UC Berkeley

UC Berkeley Previously Published Works

Title

Dry Season Transpiration and Soil Water Dynamics in the Central Amazon

Permalink

https://escholarship.org/uc/item/6x89953z

Authors

Spanner, Gustavo C Gimenez, Bruno O Wright, Cynthia L <u>et al.</u>

Publication Date

DOI 10.3389/fpls.2022.825097

Peer reviewed

Drought-driven shifts in tree water use linked to root distribution, sap flow and basal area in the Central Amazon

- 1 Gustavo C. Spanner¹, Bruno O. Gimenez^{1,2}, Cynthia L. Wright³, Valdiek Silva Menezes¹, Brent D.
- 2 Newman⁴, Adam Collins⁴, Kolby Jardine^{1,5}, Robinson I. Negrón-Juárez⁵, Adriano José Nogueira
- 3 Lima¹, Jardel Ramos Rodrigues¹, Jeffrey Chambers^{1,5,6}, Niro Higuchi¹ and Jeffrey M. Warren^{*3}
- 4 ¹National Institute of Amazonian Research (INPA), Manaus, Brazil
- 5 ²Smithsonian Tropical Research Institute (STRI), Panama City, Panama
- ⁶ ³Environmental Sciences Division and Climate Change Science Institute, Oak Ridge National
- 7 Laboratory, Oak Ridge, TN, 97831-6301, USA
- 8 ⁴Los Alamos National Laboratory, Los Alamos, NM, 87545, USA
- 9 ⁵Lawrence Berkeley National Laboratory, Berkeley, CA 94720, USA

10

- ¹¹ ⁶University of California Berkeley, Department of Geography, Berkeley, CA, USA
- 12

13 * Correspondence:

- 14 Jeffrey M. Warren
- 15 <u>warrenjm@ornl.gov</u>
- 16
- 17
- 18 Notice: This manuscript has been authored by UT-Battelle, LLC under Contract No. DE-AC05-
- 19 00OR22725 with the US Department of Energy. The United States Government retains and the
- 20 publisher, by accepting the article for publication, acknowledges that the United States Government
- 21 retains a non-exclusive, paid-up, irrevocable, world-wide license to publish or reproduce the published
- 22 form of this manuscript, or allow others to do so, for United States Government purposes. The
- 23 Department of Energy will provide public access to these results of federally sponsored research in
- 24 accordance with the DOE Public Access Plan (<u>http://energy.gov/downloads/doe-public-access-plan</u>).
- 25 26
- 27 **Keywords:** allometry, ecohydrology, root water uptake, soil water, transpiration, tropical forests
- 28

29 Abstract

- 30 With model projections of more intense and frequent droughts in the future, there is increasing
- 31 importance in understanding tree and stand-level water flux through tropical forest ecosystems during
- 32 dry periods. Here we investigate paired soil and tree water extraction dynamics in an old-growth terra-
- 33 firme forest in central Amazonia during the driest period of 2018. Species-specific tree water use was
- 34 assessed via radial patterns of sap flow in eight dominant canopy trees with a range in diameter, height
- 35 and wood density. Paired multi-sensor soil moisture probes were installed adjacent to six of those trees

36 and quantified volumetric water content dynamics at 10, 20, 30, 40, 50, 70 and 100 cm. To link depth-

37 specific water extraction patterns to root distribution, fine root biomass was also assessed through the 38 soil profile to 2.35 m. To scale water extraction to the plot level, basal area was measured for all trees

within a 5 m radius around each soil moisture probe. The sensitivity of tree transpiration to drought

40 varied by species, with some increasing and some decreasing water use during the dry period. Tree-level

- 41 water use broadly scaled with sapwood area across species, from 11-190 liters per day. Plot level water
- 42 use increased linearly with plot basal area, indicating daily transpiration from ~1.7 to 3.3 mm per day,
- 43 depending on spatial location. Upper soil water extraction was strongly dependent on root biomass,
- 44 which was dense at the surface (i.e., 45% in the upper 5 cm) and declined dramatically with depth. Yet

45 as the upper soil dried, water extraction shifted to deeper levels and model projections suggest that much

46 of the water use during the month-long dry-down could be extracted from the upper 2 m. Results

indicate wide variation in soil water extraction temporally in both the vertical and horizontal directions.
The results provide key information on the species-specific regulation of transpiration by dominant trees

49 as water availability changes and the simultaneous dynamics of soil water extraction that can be used to

50 inform mechanistic models that project tropical forest response to drought.

51

52 Introduction

53 The response of tropical forest transpiration (T) to changes in environmental conditions remains highly 54 uncertain in Earth System Models due to an unresolved understanding of both abiotic and biotic factors 55 and their interactions (e.g., Berg and Sheffield 2019). Uncertainty is amplified by the high trait diversity of tropical trees (Cardoso et al., 2017) and their differential responses to drought conditions (Cox et al., 56 57 2004; Baker et al., 2008; Malhi et al., 2009). Within tropical forests, ecosystem T is mediated by tree 58 size, and species-specific traits including stomatal sensitivity to leaf water loss driven by vapor pressure 59 deficit (VPD) (Fontes et al. 2018, Barros et al. 2019, Gimenez et al. 2019, Grossiord et al. 2019). In the 60 Amazon rainforest, studies have reported both drought-induced increases (da Rocha et al. 2009; Brum et 61 al. 2018) and decreases (da Rocha et al., 2009; Fontes et al., 2018) in the rate of T. While we know that 62 forest evapotranspiration is influenced by dominant abiotic factors including net radiation, boundary 63 layer conductance and VPD (Costa et al., 2010, Brum et al. 2018, Grossiord et al. 2019), it is plant 64 available water which may explain why evapotranspiration is limited in some regions of the Amazon, 65 but not in others (Fisher et al. 2008, Grossiord et al. 2019).

66

67 In the central Amazon, where rainfall is high and the dry season is usually less than four months, 68 evapotranspiration generally follows climatic conditions (da Rocha et al., 2009) but can become limited 69 during drought conditions (Fisher et al., 2008). Moreover, given the low soil water holding capacity of 70 the Oxisols that are common in central Amazon, only a short sequence of days without precipitation can 71 lead to the depletion of near surface soil water and as a result, a greater depth of soil has to be exploited 72 to maintain root water uptake (Hodnett et al., 1995). In the central Amazon, most root distribution is in 73 the upper 20 cm (Ferreira et al., 2002; Noguchi et al., 2014), yet water absorption has been exhibited to at least 3.6 meters (Hodnett et al., 1995, 1996), and there is evidence of rooting to 6-10 m depths 74 75 (Chauvel et al., 1992, Negrón-Juárez et al., 2020). Further east in Pará state, Brazil, tree roots have been 76 found to depths of 18 m (Nepsted et al. 1994).

77

Differences in access to and use of water are often dependent on tree height and diameter (West et al. 1997, Meinzer et al. 2005). Diameter can be used to predict characteristics related to hydraulic capacity such as sapwood area (Aparecido et al. 2019), which can then be used to estimate whole tree water use if sap velocity is known. In the tropics, the largest, emergent trees are subject to greater evaporative

82 demand than other canopy trees (Kunert et al., 2017; Motzer et al., 2005) and are a major source of

- 83 stand-level T (Brum et al. 2018). Yet, for dominant and emergent trees to maintain T and photosynthesis
- 84 during drought requires that they leverage appropriate hydraulic strategies. Amongst these, root access
- 85 to deeper water sources is one trait that the largest trees display (da Rocha et al., 2004; Lee et al., 2005;
- 86 Oliveira et al., 2005; Bruno et al., 2006), although some smaller trees have also been shown to access 87 deeper water sources (e.g., Stahl et al. 2013).
- 88

89 Other traits also influence transpiration dynamics, including relative hydraulic conductivity through the 90 xylem pathway (Fisher et al. 2006) and stomatal sensitivity to internal water availability and external 91 driving forces (Meinzer 2002, Gimenez et al. 2019). Under drying conditions, tropical trees display a 92 variety of responses, including avoidance (e.g., leaf drop, or reducing water use by closing stomata – 93 more conservative strategy) or tolerance (e.g., maintaining water use at the risk of hydraulic failure – 94 more acquisitive strategy). Either strategy may ultimately contribute to mortality under excessive 95 drought due to either carbon starvation or hydraulic failure (Rowland et al., 2015; McDowell et al. 2018 96 Aleixo et al., 2019;). Understanding species-specific variation in T and its response to drought, 97 therefore, lends insight into relative hydraulic strategies and how they may manifest at the stand level. 98 Knowledge of functional traits and how those traits scale with size can thus provide a pathway for 99 process-level understanding of the controls regulating ecosystem dynamics, including T (e.g., Kotowska 100 et al. 2021). Such data can be highly useful for forest and ecosystem modeling efforts (e.g., Scheiter et 101 al. 2013, Christoffersen et al. 2016).

102 *Objectives*

103 The objective of this work was to better understand the seasonal patterns of tree water use as linked to 104 upper soil water availability during a dry period in terra-firme soils of the Central Amazon. Our specific 105 questions are: (1) Do different co-occurring canopy species increase or decrease water use during 106 drought, when atmospheric demand for water is increasing but upper soil water supply is declining? (2) As soils dry, does tree water extraction shift to deeper layers? (3) Can we predict T based on tree size, 107 108 stand basal area or soil water content in the upper 1 m? Results will provide key insights into the 109 species-specific regulation of T by dominant trees as water availability changes, and concurrent 110 dynamics of soil water extraction that can be used to inform models of tropical forest response to 111 drought.

112 Material and methods

113 Study area

114 The study was conducted at the ZF-2 Tropical Forestry Experimental Station (02°36'33"S; 60°12'33"W; 115 ~130 m elevation) within the Biological Reserve of Cuieras in Central Amazon. According to Koppen's 116 classification the climate is type Af (Tropical) with a short dry season that generally occurs between July 117 and September (Alverez et al. 2013; Wu et al. 2016). The temperature ranges from 26.8 °C and 28.8 °C 118 but can reach up to 31 °C (Miranda et al. 2002). The predominant soil type in the plateau areas is the 119 dystrophic yellow latosol (Chauvel et al., 1992), an Oxisol. This is a very clayey soil with clay contents 120 that increase from 59% at the surface to 73% at 1 m depth, with an average porosity of about 0.58 m³ m⁻ 121 ³. Wilting point to field capacity ranged from ~0.27 to 0.41 m³ m⁻³ for the upper 1 m depth (Chauvel et 122 al., 1992; Marques et al. 2004). The closed-canopy primary forest is located on a plateau (terra-firme) 123 with heterogeneous tree sizes, ages and canopy heights (Oliveira et al., 2008). Most of the crowns are

124 globular in shape, with little direct radiation entering in lower strata. Although the height of individuals

125 varies widely, more than 50% are between 14 and 25 m, with emergent trees reaching up to 50 m

126 (Oliveira et al., 2008). Estimates of tree age based on carbon isotopes for 20 species in the area ranged

127 from 200 to 1,400 years (Chambers et al., 1998).

128 Micrometeorological variables

Micrometeorological variables were obtained from the 50 m K-34 tower managed by the Large-Scale Biosphere-Atmosphere Program – LBA (Araújo et al., 2002). Precipitation was measured by a rain gauge (EM ARG100, Environmental Measurements Limited, UK) and air temperature (T_{air}) and relative humidity (R_H) were obtained by means of a thermo-hygrometer (HMP45AC, Vaisala, Finland) at 28 m height. Vapor pressure deficit (VPD) was derived from T_{air} and R_H using Tetens equation. All micrometeorological variables were collected at a 30-minute time frequency.

135 The Accumulated Daily Water Deficit

136 As in most humid tropical forests, precipitation events during the dry season are common. We used the 137 accumulated daily water deficit (ADWD; e.g., Santos et al. 2018) metric to delimit the dry period. 138 Similar to the cumulative water deficit and maximum cumulative water deficit metrics (Malhi et al., 139 2009), the ADWD measures the duration and intensity of a drought using daily precipitation (Santos et 140 al., 2018). ADWD is calculated using the cumulative precipitation from the previous 30 days and a dry 141 period occurs when this cumulative is less than the average monthly forest evapotranspiration (~100 142 mm) (Malhi et al., 2009; Santos et al., 2018). To reinforce the delimitation of the dry period, we also 143 compared the ADWD calculations to declines in soil water content and increases in VPD.

144 Species selection

For this study, eight dominant canopy tree species were selected for sap flow monitoring and six of these 145 146 were additionally paired with soil water sensors to quantify water extraction (Figure 1; Table 1). These 147 species were selected because they are generally abundant in this forest stand (Oliveira et al., 2008) and 148 individuals were located close to one another and adjacent to the K-34 flux tower. We selected one 149 individual tree per species. For each tree, we measured DBH at a height of 1.3 m from the ground and 150 height (HT) using a laser rangefinder (Trupulse, Laser Technology, Inc, USA) targeting the most 151 emergent branch. We also estimated canopy area (CA) as the average of the canopy diameters measured 152 from the forest floor in the north-south and east-west axes. To upscale T, we measured the DBH of all 153 trees within a 10 m diameter plot of the soil water sensors depicted in Figure 1; this included the 154 dominant sap flow tree, and smaller diameter (> 5cm DBH) trees and lianas within the plot.

155 Sap velocity, sapwood area, and sap flow

To calculate water use or sap flow (Q) of each of the eight trees, we measured radial patterns of sap 156 157 velocity (v_s) , sapwood depth (SPD) and sapwood area (SA) (Table 1). The SPD was measured in the 158 field using a dye injection technique based on Goldstein et al. (1998). Briefly, a small ~6.5 mm hole was 159 drilled at 1.3 m through the sapwood at a slight downward angle through the sapwood into the 160 heartwood, and subsequently filled completely with dye (aqueous acid fucshin filtered to 0.2 µm) under 161 slight pressure. The dye was retained in a 30 ml syringe reservoir held in place by its tip inserted through 162 the outer bark and phloem tissue. Additional dye was added as needed. After 1-2 h, an increment core 163 was collected ~2-3 cm above each dye injection point and the conductive sapwood tissue depth was

164 determined from the thickness of the xylem wood where dye was visible. From SPD, we calculated the 165 total SA assuming that SPD was homogeneous across the trunk cross-section.

166

167 To measure v_s , we used heat dissipation-type sensors with a 2-cm sensing tip (TDP-GS, PlantSensors,

168 Australia; Granier, 1987). For each tree, the probes were installed at 1.3 m height at depths of 0-2 cm, 2-

169 4 cm, 4-6 cm and 6-8 cm (depending on dye-estimated sapwood depth) within active sapwood to

170 account for radial variation (Meinzer, Goldstein and Andrade, 2001). Each sensor consists of a set of

171 two temperature probes inserted into brass tubes, which protect the probes and provide good conduction

172 with the surrounding xylem. The two probes were separated by a vertical distance of 10 to 15 cm. The

173 upper probe was continuously heated, with 0.2 W, while the lower probe was used as a reference

174 temperature. The difference in temperature (or voltage) between the two probes was used to calculate v_s 175 (Oishi, Hawthorne and Oren, 2016) as:

176

$$v_s = aK^b \tag{1}$$

Where v_s is the sap velocity (g cm⁻² s⁻¹), a is a universal empirical calibration coefficient (0.0119), b is the 177 178 exponent of the scale (1.23) and K is the temperature difference between the two sensors. While studies

179 have found the calibration coefficient does not represent all species (e.g., Bush et al., 2010), it has

180 proved appropriate in some tropical trees (e.g., McCulloh et al., 2007) and for this study, we assumed

181 the calibration coefficient was suitable for our trees. K is calculated by:

182

$$K = \frac{\Delta T_{max}}{\Delta T} - 1 \tag{2}$$

Where ΔT is the temperature difference between the heated and unheated probe, and ΔT_{max} is the 183 184 temperature difference between the two probes at zero flow, which is assumed to be reached on nights 185 with low VPD and dew formation on foliage. Then, we upscaled v_s to whole tree-level sap flow (Q; g s⁻¹)

186 by summing v_s for each concentric ring of active sapwood depth (*i*):

$$Q = \sum_{i=1}^{SPD} v_{si} \times SA_i \tag{3}$$

Where SA is the sapwood area (cm^2) at each sensor position into the sapwood (0-2 cm, 2-4 cm, 4-6 cm, 187

188 6-8 cm) until maximum depth (SPD) is reached (Table 1). Since some of the innermost sap flow sensors 189 partially extended into non-functional xylem, the integrated signal at that depth could be lower than

190 exists in the adjacent functional sapwood and may create some underestimation of sap flow at this depth,

191 although lower inner sap flow reduced the magnitude of this error.

192 Allometric relationships and upscaling transpiration

193 To describe the response of Q to drought, we focused our calculations on the most severe dry period as defined by the ADWD. The trajectories of water use during this period were analyzed by linear 194 195 regressions and O was normalized as a function of dry days. We only used days with high evaporative 196 demand, defined here as days with VPD > 1 kPa, exclude wet days or those with low evaporative 197 demand and likely with low water use.

199 Then, we upscaled O to plot-level T. To do so, we first modeled tree-level SA vs DBH. We pooled data 200 from our eight trees with additional allometry data collected by Aparecido et al. (2019) at the same site. 201 In Aparecido et al. (2019), the relationship between SA and DBH was limited to trees with DBH ≤ 40 202 cm. In our study, the DBH of individual trees within the plot ranged from 29.8 to 113.8 cm, and 5 trees 203 had a DBH > 40 cm. To avoid excessively weighting outliers, we applied Cook's method (Belsley et al., 204 1980; Cook and Weisberg, 1982) to detect potential outliers within the pooled data We detected two 205 outliers: Buchenavia grandis, which was the largest tree and Ocotea nigrescens, which had the largest 206 sapwood area (Table 1). Removing these two outliers resulted in the highest correlation coefficient of SA 207 vs. DBH, as compared to the total dataset or the dataset minus each outlier, and results were used to 208 improve the Aparecido et al. (2019) allometric equation (Figure S1). In this way, we had a method to 209 estimate SA based on DBH. Second, we used our eight species with sap flow sensors to draw a 210 relationship between O and SA and then calculated O based on DBH. Then we upscaled O to plot-level 211 T according to:

$$T = \sum_{i=1}^{n} Q_i \tag{4}$$

212

213 Where *T* is the summed transpiration for all trees (n) within the 10 m diameter plot and Q_i represents the 214 *Q* of each tree in the plot estimated based on the two allometric relationships: *Q* vs *SA* and *SA* vs DBH. 215 Plot level *T* could then be directly related to plot level basal area. The six plots were centered on each of 216 the soil moisture probes that were installed one meter away from six of the sap flow trees (Figure 1). 217 We compared this estimate of plot level *T* (Eq. 4) to measured sap flow of the dominant tree (Eq. 3) in 218 each plot. Because *T* is at the plot level, we could then compare *T* to the water extraction measured by 219 the soil water sensors.

220 Volumetric soil water content

221 For six of the eight sap flow trees, we measured volumetric soil water content (Figure 1; Table 1) using 222 multi-frequency domain capacitance probes (FDC; EnviroSMART, Sentek Pty. Ltd, Stepney SA, 223 Australia). The FDC probes were inserted into PVC access tubes located at 1 m distance from each tree. 224 Initial installations reached 235 cm depth, however poor upper soil contact with the access tubes due to 225 compaction required careful reinstallations to 1 m depth. Soils from the initial installations were used for 226 root assessment (described below). Each probe consisted of seven sensors spaced at different vertical 227 intervals centered at depths 10, 20, 30, 40, 50, 70 and 100 cm, recording into a datalogger (model 228 CR1000, Campbell Scientific Inc, Logan Utah USA) at 15- minutes intervals. Each sensor provides an 229 integrated vertical measurement spanning ~10 cm; e.g., the sensor at 10 cm represents the 5-15 cm layer. 230 The output of the FDC sensors is a dimensionless frequency that is converted using a normalization 231 equation and then applied to a soil type-specific calibration equation by

232 233

$$\theta = \mathbf{i}$$
 (5)

Where f is the scaled frequency sensor output; θ is the volumetric water content (m³ m⁻³); A, B and C are the empirical coefficients of the equation. We initially used a calibration equation for a heavy cracking clay soil in Australia ('Warren' clay type, Sentek Calibration Manual, Sentek Pty Ltd), with calibration coefficients: A = 0.0254, B = 1, C = -0.150 (see the equations 1 and 2 below). This calibration resulted in reasonably, but slightly less absolute θ than other estimates at the site, thus we performed a sitespecific calibration.

241 Since the primary focus of this study was to examine relative changes in soil water content, errors in the 242 absolute calibration of the FDC sensors had minor impact on estimates of daily extraction rates. Even so, we adjusted the default factory calibration based on soil texture and using a site-specific calibration. For 243 244 calibration, two access tubes were installed and allowed to equilibrate for four months, then a vertical 245 FDC sensor was inserted into the tube and the volumetric water content using the 'Warren' type heavy, 246 cracking clay (Sentek Calibration Manual; Sentek Pty Ltd 2001) was recorded. Immediately after 247 inserting the FDC sensor, 5 cm tall metal Kopecky rings were carefully pushed vertically into the soil 248 adjacent to the access tube. Each ring was carefully excavated, excess soil was cleaned off using a knife, 249 then they were immediately wrapped in plastic for subsequent wet and dry gravimetric measurements. 250 Based on these weights, volumetric water content was calculated as the mass of water divided by ring 251 volume. Two replicate samples with seven depths were collected along either side of the tube down to 252 ~ 60 cm. Some animal burrows and soil compaction affected several of the measurements, and two samples were identified as outliers based on their interquartile range. The calibration regression between 253 254 actual and FDC θ (R² = 0.71) was then used to calculate *f* of the capacitance sensors from equation 5. 255 Data were then used to back calculate a new calibration equation with coefficients A = 0.03154, B = 1, C 256 = -0.6349, which can be used to convert f to VWC. There was one layer, 45-55 cm, that showed distinct 257 physical differences from other layers above and below. This included lower bulk density (1.01 + -0.02)258 g cm⁻¹ as compared with 1.18 ± 0.04 g cm⁻¹ for other layers in the upper 1 m) and a siltier texture (22%) 259 silt as compared with 4 - 6% for other layers in the upper 1 m) and a lower clay content (64% as 260 compared with 80% for adjacent layers). Texture was quantified by the pipette method at the Soil and 261 Plant Analysis Laboratory (LASP) of Embrapa Amazônia Ocidental, Manaus, Brazil. The differences in 262 soil properties in the 45-55 cm layer result in greater macroporosity and water availability, which has 263 been previously identified by other studies in the area (e.g., Tomasella and Hodnett 1996; Broedel et al. 264 2017). For this layer, the heavy clay calibration resulted in low estimates of water content. As such, for the 45-55 cm layer we used the manufacturer's default calibration (A = 0.1957; B = 0.404; C = 265 266 0.02852).

267 Fine root biomass

268 In order to measure fine root biomass distribution, soil samples of a known volume were collected from 269 the same locations as the FDC soil water sensors (Figure 1; Table 1). Soil was excavated at 270 progressively larger depth intervals, initially at 5 cm intervals for the root-dense upper soil, to a 35 cm 271 interval at 200-235 cm. In the field, soils were stored in plastic bags to be transported to the Forest 272 Management Laboratory at the National Institute of Amazonian Research. In the lab, roots were 273 separated manually using steel mesh sieves of various diameters and tweezers. Only the roots with a 274 diameter < 2 mm were retained; black roots which had died and decayed were discarded (Vogt and 275 Persson, 1991). The fine roots were placed in paper bags and dried at 65 °C for more than 48 hours until 276 constant weight was obtained.

277

278 Root biomass was summed by layer and for the entire 235 cm profile. Given the high diversity of 279 species, we did not separate roots by species, nor did we measure fresh root length. To describe fine root 280 biomass distribution, we fitted a non-linear model to fine root biomass and cumulative fraction as a 281 function of soil depth ($Y=1 - \beta^d$; where Y= cumulative root fraction at d=depth; Gale and Grigal, 1987).

282 Soil water extraction

283 Change in soil water content during the dry period was calculated as the difference in the initial and final 284 volumetric soil water content during each 15-minute interval for each of the six probes. Soil water

285 extraction from the upper 1 m was then estimated as the daily sum of the change in water content for each layer. At depths without sensors (e.g., 60 cm), data were interpolated based on the adjacent sensors 286 (at 50 and 70 cm). In this study, evaporation at the soil surface was assumed to be zero since the dense 287 288 forest canopy greatly limits solar radiation to the forest floor, as demonstrated by other studies in the 289 Central Amazon (Kunert et al., 2017; Salati and Vose, 1984). Thus, the change in soil water content over 290 the defined dry period was attributed exclusively to root water uptake. To relate fine root biomass to soil 291 water extraction up to 1 m depth, we used a non-linear Michaelis-Menten type regression: $V = V_{max} (R_d / I_{max})$ $(K_m + R_d)$, where V = extraction rate, R_d = root density and K_m = rate constant. Since tree water 292 extraction can shift to deeper depths during drying periods, we also fit a logarithmic function of soil 293 294 water extraction by soil depth to project potential water extraction to 2 or 3 m depth. The logarithmic fit 295 allowed for an alternate estimate of plant water use for depths > 1m. The logarithmic fit also provided an 296 estimate of water extraction between 0-10 cm depth at the beginning of the dry period. At the end of the 297 dry period, it is likely that uptake was limited at this layer, and thus uptake was set to decline to zero at 298 the surface. To compare water extraction as the soil dried, we performed the Student's t test when the 299 data were normal, and the Wilcoxon test when data were not normal. Normality was checked with the 300 Shapiro-Wilk test. Lastly, we compared soil water extraction to plot-level T upscaled using tree 301 allometric relationships.

302 **Results**

303 The Accumulated Daily Water Deficit

304 The longest ADWD period in 2018 was from October 2nd to 20th, with a cumulative water deficit of 31 mm (Figure 2A). During this two-week dry-down, the highest VPD values were observed, as well as the 305 306 lowest soil water at 1 m, making ADWD a good metric to be used to characterize drought intensity. We also included the previous two weeks (September 16th to October 1st), as they represented initial pre-307 308 drought conditions with relatively low rainfall and soil water, and high VPD (Figure 2B-D). However, 309 we discarded the last four days (October 17th to 20th) because a series of significant rain events caused an increase in soil water and a drop in VPD (Figure 2B-D). Therefore, we defined the dry period as 310 September 16th to October 16th, 2018 (shading in Figure 2). 311

312 Species-specific water use trends during the dry period

- 313 For all species, sap velocity was greatest in the outer 2 cm of sapwood, ranging from about 6-30 cm h⁻¹,
- depending on species (Figure S2a,b). For trees with deeper sapwood, sap velocity declined with depth,
- 315 with minimum values by 6 cm depth, and no sap flow at 8 cm depth. Sap velocity at different depths
- 316 declined, remained flat, or increased over the course of the dry period, depending on species. During the
- 317 dry period, Q varied by species (Figure 3; Table 2). Some species increased water use in response to the
- 318 drought, some remained unchanged, while others reduced water use during the drought. Species which
- 319 showed an increasing trend in daily water use as the dry period progressed were *Buchenavia grandis*,
- 320 Eschweilera romeu-cardosoi, Maquira sclerophylla and Ocotea nigrescens. Species which showed a
- 321 declining trend in daily water use during the same dry period include *Eriotheca globosa, Goupia glabra*
- 322 and *Scleronema micranthum*, while *Swartzia recurve* remained constant. There was a wide range in
- 323 species-specific water use, from ~11 liters day⁻¹ for *S. recurva*, up to 190 liters day⁻¹ for *O. nigrescens*
- 324 (Table 2).
- 325 Soil water extraction by depth

326 We determined soil water extraction based on changes in the volumetric soil water content over time at

each depth. For the duration of the dry period, the average water extracted from the upper 1 m was 0.86

 $\pm 0.15 \text{ mm day}^{-1}$, ranging from 0.67 to 1.01 mm day $^{-1}$. We found temporal differences in the total water extracted from top 1 m of soil (t-test; df=5, p=0.02). At the beginning of the dry period soil water

extracted from top 1 in or son (t-test; di=5, p=0.02). At the beginning of the dry period son water and extraction was $0.94 \pm 0.09 \text{ mm day}^{-1}$, but then decreased to $0.74 \pm 0.07 \text{ mm day}^{-1}$ by the end of the dry

period (Figure 4). We also found differences in the vertical pattern of water extraction within the top 1 m

- of soil. Rates of water extraction declined with depth as both the root biomass declined and there was a
- 333 change in soil texture; extraction rates declined as sand content declined from ~18% in the upper layer to
- 334 12% at 1 m depth ($R^2 = 0.93$). At the start and end of the dry period, water extraction was consistently
- 335 greatest in the shallow soil layers. However, as the drought progressed, the rate of extraction declined

for the upper layers, and increased for the deeper soil. Using the logarithmic function to extrapolate soil

water extraction patterns beyond 1 m, we estimated that by the end of the dry period, ~2.4 mm day⁻¹ of soil water was extracted from the top 2 m (Figure 5; Table 3).

339 Fine root biomass distribution and soil water extraction

Fine root biomass distribution measured for the top 235 cm of soil averaged 1.19 kg m⁻³ and ranged 340 from 0.97 kg m⁻³ and 1.37 kg m⁻³ by soil core location. Fine root biomass distribution declined 341 342 exponentially with depth, with 45% of root biomass (range 29-60%) in the upper 5 cm. Despite a decent exponential curve fit of biomass to depth using the Gale and Grigal (1987) model ($\beta = 0.8$, R² = 0.92), 343 344 the cumulative root fraction between 20-50 cm was overestimated. Therefore, we tested a more complex 345 yield-density model (Y= d × (a + b × d^c)^{-1/c}); where Y=root fraction and d=depth (Bleasdale and Nelder, 346 1960), which provided a better fit to the data ($R^2 = 0.90$; a = 1.5024, b = 0.9859, c = 0.8520). Using the 347 Bleasdale model, we fitted fine root biomass data to the depth of 235 cm (Figure 6). By integration, we 348 found that 90% of the fine root biomass was distributed in the top 22 cm, 95% in the top 45 cm, and only 349 ~4% between 45 and 150 cm soil depth. Moreover, we found a significant relationship between fine root 350 biomass and soil water extraction; fine root biomass explained 59% of the variation of vertical soil water 351 extraction (Figure 7); Michaelis-Menten kinetics model: V_{max}=0.25, K_m=0.00042, R²=0.59).

352 Plot-level transpiration and soil water extraction

353 The relationship between measured SA and DBH in this study was used to improve the allometric 354 equation of Aparecido et al. (2019) (Figure S1). SA also had a strong positive linear relationship with 355 tree T rate during the dry period (Figure 8; $R^2 = 0.91$). By application of these relationships across trees within a plot, plot level basal area could be used to derive plot level T rates (Figure 9), where increasing 356 357 basal area explained 64% of the increase in plot T. Actual water use based on measured sap flow for the 358 single dominant tree in each plot (equation 3), scaled to projected canopy ground area (Table 1), was 6-359 28% of the scaled allometric estimates of plot level water use based on the allometric scaling 360 relationship between DBH and SF for all trees within the 5 m radius plot (Table 4). Greater daily water 361 use per unit ground area of the sap flow trees resulted in a reduction in whole plot level T ($R^2 = 0.59$; 362 Figure S3). Plot level tree T rates were significantly greater than measured water extraction from the 363 upper 1 m of the soil profile; however, projected water extraction from the upper 2 m profile was equivalent to plot level T (Figure 10). 364

365 Discussion

366

367 In this study, we linked tree water use to soil water availability and extraction patterns, root biomass

368 distribution, and tree allometry to better understand drought-driven shifts in depth of water uptake. Even

369 so, the hyperdiversity within tropical forests can confound simple allometric scaling of water extraction

370 patterns due to species-specific variation in hydraulic strategies, as well as spatial patterns of tree

distribution and soil water extraction. Adequate consideration of the range of spatio-temporal responses

of stand level transpiration to drought can provide a framework for projecting water use based on

373 measured ecosystem traits, including soil properties, demographics, size, and dynamics of hydraulic 374 sensitivity and resistance, including via deeper water extraction, phenology, capacitance or hydraulic

375 redistribution. Even with such complexity, our results provide key information on the species-specific

376 regulation of transpiration by dominant trees as water availability changes, and the simultaneous

377 dynamics of soil water extraction that can be used to inform mechanistic models that project tropical

378 forest responses to drought.

379 Divergent hydraulic sensitivity to drying

380 Our first objective was to assess water use for different co-occurring canopy species as atmospheric

demand increases but upper soil water declines. We found that some tree species increased transpiration

rates during the month-long drought, while others decreased transpiration rates. There was also

differential and dynamic radial patterns of sap flow during the dry period, likely due to differential depth

of root water extraction, and timing and magnitude of transpiration and capacitance use and refilling –

topics that are currently under active investigation. Results highlight divergent species-specific hydraulic strategies. Specifically, increases in transpiration indicate that these individuals have sufficient water

387 supply to meet increased atmospheric demand. An important consideration to water demand and supply

trade-offs is tree size. Generally, we expect that taller trees that rise above the canopy experience greater

water use due to greater exposure (Kunert et al., 2017), and that larger diameter trees can have greater
 internal water storage capacity to help fulfill that demand (Scholz et al., 2011). Indeed, we found that the

391 emergent *B. grandis*, which had the greatest height and diameter, also had the steepest rate of increase in

392 daily T during drought.

393 We found a positive relationship between tree size and water use across the entire range of sampled tree 394 diameters, similar to that reported in other tropical forest studies (e.g., Meinzer et al. 2001 (Panama), 395 Kunert et al. 2017 (Central Amazon)). However, at narrower diameter ranges, this relationship can break 396 down due to differences in sapwood depth and hydraulic strategy. For example, trees larger than 30 cm 397 diameter had a wide range of sapwood area which was not well linked to diameter (Fig S1) or tree water 398 use; e.g., despite have a 60% greater diameter, the total daily Q for B. grandis was just one third that of 399 the smaller O. nigrescens, a result of the much lower SA of B. grandis. Similarly, in another nearby 400 study at ZF2, there was no relationship between diameter and water use for trees > 40 cm (Kunert et al. 401 2017). In contrast, for canopy tree species in Panama, there was a tighter relationship between SA and 402 diameter, and deep sapwood was more prevalent for the larger trees (Meinzer et al. 2001). While broad 403 scaling patterns are evident, the higher resolution differences in relationships between diameter, SA and 404 O suggest that other phenological, structural and physiological characteristics such as wood density, 405 xylem vessel size, stem hydraulic conductivity and capacitance are also important for determining plant 406 water use (e.g., Bucci et al. 2008, Santiago et al., 2018). Across 27 co-occurring canopy trees in Panama, 407 sap flux density declined dramatically with tree size, although larger trees often have greater SA, which 408 could offset the reduced rates and maintain high total tree water use (Meinzer, Goldstein and Andrade

409 2001).

410 Tree water use and xylem transport rates also depend on sapwood water storage, or capacitance, which

411 increases with SA and can vary with wood physical properties; e.g., declining with wood density

412 (Meinzer et al. 2003). Leveraging capacitance during dry periods requires continued access to soil water.

413 Thus, deeply rooted trees may maintain high transpiration rates during drought (Lee et al., 2005; Baker

414 et al., 2008) when buffered by diurnal use and refilling of stored water in stem, or in the upper roots and

soils via hydraulic redistribution. Indeed, the importance of stem water storage (e.g., Goldstein et al.

416 1998, Sholtz et al. 2008, Yan et al. 2020) and hydraulic redistribution has been shown to be a significant

417 component of diurnal and seasonal water use during dry periods (e.g., Lee et al. 2005, Neumann and

418 Cardon 2012). Because both processes would provide a buffer for net daily declines in upper soil water

419 availability, apparent upper soil water stress can be delayed.

420 Alternately, decrease in T indicate that these individuals are sensitive to the drying of near surface soils.

- 421 After B. *grandis*, the next four largest trees were co-dominant, with similar heights. Still, drought-
- responses differed. There was an increase in T for two trees, but a decline for the other two. Moreover,
 two of the three smallest trees had flat or declining rate of T. The species-specific sensitivity to drought
- 424 is likely related to their hydraulic strategy, including stomatal sensitivity to VPD, phenology or leaf loss,
- 425 capacitance or rooting depth and ability to extract soil water held at high tensions. Because we found
- that even a week without precipitation was sufficient to reduce much of the plant available water near
- the soil surface, 0 cm to 40 cm deep, we might expect that some shallowly rooted species tolerate the

428 low leaf water potentials necessary to continue to extract water and maintain or increase T during short

- 429 dry periods, while others that cannot tolerate low water potentials would decrease T. One of the smallest 420 trace $E_{\rm rel}$ is the smallest field of the smallest field
- trees, *Eschweilera romeu-cardosoi*, displayed large increases in T during the dry period. This species
 could meet increased water demands if it either tolerates low leaf water potentials, is deeply rooted, or
- 431 could meet increased water demands if it either tolerates low leaf water potentials, is deeply rooted, or
 432 alternately, has more extensive lateral rooting that may be able to access pockets of higher water content
- 433 spatially across the landscape. Root distributions are remarkably diverse, and in some cases can extend
- 434 long distances across the landscape to access resources where available; e.g., a neotropical fig (*Ficus*
- 435 *schultesii*) in a Peruvian floodplain was found to have a 102 m long lateral surface root (Silman and
- 436 Krisel 2006).

437 The forest shifts to deeper soil water extraction as soil dries

438 Our second objective was to track shifts in the depth of water extraction as soils dry. Yet the ability to 439 extract soil water at deeper depths is constrained by rooting depth. We found that fine root biomass was 440 greatest near the surface. In fact, field observations indicated a root mat at the surface (<2 cm depth) 441 intermixed with a minimal leaf litter layer. After that, root density declined rapidly with depth. These 442 results are not surprising and agree with most studies in Central Amazon. For example, Chauvel et al. 443 (1992) showed that root length was concentrated in the superficial soil horizons but then decreased 444 rapidly with depth; i.e., 70 km m⁻³ at 0-30 cm depth, while deeper layers (1-6 m) were just 3.5-6.5 km m⁻ 445 ³. Ferreira et al. (2002) showed that 74% to 93% of fine root biomass was found in the top 20 to 40 cm 446 respectively; and Noguchi et al. (2014) showed that 89% of the fine root biomass in the plateau was 447 concentrated in the first 20 cm, in agreement with our results. The higher concentration of biomass of 448 fine roots near the surface is likely due to the higher levels of organic matter, litter decomposition and 449 nutrient cycling found in these layers, in addition to the physical structure of the soil (lower density, 450 higher porosity, higher concentration of sand). Observations at a nearby site with minirhizotrons 451 reported root length production and mortality were also much greater in upper soils, and more sensitive 452 to precipitation events (Cordeiro et al. 2020). Since these soils are generally nutrient poor, primarily low 453 in phosphorus (Lugli et al. 2020; B. Newman, unpublished data), root distribution at the surface can 454 convey competitive advantage for interception of nutrients released from surface litter decomposition.

455

In addition, high root density at the surface would allow for rapid uptake of precipitation, which couldbe important to alleviating drought stress during periods when the frequency or magnitude of

458 precipitation is limited. While we did not assess surface root water uptake, we projected an exponential 459 increase in uptake towards the surface. The importance of this finding is that, despite the high 460 concentration of fine roots near the surface, it is the deeper roots which will become critical for 461 sustaining plant function during extended dry periods.

462

463 We found that at the stand-level, the forest generally shifted to deeper soil water sources. For another 464 study at K34, Broedel et al. (2017) reported a significant increase in the depth of water extraction during 465 the evolution of a drought. This pattern has also been found in other forest types with seasonal dry 466 periods (Nepstad et al., 1994, Hodnett et al., 1995; da Rocha et al., 2004; Warren et al., 2005; Bruno et al., 2006: Nepstad et al., 2007: Tomasella et al., 2008). Hodnett et al. (1995) showed only 80 mm of 467 468 water were available in the soil to plants between 2 to 3.6 m. Even so, these deep layers provide 469 additional volume of water to the forest during dry periods as a way to compensate for the higher 470 evaporative demand, reflections of low levels of precipitation, high air temperatures and VPD, and low 471 air and soil humidity. Our data demonstrate that at the plot-level, there was an increase in total water 472 extraction and in water extraction with soil depth.

473

474 Soil water availability has been shown to be limited for the oxisols of the Central Amazon, with only 475 ~18% availability of total water between field capacity and the permanent wilting point in the upper 1 m 476 (Ferreira et al. 2002). Given the high clay content present at our site (Corrêa, 1984; Chauvel et al., 1992, 477 Margues et al., 2004), we would expect that in an extended dry period, root extraction exhausts the soil 478 water pool of the top 1 m and deeper soil water reserves must be tapped to sustain transpiration. If 479 deeper soil water is not available because soil is dry or there is a lack of roots, transpiration must 480 decline. In the more seasonal eastern Amazon, tree water use was reduced after a 21-day dry period 481 characterized by low soil moisture and high VPD, indicating a lack of water extraction from deeper 482 depths (Brum et al., 2018). At our site, upper soil water content declined substantially during the month-483 long dry period, likely reaching the point where water was no longer available for plant uptake. For 484 example, at 15 cm depth (Figure 2B), water content declined from ~43% to 38% by the end of the 485 drought. Earlier work in a nearby plateau that modeled van Genuchten soil water release parameters 486 indicated water content at the wilting point (1500 kPa) was ~37% for the 10-20 cm layer (Ferreira et al. 487 2002), suggesting a substantial loss of water available for extraction. In our study, the lack of water 488 availability in the upper layers toward the end of the dry period illustrates the necessity of progressively 489 deeper water extraction to fulfill the transpirational demand of the trees. Since we observed divergent 490 trends in T rates during dry conditions, we expect that species which were heavy water users—increased 491 rate of T and a high daily T—were important in driving the shifted to greater plant-water-uptake depth.

492 Discrepancy between stand level T and soil water extraction

493 Our last objective was to predict T based on tree size, stand basal area, and soil water extraction in the 494 top 1 m. Specifically, we attempted to use allometric equations to scale tree size into stand-scale T and 495 then compared to total soil water extraction. There are several reasons that could lead to a discrepancy 496 between these methods. One reason is related to the assumptions we make to estimate stand-level T. 497 Essentially, we are asking, can we use diameter to estimate T? This question has previously been posed 498 by others. For example, Meinzer et al. (2001) found that diameter alone accounted for more than 90% of 499 the variation in both maximum and total daily sap flux density. Still, this study was conducted in 500 seasonally dry tropical forest, and unlike our study, is based on sap flux density measurements of the 501 outer 2 cm of sapwood, without accounting for radial variation which could lead to underestimations of 502 T (Moore et al., 2020). Other research has shown that stand-level, time-averaged scaling overlooks

503 species-level differences and fluctuation in environmental conditions (Anderson-Teixeira et al., 2015). 504 This is supported by species-specific differences in T rate, and hence the need for outlier detection when 505 performing allometric relationships between DBH, SAP, and T. That is, if we were to integrate over 506 different species, a longer timeframe, or contrasting environmental conditions, we would expect our 507 allometric relationships to change.

508

509 Another reason for the discrepancy between scaling methods is that there are inherent difference 510 between T estimated via sap flux sensors vs. estimated water extraction patterns in the top 1 m of soil. 511 Practically speaking, the point at which these water fluxes are measured and how measurements are 512 scaled is different. Measurements of tree water use via sap flux density will be skewed according to 513 wood physical properties and stem water content (Vergeynst et al., 2014), sensors placement in inactive 514 xylem (Clearwater et al., 1999), the accuracy of sapwood area estimates (Kumagai et al., 2005), and the 515 spatial and radial variations within the xylem (Tseng et al., 2017). Scaling thermal dissipation probe 516 voltage output to sap flux density also relies on an accurate calibration equation, which is not easily or 517 often measured, thus leading to potential bias in scaling using the standard equation (e.g., Bush et al. 518 2010). Moreover, sap flux-based estimates at the tree-level integrate across the soil depths from which 519 water is accessed. Measurements of soil water extraction, on the other hand, are limited to the depths at 520 which soil water is measured. Although we aimed to overcome some depth limitations by extrapolating 521 soil water extraction patterns to 2+ m depth, tree water access at even greater depths is likely, given the 522 presences of some roots (e.g., Negrón-Juárez et al. 2020), albeit at very low densities, and the necessity for access to free water. Since stand-level T estimates were consistently greater than soil water 523 524 extraction estimates, and since we have seen a general shift towards deeper soil water during the dry 525 period, it is likely that soil water extraction extended beyond the 2 m depth during the study period, and 526 would certainly be deeper than 2 m under a more extended drought.

527 Conclusions

528 Our results provide field-based empirical evidence for a dependence on deep soil water sourcing due to 529 drought-driven shifts in tree water use in Central Amazon. Such findings are imperative given the 530 sensitivity of tropical forests to changes in climatic conditions and considering that in the past the humid 531 tropics have only experienced extreme water stress periodically, a pattern that is likely to change in the 532 future. Here we show that even a month of drought is sufficient to warrant a transpiration response. We 533 found that differential transpiration patterns for co-occurring canopy species, the major water users 534 contributing to stand-level signals, reflects divergent hydraulic strategies. And that even though most of 535 the tree root biomass is found near the soil surface, it is an increased dependence on deep soil beyond 536 the top 1 m or more that sustained transpiration for some species, and more generally, at the forest stand-537 level. Moreover, discrepancies between sap-flux based allometric estimates of T and total soil water 538 extraction highlights challenges when scaling tree-based water use to stand-level estimates of ecosystem 539 water use at higher spatial scales, and thus provides impetus for a holistic assessment of contributing 540 soil, tree and environmental components in a modeling framework. Understanding how different species 541 respond to and cope with drought and quantifying this difference in terms of transpiration fluxes is 542 critical as a first step to upscaling forest dynamics and understanding validity of model representation 543 and community-scale generalizations. We expect that these findings will inform and help better 544 constrain mechanistic Earth Systems Models that are used to project functional responses of tropical 545 forests to drought.

547 Author Contributions

548 JW planned and designed the experiments. GS, BG, VM and JW performed the experiments. GS, CW,

549 BG, VM and JW analyzed the data. GS, CW, BG, and JW wrote the manuscript. All authors contributed 550 to fieldwork, analysis or improved the manuscript.

551 Funding

552 This material is based upon work supported as part of the Next Generation Ecosystem Experiments-553 Tropics (NGEE-Tropics) funded by the U.S. Department of Energy, Office of Science, Office of 554 Biological and Environmental Research through contract No. DE-AC02-05CH11231 to LBNL, as part 555 of DOE's Terrestrial Ecosystem Science Program. Additional funding for this research was provided by 556 the Conselho Nacional de Desenvolvimento Científico e Tecnológico (CNPq).

557

558 Acknowledgements

559 The authors are thankful for the logistical and scientific support provided by the Laboratório de Manejo

560 Florestal (LMF) and the Large-Scale Biosphere-Atmosphere Program (LBA) at the National Institute of

561 Amazonian Research (INPA). We appreciate discussions with LBA hydrologist Dr. Sávio Ferreira and

562 his team, and the meteorological data provided by the scientific manager of LBA, Dr. Alessandro C.

563 Araújo.

564 **References**

- Aleixo, I., Norris, D., Hemerik, L., Barbosa, A., Prata, E., Costa, F., Poorter, L. (2019). Amazonian
 rainforest tree mortality driven by climate and functional traits. *Nat. Clim. Chang.* 9, 384.
- Aparecido, LMT, dos Santos, J, Higuchi, N, Kunert, N. (2019). Relevance of wood anatomy and size of
 Amazonian trees in the determination and allometry of sapwood area. *Acta Amazon* 49(1):1-10.
- Araújo, A.C., Nobre, A.D., Kruijt, B., Elbers, J.A., Dallarosa, R., ..., Kabat, P. (2002). Comparative
 measurements of carbon dioxide fluxes from two nearby towers in a central Amazonian rainforest:
 The Manaus LBA site. *J Geophys. Res.: Atmospheres* 107, LBA 58-51-LBA 58-20.
- Baker, I. T., Prihodko, L., Denning, A. S., Goulden, M., Miller, S., Rocha, H. R. (2008). Seasonal
 drought stress in the Amazon: Reconciling models and observations. *J Geophys. Res.* 113, 1–10.
 https://doi.org/10.1029/2007JG000644
- Barros, F.d.V., Bittencourt, P.R.L., Brum, M., Restrepo-Coupe, N., Pereira, L., ..., Oliveira, R.S.
 (2019). Hydraulic traits explain differential responses of Amazonian forests to the 2015 El Niñoinduced drought. *New Phytol.* 223: 1253-1266. https://doi.org/10.1111/nph.15909
- Berg, A., Sheffield, J. (2019). Evapotranspiration partitioning in CMIP5 models: uncertainties and
 future changes. J. Clim. 32(10), 2653-2671. https://doi.org/10.1175/JCLI-D-18-0583.1
- 580 Bleasdale, J., Nelder, J. (1960). Plant population and crop yield. *Nature* 188: 342.
- Broedel, E., Tomasella, J., Cândido, L. A., & von Randow, C. (2017). Deep soil water dynamics in an
 undisturbed primary forest in central Amazonia: Differences between normal years and the 2005
 drought. *Hydrol. Proc.* 31(9), 1749–1759. https://doi.org/10.1002/hyp.11143
- Brum, M., Gutiérrez López, J., Asbjornsen, H., Licata, J., Pypker, T., Sanchez, G., Oiveira, R.S., 2018.
 ENSO effects on the transpiration of eastern Amazon trees. *Philos. Trans. R. Soc. B Biol. Sci.* 373, 20180085.
- 587 Bruno, R. D., Rocha, H. R., Freitas, H. C. De, Goulden, M. L., Miller, S. D. (2006). Soil water

- 588 dynamics in an eastern Amazonian tropical forest. *Hydrol. Proc.* 2489, 2477–2489.
- Bucci, S. J., Scholz, F. G., Goldstein, G., Meinzer, F. C., Franco, A. C., Zhang, Y., Hao, G. Y. (2008).
 Water relations and hydraulic architecture in Cerrado trees: Adjustments to seasonal changes in
 water availability and evaporative demand. *Braz. J. Plant Physiol.* 20(3), 233–245.
- 592 Bush, S.E., Hultine, K.R., Sperry, J.S., and Ehleringer, J.R. (2010). Calibration of thermal dissipation 593 sap flow probes for ring- and diffuse-porous trees. *Tree Physiol.* 30, 1545-1554.
- Cardoso, D., Särkinen, T., Alexander, S., Amorim, A.M., Bittrich, V., ..., Forzza, R.C., 2017. Amazon
 plant diversity revealed by a taxonomically verified species list. *Proc. Natl. Acad. Sci.* 114, 10695 10700. https://doi.org/10.1073/pnas.1706756114
- 597 Chambers, J.Q., Higuchi, N., Schimel, J.P. (1998). Ancient trees in Amazonia. *Nature* 391, 135–136.
- 598 Chauvel, A., Vital, A. R. T., Lucas, Y., Desjardins, T., Franken, W. K., Luizão, F. J., ... Bedmar, A. P.
 599 (1992). O papel das raízes no ciclo hidrológico da floresta amazônica. *In: Anais do VII Congresso*600 *Brasileiro de Meteorologia*, São Paulo, SP, pp. 298-302.
- 601 Christoffersen, B.O., Gloor, M., Fauset, S., Fyllas, N.M., Galbraith, D.R., ..., Meir, P. (2016). Linking
 602 hydraulic traits to tropical forest function in a size-structured and trait-driven model (TFS v.1603 Hydro). *Geosci. Model Dev.* 9, 4227-4255.
- Clearwater, M.J., Meinzer, F.C., Andrade, J.L., Goldstein, G., Holbrook, N.M. (1999). Potential errors
 in measurement of nonuniform sap flow using heat dissipation probes. *Tree Physiol.* 19(10), 681–
 <u>687. https://doi.org/10.1093/treephys/19.10.681</u>
- Cordeiro, AL, Norby, RJ, Andersen, KM, Valverde-Barrantes, O, ..., Quesada, CA. (2020). Fine-root
 dynamics vary with soil depth and precipitation in a low-nutrient tropical forest in the Central
 Amazonia. *Plant-Envir. Inter.* 1: 3-16. <u>https://doi.org/10.1002/pei3.10010</u>
- 610 Corrêa, J. C. (1984). Características físico-hídricas dos solos latossolo amarelo, podzõlico vermelho611 amarelo e podzol hidromórfico do estado do amazonas'. *Pesq. Agropec. Bras.*, 19(3), 347–360.
- Costa, M.H., Biajoli, M.C., Sanches, L., Malhado, A.C.M., Hutyra, L.R., ..., de Araújo, A.C. (2010),
 Atmospheric versus vegetation controls of Amazonian tropical rain forest evapotranspiration: Are
 the wet and seasonally dry rain forests any different? *J. Geophys. Res.* 115, G04021,
 doi:10.1029/2009JG001179.
- Cox PM, Betts RA, Collins M, Harris PP, Huntingford C, Jones CD (2004) Amazonian forest dieback
 under climate-carbon cycle projections for the 21st century. *Theor Appl Climatol* 78:137–156
- da Rocha, H., Goulden, M., Miller, S., Menton, M., Pinto, L., de Freitas, H., & e Silva Figueira, A.
 (2004). Seasonality of water and heat fluxes over a tropical forest in eastern Amazonia. *Ecol. Appl.*14:S22-S32.
- da Rocha, H.R., Manzi, A. O., Cabral, O. M., Miller, S. D., Goulden, M. L., Saleska, S. R., ... Maia, J.
 F. (2009). Patterns of water and heat flux across a biome gradient from tropical forest to savanna in brazil. *J Geophys. Res.: Biogeosci.*, 114(1), 1–8. https://doi.org/10.1029/2007JG000640
- Ferreira, S. J. F., Luizão, F. J., Melloivo, W., Ross, S. M., Biot, Y. (2002). Propriedades físicas do solo
 após extração seletiva de madeira na Amazônia Central. *Acta Amazôn*, 32, 449–466. <u>https://doi.org/</u>
 10.1590/1809-43922002323466
- Fisher RA, Williams M, Do Vale RL, Da Costa AL, Meir P. Evidence from Amazonian forests is
 consistent with isohydric control of leaf water potential. (2006) *Plant Cell Environ* 29(2), 151-65.
 doi: 10.1111/j.1365-3040.2005.01407.x. PMID: 17080631.
- Fisher, R. A., Williams, M., Lourdes, M. De, Lola, A., Costa, D., Meir, P. (2008). Evaluating climatic
 and soil water controls on evapotranspiration at two Amazonian rainforest sites. *Agric. For. Meteorol.* 148, 850–861. https://doi.org/10.1016/j.agrformet.2007.12.001
- 633 Fontes, C.G., Dawson, T.E., Jardine, K., McDowell, N., Gimenez, B.O., ..., Araújo, A.C. (2018). Dry
- and hot: the hydraulic consequences of a climate change–type drought for Amazonian trees. Philos.

635 Trans. R. Soc. B Biol. Sci. 373, 20180209.

- Gale MR, Grigal DK. (1987) Vertical root distributions of northern tree species in relation to
 successional status. *Can J For Res* 17:829-834
- Gimenez, B.O., Jardine, K.J., Higuchi, N., Negrón-Juárez, R.I., Sampaio-Filho, I.D.J., ..., Chambers,
 J.Q. (2019). Species-specific shifts in diurnal sap velocity dynamics and hysteretic behavior of
 ecophysiological variables during the 2015–2016 El Niño event in the Amazon forest. *Front. Plant Sci.* 10, 830.
- Goldstein, G., Andrade, J.L., Meinzer, F.C., Holbrook, N.M., Jackson, P., Cavelier, J., and Celis, A.
 (1998). Stem water storage and diurnal patterns of water use in tropical forest trees. *Plant Cell Environ*. 21:397–406.
- Grossiord, C., Christoffersen, B., Alonso-Rodríguez, A.M., Anderson-Teixeira, K., Asbjornsen, H., ...,
 McDowell, N.G. (2019). Precipitation mediates sap flux sensitivity to evaporative demand in the
 neotropics. *Oecologia*. https://doi.org/10.1007/s00442-019-04513-x
- Hodnett, M.G., da Silva, L.P., da Rocha, H.R., Senna, R.C. (1995). Seasonal soil water storage changes
 beneath central Amazonian rainforest and pasture. *J. Hydrol.* 170, 233–254.
- Hodnett, M.G., Oyama, M.D., Tomasella, J., Marques Filho, A. de O. (1996). Comparisons of longterm soil water storage behaviour under pasture and forest in three areas of Amazonia. *In:* J.H.C.
 Gash et al. (Editors), *Amazonian Deforestation and Climate*. J. Wiley and Sons, New York, pp. 57-
- 653

- Kotowska, M.M., Link, R.M., Röll, A., Hertel, D., Hölscher, D., ..., Schuldt, B. (2021). Effects of wood
 hydraulic properties on water use and productivity of tropical rainforest trees. *Front. For. Glob. Change* 3:598759. doi: 10.3389/ffgc.2020.598759
- Kumagai, T., Nagasawa, H., Mabuchi, T., Ohsaki, S., Kubota, K., Kogi, K., Utsumi, Y., Koga, S.,
 Otsuki, K. (2005). Sources of error in estimating stand transpiration using allometric relationships
 between stem diameter and sapwood area for *Cryptomeria japonica* and *Chamaecyparis obtusa*. *For. Ecol. Manag.* 206(1–3), 191–195. https://doi.org/10.1016/j.foreco.2004.10.066
- Kunert, N., Aparecido, L.M.T., Wolff, S., Higuchi, N., dos Santos, J., de Araujo, A.C., Trumbore, S.,
 2017. A revised hydrological model for the Central Amazon: the importance of emergent canopy
 trees in the forest water budget. *Agric. For. Meteorol.* 239, 47–57.
- Lee, J.-E., R. S. Oliveira, T. Dawson, and I. Fung (2005), Root functioning modifies seasonal climate,
 Proc. Natl. Acad. Sci. 102(49), 17576–17581.
- Lugli, L.F., Andersen, K.M., Aragão, L.E.O.C., Cordeiro, A.L., Cunha, H.F.V, ..., Hartley, I.P. (2020).
 Multiple phosphorus acquisition strategies adopted by fine roots in low-fertility soils in Central
 Amazonia. *Plant Soil*. 50, 49–63. <u>https://doi.org/10.1007/s11104-019-03963-9</u>
- Malhi, Y., Aragão, L. E. O. C., Galbraith, D., Huntingford, C., Fisher, R., Zelazowski, P., ... Meir, P.
 (2009). Exploring the likelihood and mechanism of a climate-change-induced dieback of the
 Amazon rainforest. *Proc. Natl. Acad. Sci.* 16(49) 20610-20615; DOI: 10.1073/pnas.0804619106.
- Marques, J.D. de O., Libardi, P.L., Teixeira, W.G., Reis, A.M. (2004). Estudo de parâmetros físicos,
 químicos e hídricos de um Latossolo Amarelo, na região Amazônica. *Acta Amazon* 34(2), 145–154.
 https://doi.org/10.1590/s0044-59672004000200002
- McCulloh, K.A., Winter, K., Meinzer, F.C., Garcia, M., Aranda, J., and Lachenbruch, B. (2007). A
 comparison of daily water use estimates derived from constant-heat sap-flow probe values and
 gravimetric measurements in pot-grown saplings. *Tree Physiol.* 27, 1355-1360.
- McDowell, N., Allen, C.D., Anderson-Teixeira, K., Brando, P., Brienen, R., ..., Xu, X., 2018. Drivers
 and mechanisms of tree mortality in moist tropical forests. *New Phytol.* 219, 851–869.
 https://doi.org/10.1111/nph.15027
- 681 Meinzer F.C. Co-ordination of vapour and liquid phase water transport properties in plants. (2002)

682 Plant Cell Environ. 25(2), 265-274. doi: 10.1046/j.1365-3040.2002.00781.x.

- Meinzer, F.C., Bond, B.J., Warren, J.M., Woodruff, D.R. (2005). Does water transport scale universally
 with tree size? *Funct. Ecol.* 19(4), 558–565. https://doi.org/10.1111/j.1365-2435.2005.01017.x
- 685 Meinzer, F.C., Goldstein, G., and Andrade, J.L. (2001). Regulation of water flux through tropical forest 686 canopy trees: Do universal rules apply? *Tree Physiol.* 21, 19-26.
- Meinzer, F.C., James, S.A., Goldstein, G., Woodruff, D. (2003) Whole-tree water transport scales with
 sapwood capacitance in tropical forest canopy trees. *Plant Cell Environ*. 26 (1147–1155).
- Moore, G.W., Adkison, C., Aparecido, L.M.T., Basant, S., Cooper, ..., Wright, C. (2020). Thermal
 dissipation sensors enter a new age: navigating frontiers in transpiration and hydrologic function.
 Acta Hortic. 1300, 37-46. https://doi.org/10.17660/ActaHortic.2020.1300.6
- Motzer, T., Munz, N., Küppers, M., Schmitt, D., Anhuf, D. (2005). Stomatal conductance, transpiration
 and sap flow of tropical montane rain forest trees in the southern Ecuadorian Andes. *Tree Physiol.*25, 1283–1293.
- Negrón-Juárez R., Ferreira, S.J.F., Crestani Mota, M., Faybishenko, B., Monteiro, M.T.F., ...,
 Chambers, J.Q. (2020). Measurement and characterization of soil moisture dynamics in a Central
 Amazonian tropical forest. *Vadoze Zone J* 19:e20070. https://doi.org/10.1002/vzj2.20070
- Nepstad, D.C., de Carvalho, C.R., Davidson, E.A., Jipp, P.H., Lefebvre, P.A., ..., Vieira, S. (1994). The
 role of deep roots in the hydrological and carbon cycles of Amazonian forests and pastures. *Nature*372, 666–669.
- Nepstad, D.C., Tohver, I.M., Ray, D., Moutinho, P. and Cardinot, G. (2007). Mortality of large trees
 and lianas following experimental drought in an Amazon forest. *Ecology* 88(9), 2259–2269.
- Neumann, R.B., Cardon, Z.G. (2012). The magnitude of hydraulic redistribution by plant roots: a
 review and synthesis of empirical and modeling studies. *New Phytol.* 194: 337-352.
 https://doi.org/10.1111/j.1469-8137.2012.04088.x
- Noguchi, H., Suwa, R., Adélia, C., Souza, S. De, Kajimoto, T., & Ishizuka, M. (2014). Examination of
 vertical distribution of fine root biomass in a tropical moist forest of the Central Amazon, Brazil. *JARQ*, 48, 231–235.
- Oishi, A.C., Hawthorne, D.A., Oren, R. (2016). Baseliner: An open-source, interactive tool for
 processing sap flux data from thermal dissipation probes. *SoftwareX* 5, 139-143.
- Oliveira, A.N. de, Amaral, I.L. do, Ramos, M.B.P., Nobre, A.D., Couto, L.B., Sahdo, R.M., 2008.
 Composição e diversidade florístico-estrutural de um hectare de floresta densa de terra firme na Amazônia Central, Amazonas, Brasil. *Acta Amazôn.* 38, 627–641.
- Oliveira, R.S., Dawson, T.E., Burgess, S.S.O., Nepsted, D.C. (2005). Hydraulic redistribution in three
 Amazonian trees. Oecologia 145, 354–363. <u>https://doi.org/10.1007/s00442-005-0108-2</u>
- Rowland, L., da Costa, A.C.L., Galbraith, D.R., Oliveira, R.S., Binks, O.J., ..., Vasconcelos, S.S.
 (2015). Death from drought in tropical forests is triggered by hydraulics not carbon starvation. *Nature* 528, 119–122.
- 719 Salati, E., Vose, P.B. (1984). Amazon basin: a system in equilibrium. *Science* 225, 129–138.
- Santiago, L.S., De Guzman, M.E., Baraloto, C., Vogenberg, J.E., Brodie, ..., Bonal, D. (2018).
 Coordination and trade-offs among hydraulic safety, efficiency and drought avoidance traits in
 Amazonian rainforest canopy tree species. *New Phytol.* 218(3), 1015–1024. <u>https://doi.org/10.1111/</u>
 nph.15058
- Santos, V.A.H.F., Ferreira, M.J., Rodrigues, J.V.F.C., Garcia, M.N., Ceron, J.V.B., Nelson, B.W.,
 Saleska, SR. (2018). Causes of reduced leaf-level photosynthesis during strong El Niño drought in a
 Central Amazon forest. *Glob Change Biol.* 24: 4266–4279. https://doi.org/10.1111/gcb.14293
- 727 Scheiter, S., Langan, L., Higgins, S.I. (2013). Next-generation dynamic global vegetation models:
- learning from community ecology. *New Phytol* 198: 957-969. https://doi.org/10.1111/nph.12210

- Scholz, F.G., Phillips, N.G., Bucci, S.J., Meinzer, F.C., Goldstein, G. (2011). Hydraulic Capacitance:
 Biophysics and Functional Significance of Internal Water Sources in Relation to Tree Size. In F.C.
- Meinzer, B. Lachenbruch, & T.E. Dawson (Eds.), *Size- and Age-Related Changes in Tree Structure and Function* (Vol. 4, pp. 341–361). Springer Netherlands. <u>https://doi.org/10.1007/978-94-007-</u>
 1242-3 13
- Silman, M.R., Krisel, C. (2006). Getting to the Root of Tree Neighbourhoods: Hectare-Scale Root
 Zones of a Neotropical Fig. *J Trop. Ecol.* 22(6), 727–730. <u>http://www.jstor.org/stable/4092134</u>
- Stahl, C., Hérault, B., Rossi, V., Burban, B., Bréchet, C., Bonal, D. (2013). Depth of soil water uptake
 by tropical rainforest trees during dry periods: does tree dimension matter?. Oecologia 173, 1191–
 1201 (2013). https://doi.org/10.1007/s00442-013-2724-6
- Tomasella, J., Hodnett, M.G. (1996). Soil hydraulic properties and van Genuchten parameters for a
 oxisol under pasture in central Amazonia. *In:* J.H.C. Gash et al. (Editors), *Amazonian Deforestation and Climate*. J. Wiley and Sons, New York, pp. 101–124.
- Tseng, H., Chiu, C.-W., Laplace, S., Kume, T. (2017). Can we assume insignificant temporal changes
 in spatial variations of sap flux for year-round individual tree transpiration estimates? A case study
 on *Cryptomeria japonica* in central Taiwan. *Trees* 31(4), 1239–1251.
 https://doi.org/10.1007/s00468-017-1542-6
- Vergeynst, L.L., Vandegehuchte, M.W., McGuire, M.A., Teskey, R.O., Steppe, K. (2014). Changes in
 stem water content influence sap flux density measurements with thermal dissipation probes. *Trees*28(3), 949–955. <u>https://doi.org/10.1007/s00468-014-0989-y</u>
- Vogt K.A., Persson, H. 1991. Measuring growth and development of roots. *In:* Lassoie JP, Hinckley
 TM (eds) *Techniques and Approaches in Forest Tree Ecophysiology*. CRC Press, Inc., Boca Raton,
 FL. 478–493.
- Warren JM, FC Meinzer, JR Brooks and J-C Domec. 2005. Vertical stratification of soil water storage
 and release dynamics in Pacific Northwest coniferous forests. *Agric. For Meteorol.* 130:39-58.
- Yan, B., Mao, J., Dickinson, R.E., Thornton, P.E., Shi, X., Ricciuto, D.M., Warren, J.M., Hoffman,
 F.M. (2020). Modeling tree stem-water dynamics over an Amazonian rainforest. *Ecohydrol*.
 13:e2180. <u>https://doi.org/10.1002/eco.2180</u>
- Yi, K., Dragoni, D., Phillips, R.P., Roman, D.T., & Novick, K.A. (2017). Dynamics of stem water
 uptake among isohydric and anisohydric species experiencing a severe drought. *Tree Physiol.*37(10), 1379-1392.

760 Data Availability Statement

761 The datasets generated for this study can be found in the NGEE-Tropics Data Archive and can be made

- 762 available upon request.
- 763