

# UC Davis

## UC Davis Previously Published Works

### Title

Prenatal Exposure to Source-Specific Fine Particulate Matter and Autism Spectrum Disorder.

### Permalink

<https://escholarship.org/uc/item/6gr08926>

### Journal

Environmental Science and Technology, 58(42)

### Authors

Luglio, David

Kleeman, Michael

Yu, Xin

[et al.](#)

### Publication Date

2024-10-22

### DOI

10.1021/acs.est.4c05563

Peer reviewed

# Prenatal Exposure to Source-Specific Fine Particulate Matter and Autism Spectrum Disorder

David G. Luglio, Michael J. Kleeman, Xin Yu, Jane C. Lin, Ting Chow, Mayra P. Martinez, Zhanghua Chen, Jiu-Chiuan Chen, Sandrah Proctor Eckel, Joel Schwartz, Frederick Lurmann, Rob McConnell, Anny H. Xiang,\* and Md Mostafijur Rahman\*



Cite This: *Environ. Sci. Technol.* 2024, 58, 18566–18577



Read Online

ACCESS |

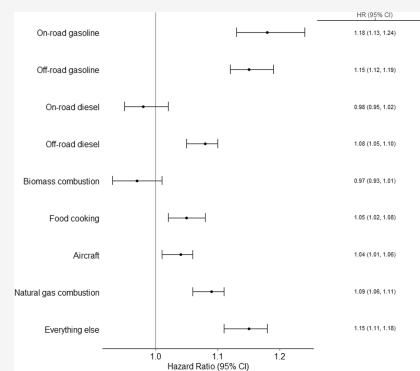
Metrics & More

Article Recommendations

Supporting Information

**ABSTRACT:** In this study, associations between prenatal exposure to fine particulate matter (PM<sub>2.5</sub>) from 9 sources and development of autism spectrum disorder (ASD) were assessed in a population-based retrospective pregnancy cohort in southern California. The cohort included 318,750 mother–child singleton pairs. ASD cases ( $N = 4559$ ) were identified by ICD codes. Source-specific PM<sub>2.5</sub> concentrations were estimated from a chemical transport model with a  $4 \times 4$  km<sup>2</sup> resolution and assigned to maternal pregnancy residential addresses. Cox proportional hazard models were used to estimate the hazard ratios (HR) of ASD development for each individual source. We also adjusted for total PM<sub>2.5</sub> mass and in a separate model for all other sources simultaneously. Increased ASD risk was observed with on-road gasoline (HR [CI]: 1.18 [1.13, 1.24]), off-road gasoline (1.15 [1.12, 1.19]), off-road diesel (1.08 [1.05, 1.10]), food cooking (1.05 [1.02, 1.08]), aircraft (1.04 [1.01, 1.06]), and natural gas combustion (1.09 [1.06, 1.11]), each scaled to standard deviation increases in concentration. On-road gasoline and off-road gasoline were robust for other pollutant groups. PM<sub>2.5</sub> emitted from different sources may have different impacts on ASD. The results also identify PM source mixtures for toxicological investigations that may provide evidence for future public health policies.

**KEYWORDS:** autism spectrum disorders, PM<sub>2.5</sub>, gasoline, pregnancy, air pollution sources, prenatal exposures



## INTRODUCTION

Autism spectrum disorder (ASD) is a set of social communicative and behavioral challenges related to atypical neurodevelopment. These characteristics range from repetitive behaviors and patterns to difficulty interacting with others and even an inability to talk.<sup>1–3</sup> ASD is diagnosed at an early age and persists throughout adulthood;<sup>4–6</sup> therefore, ASD etiologic research focuses mainly on prenatal exposures.

Prior research has primarily focused on genetic risk factors for ASD, identifying over 100 genes associated with ASD. These genes account for a large fraction of ASD variation.<sup>7,8</sup> Recent literature, additionally, has revealed associations between several environmental exposures and increased risk of ASD. Prenatal exposure to air pollution, most consistently to fine particulate matter (PM<sub>2.5</sub>), has been associated with ASD among children.<sup>9–20</sup> Other associated pollutants were ozone (O<sub>3</sub>) and nitrogen dioxide (NO<sub>2</sub>).<sup>9,13,17</sup> PM<sub>2.5</sub> is composed of different species.<sup>21–25</sup>

PM<sub>2.5</sub> is regulated based on total concentration, but studies indicate that different compositions have different effects on several health outcomes.<sup>26–31</sup> We recently reported PM<sub>2.5</sub>-composition-specific associations, such as black carbon (BC), organic matter (OM), nitrate (NO<sub>3</sub><sup>-</sup>), and sulfate (SO<sub>4</sub><sup>2-</sup>), on ASD.<sup>15</sup> In another study, we reported strong association of

ASD with components related to nontailpipe emissions, such as copper (Cu), iron (Fe), and manganese (Mn).<sup>16</sup>

New models of spatial variability in sources of PM<sub>2.5</sub> are relatively recently available. To the best of our knowledge, only a single study from Sweden examined the effect of different PM sources on ASD.<sup>32</sup> Identification of the risk associated with increased exposure to sources of PM would have important implications. Different sources of PM may emit different mixtures of components or, at times, uniquely prominent components or forms of components (e.g., different oxides of the same metal). Given this, it may be expected that different PM<sub>2.5</sub> sources have different toxicities, which would have bearings on emergent outcomes, such as ASD development. In studies of other outcomes, PM<sub>2.5</sub> components derived from fossil-fuel combustion are consistently found to be the most toxic. Therefore, it has been proposed that regulation of PM<sub>2.5</sub>

**Received:** June 5, 2024

**Revised:** October 1, 2024

**Accepted:** October 3, 2024

**Published:** October 11, 2024



should be tailored to the component,<sup>33</sup> although there are arguments against this given the mixed findings about the relative potencies of different components and challenges associated with operationalizing source-specific regulations.<sup>34–36</sup> Here, we focus on sources relevant to the ASD.

Our aim of this study was to assess the differential effects of prenatal exposure to ambient PM<sub>2.5</sub> sources at maternal addresses on ASD development among children in a large population-based southern California pregnancy cohort. In this cohort, we have previously reported associations of ASD with aircraft-sourced PM<sub>2.5</sub>, using the same exposure modeling approach we use in this study.<sup>10</sup> We now examined the associations of eight other sources with ASD. We also assessed whether the source-specific exposure associations were independent of each other and of associations with PM<sub>2.5</sub> mass.

## MATERIALS AND METHODS

**Study Population.** We used a large, multiethnic population-based retrospective pregnancy cohort study to address the study questions. The cohort included mother–child pairs of singletons delivered at Kaiser Permanente Southern California (KPSC) hospitals between January 1, 2001 and December 31, 2014. KPSC is a large integrated healthcare system with over 4.5 million members and is representative, demographically, of the population of the region.<sup>37</sup> Information on maternal social and demographic characteristics, as well as pregnancy health information, was extracted from KPSC's electronic medical records (EMR) system. Maternal addresses during pregnancy were also extracted from the EMR and were geocoded using ArcGIS. We further assessed the geocodes for exposure assignment suitability. Mother–child pairs with addresses with only a street name, locality, administrative unit, or 5-digit postal code were excluded from the study since they provided a location too uncertain for exposure assignment.

The starting cohort size used in this study was 370,723 maternal–child pairs, inclusive of singleton births with continued KPSC membership at age 1. A total of 51,973 births were excluded, including those with implausible age of delivery or birth weight and missing covariates ( $n = 666$ ), out-of-range maternal age of delivery (i.e., age <15 years or >55 years;  $n = 159$ ), and missing/incomplete addresses or geocodes not suitable for exposure assignment ( $n = 51,148$ ). Detailed derivation of the sample size, including other exclusionary factors, is shown in Figure S1. The final cohort size was 318,750 mother–child pairs. Characteristics of this group are listed in Table 1. Screenings for potential ASD risk started at 18 months of life during well-child visits, and the median age of ASD diagnosis was 3 years. Children were followed from birth through EMR until the clinical diagnosis of ASD, death, loss to follow-up, or age 5, whichever came first. The follow-up period for this study concluded in December 2019, ensuring that children born in 2014 received a minimum of 5 years of follow-up time. To ensure uniform follow-up duration for all children, we applied censoring at the age of 5 years.

Both the KPSC and University of Southern California Institutional Review Board approved this study with a waiver of individual subject consent.

**ASD Ascertainment.** ASD diagnoses were considered valid if identified by the ICD-9 codes 299.0, 299.1, 299.8, and 299.9 (for EMR records before October 1, 2015) and ICD-10 codes F84.0, F84.3, F84.5, F84.8, and F84.9 (after October 1,

**Table 1. Characteristics of Children, with and without Autism Spectrum Disorder (ASD)**

characteristics	children, no. (%) or median (interquartile range)		
	overall ( $n = 318,750$ )	with ASD ( $n = 4559$ )	without ASD ( $n = 314,191$ )
sex			
male (%)	163 181 (51.2)	3703 (81.2)	159 428 (50.7)
female (%)	155 569 (49.8)	856 (18.8)	154 763 (49.3)
follow-up year after birth, median [IQR <sup>c</sup> ], years	4.0 [4.0, 4.0]	3.0 [2.3, 3.7]	4.0 [4.0, 4.0]
maternal age at delivery, median [IQR <sup>c</sup> ], years	30.4 [26.3, 34.3]	31.3 [27.5, 35.2]	30.4 [26.2, 34.3]
parity, $N$ (%)			
0	111 981 (35.1)	1844 (40.4)	110 137 (35.1)
1	104 561 (32.8)	1495 (32.8)	103 066 (32.8)
>2	84 176 (26.4)	903 (19.8)	83 273 (26.5)
unknown	18 032 (5.7)	317 (7.0)	17 715 (5.6)
maternal education, $N$ (%)			
high school or lower	112 096 (35.2)	1335 (29.3)	110 761 (35.3)
some college	94 524 (29.7)	1477 (32.4)	93 047 (29.6)
college graduate or higher	109 087 (34.2)	1713 (37.6)	107 374 (34.2)
unknown	3043 (1.0)	43 (0.7)	3009 (1.0)
household annual income, <sup>a</sup> $N$ (%)			
<\$30,000	24 027 (7.5)	325 (7.1)	23 710 (7.5)
\$30,000–\$49,999	100 575 (31.6)	1436 (31.5)	99 139 (31.6)
\$50,000–\$69,999	98 015 (30.7)	1415 (31.0)	96 593 (30.7)
\$70,000–\$89,999	55 611 (17.4)	801 (17.5)	54 816 (17.4)
>\$90,000	40 512 (12.7)	582 (12.8)	39 933 (12.7)
race/ethnicity, $N$ (%)			
non-Hispanic white	81 050 (25.4)	956 (21.0)	80 094 (25.5)
non-Hispanic black	29 773 (9.3)	477 (9.8)	29 326 (9.3)
Hispanic	161 414 (50.6)	2300 (50.4)	159 114 (50.6)
Asian/Pacific Islander	39 974 (12.5)	744 (16.3)	39 230 (12.5)
other	6539 (2.1)	112 (2.5)	6427 (2.0)
any history of maternal comorbidity, <sup>b</sup> $N$ (%)	46 717 (14.6)	839 (18.4)	45 878 (14.6)
prepregnancy diabetes, <sup>c</sup> $N$ (%)	10 248 (3.2)	242 (5.3)	10 006 (3.2)
prepregnancy obesity, <sup>d</sup> $N$ (%)	53 354 (16.7)	1049 (23.0)	52 305 (16.6)
year of birth, $N$ (%)			
2001–2007	152 750 (47.9)	1802 (39.5)	164 198 (52.2)
2008–2014	166 000 (52.1)	2757 (60.5)	149 993 (47.2)

<sup>a</sup>Census tract-level median household income. <sup>b</sup> $\geq 1$  diagnosis of heart, lung, kidney, or liver disease; cancer. <sup>c</sup>Type I and Type II diabetes diagnosed before pregnancy. <sup>d</sup>Prepregnancy BMI  $\geq 30$ . <sup>e</sup>Abbreviations: IQR, interquartile range.

2015) in at least two separate visits, as described in previous studies.<sup>38–41</sup>

**Air Pollution Exposure Assessment.** A source-oriented chemical transport model developed by the University of California Davis/California Institute of Technology (UCD/CIT) was employed to assess prenatal source-specific PM<sub>2.5</sub>

exposures.<sup>42</sup> This model tracked 22 PM constituents from emission through atmospheric transport and deposition, incorporating calculations on coagulation, gas- and particle-phase chemistry, and gas-to-particle conversion. Emission rates were derived from emission inventories provided by the California Air Resources Board (CARB), and meteorological data from the Weather Research and Forecast (WRF) model was incorporated to project particle chemical activity, movement, and fate. PM data were estimated in 3D atmospheric grid cells. The UCD/CIT CTM was configured to predict mass and number concentrations of particles ranging in diameter from 0.01 to 10  $\mu\text{m}$  with  $4 \times 4 \text{ km}^2$  horizontal grid resolution at the ground level in the current study. Results were generated with hourly time resolution but were averaged to monthly time resolution prior to use in the exposure analysis.<sup>43,44</sup> The UCD/CIT model has been applied in multiple studies across the United States,<sup>45</sup> but results from California are used in the current study.

PM concentrations emitted from 9 different source groups were tagged and tracked through the simulation of emissions, transport, and deposition. These source groups included on-road gasoline, on-road diesel, off-road gasoline, off-road diesel, natural gas combustion, food cooking, biomass burning, aircraft, and an all-other source category. We previously reported associations between aircraft-sourced PM<sub>2.5</sub> and ASD development within this cohort,<sup>10</sup> but not in the context of other sources. This current study builds on that work by adjusting for other specific-source PM<sub>2.5</sub>. Only primary particle contributions were tracked from each source group. Secondary coatings on particles were not tagged for source apportionment. Measured PM size and composition profiles were applied to PM emissions<sup>46–51</sup> and followed through the atmospheric simulation. This model is based on updated work from ref 52 and is described in more detail in refs 43,44,53. Monthly averaged modeled concentrations were compared against all of the monthly averaged ambient monitoring data assembled by the EPA at all available locations and times. The bias between predicted and measured monthly average PM<sub>2.5</sub> mass concentrations was used as a target for a constrained multilinear regression model based on the primary PM<sub>2.5</sub> concentrations from on-road gasoline vehicles, off-road gasoline vehicles, on-road diesel vehicles, off-road diesel vehicles, biomass combustion, food cooking, aircraft, natural gas combustion, and all other sources. Additional independent variables were based on secondary nitrate and sulfate. Regression coefficients were constrained to range from  $\pm 5$ . Exposure concentrations were adjusted to remove bias at all locations across the CTM grid. Strong correlations ( $r > 0.8$ ) were observed between the predicted and measured PM<sub>2.5</sub> mass concentrations at most of the monitoring stations. Previous studies have compared the source-specific CTM predictions to measurements at receptor sites across California. Reasonable agreement between predictions and measurements is generally observed across all locations.<sup>42</sup>

Monthly average exposures to each of the sources and total PM<sub>2.5</sub> mass were assigned to maternal addresses during the entire pregnancy. Monthly exposure estimates that did not correspond exactly to a pregnancy start/end month were assigned proportionally based on an overlap of the months. Exposures were time-weighted to account for changes in subject addresses during pregnancy.

**Covariates.** Covariates were selected based on associations with ASD in previous studies and on expert knowledge.<sup>39,54</sup>

These variables included characteristics of the mother: self-reported race/ethnicity, age at delivery, parity, education levels, estimated household income based on census tract expressed as per 10k, and history of medical co-occurring conditions (e.g.,  $\geq$  diagnosis of heart, lung, kidney, or liver disease; cancer). Also included were child sex, birth year, and an indicator variable for season (i.e., dry from April to October; wet from November to March). ASD incidence has been observed to have increased over time, whereas PM concentrations have decreased. To control for this potential confounding time effect, models were adjusted for birth year. It was included as a nonlinear term with a penalized natural spline with four degrees of freedom as selected through the Akaike Information Criterion (AIC). Prepregnancy diabetes mellitus and obesity of the mother, but not birth weight or gestational age, were additionally included as covariates. The former two have been demonstrated risk factors for ASD,<sup>41</sup> whereas the latter two were not adjusted for since they may be directly in the causal pathway.<sup>55–57</sup> Finally, neighborhood socioeconomic status (SES) and urbanicity indicator variables were included in sensitivity analyses. The SES indicator was defined as the neighborhood disadvantage as in Yu et al.<sup>58</sup> The urbanicity variable was classified into two levels (urban and suburban/rural) and assigned to each participant based on the USDA rural–urban commuting area codes based on the year 2000 census tracts and the mapping in a previous report.<sup>59,60</sup>

The total list of covariates is birth year, season, maternal age at delivery, maternal parity, maternal race/ethnicity, maternal education, maternal comorbidities (yes/no), maternal prepregnancy obesity (yes/no), maternal prepregnancy diabetes (yes/no), household income, and child gender.

**Statistical Analysis.** Associations of pregnancy average exposure to PM from the 9 sources and the development of ASD were assessed using Cox proportional hazard models. A base model included all of the covariates described above in each analysis. Schoenfeld residual plots were constructed to assess whether the proportional hazards assumption was met, and no clear nonrandom pattern against follow-up time was discerned. Since the mean concentrations of PM varied from source to source, we standardized them by subtracting by their means and then dividing the values by their standard deviation when included in the model. Hazard ratios (HR) and 95% confidence intervals (CIs) associated with the ASD outcome were reported as a standard deviation (SD) increase in their respective source concentration so that the population HRs for each source were comparable relative to each other. A secondary analysis was conducted per 1  $\mu\text{g}/\text{m}^3$  increase of source PM<sub>2.5</sub> concentration to assess the inherent potency of each particle type. This was conducted only for those sources with significant positive associations in the population-relevant analysis. We first fitted the single-source model. In sensitivity analyses, to assess the independence of PM<sub>2.5</sub> source association, we adjusted the single-source model with total PM<sub>2.5</sub> mass, as well as remainder PM<sub>2.5</sub> mass (i.e., PM<sub>2.5</sub> mass minus source-specific mass), as described previously.<sup>15</sup> The correlations between sources were not exceptionally high (see Table 3); therefore, we additionally ran a multisource model including all sources simultaneously to assess the relative strength of the association of each source. Each of the covariates listed above were included into each of the models. In a previous study, SES (defined as neighborhood disadvantage) was found to be associated with ASD<sup>58</sup> in this cohort but was not included in the current main model since

related variables such as income, race/ethnicity, and maternal education were already included. Additionally, prior work has found that autism diagnosis rates vary by urbanicity with higher numbers in urban areas;<sup>61,62</sup> therefore, sensitivity analyses were conducted with inclusion of these two variables separately and simultaneously to test whether the results of our primary analyses are robust to these adjustments. Since 45,680 participants were excluded due to imprecise residential addresses (as mentioned above), which may lead to selection bias, we conducted an additional analysis including these excluded participants to assess if there was any change in the effect estimates. Robust standard errors were used to correct for potential correlations within siblings born to the same mothers.  $\alpha$  value for statistical tests was set at 0.05. All analyses were performed using R Statistical Software (v4.1.2).<sup>63</sup>

## RESULTS

A total of 4559 children were diagnosed with ASD by the age of 5, with males making up more than 80% of the cases/diagnoses (Table 1). The ASD and non-ASD groups showed small differences in maternal age at delivery and household income distributions. Furthermore, children with ASD were more likely to have mothers with co-occurring medical conditions, as well as prepregnancy diabetes and obesity. In addition, these mothers were more often nulliparous.

Table 2 shows the pregnancy average and median exposures to PM<sub>2.5</sub> concentrations, in total and by source. Note that the

**Table 2. Descriptive Statistics of PM<sub>2.5</sub> Concentrations According to the Source of Emission<sup>a</sup>**

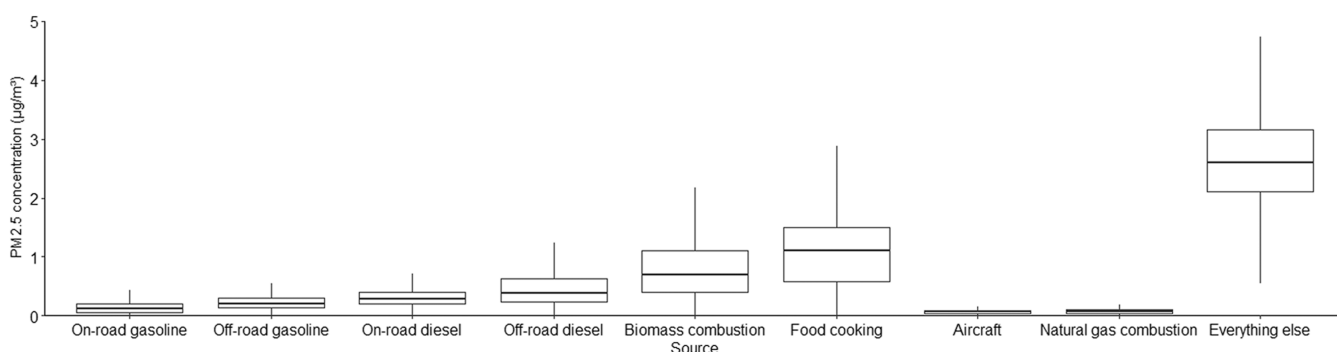
sources	PM <sub>2.5</sub> concentration (ng/m <sup>3</sup> )			
	mean	SD	median	IQR
on-road gasoline	148.9	135.1	115.0	151.3
off-road gasoline	232.8	120.9	204.1	161.4
on-road diesel	315.1	161.5	285.1	205.4
off-road diesel	473.0	346.2	385.7	395.0
biomass combustion	858.1	779.1	704.2	718.6
food cooking	1117.8	626.7	1101.5	920.7
aircraft	83.9	95.4	66.8	42.8
natural gas combustion	80.8	51.4	75.0	56.0
other/everything else	2659.8	811.5	2610	1050.4
all-source PM <sub>2.5</sub> mass	5970.2	2106.2	5850.7	2824.9
total PM <sub>2.5</sub> mass	14242.1	4097.8	13575	5459.3

<sup>a</sup>Abbreviations: SD, standard deviation; IQR, interquartile range.

all-source PM<sub>2.5</sub> is a summation of all the sources, whereas the total PM<sub>2.5</sub> concentration is the estimated measurement by the model as a group. This latter term includes secondary particles that are excluded from each of the source groups. The largest identified air pollution source affecting this population was food cooking and biomass combustion with average exposures of 1117.8 and 858.1 ng/m<sup>3</sup>, respectively. The pregnancy averages of other common sources such as on-road gasoline, on-road diesel, and natural gas combustion were 148.9, 315.1, and 80.8 ng/m<sup>3</sup>, respectively. Since the UCD/CIT model only tags primary PM emissions, each of these sources composes a small percentage of total PM<sub>2.5</sub> mass. The variability in concentrations in each of the sources differed, with the largest range observed in the “other”/“everything else” and food cooking categories. The typical spread of exposures is displayed in Figure 1, excluding all outliers (i.e., data points than 1.5 × IQR greater than the third quartile), which are shown in Figure S2. Notably, the peak exposures are to biomass combustion particles despite having a lower overall median or mean than food cooking or the “other”/“everything else” categories. Sources in the other category include windblown dust and tire and brake wear. SD increases were highest proportionally for aircraft (coefficient of variation [CV]: 1.14) and lowest proportionally for “other”/“everything else” (0.31). Spatial distribution patterns for different sources estimated by this model are shown by Yu et al.<sup>42</sup>

There were only two instances of high correlations between source and total PM<sub>2.5</sub> ( $r > 0.8$ ), e.g., on-road gasoline and off-road diesel. Correlations among other sources were moderate, all less than 0.8. The highest correlation was 0.77 between the natural gas combustion and food cooking (Table 3).

In the single-source model analysis for the associations with ASD, pregnancy exposure to all sources, except for on-road diesel and biomass combustion, were significantly associated with increased risk of ASD (Figure 2). The HRs [95% CI] were highest for on-road gasoline (1.18 [1.13, 1.24]; per 135.1 ng/m<sup>3</sup>), off-road gasoline (1.15 [1.12, 1.19]; per 120.9 ng/m<sup>3</sup>), and “other”/“everything else” (1.15 [1.11, 1.18]; per 811.5 ng/m<sup>3</sup>). HRs for the other significant sources include off-road diesel (1.08 [1.05, 1.10]; per 346.2 ng/m<sup>3</sup>), food cooking (1.05 [1.02, 1.08]; per 626.7 ng/m<sup>3</sup>), natural gas combustion (1.09 [1.06, 1.11]; per 51.4 ng/m<sup>3</sup>), and previously reported aircraft-sourced PM<sub>2.5</sub> (1.04 [1.01, 1.06]; per 95.4 ng/m<sup>3</sup>). These source-specific associations of increased ASD risks remained largely unchanged when adjusting the single-source model for the total PM<sub>2.5</sub> and remainder PM<sub>2.5</sub> mass. Notably, however, on-road diesel goes from a null association



**Figure 1.** Pregnancy-averaged source-specific PM<sub>2.5</sub> exposure concentrations across participants (outliers removed). Outliers were selected as data points 1.5 × IQR greater than the third quartile.

Table 3. Spearman Rank Correlations between PM<sub>2.5</sub> Sources and Total Mass

	on-road gasoline	off-road gasoline	on-road diesel	off-road diesel	biomass combustion	food cooking	aircraft	natural gas combustion	other	total PM <sub>2.5</sub> mass
on-road gasoline	1.00	0.54	0.31	0.74	0.51	0.44	0.22	0.54	0.68	0.83
off-road gasoline		1.00	0.24	0.74	0.53	0.46	0.21	0.47	0.72	0.62
on-road diesel			1.00	0.41	0.06	0.70	0.33	0.72	0.52	0.29
off-road diesel				1.00	0.55	0.55	0.17	0.65	0.76	0.82
biomass combustion					1.00	0.18	0.13	0.19	0.39	0.65
food cooking						1.00	0.15	0.77	0.54	0.37
aircraft							1.00	0.33	0.46	0.20
natural gas combustion								1.00	0.75	0.48
other									1.00	0.72
total PM <sub>2.5</sub> mass										1.00

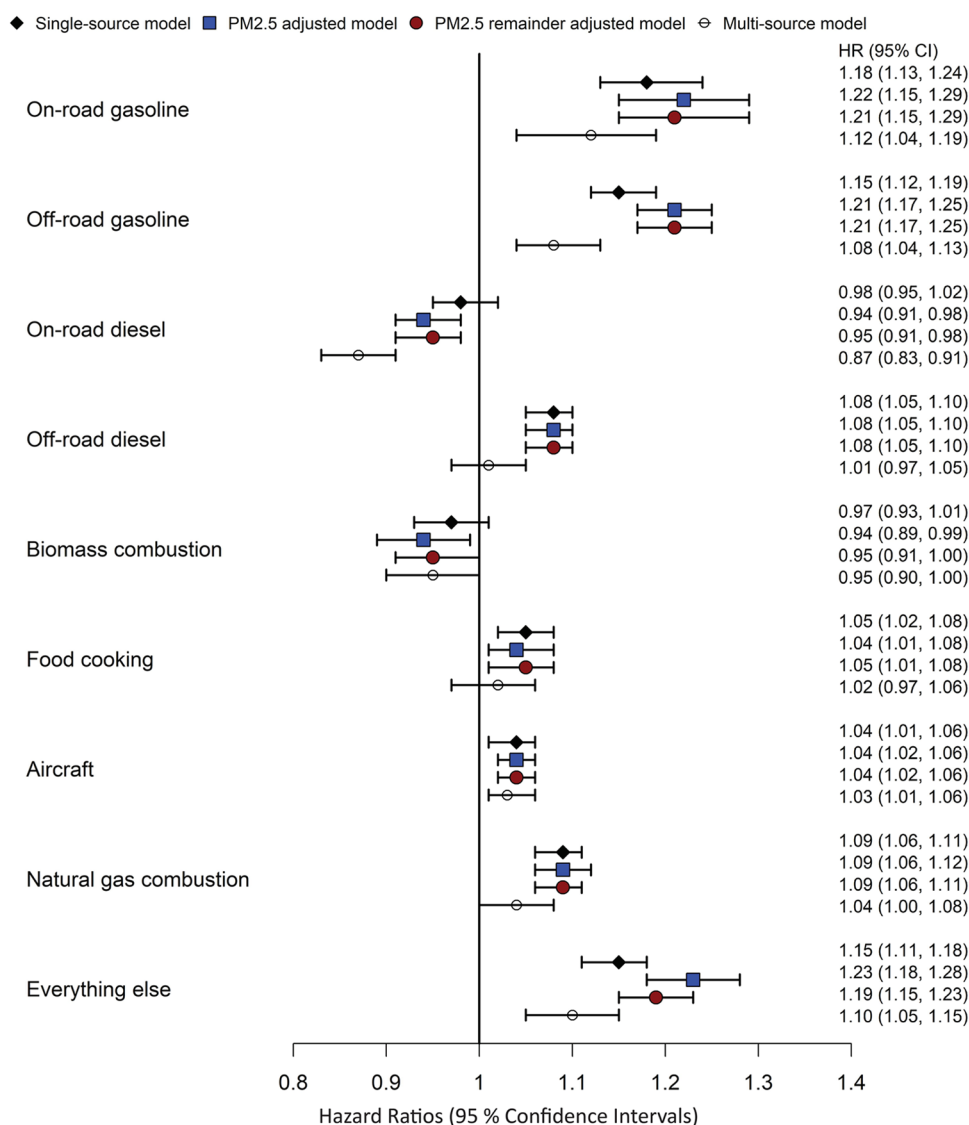


Figure 2. Hazard ratios (HR) of ASD scaled to standard deviation increases in each of the 9 source categories, estimated in single-, PM<sub>2.5</sub> adjusted-, and multipollutant models. Standard deviations for each of the sources are given in Table 2.

in the single-source model to a protective one in the PM<sub>2.5</sub> adjusted and multisource models. The HR for the total and remainder PM<sub>2.5</sub> variables are listed by the source model in Table S1. The HR for on-road gasoline, off-road diesel, food cooking, and natural gas combustion models were inclusive of

1.00. Furthermore, total or remainder PM<sub>2.5</sub> had an inverse association with ASD when included in the off-road gasoline and “other”/“everything else” source models. In contrast, total and remainder PM<sub>2.5</sub> maintained a significant association with ASD development in on-road diesel, biomass combustion, and

aircraft-source models. Note that on-road diesel and biomass combustion exposures themselves had an inverse association or null association with the ASD in the PM<sub>2.5</sub> adjusted models.

Including all 9 of the sources specified in this study in the same model (i.e., multipollutant model with different sources) resulted in noticeable attenuation of the HRs associated with some source-specific PM<sub>2.5</sub> exposures (Figure 2). The association of ASD with on-road gasoline, off-road gasoline, aircraft, and “other”/“everything else” remained statistically significant; the associations of ASD with off-road diesel, food cooking, and natural gas combustion all weakened and lost statistical significance. The inverse association with on-road diesel became stronger and more significant. In the multisource model, the ASD HRs associated with on-road gasoline, off-road gasoline, aircraft, and “other”/“everything else” were 1.12 (1.04, 1.19), 1.08 (1.04, 1.13), 1.03 (1.01, 1.06), and 1.10 (1.05, 1.15), respectively.

In the multisource model, the variance inflation factors (VIF) for each of the sources ranged from 1.16 to 4.77 (Table S2).

In sensitivity analyses, inclusion of SES and urbanicity factors did not materially change the effects estimates (Table S3) nor did analyses that included subjects with imprecise addresses excluded from the study population (Table S4).

Associations between aircraft, on-road, and off-road gasoline (i.e., the sources with positive significant associations in the multisource model) and ASD development were also examined per 1  $\mu\text{g}/\text{m}^3$  source concentration, which allows for comparisons of the potency of each aerosol type (Table S5). The HR results for these sources in a multisource model remain significant but are generally larger, as 1  $\mu\text{g}/\text{m}^3$  was larger than common population-relevant (SD) exposures.

## DISCUSSION

In this retrospective birth cohort of over 300,000 mother–child pairs in southern California, we found that prenatal exposures to PM<sub>2.5</sub> concentrations from various sources were associated with an increased ASD risk among children. These sources include on-road and off-road gasoline, aircraft, off-road diesel, food cooking, and natural gas combustion. Among these, the associations with on-road gasoline, off-road gasoline, and (previously reported) aircraft-sourced PM<sub>2.5</sub> remained significant in multipollutant models. Biomass combustion showed no associations, whereas on-road diesel showed an inverse effect in the multisource and PM<sub>2.5</sub> adjusted models.

Prior studies using different exposure modeling approaches, reported association between near-roadway air pollution (NRAP) and ASD,<sup>19</sup> but only for NRAP from nonfreeway sources.<sup>11</sup> Nonfreeway vehicular traffic was overwhelmingly gasoline-powered during the period of the study, and on-road diesel was largely used in heavy-duty vehicles on freeways. Therefore, the observed increased risk from gasoline sources is consistent with those previous findings.

In the single-source models, off-road, but not on-road, diesel was strongly associated with the risk of ASD development. This could be due to off-road diesel sources being more likely to be located in closer proximity to residences compared with on-road diesel sources in this cohort. Source-specific exposures across southern California using the UCD/CIT CTM model showed concurrent on-road diesel and gasoline peaks across downtown Los Angeles.<sup>42</sup> A secondary peak for the on-road diesel was found to the east of the city in an area notable for a heavy concentration of warehouses and heavy-duty truck

traffic. It is unclear why the on-road diesel combustion particles had an inverse association with ASD development in the multisource model, as evidence is strong that diesel exhaust particulate is neurotoxic.<sup>64,65</sup> Effects of multicollinearity is a potential reason for the discrepancy between the single-source and multisource models. In a previous study using exposure estimated from hazardous air pollutant (HAP) inventories, higher levels of diesel PM exposure, nondistinguishing between on- vs off-road sources, were associated with increased risk of ASD.<sup>66</sup>

Natural gas combustion and food cooking sources have not been previously shown to be associated with ASD, which we observed in all but the multisource model. A number of studies have examined the associations between natural gas production and other health outcomes;<sup>67–70</sup> the exposures in most studies, however, are not limited to natural gas combustion directly but include emissions from diesel generators, drilling processes, and fugitive releases of various chemicals. In other studies, exposure specifically to gas flaring was associated with adverse health effects.<sup>69,71</sup> Our study is the first to report the association of ASD (or any other child neurodevelopmental outcome) with prenatal exposure to natural gas combustion. Most studies on the effects of food cooking have focused on exposures in indoor settings, with inconsistent results across various health end points.<sup>72–74</sup> We found no significant effects of either natural gas or food PM<sub>2.5</sub> that were independent of the other sources (i.e., based on the results of the multisource model), although multicollinearity effects in the model could have attenuated the effect estimate and reduced the statistical significance. Aircraft-sourced PM, on the other hand, showed robust associations across all models. Associations of ultrafine aircraft PM and ASD were previously demonstrated in this cohort. These novel associations were described by Carter et al.<sup>10</sup> We additionally showed that this association is robust to adjustment with other sources.

We did not see any positive association between prenatal exposure to biomass combustion-derived PM<sub>2.5</sub> and ASD. Biomass combustion emission levels are driven mainly by wildfires, residential heating, and crop burning.<sup>75–77</sup> A prior California study, which utilized PM metrics from the same UCD/CIT CTM, found no association between long-term exposure to biomass combustion PM and mortality in the California Teachers Study Cohort.<sup>78</sup> Yet, a significant body of research found health effects of biomass combustion PM.<sup>79–82</sup> This includes a recent study in Sweden, which found increased ASD risk in association with residential wood burning.<sup>32</sup> Therefore, the absence of an association between biomass PM derived from the UCD/CIT CTM with ASD in this study, and with mortality previously, could stem from the model's inability to adequately represent the plume rise of wildfires or the sharp spatial gradients associated with residential wood combustion.

The “everything else”/“other” category, which was the largest single contributor to total PM<sub>2.5</sub> mass, had a strong association with ASD development across all models. This category includes windblown dust, construction dust, tire wear, brake wear, and other types of fugitive dust emissions. The Swedish study related exposure to a smaller set of sources to ASD diagnosis.<sup>32</sup> It was found that small-scale residential heating, tailpipe exhaust, and vehicle wear-and-tear were all associated with ASD risk. The traffic-related associations match well with our results and with previous work examining association with near-roadway air pollution and constituents of

tailpipe and nontailpipe emissions from roadways in this cohort using different exposure modeling strategies.<sup>11,16</sup> In those studies, tailpipe combustion tracers black and elemental carbons (BC and EC) had significant positive associations with ASD. Moreover, nontailpipe tracers Cu, Fe, and Mn also showed positive associations consistent with the results from the Swedish source study. An earlier study of the effects of PM components in this cohort also found associations of EC/BC and organic matter with ASD.<sup>15</sup> A few other studies have examined ASD association with several air toxics.<sup>66,83,84</sup>

In our study, gasoline, both on- and off-road, and aircraft were robustly associated with ASD. The effect sizes were based on population-relevant exposures (SD) specific to each source. To evaluate the potency of the particles for each of these three sources, HR results were examined per 1  $\mu\text{g}/\text{m}^3$  increase in PM<sub>2.5</sub> concentration (Table S2). The variability in effect size indicates that potency (per 1  $\mu\text{g}/\text{m}^3$ ) was larger for on-road gasoline and the least for aircraft particles. Nevertheless, these HR values are not substantially different from one another, suggesting that the potency of the particles was not markedly different.

There are numerous studies evaluating the toxicology of PM from various sources. Previous work has evaluated the effects of gasoline and aircraft particles in lung epithelial cells,<sup>85–89</sup> with evidence for cytotoxicity, oxidative stress, and alteration of cytokine production. Studies have focused on toxicological comparisons across different vehicle fuel types, namely, between gasoline, diesel, and biofuels.<sup>90–92</sup> One in China reported greater cytotoxicity and oxidative stress potential in gasoline vs diesel combustion particles.<sup>92</sup> Comparative studies have also examined other sets of sources.<sup>93–96</sup> For example, Karlsson et al.<sup>96</sup> compared DNA damage and cytokine production in lung epithelial cells exposed to wood combustion, tire and road wear, subway, and street-side particles. The strongest cytokine responses were driven by the street particles and DNA damage by the subway particles.

Total PM<sub>2.5</sub> mass by itself was associated with ASD, consistent with previous findings from this cohort.<sup>15–17,39</sup> When coadjusted with some source-specific PM<sub>2.5</sub>, however, both total PM<sub>2.5</sub> and remainder PM<sub>2.5</sub> had generally smaller coefficients of effect. This occurred when PM<sub>2.5</sub> was coadjusted for those sources with markedly large effects (e.g., on-road gasoline), but not for those with weaker effects (e.g., aircraft, biomass combustion, on-road diesel). Surprisingly, total and remainder PM<sub>2.5</sub> were inversely associated with ASD risk when adjusted for off-road gasoline and the “other”/“everything else” sources. Protective effects are not biologically plausible given robust toxicological and epidemiological evidence on neurodevelopmental toxicity, including on ASD. Collinearity of the source PM with the total (and remainder) PM<sub>2.5</sub> may explain the loss of significance in the coadjusted models. On-road gasoline and off-road diesel PM<sub>2.5</sub> had correlation coefficients with total PM<sub>2.5</sub> of 0.83 and 0.82, respectively. However, the variation inflation factors (VIF) for each source in the multisource model (Table S2) were not remarkably high (less than 5). Further investigation is warranted.

This study assessed associations of ASD with primary source-specific particulate matter (total PM<sub>2.5</sub> contained secondary PM<sub>2.5</sub> as well). Secondary particulate matter forms in the atmosphere from the reaction of gaseous precursor species, which can be emitted from many sources. Hypotheses about the health effects of primary particulate

matter are easier to test, and the sources that are associated with negative health effects are potential targets for regulations. Hypotheses about the health effects of secondary particulate matter are more difficult to evaluate, and emissions controls to limit secondary particulate matter require a thorough understanding of formation mechanisms. A thorough evaluation of the sources and health effects of secondary particulate matter is beyond the scope of our study analysis.

A major strength of the study is the large and diverse cohort, representative of a major region of the United States.<sup>37</sup> A wealth of demographic and health data are available from the established EMR system maintained by Kaiser Permanente. The CTM model has been employed in several epidemiological studies.<sup>15,16,42,78,97–99</sup> This is a locally constructed and validated model used extensively in California and is likely to provide the most accurate and relevant results for this area. We were also able to adjust the estimated exposure to the residential movement of the subjects.

There are some limitations to the study. The CTM estimates outdoor ambient exposures for all locations; it does not estimate indoor exposures, which may be a significant contributor to total PM exposure. A lack of information on time spent in indoor and outdoor activities may have resulted in misclassification of the exposure estimate. However, unless there were consistent patterns of time–activity or indoor exposures that were different in the ASD and non-ASD groups during pregnancy, it is unlikely that missing this information would bias the effects differentially. PM<sub>2.5</sub> source data used in this analysis were available at 4 km spatial resolution, which limited the ability to assess fine-scale variability of sources. The consequence of this exposure misclassification for epidemiological investigations due to coarser spatial resolution is likely to underestimate effects because the misclassification will likely be nondifferential to the outcome. Finally, there were additional covariates for which we have not adjusted in the models. The consumption of folic acid has been related to ASD development<sup>100</sup> but likely has no association with air pollution levels. Detailed dietary information was not available in the EMR. Furthermore, effects of maternal smoking have been previously assessed in this cohort, finding no association with ASD.<sup>41</sup> Accordingly, smoking was left out of the analysis because it is an unlikely confounder. Outside work on this relationship has yielded mixed results.<sup>101</sup> Other factors such as maternal medication use and the occurrence of various birth-related events (e.g., cesarean section) previously investigated<sup>102,103</sup> were not assessed, as these variables may also be mediators instead of confounders.

Impacts of prenatal gasoline and aircraft PM exposure on ASD in this study provide targets for future toxicological and other research. A focus could be on how the individual components or the whole mixture of these sources affect prenatal neurodevelopment in toxicological and population studies. Our results are based on population-relevant exposures, demonstrating the public health importance of these exposures and effects with regulatory implications. Overall, these results are consistent with emerging evidence that combustion products from different sources have different biological effects. This suggests that regulation of PM<sub>2.5</sub> might provide greater benefits if it is focused on reducing more toxic sources. However, there is strong evidence that PM<sub>2.5</sub> mass, irrespective of components or sources, has health effects and an emerging body of literature that reductions in general PM<sub>2.5</sub> result in improvements in a host of health outcomes.



In summary, this study provides evidence that the impacts of prenatal exposure to PM<sub>2.5</sub> on ASD, an important neurodevelopmental outcome, vary by the source. The most consistent associations were observed for on- and off-road gasoline in southern California. Associations were also observed with food cooking and natural gas combustion in the single-source and PM<sub>2.5</sub> adjusted models; however, these associations were weakened in the models adjusted for multiple sources. On-road diesel and biomass combustion were not associated with ASD. Results also identify PM source mixtures for toxicological investigations that may provide evidence for future public health policies.

## ■ ASSOCIATED CONTENT

### Data Availability Statement

Dr. Xiang had full access to all of the data in the study. Drs. Luglio, Xiang, and Rahman take responsibility for the integrity of the data and accuracy of the data analysis.

### Supporting Information

The Supporting Information is available free of charge at <https://pubs.acs.org/doi/10.1021/acs.est.4c05563>.

Additional results in the forms of tables and figures, including sensitivity analysis and single figure showing the study sample derivation (PDF)

## ■ AUTHOR INFORMATION

### Corresponding Authors

**Anny H. Xiang** – Department of Research & Evaluation, Kaiser Permanente Southern California, Pasadena, California 91101, United States; [orcid.org/0000-0003-2786-1268](https://orcid.org/0000-0003-2786-1268); Email: [anny.h.xiang@kp.org](mailto:anny.h.xiang@kp.org)

**Md Mostafijur Rahman** – Department of Environmental Health Sciences, Tulane University School of Public Health and Tropical Medicine, New Orleans, Louisiana 70118, United States; Department of Population and Public Health Sciences, Keck School of Medicine, University of Southern California, Los Angeles, California 90089, United States; [orcid.org/0000-0002-2405-3276](https://orcid.org/0000-0002-2405-3276); Email: [mrahman8@tulane.edu](mailto:mrahman8@tulane.edu)

### Authors

**David G. Luglio** – Department of Environmental Health Sciences, Tulane University School of Public Health and Tropical Medicine, New Orleans, Louisiana 70118, United States; [orcid.org/0000-0003-4669-547X](https://orcid.org/0000-0003-4669-547X)

**Michael J. Kleeman** – Department of Civil and Environmental Engineering, University of California, Davis, Davis, California 95616, United States; [orcid.org/0000-0002-0347-7512](https://orcid.org/0000-0002-0347-7512)

**Xin Yu** – Spatial Science Institute, University of Southern California, Los Angeles, California 90089, United States; [orcid.org/0000-0001-5118-2324](https://orcid.org/0000-0001-5118-2324)

**Jane C. Lin** – Department of Research & Evaluation, Kaiser Permanente Southern California, Pasadena, California 91101, United States

**Ting Chow** – Department of Research & Evaluation, Kaiser Permanente Southern California, Pasadena, California 91101, United States

**Mayra P. Martinez** – Department of Research & Evaluation, Kaiser Permanente Southern California, Pasadena, California 91101, United States

**Zhanghua Chen** – Department of Population and Public Health Sciences, Keck School of Medicine, University of

Southern California, Los Angeles, California 90089, United States

**Jiu-Chiuan Chen** – Department of Population and Public Health Sciences, Keck School of Medicine, University of Southern California, Los Angeles, California 90089, United States

**Sandrah Proctor Eckel** – Department of Population and Public Health Sciences, Keck School of Medicine, University of Southern California, Los Angeles, California 90089, United States

**Joel Schwartz** – Department of Environmental Health, Harvard T.H. Chan School of Public Health, Boston, Massachusetts 02115, United States; Department of Epidemiology, Harvard T.H. Chan School of Public Health, Boston, Massachusetts 02115, United States

**Frederick Lurmann** – Sonoma Technology, Inc., Petaluma, California 94954, United States

**Rob McConnell** – Department of Population and Public Health Sciences, Keck School of Medicine, University of Southern California, Los Angeles, California 90089, United States

Complete contact information is available at: <https://pubs.acs.org/10.1021/acs.est.4c05563>

### Funding

This study was supported by National Institutes of Environmental Health Sciences (R01 ES029963 (Xiang, McConnell); R56ES028121 (Xiang); P30ES007048 and P2CES033433 (McConnell)) and by Kaiser Permanente Southern California Direct Community Benefit Funds. Joel Schwartz was supported by EPA Grant RD-835872.

### Notes

The authors declare no competing financial interest.

**Role of the Funder/Sponsor** The funders had no role in the design and conduct of the study; collection, management, analysis, and interpretation of the data; preparation, review, or approval of the manuscript; and decision to submit the manuscript for publication.

J.S. has testified on behalf of the U.S. Department of Justice in a case involving a Clean Air Act violation.

**Data Sharing Statement** KPSC Institutional Review Board approved this study, with waiver of informed consent with the condition that raw data remain confidential and would not be shared. Thus, due to the sensitive nature of these data, the data are not available to be shared.

## ■ ACKNOWLEDGMENTS

The authors thank the patients of Kaiser Permanente for helping them improve care through the use of information collected through their electronic health record systems, and the Kaiser Permanente and the Utility for Care Data Analysis (UCDA) team within Kaiser Permanente for creating the GEMS Datamart with consolidated addresses histories available to facilitate the research.

## ■ REFERENCES

(1) Neurodevelopmental Disorders. In *Diagnostic and Statistical Manual of Mental Disorders*; American Psychiatric Association: Arlington, VA, 2022.

(2) Hodges, H.; Fealko, C.; Soares, N. Autism spectrum disorder: definition, epidemiology, causes, and clinical evaluation. *Transl. Pediatr.* **2020**, *9*, S55–S65.

- (3) Lord, C.; Elsabbagh, M.; Baird, G.; Veenstra-Vanderweele, J. Autism spectrum disorder. *Lancet* **2018**, *392* (10146), 508–520.
- (4) Bacon, E. C.; Courchesne, E.; Barnes, C. C.; Cha, D.; Pence, S.; Schreibman, L.; Stahmer, A. C.; Pierce, K. Rethinking the idea of late autism spectrum disorder onset. *Dev. Psychopathol.* **2018**, *30* (2), 553–569.
- (5) Charman, T.; Baird, G. Practitioner review: Diagnosis of autism spectrum disorder in 2- and 3-year-old children. *J. Child Psychol. Psychiatry* **2002**, *43* (3), 289–305.
- (6) van 't Hof, M.; Tisseur, C.; van Berckeleer-Onnes, I.; van Nieuwenhuyzen, A.; Daniels, A. M.; Deen, M.; Hoek, H. W.; Ester, W. A. Age at autism spectrum disorder diagnosis: A systematic review and meta-analysis from 2012 to 2019. *Autism* **2021**, *25* (4), 862–873.
- (7) Eyring, K. W.; Geschwind, D. H. Three decades of ASD genetics: building a foundation for neurobiological understanding and treatment. *Hum. Mol. Genet.* **2021**, *30* (20), R236–R244.
- (8) Tick, B.; Bolton, P.; Happé, F.; Rutter, M.; Rijdsdijk, F. Heritability of autism spectrum disorders: a meta-analysis of twin studies. *J. Child Psychol. Psychiatry* **2016**, *57* (5), 585–595.
- (9) Becerra, T. A.; Wilhelm, M.; Olsen, J.; Cockburn, M.; Ritz, B. Ambient Air Pollution and Autism in Los Angeles County, California. *Environ. Health Perspect.* **2013**, *121* (3), 380–386.
- (10) Carter, S. A.; Rahman, M. M.; Lin, J. C.; Chow, T.; Yu, X.; Martinez, M. P.; Levitt, P.; Chen, Z.; Chen, J.-C.; Eckel, S. P.; Schwartz, J.; Lurmann, F. W.; Kleeman, M. J.; McConnell, R.; Xiang, A. H. Maternal exposure to aircraft emitted ultrafine particles during pregnancy and likelihood of ASD in children. *Environ. Int.* **2023**, *178*, No. 108061.
- (11) Carter, S. A.; Rahman, M. M.; Lin, J. C.; Shu, Y. H.; Chow, T.; Yu, X.; Martinez, M. P.; Eckel, S. P.; Chen, J. C.; Chen, Z.; Schwartz, J.; Pavlovic, N.; Lurmann, F. W.; McConnell, R.; Xiang, A. H. In utero exposure to near-roadway air pollution and autism spectrum disorder in children. *Environ. Int.* **2022**, *158*, No. 106898.
- (12) Kalkbrenner, A. E.; Windham, G. C.; Serre, M. L.; Akita, Y.; Wang, X.; Hoffman, K.; Thayer, B. P.; Daniels, J. L. Particulate matter exposure, prenatal and postnatal windows of susceptibility, and autism spectrum disorders. *Epidemiology* **2015**, *26* (1), 30–42.
- (13) McGuinn, L. A.; Windham, G. C.; Kalkbrenner, A. E.; Bradley, C.; Di, Q.; Croen, L. A.; Fallin, M. D.; Hoffman, K.; Ladd-Acosta, C.; Schwartz, J.; Rappold, A. G.; Richardson, D. B.; Neas, L. M.; Gammon, M. D.; Schieve, L. A.; Daniels, J. L. Early Life Exposure to Air Pollution and Autism Spectrum Disorder: Findings from a Multisite Case-Control Study. *Epidemiology* **2020**, *31* (1), 103–114.
- (14) Pagalan, L.; Bickford, C.; Weikum, W.; Lanphear, B.; Brauer, M.; Lanphear, N.; Hanley, G. E.; Oberlander, T. F.; Winters, M. Association of Prenatal Exposure to Air Pollution With Autism Spectrum Disorder. *JAMA Pediatr.* **2019**, *173* (1), 86–92.
- (15) Rahman, M. M.; Carter, S. A.; Lin, J. C.; Chow, T.; Yu, X.; Martinez, M. P.; Chen, Z.; Chen, J. C.; Rud, D.; Lewinger, J. P.; van Donkelaar, A.; Martin, R. V.; Eckel, S. P.; Schwartz, J.; Lurmann, F.; Kleeman, M. J.; McConnell, R.; Xiang, A. H. Associations of Autism Spectrum Disorder with PM(2.5) Components: A Comparative Study Using Two Different Exposure Models. *Environ. Sci. Technol.* **2023**, *57* (1), 405–414.
- (16) Rahman, M. M.; Carter, S. A.; Lin, J. C.; Chow, T.; Yu, X.; Martinez, M. P.; Levitt, P.; Chen, Z.; Chen, J. C.; Rud, D.; Lewinger, J. P.; Eckel, S. P.; Schwartz, J.; Lurmann, F. W.; Kleeman, M. J.; McConnell, R.; Xiang, A. H. Prenatal exposure to tailpipe and non-tailpipe tracers of particulate matter pollution and autism spectrum disorders. *Environ. Int.* **2023**, *171*, No. 107736.
- (17) Rahman, M. M.; Shu, Y.-H.; Chow, T.; Lurmann, F. W.; Yu, X.; Martinez, M. P.; Carter, S. A.; Eckel, S. P.; Chen, J.-C.; Chen, Z.; Levitt, P.; Schwartz, J.; McConnell, R.; Xiang, A. H. Prenatal Exposure to Air Pollution and Autism Spectrum Disorder: Sensitive Windows of Exposure and Sex Differences. *Environ. Health Perspect.* **2022**, *130* (1), No. 017008.
- (18) Raz, R.; Roberts, A. L.; Lyall, K.; Hart, J. E.; Just, A. C.; Laden, F.; Weisskopf, M. G. Autism spectrum disorder and particulate matter air pollution before, during, and after pregnancy: a nested case-control analysis within the Nurses' Health Study II Cohort. *Environ. Health Perspect.* **2015**, *123* (3), 264–270.
- (19) Volk, H. E.; Lurmann, F.; Penfold, B.; Hertz-Picciotto, I.; McConnell, R. Traffic-related air pollution, particulate matter, and autism. *JAMA Psychiatry* **2013**, *70* (1), 71–77.
- (20) Yousefian, F.; Mahvi, A. H.; Yunesian, M.; Hassanvand, M. S.; Kashani, H.; Amini, H. Long-term exposure to ambient air pollution and autism spectrum disorder in children: A case-control study in Tehran, Iran. *Sci. Total Environ.* **2018**, *643*, 1216–1222.
- (21) Bell, M. L.; Dominici, F.; Ebisu, K.; Zeger, S. L.; Samet, J. M. Spatial and Temporal Variation in PM2.5 Chemical Composition in the United States for Health Effects Studies. *Environ. Health Perspect.* **2007**, *115* (7), 989–995.
- (22) Hasheminassab, S.; Daher, N.; Saffari, A.; Wang, D.; Ostro, B. D.; Sioutas, C. Spatial and temporal variability of sources of ambient fine particulate matter (PM2.5) in California. *Atmos. Chem. Phys.* **2014**, *14* (22), 12085–12097.
- (23) Martins, V.; Faria, T.; Diapouli, E.; Manousakas, M. I.; Eleftheriadis, K.; Viana, M.; Almeida, S. M. Relationship between indoor and outdoor size-fractionated particulate matter in urban microenvironments: Levels, chemical composition and sources. *Environ. Res.* **2020**, *183*, No. 109203.
- (24) Thurston, G. D.; Ito, K.; Lall, R. A source apportionment of U.S. fine particulate matter air pollution. *Atmos. Environ.* **2011**, *45* (24), 3924–3936.
- (25) Zhang, X. Y.; Wang, Y. Q.; Niu, T.; Zhang, X. C.; Gong, S. L.; Zhang, Y. M.; Sun, J. Y. Atmospheric aerosol compositions in China: spatial/temporal variability, chemical signature, regional haze distribution and comparisons with global aerosols. *Atmos. Chem. Phys.* **2012**, *12* (2), 779–799.
- (26) Dai, L.; Zanutti, A.; Koutrakis, P.; Schwartz, J. D. Associations of fine particulate matter species with mortality in the United States: a multicity time-series analysis. *Environ. Health Perspect.* **2014**, *122* (8), 837–842.
- (27) Karri, V.; Kumar, V.; Ramos, D.; Oliveira, E.; Schuhmacher, M. An in vitro cytotoxic approach to assess the toxicity of heavy metals and their binary mixtures on hippocampal HT-22 cell line. *Toxicol. Lett.* **2018**, *282*, 25–36.
- (28) Kim, S.; Kang, K.; Kim, H.; Seo, M. In Vitro Toxicity Screening of Fifty Complex Mixtures in HepG2 Cells. *Toxics* **2024**, *12* (2), No. 126.
- (29) Kioumourtzoglou, M. A.; Austin, E.; Koutrakis, P.; Dominici, F.; Schwartz, J.; Zanobetti, A. PM2.5 and survival among older adults: effect modification by particulate composition. *Epidemiology* **2015**, *26* (3), 321–327.
- (30) Krall, J. R.; Anderson, G. B.; Dominici, F.; Bell, M. L.; Peng, R. D. Short-term Exposure to Particulate Matter Constituents and Mortality in a National Study of U.S. Urban Communities. *Environ. Health Perspect.* **2013**, *121* (10), 1148–1153.
- (31) Thurston, G. D.; Burnett, R. T.; Turner, M. C.; Shi, Y.; Krewski, D.; Lall, R.; Ito, K.; Jerrett, M.; Gapstur, S. M.; Diver, W. R.; Pope, C. A. Ischemic Heart Disease Mortality and Long-Term Exposure to Source-Related Components of U.S. Fine Particle Air Pollution. *Environ. Health Perspect.* **2016**, *124* (6), 785–794.
- (32) Flanagan, E.; Malmqvist, E.; Rittner, R.; Gustafsson, P.; Källén, K.; Oudin, A. Exposure to local, source-specific ambient air pollution during pregnancy and autism in children: a cohort study from southern Sweden. *Sci. Rep.* **2023**, *13* (1), No. 3848.
- (33) Lippmann, M.; Chen, L. C.; Gordon, T.; Ito, K.; Thurston, G. D. National Particle Component Toxicity (NPACT) Initiative: integrated epidemiologic and toxicologic studies of the health effects of particulate matter components. *Res. Rep. Health Eff. Inst.* **2013**, No. 177, 5–13.
- (34) Vedal, S.; Campen, M.; McDonald, J.; Larson, T.; Sampson, P.; Sheppard, L.; Simpson, C.; Szpiro, A. *National Particle Component Toxicity (NPACT) Initiative Report on Cardiovascular Effects*, Research Report 178; Health Effects Institute: Boston, MA, 2013.
- (35) Adams, K.; Greenbaum, D. S.; Shaikh, R.; van Erp, A. M.; Russell, A. G. Particulate matter components, sources, and health:

- Systematic approaches to testing effects. *J. Air Waste Manage. Assoc.* **2015**, *65* (5), 544–558.
- (36) Stanek, L. W.; Sacks, J. D.; Dutton, S. J.; Dubois, J.-J. B. Attributing health effects to apportioned components and sources of particulate matter: An evaluation of collective results. *Atmos. Environ.* **2011**, *45* (32), 5655–5663.
- (37) Koebnick, C.; Langer-Gould, A. M.; Gould, M. K.; Chao, C. R.; Iyer, R. L.; Smith, N.; Chen, W.; Jacobsen, S. J. Sociodemographic Characteristics of Members of a Large, Integrated Health Care System: Comparison with US Census Bureau Data. *Perm. J.* **2012**, *16* (3), 37–41.
- (38) Coleman, K. J.; Lutsky, M. A.; Yau, V.; Qian, Y.; Pomichowski, M. E.; Crawford, P. M.; Lynch, F. L.; Madden, J. M.; Owen-Smith, A.; Pearson, J. A.; Pearson, K. A.; Rusinak, D.; Quinn, V. P.; Croen, L. A. Validation of Autism Spectrum Disorder Diagnoses in Large Healthcare Systems with Electronic Medical Records. *J. Autism Dev. Disord.* **2015**, *45* (7), 1989–1996.
- (39) Jo, H.; Eckel, S. P.; Wang, X.; Chen, J.-C.; Cockburn, M.; Martinez, M. P.; Chow, T.; Molshatzki, N.; Lurmann, F. W.; Funk, W. E.; Xiang, A. H.; McConnell, R. Sex-specific associations of autism spectrum disorder with residential air pollution exposure in a large Southern California pregnancy cohort. *Environ. Pollut.* **2019**, *254*, No. 113010.
- (40) Xiang, A. H.; Wang, X.; Martinez, M. P.; Page, K.; Buchanan, T. A.; Feldman, R. K. Maternal Type 1 Diabetes and Risk of Autism in Offspring. *JAMA* **2018**, *320* (1), 89–91.
- (41) Xiang, A. H.; Wang, X.; Martinez, M. P.; Walthall, J. C.; Curry, E. S.; Page, K.; Buchanan, T. A.; Coleman, K. J.; Getahun, D. Association of Maternal Diabetes With Autism in Offspring. *JAMA* **2015**, *313* (14), 1425–1434.
- (42) Yu, X.; Venecek, M.; Kumar, A.; Hu, J.; Tanrikulu, S.; Soon, S. T.; Tran, C.; Fairley, D.; Kleeman, M. J. Regional sources of airborne ultrafine particle number and mass concentrations in California. *Atmos. Chem. Phys.* **2019**, *19* (23), 14677–14702.
- (43) Hu, J.; Zhang, H.; Chen, S.-H.; Wiedinmyer, C.; Vandenberghe, F.; Ying, Q.; Kleeman, M. J. Predicting Primary PM<sub>2.5</sub> and PM<sub>0.1</sub> Trace Composition for Epidemiological Studies in California. *Environ. Sci. Technol.* **2014**, *48* (9), 4971–4979.
- (44) Hu, J.; Zhang, H.; Ying, Q.; Chen, S. H.; Vandenberghe, F.; Kleeman, M. J. Long-term particulate matter modeling for health effect studies in California – Part 1: Model performance on temporal and spatial variations. *Atmos. Chem. Phys.* **2015**, *15* (6), 3445–3461.
- (45) Venecek, M. A.; Yu, X.; Kleeman, M. J. Predicted ultrafine particulate matter source contribution across the continental United States during summertime air pollution events. *Atmos. Chem. Phys.* **2019**, *19* (14), 9399–9412.
- (46) Kleeman, M. J.; Robert, M. A.; Riddle, S. G.; Fine, P. M.; Hays, M. D.; Schauer, J. J.; Hannigan, M. P. Size distribution of trace organic species emitted from biomass combustion and meat charbroiling. *Atmos. Environ.* **2008**, *42* (13), 3059–3075.
- (47) Riddle, S. G.; Robert, M. A.; Jakober, C. A.; Hannigan, M. P.; Kleeman, M. J. Size Distribution of Trace Organic Species Emitted from Light-Duty Gasoline Vehicles. *Environ. Sci. Technol.* **2007**, *41* (21), 7464–7471.
- (48) Riddle, S. G.; Robert, M. A.; Jakober, C. A.; Hannigan, M. P.; Kleeman, M. J. Size Distribution of Trace Organic Species Emitted from Heavy-Duty Diesel Vehicles. *Environ. Sci. Technol.* **2007**, *41* (6), 1962–1969.
- (49) Robert, M. A.; Kleeman, M. J.; Jakober, C. A. Size and composition distributions of particulate matter emissions: part 2–heavy-duty diesel vehicles. *J. Air Waste Manage. Assoc.* **2007**, *57* (12), 1429–1438.
- (50) Robert, M. A.; VanBergen, S.; Kleeman, M. J.; Jakober, C. A. Size and composition distributions of particulate matter emissions: part 1–light-duty gasoline vehicles. *J. Air Waste Manage. Assoc.* **2007**, *57* (12), 1414–1428.
- (51) Xue, J.; Li, Y.; Peppers, J.; Wan, C.; Kado, N. Y.; Green, P. G.; Young, T. M.; Kleeman, M. J. Ultrafine Particle Emissions from Natural Gas, Biogas, and Biomethane Combustion. *Environ. Sci. Technol.* **2018**, *52* (22), 13619–13628.
- (52) Kleeman, M. J.; Cass, G. R. Source contributions to the size and composition distribution of urban particulate air pollution. *Atmos. Environ.* **1998**, *32* (16), 2803–2816.
- (53) Hu, J.; Zhang, H.; Chen, S.; Ying, Q.; Wiedinmyer, C.; Vandenberghe, F.; Kleeman, M. J. Identifying PM<sub>2.5</sub> and PM<sub>0.1</sub> Sources for Epidemiological Studies in California. *Environ. Sci. Technol.* **2014**, *48* (9), 4980–4990.
- (54) Ritz, B.; Liew, Z.; Yan, Q.; Cuia, X.; Virk, J.; Ketznel, M.; Raaschou-Nielsen, O. Air pollution and autism in Denmark. *Environ. Epidemiol.* **2018**, *2* (4), No. e028, DOI: 10.1097/EE9.000000000000028.
- (55) Bekkar, B.; Pacheco, S.; Basu, R.; DeNicola, N. Association of Air Pollution and Heat Exposure With Preterm Birth, Low Birth Weight, and Stillbirth in the US: A Systematic Review. *JAMA Network Open* **2020**, *3* (6), No. e208243.
- (56) Chang, Y.-S.; Chen, L.-W.; Yu, T.; Lin, S.-H.; Kuo, P.-L. Preterm birth and weight-for-gestational age for risks of autism spectrum disorder and intellectual disability: A nationwide population-based cohort study. *J. Formosan Med. Assoc.* **2023**, *122* (6), 493–504.
- (57) Gardener, H.; Spiegelman, D.; Buka, S. L. Perinatal and Neonatal Risk Factors for Autism: A Comprehensive Meta-analysis. *Pediatrics* **2011**, *128* (2), 344–355.
- (58) Yu, X.; Rahman, M. M.; Carter, S. A.; Lin, J. C.; Chow, T.; Lurmann, F. W.; Chen, J.-C.; Martinez, M. P.; Schwartz, J.; Eckel, S. P.; Chen, Z.; McConnell, R.; Xiang, A. H.; Hackman, D. A. Neighborhood Disadvantage and Autism Spectrum Disorder in a Population With Health Insurance. *JAMA Psychiatry* **2024**, *81* (2), 209–213.
- (59) Paykin, S.; Menghaney, M.; Lin, Q.; Kolak, M. Rural, Suburban, Urban Classification for Small Area Analysis. *Healthy Reg. Policies Lab* **2021**, 1–13.
- (60) USDA. Rural-Urban Commuting Area Codes 2019. <https://www.ers.usda.gov/data-products/rural-urban-commuting-area-codes/>. 09/25/23 [cited 2024 07/29/24].
- (61) Hsu, Y.-H.; Chen, C.-W.; Lin, Y.-J.; Li, C.-Y. Urban–Rural Disparity in the Incidence of Diagnosed Autism Spectrum Disorder in Taiwan: A 10-Year National Birth Cohort Follow-up Study. *J. Autism Dev. Disord.* **2023**, *53* (5), 2127–2137.
- (62) Lauritsen, M. B.; Astrup, A.; Pedersen, C. B.; Obel, C.; Schendel, D. E.; Schieve, L.; Yeargin-Allsopp, M.; Parner, E. T. Urbanicity and autism spectrum disorders. *J. Autism Dev. Disord.* **2014**, *44* (2), 394–404.
- (63) R Core Team. *R: A Language and Environment for Statistical Computing*; R Foundation for Statistical Computing: Vienna, Austria, 2021.
- (64) Costa, L. G.; Cole, T. B.; Coburn, J.; Chang, Y.-C.; Dao, K.; Roqué, P. J. Neurotoxicity of traffic-related air pollution. *Neurotoxicology* **2017**, *59*, 133–139.
- (65) Weitekamp, C. A.; Kerr, L. B.; Dishaw, L.; Nichols, J.; Lein, M.; Stewart, M. J. A systematic review of the health effects associated with the inhalation of particle-filtered and whole diesel exhaust. *Inhalation Toxicol.* **2020**, *32* (1), 1–13.
- (66) Windham, G. C.; Zhang, L.; Gunier, R.; Croen, L. A.; Grether, J. K. Autism spectrum disorders in relation to distribution of hazardous air pollutants in the san francisco bay area. *Environ. Health Perspect.* **2006**, *114* (9), 1438–1444.
- (67) Adgate, J. L.; Goldstein, B. D.; McKenzie, L. M. Potential Public Health Hazards, Exposures and Health Effects from Unconventional Natural Gas Development. *Environ. Sci. Technol.* **2014**, *48* (15), 8307–8320.
- (68) Werner, A. K.; Vink, S.; Watt, K.; Jagals, P. Environmental health impacts of unconventional natural gas development: A review of the current strength of evidence. *Sci. Total Environ.* **2015**, *505*, 1127–1141.

- (69) Willis, M.; Hystad, P.; Denham, A.; Hill, E. Natural gas development, flaring practices and paediatric asthma hospitalizations in Texas. *Int. J. Epidemiol.* **2021**, *49* (6), 1883–1896.
- (70) McKenzie, L. M.; Guo, R.; Witter, R. Z.; Savitz, D. A.; Newman, L. S.; Adgate, J. L. Birth Outcomes and Maternal Residential Proximity to Natural Gas Development in Rural Colorado. *Environ. Health Perspect.* **2014**, *122* (4), 412–417.
- (71) Motte, J.; Alvarenga, R. A. F.; Thybaut, J. W.; Dewulf, J. Quantification of the global and regional impacts of gas flaring on human health via spatial differentiation. *Environ. Pollut.* **2021**, *291*, No. 118213.
- (72) Cole-Hunter, T.; Dhingra, R.; Fedak, K. M.; Good, N.; L'Orange, C.; Luckasen, G.; Mehaffy, J.; Walker, E.; Wilson, A.; Balmes, J.; Brook, R. D.; Clark, M. L.; Devlin, R. B.; Volckens, J.; Peel, J. L. Short-term differences in cardiac function following controlled exposure to cookstove air pollution: The subclinical tests on volunteers exposed to smoke (STOVES) study. *Environ. Int.* **2021**, *146*, No. 106254.
- (73) Fedak, K. M.; Good, N.; Walker, E. S.; Balmes, J.; Brook, R. D.; Clark, M. L.; Cole-Hunter, T.; Devlin, R.; L'Orange, C.; Luckasen, G.; Mehaffy, J.; Shelton, R.; Wilson, A.; Volckens, J.; Peel, J. L. Acute changes in lung function following controlled exposure to cookstove air pollution in the subclinical tests of volunteers exposed to smoke (STOVES) study. *Inhalation Toxicol.* **2020**, *32* (3), 115–123.
- (74) Lee, T.; Gany, F. Cooking oil fumes and lung cancer: a review of the literature in the context of the U.S. population. *J. Immigr. Minor. Health* **2013**, *15* (3), 646–652.
- (75) Burke, M.; Childs, M. L.; de la Cuesta, B.; Qiu, M.; Li, J.; Gould, C. F.; Heft-Neal, S.; Wara, M. The contribution of wildfire to PM<sub>2.5</sub> trends in the USA. *Nature* **2023**, *622* (7984), 761–766.
- (76) Wagner, J.; Naik-Patel, K.; Wall, S.; Harnly, M. Measurement of ambient particulate matter concentrations and particle types near agricultural burns using electron microscopy and passive samplers. *Atmos. Environ.* **2012**, *54*, 260–271.
- (77) Yap, P. S.; Garcia, C. Effectiveness of residential wood-burning regulation on decreasing particulate matter levels and hospitalizations in the San Joaquin Valley Air Basin. *Am. J. Public Health* **2015**, *105* (4), 772–778.
- (78) Ostro, B.; Hu, J.; Goldberg, D.; Reynolds, P.; Hertz, A.; Bernstein, L.; Kleeman, M. J. Associations of mortality with long-term exposures to fine and ultrafine particles, species and sources: results from the California Teachers Study Cohort. *Environ. Health Perspect.* **2015**, *123* (6), 549–556.
- (79) Aguilera, R.; Corringham, T.; Gershunov, A.; Benmarhnia, T. Wildfire smoke impacts respiratory health more than fine particles from other sources: observational evidence from Southern California. *Nat. Commun.* **2021**, *12* (1), No. 1493.
- (80) Chen, K.; Ma, Y.; Bell, M. L.; Yang, W. Canadian Wildfire Smoke and Asthma Syndrome Emergency Department Visits in New York City. *JAMA* **2023**, *330* (14), 1385–1387.
- (81) Thurston, G.; Yu, W.; Luglio, D. An Evaluation of the Asthma Impact of the June 2023 New York City Wildfire Air Pollution Episode. *Am. J. Respir. Crit. Care Med.* **2023**, *208* (8), 898–900.
- (82) Zhang, B.; Weuve, J.; Langa, K. M.; D'Souza, J.; Szpiro, A.; Faul, J.; de Leon, C. M.; Gao, J.; Kaufman, J. D.; Sheppard, L.; Lee, J.; Kobayashi, L. C.; Hirth, R.; Adar, S. D. Comparison of Particulate Air Pollution From Different Emission Sources and Incident Dementia in the US. *JAMA Intern. Med.* **2023**, *183* (10), 1080–1089.
- (83) Kalkbrenner, A. E.; Daniels, J. L.; Chen, J. C.; Poole, C.; Emch, M.; Morrissey, J. Perinatal exposure to hazardous air pollutants and autism spectrum disorders at age 8. *Epidemiology* **2010**, *21* (5), 631–641.
- (84) von Ehrenstein, O. S.; Aralis, H.; Cockburn, M.; Ritz, B. In utero exposure to toxic air pollutants and risk of childhood autism. *Epidemiology* **2014**, *25* (6), 851–858.
- (85) Durga, M.; Nathiya, S.; Rajasekar, A.; Devasena, T. Effects of ultrafine petrol exhaust particles on cytotoxicity, oxidative stress generation, DNA damage and inflammation in human A549 lung cells and murine RAW 264.7 macrophages. *Environ. Toxicol. Pharmacol.* **2014**, *38* (2), 518–530.
- (86) He, R.-W.; Gerlofs-Nijland, M. E.; Boere, J.; Fokkens, P.; Leseman, D.; Janssen, N. A. H.; Cassee, F. R. Comparative toxicity of ultrafine particles around a major airport in human bronchial epithelial (Calu-3) cell model at the air–liquid interface. *Toxicol. In Vitro* **2020**, *68*, No. 104950.
- (87) Jonsdottir, H. R.; Delaval, M.; Leni, Z.; Keller, A.; Brem, B. T.; Siegerist, F.; Schönerberger, D.; Durdina, L.; Elser, M.; Burtscher, H.; Liati, A.; Geiser, M. Non-volatile particle emissions from aircraft turbine engines at ground-idle induce oxidative stress in bronchial cells. *Commun. Biol.* **2019**, *2*, No. 90.
- (88) Künzi, L.; Krapf, M.; Daher, N.; Dommén, J.; Jeannot, N.; Schneider, S.; Platt, S.; Slowik, J. G.; Baumlin, N.; Salathe, M.; Prévôt, A. S. H.; Kalberer, M.; Strähle, C.; Dümmbgen, L.; Sioutas, C.; Baltensperger, U.; Geiser, M. Toxicity of aged gasoline exhaust particles to normal and diseased airway epithelia. *Sci. Rep.* **2015**, *5* (1), No. 11801.
- (89) Yang, J.; Roth, P.; Ruehl, C. R.; Shafer, M. M.; Antkiewicz, D. S.; Durbin, T. D.; Cocker, D.; Asa-Awuku, A.; Karavalakis, G. Physical, chemical, and toxicological characteristics of particulate emissions from current technology gasoline direct injection vehicles. *Sci. Total Environ.* **2019**, *650*, 1182–1194.
- (90) Jaramillo, I. C.; Sturrock, A.; Ghiassi, H.; Woller, D. J.; Deering-Rice, C. E.; Lighty, J. S.; Paine, R.; Reilly, C.; Kelly, K. E. Effects of fuel components and combustion particle physicochemical properties on toxicological responses of lung cells. *J. Environ. Sci. Health, Part A* **2018**, *53* (4), 295–309.
- (91) Juárez-Facio, A. T.; Rogez-Florent, T.; Méausoone, C.; Castilla, C.; Mignot, M.; Devouge-Boyer, C.; Lavanant, H.; Afonso, C.; Morin, C.; Merlet-Machour, N.; Chevalier, L.; Ouf, F.-X.; Corbière, C.; Yon, J.; Vaugeois, J.-M.; Monteil, C. Ultrafine Particles Issued from Gasoline-Fuels and Biofuel Surrogates Combustion: A Comparative Study of the Physicochemical and In Vitro Toxicological Effects. *Toxics* **2023**, *11* (1), No. 21.
- (92) Wu, D.; Zhang, F.; Lou, W.; Li, D.; Chen, J. Chemical characterization and toxicity assessment of fine particulate matters emitted from the combustion of petrol and diesel fuels. *Sci. Total Environ.* **2017**, *605–606*, 172–179.
- (93) Jin, T.; Amini, H.; Kosheleva, A.; Yazdi, M. D.; Wei, Y.; Castro, E.; Di, Q.; Shi, L.; Schwartz, J. Associations between long-term exposures to airborne PM<sub>2.5</sub> components and mortality in Massachusetts: mixture analysis exploration. *Environ. Health* **2022**, *21* (1), No. 96.
- (94) Luo, X. S.; Huang, W.; Shen, G.; Pang, Y.; Tang, M.; Li, W.; Zhao, Z.; Li, H.; Wei, Y.; Xie, L.; Mehmood, T. Source differences in the components and cytotoxicity of PM<sub>2.5</sub> from automobile exhaust, coal combustion, and biomass burning contributing to urban aerosol toxicity. *Atmos. Chem. Phys.* **2024**, *24* (2), 1345–1360.
- (95) Kaivosoja, T.; Jalava, P. I.; Lamberg, H.; Virén, A.; Tapanainen, M.; Torvela, T.; Tapper, U.; Sippula, O.; Tissari, J.; Hillamo, R.; Hirvonen, M. R.; Jokiniemi, J. Comparison of emissions and toxicological properties of fine particles from wood and oil boilers in small (20–25 kW) and medium (5–10 MW) scale. *Atmos. Environ.* **2013**, *77*, 193–201.
- (96) Karlsson, H. L.; Ljungman, A. G.; Lindbom, J. and L. Möller, Comparison of genotoxic and inflammatory effects of particles generated by wood combustion, a road simulator and collected from street and subway. *Toxicol. Lett.* **2006**, *165* (3), 203–211.
- (97) Jerrett, M.; Nau, C. L.; Young, D. R.; Butler, R. K.; Batteate, C. M.; Padilla, A.; Tartof, S. Y.; Su, J.; Burnett, R. T.; Kleeman, M. J. Air pollution and the sequelae of COVID-19 patients: A multistate analysis. *Environ. Res.* **2023**, *236*, No. 116814.
- (98) Wang, M.; Sampson, P. D.; Hu, J.; Kleeman, M.; Keller, J. P.; Olives, C.; Szpiro, A. A.; Vedal, S.; Kaufman, J. D. Combining Land-Use Regression and Chemical Transport Modeling in a Spatiotemporal Geostatistical Model for Ozone and PM<sub>2.5</sub>. *Environ. Sci. Technol.* **2016**, *50* (10), 5111–5118.

(99) Ying, Q.; Lu, J.; Allen, P.; Livingstone, P.; Kaduwela, A.; Kleeman, M. Modeling air quality during the California Regional PM10/PM2.5 Air Quality Study (CRPAQS) using the UCD/CIT source-oriented air quality model – Part I. Base case model results. *Atmos. Environ.* **2008**, *42* (39), 8954–8966.

(100) Liu, X.; Zou, M.; Sun, C.; Wu, L.; Chen, W.-X. Prenatal Folic Acid Supplements and Offspring's Autism Spectrum Disorder: A Meta-analysis and Meta-regression. *J. Autism Dev. Disord.* **2022**, *52* (2), 522–539.

(101) Jung, Y.; Lee, A. M.; McKee, S. A.; Picciotto, M. R. Maternal smoking and autism spectrum disorder: meta-analysis with population smoking metrics as moderators. *Sci. Rep.* **2017**, *7* (1), No. 4315.

(102) Modabbernia, A.; Velthorst, E.; Reichenberg, A. Environmental risk factors for autism: an evidence-based review of systematic reviews and meta-analyses. *Mole. Autism* **2017**, *8*, No. 13.

(103) Nakahara, K.; Michikawa, T.; Morokuma, S.; Hamada, N.; Ogawa, M.; Kato, K.; Sanefuji, M.; Shibata, E.; Tsuji, M.; Shimono, M.; Kawamoto, T.; Ohga, S.; Kusuhara, K.; Kamijima, M.; Yamazaki, S.; Ohya, Y.; Kishi, R.; Yaegashi, N.; Hashimoto, K.; Mori, C.; Ito, S.; Yamagata, Z.; Inadera, H.; Nakayama, T.; Iso, H.; Shima, M.; Kurozawa, Y.; Sugauma, N.; Katoh, T.; Japan Environment and Children's Study Group. Association of physical activity and sleep habits during pregnancy with autistic spectrum disorder in 3-year-old infants. *Commun. Med.* **2022**, *2*, No. 35.