact during the simulations (Fig. 3). When the superior interference competitor failed

to converge rapidly enough to maintain a high probability of heterospecific recognition, it
usually went extinct. The heterospecific recognition probability required to prevent competitive
exclusion of the superior interference competitor decreased as the fraction of refuge habitat
decreased and as the species difference in fighting ability increased (Fig. 3). However, at the
lowest fraction of refuge habitat, the positive effect of heterospecific recognition on coexistence
disappeared when the species differed greatly in fighting ability, because in this situation the
superior exploitative competitor was at risk of extinction (Fig. 3, bottom panels).

308

309 Discussion

Our overall conclusion is that species that engage in costly forms of interference competition can 310 311 coexist (or at least co-occur for thousands of generations) under biologically realistic conditions. Furthermore, coevolution allows interference competition to actually stabilize coexistence. An 312 important corollary is that factors that impede natural selection could destabilize coexistence, 313 314 resulting in extinction. Many species today have declining population sizes, increasing levels of inbreeding, and decreasing heterozygosity due to habitat loss and fragmentation (Andersen et al. 315 2004; Keyghobadi 2007; Aguilar et al. 2008; Allentoft & O'Brien 2010). While it is generally 316 317 understood that genetically depauperate species might fail to adapt when the environment 318 changes (Feiner et al. 2021), our model shows that species engaging in interference competition 319 could especially be in peril, even if the abiotic environment stays the same. Taking both 320 interference competition and selection into account would likely to lead to better predictions about the effects of climate change and other anthropogenic disturbances on species ranges and 321 322 biodiversity.

323 Our eco-evolutionary model also validates the original, interspecific Red Queen hypothesis, 324 proposed nearly 50 years ago to explain a striking pattern in the fossil record. In most groups of organisms, the probability of extinction is independent of a taxon's age and effectively constant 325 326 for millions of years (with the exception of mass extinction events) (Van Valen 1973). Van 327 Valen (1973) deduced that the most plausible explanation for this pattern is that competing 328 species can only increase in fitness by evolving at each other's expense, resulting in a zero-sum 329 game in which none of the coevolving species increases in fitness in the long term. In other 330 words, species have to keep evolving just to stay in the same "place" with their competitors. Van 331 Valen formulated the Red Queen as a group selection hypothesis, but our simulations show the zero-sum game emerging from selection at the individual level, which puts the hypothesis on a 332 333 solid theoretical foundation.

Our modeling approach applies to most, but not all, forms of interspecific territoriality. In our 334 335 model, individuals have a single competitor recognition function that determines how they 336 respond to both species. This is realistic for closely related species like *Ficedula* flycatchers (Nadachowska-Brzyska et al. 2013), because interspecific territorial aggression at the time of 337 secondary contact would almost certainly result from homology in the phenotypic traits and 338 339 neural templates used to recognize conspecifics (Grether 2011). Most cases of interspecific 340 territoriality in birds do involve phenotypically and genetically similar species, but there are 341 some well-documented cases involving phenotypically distinct species from different families 342 (Drury et al. 2020). Modeling heterospecific recognition between such distantly related species 343 would probably require two separate competitor recognition functions (i.e., one for conspecifics 344 and the other for heterospecifics). It would be unrealistic, however, to assume species possess 345 recognition functions for other species they have yet to encounter. Innate recognition of

phenotypically dissimilar species likely evolves in sympatry. Modeling the *de novo* evolution of 346 347 heterospecific recognition may require integrating neural network models with explicitly genetic 348 models of selection (Strand, E., Huse, G., & Giske 2002; Watson & Szathmáry 2016). 349 We refer to the resource that individuals compete for exploitatively in our model as food, but the results likely hold for any depletable limiting resources to which territory holders have 350 351 priority of access. In Ficedula flycatchers, nest sites may be the key limiting resource (Slagsvold 352 1978; Gustafsson 1987). Interspecific competition for food occurs during the nestling period in 353 other cavity nesting birds (reviewed by Dhondt 2012), but *Ficedula* flycatchers have relatively 354 small territories and frequently forage elsewhere (von Haartman 1956; Rybinski et al. 2016). While the most intensively studied *Ficedula* populations are provided with nest boxes, natural 355 356 tree cavities vary in size and other characteristics that affect their suitability and accessibility to nest predators (Walankiewicz et al. 2007). Males with larger territories and more nest sites may 357 have a higher probability of attracting a mate (or multiple mates; Alatalo et al. 1986; Dale & 358 359 Slagsvold 1996), and with more nest sites to choose from, a lower probability of nest failure. Thus, for this study system, the model parameter that determines the relative strength of 360 interspecific exploitative competition could be called "similarity in nest site requirements" 361 362 instead of "dietary overlap". Pied and collared flycatchers evidently have very similar nest site 363 requirements, and collared flycatchers are dominant over pied flycatchers (Qvarnström et al. 364 2010; Vallin et al. 2012), which suggests this species pair's position in parameter space is the 365 far-right side of the right-hand panels in figure 1, where coexistence is precarious, which species 366 is at risk of extinction depends on the availability of refuge habitat for pied flycatchers, and when 367 the species do coexist, they are locked in a coevolutionary stalemate.

368 We have shown that agonistic character displacement (ACD) can rescue a superior 369 interference competitor from extinction, without driving a superior exploitative competitor 370 extinct, in theory. Whether ACD enables interference competitors to coexist in nature is a 371 different question. An alternative hypothesis is that learned competitor recognition somehow 372 enables coexistence. Competitor recognition often involves learning (Grether 2011), as the 373 following examples illustrate. There is experimental evidence that territorial damselfish learn to 374 distinguish between competitors and non-competitors by observing feeding habits (Losey 1982). 375 After Daphne Major was colonized by the large ground finch, the territorial songs of two smaller 376 Darwin's finches shifted away from that of the larger competitor, apparently due to an imprinting bias (Grant & Grant 2010). Imprinting also appears to be how ants distinguish colony mates from 377 intruders based on cuticular hydrocarbon profiles (reviewed in Grether 2011). However, in many 378 situations, learning would be an inefficient and error-prone way to develop the ability to 379 380 distinguish competitors from non-competitors (Grether et al. 2013, 2017). Many species overlap 381 in resource use, but compete weakly. Individuals that innately distinguished competitors from non-competitors would have a selective advantage over those that had to learn which species to 382 383 challenge or ignore. Whether damselfish have to learn which species to attack by watching them 384 forage has not been established; the learning experiment was performed with a species 385 damselfish do not encounter in the wild (Losey 1982). The song shift in Darwin's finches was 386 also in response to a novel species, and whether it was effective in reducing aggression from that 387 species is unknown (Grant & Grant 2010). In ants, innate recognition would not work because 388 cuticular hydrocarbon profiles depend on a colony's diet, which changes over time (Leonhardt et 389 al. 2007).

390 If learned competitor recognition and innate competitor recognition were different strategies 391 in an evolutionary model based on territorial birds, we think innate competitor recognition would 392 prevail. To be sure, how an individual responds to other species is probably a product of both innate biases and learning. For example, the territorial calls of crested and Thekla larks converge 393 394 in syntopy, presumably because of learning, and this may facilitate defense of interspecific 395 territories, but the birds also recognize each other's allopatric calls (Laiolo 2012, 2013). Learning 396 could facilitate coexistence if it allowed a superior interference competitor's recognition function 397 to track changes in a superior exploitative competitor's phenotype. But learning could also 398 increase the probability of competitive exclusion if it enabled a superior exploitative competitor to avoid interspecific encounters. Thus, while we have shown in this paper that learning is not 399 400 required for interspecific territoriality to stabilize consistence, learning likely affects the outcome in nature. Incorporating learning into individual-based eco-evolutionary models is an important 401 402 challenge for future research (Romero-Mujalli et al. 2019).

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411 References

412 Aguilar, R., Quesada, M., Ashworth, L., Herrerias-Diego, Y. & Lobo, J. (2008). Genetic

- 413 consequences of habitat fragmentation in plant populations: susceptible signals in plant
- 414 traits and methodological approaches. *Mol. Ecol.*, 17, 5177–5188.
- 415 Alatalo, R. V, Lundberg, A. & Glynn, C. (1986). Female pied flycatchers choose territory quality
- and not male characteristics. *Nature*, 323, 152–153.
- 417 Allentoft, M.E. & O'Brien, J. (2010). Global amphibian declines, loss of genetic diversity and
- 418 fitness: A review. *Diversity*, 2, 47–71.
- Amarasekare, P. (2002). Interference competition and species coexistence. *Proc. R. Soc. B*, 269,
 2541–2550.
- 421 Andersen, L.W., Fog, K. & Damgaard, C. (2004). Habitat fragmentation causes bottlenecks and
- 422 inbreeding in the European tree frog (*Hyla arborea*). *Proc. R. Soc. B Biol. Sci.*, 271, 1293–
 423 1302.
- 424 Case, T.J. & Gilpin, M.E. (1974). Interference competition and niche theory. *Proc. Natl. Acad.*425 *Sci. USA*, 71, 3073–3077.
- 426 Chesson, P. (2000). Mechanisms of maintenance of species diversity. *Annu. Rev. Ecol. Syst.*, 31,
 427 343–358.
- 428 Cordero, O.X. & Datta, M.S. (2016). Microbial interactions and community assembly at
- 429 microscales. *Curr. Opin. Microbiol.*, 31, 227–234.
- 430 Cowen, M.C., Drury, J.P. & Grether, G.F. (2020). Multiple routes to interspecific territoriality in
- 431 sister species of North American perching birds. *Evolution*
- 432 Dale, S. & Slagsvold, T. (1996). Mate choice on multiple cues, decsion rules and sampling
- 433 strategies in female pied flycatchers. *Behaviour*, 133, 903–944.
- 434 Darwin, C. (1859). On the Origin of Species by Means of Natural Selection, or, the Preservation
- 435 *of Favored Races in the Struggle for Life*. John Murray, London.

- 436 Dhondt, A.A. (2012). Interspecific Competition in Birds. Oxford University Press, Oxford.
- 437 Diamond, J.M. (1978). Niche shifts and the rediscovery of interspecific competition. *Am. Sci.*,
 438 66, 322–331.
- 439 Dobson, A.P. (1985). The population dynamics of competition between parasites. *Parasitology*,
- **440** 91, 317–347.
- Drury, J.P., Cowen, M.C. & Grether, G.F. (2020). Competition and hybridization drive
 interspecific territoriality in birds. *Proc. Natl. Acad. Sci.*, 117, 12923–12930.
- 443 Feiner, N., Brun-Usan, M. & Uller, T. (2021). Evolvability and evolutionary rescue. Evol. Dev.,
- 444 23, 308–319.
- 445 Freshwater, C., Ghalambor, C.K. & Martin, P.R. (2014). Repeated patterns of trait divergence
- between closely related dominant and subordinate bird species. *Ecology*, 95, 2334–2345.
- García-Bayona, L. & Comstock, L.E. (2018). Bacterial antagonism in host-associated microbial
 communities. *Science*, 361, 1215, eaat2456.
- 449
- 450 Ghoul, M. & Mitri, S. (2016). The ecology and evolution of microbial competition. *Trends*
- 451 *Microbiol.*, 24, 833–845.
- 452 Grant, B.R. & Grant, P.R. (2010). Songs of Darwin's finches diverge when a new species enters
 453 the community. *Proc. Natl. Acad. Sci.*, 107, 20156–20163.
- 454 Grether, G. (2011). The neuroecology of competitor recognition. *Integr. Comp. Biol.*, 51, 807–
 455 818.
- 456 Grether, G.F., Anderson, C.N., Drury, J.P., Kirschel, A.N.G., Losin, N., Okamoto, K., et al.
- 457 (2013). The evolutionary consequences of interspecific aggression. Ann. N. Y. Acad. Sci.,
- 458 1289, 48–68.

- 459 Grether, G.F., Drury, J.P., Okamoto, K.W., McEachin, S. & Anderson, C.N. (2020). Predicting
- 460 evolutionary responses to interspecific interference in the wild. *Ecol. Lett.*, 23, 221–230.
- 461 Grether, G.F., Losin, N., Anderson, C.N. & Okamoto, K. (2009). The role of interspecific
- 462 interference competition in character displacement and the evolution of competitor
- 463 recognition. *Biol. Rev.*, 84, 617–635.
- Grether, G.F., Peiman, K.S., Tobias, J.A. & Robinson, B.W. (2017). Causes and consequences of
 behavioral interference between species. *Trends Ecol. Evol.*, 32, 760–772.
- 466 Gustafsson, L. (1987). Interspecific competition lowers fitness in collared flycatchers Ficedula
- 467 *albicollis*: an experimental demonstration. *Ecology*, 68, 291.
- 468 von Haartman, L. (1956). Territory in the pied flycatcher. *Ibis*, 98, 460–475.
- 469 Hierro, J.L. & Callaway, R.M. (2021). The ecological importance of allelopathy. *Annu. Rev.*470 *Ecol. Evol. Syst.*, 52, 25–45.
- 471 HilleRisLambers, J., Adler, P.B., Harpole, W.S., Levine, J.M. & Mayfield, M.M. (2012).
- 472 Rethinking community assembly through the lens of coexistence theory. *Annu. Rev. Ecol.*
- 473 *Evol. Syst.*, 43, 227–248.
- 474 Inderjit, Wardle, D.A., Karban, R. & Callaway, R.M. (2011). The ecosystem and evolutionary
 475 contexts of allelopathy. *Trends Ecol. Evol.*, 26, 655–662.
- 476 Keyghobadi, N. (2007). The genetic implications of habitat fragmentation for animals. *Can. J.*477 *Zool.*, 85, 1049–1064.
- 478 Krebs, C.J. (2001). Ecology: the experimental analysis of distribution and abundance. 5th edn.
- 479 Benjamin Cummings, San Francisco.
- 480 Laiolo, P. (2012). Interspecific interactions drive cultural co-evolution and acoustic convergence
- 481 in syntopic species. J. Anim. Ecol., 81, 594–604.

- 482 Laiolo, P. (2013). From inter-specific behavioural interactions to species distribution patterns
 483 along gradients of habitat heterogeneity. *Oecologia*, 171, 207–215.
- 484 Leonhardt, S.D., Brandstaetter, A.S. & Kleineidam, C.J. (2007). Reformation process of the
- 485 neuronal template for nestmate-recognition cues in the carpenter ant *Camponotus*
- 486 floridanus. J. Comp. Physiol. a-Neuroethology Sens. Neural Behav. Physiol., 193, 993–
- 487 1000.
- 488 Levy, O., Dayan, T. & Kronfeld-Schor, N. (2011). Interspecific competition and torpor in golden
- spiny mice: two sides of the energy-acquisition coin. *Integr. Comp. Biol.*, 51, 441–448.
- 490 Losey, G.S.J. (1982). Ecological cues and experience modify interspecific aggression by the
- damselfish *Stegastes fasciolatus*. *Behaviour*, 81, 14–37.
- Martin, P.R., Freshwater, C. & Ghalambor, C.K. (2017). The outcomes of most aggressive
 interactions among closely related bird species are asymmetric. *PeerJ*, 5, e2487.
- 494 Martin, P.R. & Ghalambor, C.K. (2014). When David beats Goliath: the advantage of large size
- in interspecific aggressive contests declines over evolutionary time. *PLoS One*, 9, e108741.
- 496 McEachin, S., Drury, J.P., Anderson, C.N. & Grether, G.F. (2021). Mechanisms of reduced
- 497 interspecific interference between territorial species. *Behav. Ecol.*, 33, 126–136.
- 498 Meiners, S.J., Kong, C.-H., Ladwig, L.M., Pisula, N.L. & Lang, K.A. (2012). Developing an
- 499 ecological context for allelopathy. *Plant Ecol.*, 213, 1221–1227.
- 500 Nadachowska-Brzyska, K., Burri, R., Olason, P.I., Kawakami, T., Smeds, L. & Ellegren, H.
- 501 (2013). Demographic divergence history of pied flycatcher and collared flycatcher inferred
 502 from whole-genome re-sequencing data. *PLoS Genet.*, 9, e1003942.
- 503 Okamoto, K.W. & Amarasekare, P. (2018). A framework for high-throughput eco-evolutionary
- simulations integrating multilocus forward-time population genetics and community

- 506 Okamoto, K.W. & Grether, G.F. (2013). The evolution of species recognition in competitive and
- 507 mating contexts: the relative efficacy of alternative mechanisms of character displacement.
- 508 *Ecol. Lett.*, 16, 670–678.
- 509 Pasch, B., Bolker, B.M. & Phelps, S.M. (2013). Interspecific dominance via vocal interactions
- 510 mediates altitudinal zonation in neotropical singing mice. *Am. Nat.*, 182, E161-73.
- 511 Peiman, K.S. & Robinson, B.W. (2010). Ecology and evolution of resource-related
- 512 heterospecific aggression. Q. Rev. Biol., 85, 133–158.
- 513 Qvarnström, A., Rice, A.M. & Ellegren, H. (2010). Speciation in *Ficedula* flycatchers. *Philos*.
- 514 Trans. R. Soc. B Biol. Sci., 365, 1841–1852.
- Qvarnström, A., Wiley, C., Svedin, N. & Vallin, N. (2009). Life-history divergence facilitates
 regional coexistence of competing *Ficedula* flycatchers. *Ecology*, 90, 1948–1957.
- 517 Romero-Mujalli, D., Jeltsch, F. & Tiedemann, R. (2019). Individual-based modeling of eco-
- evolutionary dynamics: state of the art and future directions. *Reg. Environ. Chang.*, 19, 1–

519 12.

- 520 Rybinski, J., Sirkiä, P.M., Mcfarlane, S.E., Vallin, N., Wheatcroft, D., Ålund, M., et al. (2016).
- 521 Competition-driven build-up of habitat isolation and selection favoring modified dispersal
 522 patterns in a young avian hybrid zone. *Evolution*, 70, 2226–2238.
- 523 Schoener, T.W. (1983). Field experiments on interspecific competition. Am. Nat., 122, 240–285.
- 524 Slagsvold, T. (1978). Competition between the great tit *Parus major* and the pied flycatcher
- 525 *Ficedula hypoleuca*: an experiment. *Ornis Scand.*, 9, 46–50.
- 526 Strand, E., Huse, G., & Giske, J. (2002). Artificial evolution of life history and behavior. Am.
- 527 *Nat.*, 159, 624–644.

⁵⁰⁵ ecology. *Methods Ecol. Evol.*, 9, 525–534.

- 528 Tobias, J.A., Cornwallis, C.K., Derryberry, E.P., Claramunt, S., Brumfield, R.T. & Seddon, N.
- 529 (2014). Species coexistence and the dynamics of phenotypic evolution in adaptive radiation.
 530 *Nature*, 506, 359–363.
- 531 Van Valen, L. (1973). A new evolutionary law. *Evol. Theory*, 1, 1–30.
- 532 Vallin, N., Rice, A.M., Arntsen, H., Kulma, K. & Qvarnström, A. (2012). Combined effects of
- 533 interspecific competition and hybridization impede local coexistence of *Ficedula*
- flycatchers. *Evol. Ecol.*, 26, 927–942.
- 535 Vance, R.R. (1984). Interference competition and the coexistence of 2 competitors on a single
- 536 limiting resource. *Ecology*, 65, 1349–1357.
- 537 Veen, T., Sheldon, B.C., Weissing, F.J., Visser, M.E., Qvarnström, A. & Sætre, G.P. (2010).
- 538 Temporal differences in food abundance promote coexistence between two congeneric
 539 passerines. *Oecologia*, 162, 873–884.
- 540 Walankiewicz, W., Czeszczewik, D. & Mitrus, C. (2007). Natural nest sites of the collared
- 541 flycatcher *Ficedula albicollis* in lime-hornbeam-oak stands of a primeval forest. *Ornis*
- 542 *Fenn.*, 84, 155–162.
- 543 Watson, R.A. & Szathmáry, E. (2016). How can evolution learn? *Trends Ecol. Evol.*, 31, 147–
 544 157.
- Ziv, Y. & Kotler, B.P. (2003). Giving-up densities of foraging gerbils: the effect of interspecific
 competition on patch use. *Evol. Ecol.*, 17, 333–347.
- 547



Fig. 1. Coexistence and evolutionary outcomes in relation to dietary overlap, the fraction of the 549 habitat where species 2 could breed without interference from species 1, and the probability of 550 species 1 winning interspecific fights (Pw_1 , based on the species difference in fighting ability). 551 Each point represents a unique parameter set. In the top panels, symbols correspond to the 552 553 observed coexistence outcomes (see legend). For example, "Coexist or Sp1 extinct" means that the species coexisted to the end or species 1 went extinct, in different simulation runs. In the 554 555 lower panels, symbols correspond to the evolutionary outcomes when both species persisted to 556 the end (see legend): "Stasis", neither species' mean values of z and μ changed significantly; "Sp2 diverged", species 2 shifted away from species 1 and species 1 exhibited stasis; 557 558 "Divergence", both species shifted away from the other; "Sp1 chased Sp2", species 1 converged 559 and species 2 diverged; "Divergence/Sp2 diverged", both outcomes occurred in different 560 simulation runs; etc. This figure summarizes a subset of the simulations in which the initial mean 561 difference between the species in traits z and μ was 1. Figs. S9 – S12 summarize all of the 562 simulations.



Fig. 2. Probability of coexistence in relation to dietary overlap, the fraction of the habitat where species 2 could breed without interference from species 1 (color scale), and the probability of species 1 winning interspecific fights (Pw_1). The lines were generated by a non-parametric smoothing function bounded by 0 and 1 (geom_smooth in the R package ggplot2). Here, the initial mean difference between the species in traits *z* and μ was 1. Fig. S15 shows results for two additional levels of Pw_1 , and Fig. S16 shows the probability of coexistence for simulations in which the initial mean difference was 2.



Fig. 3. Probability of coexistence under high dietary overlap (≥ 0.75) in relation to the average probability of heterospecific males recognizing each other as competitors in the final years of contact (i.e., just before one species went extinct or the simulation ended). Each point represents a unique parameter set. The color scale indicates the fraction of the habitat where species 2 could breed without interference from species 1 (limited to the range shown in the legend). The lines were generated as in Fig. 2. The initial mean difference between the species in traits *z* and μ was 1 or 2 in these simulations.