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Omnivory does not preclude strong trophic cascades

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Abstract. Omnivory has been cited as an explanation for why trophic cascades are weak in many ecosystems, but empirical support for this prediction is equivocal. Compared to predators that feed only on herbivores, top omnivores—species that feed on both herbivores and primary producers—have been observed generating cascades ranging from strong to moderate, null, and negative. To gain intuition about the sensitivity of cascades to omnivory, we analyzed models describing systems with top omnivores that display either fixed or flexible diets, two foraging strategies that are supported by empirical observations. We identified regions of parameter space, wherein omnivores following a fixed foraging strategy, with herbivores and producers comprising a constant proportion of the diet, non-intuitively generate stronger cascades than predators that are otherwise demographically identical: (1) high productivity relative to herbivore mortality and (2) small discrepancies in producer vs. herbivore reward create conditions in which cascades are stronger with moderate omnivory. In contrast, flexible omnivores that attempt to optimize per capita growth rates during search never induce cascades that are stronger than the case of predators. Although we focus on simple models, the consistency of these general patterns together with prior empirical evidence suggests that omnivores should not be uniformly ruled out as agents of strong trophic cascades.

Key words: adaptive foraging; diet; omnivore; predator; productivity; trophic cascade.

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INTRODUCTION

Trophic cascades occur when top predators indirectly effect change in primary producer biomass by directly reducing populations of intermediate herbivores (Paine 1980, Strong 1992, Terborgh and Estes 2013). A growing number of factors that control the strength of trophic cascades continue to surface from model-based and experimental studies, and their identification has improved our understanding of processes that dampen or enhance indirect effects between species in ecological networks and ecosystem responses to disturbance (Pace et al. 1999, Shurin

et al. 2002, 2010, Borer et al. 2005, Estes et al. 2011, Heath et al. 2014, Fahimipour et al. 2017, Piovia-Scott et al. 2017). Theories for cascades have traditionally focused on top-down effects in tritrophic food chain models comprising predators that do not directly interact with primary producers (Oksanen et al. 1981, Schmitz et al. 2000, Heath et al. 2014). In many communities however, omnivores that additionally feed on producers occupy top trophic levels (Arim and Marquet 2004, Thompson et al. 2007). This potential for direct consumption of both producer and herbivore species has led to the prediction that omnivory should override

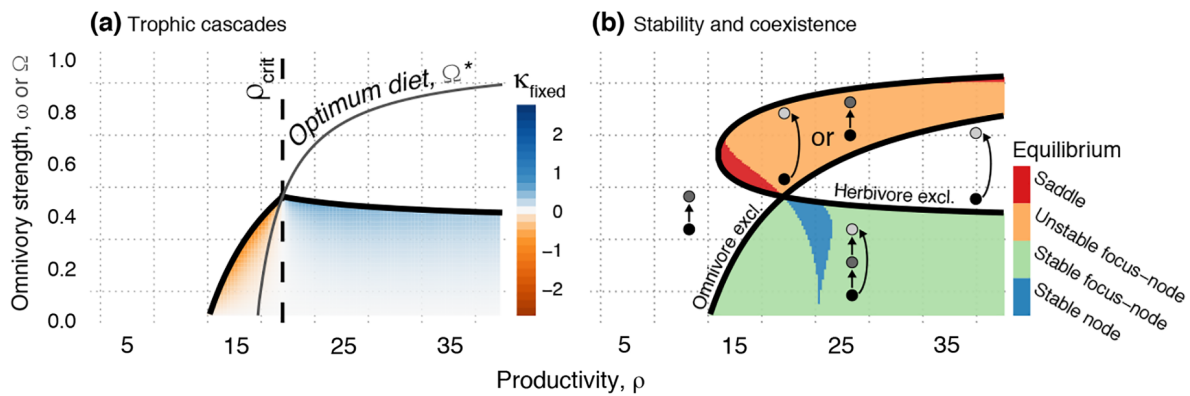


Fig. 1. (a) Numerical summary of relationships between relative cascade strength κ_{fixed} , productivity ρ , and omnivory strength ω . Colors show κ_{fixed} values for combinations of ρ and ω within the three-species coexistence region. Blue represents stronger cascades with omnivory, and orange represents weaker cascades with omnivory. Black curves mark extinction boundaries for either the omnivore or herbivore species (see panel b). The gray curve shows the equilibrium foraging strategy for the flexible omnivore, $\Omega_{\text{flexible}}^*$. A vertical dashed line marks the critical productivity ρ_{crit} . (b) Bifurcation curves identify stability boundaries separating steady states with different dynamics. Colors are different types of equilibria, determined by eigenvalues of the Jacobian matrix $\mathbf{J}_{\text{fixed}}$. Non-coexistence regions are labeled and identified by the inset networks of producers (black), herbivores (dark gray), and omnivores (light gray). Parameter values are $\alpha = 7.5$, $\beta = 5.5$, $f = 0.25$, $\delta = 2$, and $v = 1.05$.

indirect beneficial effects on producer biomass, thereby dampening or disrupting cascades in most cases (Polis and Strong 1996, Pace et al. 1999, Bruno and O'Connor 2005, Duffy et al. 2007, Kratina et al. 2012, Wootton 2017).

Documented instances of weakened and even reversed trophic cascades in food webs with omnivory (Flecker 1996, Pringle and Hamazaki 1998, Snyder and Wise 2001, Bruno and O'Connor 2005, Finke and Denno 2005, Denno and Finke 2006, Johnson et al. 2014, Fahimipour and Anderson 2015, Visakorpi et al. 2015), compared to those typically induced by predators (Shurin et al. 2002), are not uncommon and provide support to the intuitive hypothesis that omnivory precludes strong trophic cascades. However, a large meta-analysis of 114 experimental predator and omnivore manipulations in terrestrial, freshwater, and marine systems could not identify differences in the magnitudes of trophic cascades between the two groups (Borer et al. 2005). Empirical evidence to the contrary—namely, examples of strong or comparable cascades that are generated by omnivores (Power 1990, Power et al. 1992, Okun et al. 2008, France 2012)—implies that weak cascades may not be a guaranteed outcome of omnivory in food webs. Despite a growing body of theoretical and empirical work, an understanding

of when omnivores occupying top trophic positions will generate strong or weak cascading effects is lacking and likely depends on multiple population- or community-level factors (Wootton 2017).

We analyzed mathematical models describing trophic interactions between basal producers, intermediate herbivores, and top omnivores to systematically evaluate the effects of omnivory on the strength of trophic cascades. We consider two types of empirically observed foraging behaviors, namely *fixed* (Diehl and Feiel 2000) and *flexible* (Fahimipour and Anderson 2015) omnivory (see *Model Formulations* for definitions), and present a comparison between trophic cascades in these systems and traditional ones induced by analogous predators. We have chosen to study minimally detailed models to focus on coarse-grained system features that may point to potential future directions for experimental work, as opposed to making predictions about the behavior of a particular ecosystem (Anderson et al. 2009). We draw two primary conclusions based on numerical and analytical results: Stronger trophic cascades with omnivory are at least possible in high productivity systems if omnivores forage according to a fixed strategy, whereas cascades are never stronger when omnivores forage according to a flexible strategy.

MODEL FORMULATIONS

Models were analyzed with a focus on equilibrium outcomes to gain insight into how differences in the foraging strategies of species occupying top trophic levels (i.e., predators vs. fixed or adaptive omnivores) influence long-term community structure as measured by the trophic cascade. We modeled the population dynamics of three species: (1) basal producers that are eaten by (2) intermediate herbivores and (3) top omnivores that consume both producers and herbivores (Diehl and Feiel 2000). Analyses of similar three-node motifs have demonstrated how the coexistence of all species and community stability are sensitive to variation in system's primary productivity and the strength of omnivory (parameters ρ and ω in Eqs. 1a–c and 3a–c below; discussed extensively by McCann and Hastings 1997, Diehl and Feiel 2000, Gellner and McCann 2011). For this reason, a primary goal of our analysis was to elucidate how primary productivity and omnivory strength interact to influence trophic cascades in three-species motifs with and without true omnivory.

Two omnivore foraging strategies with empirical support were considered. We refer to the first as a *fixed* foraging strategy, indicating that foraging effort toward either producers or herbivores comprises constant proportions of the *fixed omnivores'* total foraging effort (McCann and Hastings 1997, Diehl and Feiel 2000). The second strategy, which we refer to as *flexible* foraging, indicates that the foraging effort apportioned toward either producers or herbivores by the *flexible omnivore* varies in time and depends on the availability and reward associated with each resource species (Kondoh 2003).

Fixed foragers

We assume a linear (type I; Holling 1959) functional response relating resource densities to per capita consumption rates, so that the dynamics of species' biomasses are represented by the system of equations

$$\frac{d\hat{r}}{dt} = \hat{\rho}\hat{r}\left(1 - \frac{\hat{r}}{\hat{k}}\right) - \hat{\alpha}\hat{r}\hat{n} - \omega\hat{\beta}\hat{r}\hat{p} \quad (1a)$$

$$\frac{d\hat{n}}{dt} = \hat{e}_{r,n}\hat{\alpha}\hat{r}\hat{n} - \hat{\mu}_n\hat{n} - \tilde{\omega}\hat{\beta}\hat{n}\hat{p} \quad (1b)$$

$$\frac{d\hat{p}}{dt} = \hat{e}_{r,p}\omega\hat{\beta}\hat{r}\hat{p} + \hat{e}_{n,p}\tilde{\omega}\hat{\beta}\hat{n}\hat{p} - \hat{\mu}_p\hat{p} \quad (1c)$$

where hats over terms indicate that they have dimensions, and \hat{r} , \hat{n} , and \hat{p} are the biomasses of producers, herbivores, and omnivores.

Here, $\hat{\rho}$ and \hat{k} are the producer productivity rate and carrying capacity, $\hat{\alpha}$ is the herbivore foraging rate, $\hat{\mu}_i$ is the per capita mortality rate of species i , and $\hat{e}_{i,j}$ is the resource i assimilation efficiency for consumer j . We assumed a total foraging rate $\hat{\beta}$ for omnivores, that is apportioned toward herbivores proportionately to $\tilde{\omega}$, where $\tilde{\omega} = 1 - \omega$. We therefore interpret ω as a nondimensional parameter describing omnivory strength (McCann and Hastings 1997); the system reduces to a food chain when $\omega = 0$. See Table 1 for a summary of all model parameters.

Flexible foragers

Eqs. 1a–c can be modified to include flexible foraging behavior by the omnivore, by substituting the omnivory strength parameter ω with the dynamical state variable Ω . Flexible foraging

Table 1. Parameter descriptions for Eqs. 1a–c, 2.

Parameter	Description	Units	Range or Value
$\hat{\rho}$	Producer productivity rate	time ⁻¹	$\hat{\rho} > 0$
\hat{k}	Producer carrying capacity	producer·area ⁻¹	$\hat{k} > 0$
$\hat{\alpha}$	Herbivore foraging rate	area·herbivore ⁻¹ ·time ⁻¹	$\hat{\alpha} > 0$
$\hat{\beta}$	Omnivore foraging rate	area·omnivore ⁻¹ ·time ⁻¹	$\hat{\beta} > 0$
ω	Fraction of omnivore foraging effort toward producers	dimensionless	$0 < \omega < 1$
$\tilde{\omega}$	Fraction of omnivore foraging effort toward herbivores	dimensionless	$1 - \omega$
$\hat{e}_{i,j}$	Conversion efficiency of resource i to consumer j	units of j ·units of i ⁻¹	$0 < \hat{e}_{ij} < 1$
$\hat{\mu}_i$	Mortality rate of species i	time ⁻¹	$\hat{\mu}_i > 0$
v	Time scale of behavioral change	dimensionless	$v > 0$

behavior was modeled using a replicator-like equation (Kondoh 2003), which provides a reasonable representation of flexible omnivory in real food webs (Fahimipour and Anderson 2015). The behavioral equation is

$$\frac{d\Omega}{dt} = v\Omega \left[\frac{\partial\gamma}{\partial\Omega} - \left(\Omega \frac{\partial\gamma}{\partial\Omega} + \tilde{\Omega} \frac{\partial\gamma}{\partial\tilde{\Omega}} \right) \right] \quad (2)$$

where $\tilde{\Omega} = 1 - \Omega$, $\gamma = \hat{e}_{r,p}\Omega\hat{\beta}\hat{r} + \hat{e}_{n,p}\tilde{\Omega}\hat{\beta}\hat{n}$ is the flexible omnivore's instantaneous per capita biomass production rate, and the constant v is a nondimensional ratio between the time scales of foraging adaptation and omnivore population dynamics (Heckmann et al. 2012). Values of $v > 1$ represent behavioral changes that occur on faster time scales than omnivore generations. This behavioral model implies that omnivores gradually adjust their foraging strategy during search if behavioral changes yield a higher instantaneous per capita biomass production rate than the current diet (Kondoh 2003).

Model nondimensionalizations and assumptions

The parameters in Eqs. 1a–c, 2 were transformed into nondimensional parameters using scaled quantities, reducing the total number of model parameters to those with values having clear interpretations (Murray 1993, Nisbet and Gurney 2003). We use substitutions similar to Amarasekare (2006, 2007): $r = \hat{r}/\hat{k}$, $n = \hat{n}/\hat{e}_{r,n}\hat{k}$, $p = \hat{p}/\hat{e}_{r,p}\hat{k}$, $\rho = \hat{\rho}/\hat{\mu}_n$, $\alpha = \hat{\alpha}\hat{e}_{r,n}\hat{k}/\hat{\mu}_n$, $\beta = \hat{\beta}\hat{e}_{r,p}\hat{k}/\hat{\mu}_n$, $f = \hat{e}_{r,n}\hat{e}_{n,p}/\hat{e}_{r,p}$, $\delta = \hat{\mu}_p/\hat{\mu}_n$, and $\tau = \hat{\mu}_n t$. Substituting into Eqs. 1a–c, 2, we obtain the nondimensional system

$$\frac{dr}{d\tau} = \rho r(1 - r) - \alpha r n - \omega \beta r p \equiv F_r(r, n, p) \quad (3a)$$

$$\frac{dn}{d\tau} = \alpha r n - n - \tilde{\omega} \beta n p \equiv F_n(r, n, p) \quad (3b)$$

$$\frac{dp}{d\tau} = \omega \beta r p + \tilde{\omega} \beta f n p - \delta p \equiv F_p(r, n, p) \quad (3c)$$

for fixed omnivory, with

$$\frac{d\Omega}{d\tau} = v\beta\Omega(r - \Omega r - \tilde{\Omega} f n) \equiv F_\Omega(r, n, p, \Omega) \quad (4)$$

representing flexible foraging behavior. For the fixed model, scaled producer, herbivore, and omnivore biomasses are represented as $x = [r, n,$

$p]$. The vector field which maps $[r, n, p]$ to $[F_r(r, n, p), F_n(r, n, p), F_p(r, n, p)]$ is denoted by $F_{\text{fixed}} : \mathbf{R}^3 \rightarrow \mathbf{R}^3$, and the coexistence equilibrium of the fixed foraging model Eqs. 3a–c is denoted by $x_{\text{fixed}}^* = [r_{\text{fixed}}^*, n_{\text{fixed}}^*, p_{\text{fixed}}^*]$. We considered trophic cascades in systems with stable equilibria, satisfying

$$[r_{\text{fixed}}^*, n_{\text{fixed}}^*, p_{\text{fixed}}^*] > 0 \quad \begin{array}{l} \text{positivity, all} \\ \text{species coexist} \end{array} \quad (5a)$$

$$F_{\text{fixed}}(x_{\text{fixed}}^*) = 0 \quad \text{equilibrium} \quad (5b)$$

$$\max_{1 \leq i \leq 3} \text{Re}\lambda_i < 0 \quad \text{stability} \quad (5c)$$

where $\dot{x}_{\text{fixed}} = F_{\text{fixed}}(x_{\text{fixed}})$ describes the system of Eqs. 3a–c, and λ_i are the eigenvalues of the Jacobian matrix evaluated at equilibrium $J_{\text{fixed}} = DF_{\text{fixed}}(x_{\text{fixed}}^*)$. These conditions ensure a straightforward comparison of trophic cascades, which in the case of nonstationary steady states would depend on the time scales under consideration (Borer et al. 2005).

Eqs. 3a–c is extended to the case of flexible foraging by replacing the fixed foraging parameter ω with a quantity satisfying Eq. 4. The system of equations is now four-dimensional and is defined by the vector field $F_{\text{flexible}} : \mathbf{R}^4 \rightarrow \mathbf{R}^4$ which maps $[r, n, p, \Omega]$ to $[F_r(r, n, p, \Omega), F_n(r, n, p, \Omega), F_p(r, n, p, \Omega), F_\Omega(r, n, p, \Omega)]$. We demand that the flexible model likewise has a coexistence equilibrium $x_{\text{flexible}}^* = [r_{\text{flexible}}^*, n_{\text{flexible}}^*, p_{\text{flexible}}^*, \Omega_{\text{flexible}}^*]$, so that $x_{\text{flexible}}^* > 0$, $F_{\text{flexible}}(x_{\text{flexible}}^*) = 0$, and all eigenvalues of the system's Jacobian matrix $J_{\text{flexible}} = DF_{\text{flexible}}(x_{\text{flexible}}^*)$ have negative real parts. Finally, for the case of predators in a food chain that do not feed on primary producers, we denote by $[r_{\text{chain}}^*, n_{\text{chain}}^*, p_{\text{chain}}^*]$ the stable and positive solution satisfying Eqs. 5a–c when setting $\omega = 0$.

A comparison of trophic cascades

We quantified differences in trophic cascade strengths between systems with omnivores (i.e., $\omega > 0$) and predators (i.e., $\omega = 0$) and examined the dependencies of these differences on model parameters. We denote by $[r_\chi^*, n_\chi^*, 0]$ the non-positive equilibrium solution to Eqs. 3a–c in the absence of predators, so that $F_{\text{fixed}}(r_\chi^*, n_\chi^*, 0) = 0$. A traditional measure of trophic cascade strength (Shurin et al. 2002, Borer et al. 2005) applied to

omnivory systems at equilibrium is therefore $\log_2(r_{\text{fixed}}^*/r_{\chi}^*)$. Likewise, cascade strength in the analogous food chain can be calculated as $\log_2(r_{\text{chain}}^*/r_{\chi}^*)$. The difference in trophic cascade strengths induced by a fixed omnivore and the predator in its analogous food chain, κ_{fixed}

$$\kappa_{\text{fixed}} := \log_2 \frac{r_{\text{fixed}}^*}{r_{\chi}^*} - \log_2 \frac{r_{\text{chain}}^*}{r_{\chi}^*} = \log_2 \frac{r_{\text{fixed}}^*}{r_{\text{chain}}^*}. \quad (6)$$

This measure κ_{fixed} of the relative cascade strength is similar to the “proportional response” measure of Heath et al. (2014), and equals 1 (or -1) if the trophic cascade induced by omnivores is twice as strong (or half the strength) as in the analogous food chain. Likewise, the difference in cascade strengths between flexible omnivory systems and food chains, $\kappa_{\text{flexible}} := \log_2 r_{\text{flexible}}^*/r_{\text{chain}}^*$. Closed-form equilibrium solutions for all variables in Eqs. 3a–c, 4 are provided in Appendix S1: Table S1.

RESULTS

Fixed omnivores can generate strong trophic cascades

We first considered the case of fixed omnivores that do not exhibit diet flexibility. In Fig. 1, we summarize changes in the relative cascade strengths induced by fixed omnivores κ_{fixed} (Fig. 1a), community dynamics, and species coexistence (Fig. 1b) as primary productivity ρ and omnivory strength ω —two key determinants in the behavior of omnivory systems (McCann and Hastings 1997, Diehl and Feiel 2001, Amarasekare 2007, Gellner and McCann 2011)—are varied, while other parameters are held constant. Consistent with prior analyses of the three-species omnivory motif, increasing omnivory strength ω causes the system to undergo a transcritical bifurcation (McCann and Hastings 1997) resulting in extinction of either omnivores at low productivities or herbivores at high productivities (Fig. 1b; Diehl and Feiel 2001, Amarasekare 2007). These occur when the determinant of the Jacobian matrix vanishes, $\det \mathbf{J}_{\text{fixed}} = 0$, for a combination of ρ and ω , marking the presence of a zero eigenvalue (Kuznetsov 2013).

Within the stable coexistence region (Fig. 1b, green and blue regions), predictions of weaker

trophic cascades with omnivory (Pace et al. 1999, Shurin et al. 2010, Kratina et al. 2012, Wootton 2017) held when primary productivity, ρ , was below a threshold value (Fig. 1a, orange region). Productivities above this threshold, however, yield omnivory cascades that are non-intuitively stronger compared to predators (Fig. 1a, blue region). This critical transition in relative trophic cascade strengths with increasing productivity occurs at a point which we refer to as ρ_{crit} or the *critical productivity* for convenience (Fig. 1a). A vertical dashed line marks the critical productivity, which is the value of ρ at which $\kappa_{\text{fixed}} = 0$, given parametrically by $\rho_{\text{crit}} = \delta\alpha^2(f-1)/f[\alpha(\delta-\beta) + \beta f(\alpha-1)]$ (Appendix S1: Fig. S1). Note that the transition from weaker ($\kappa_{\text{fixed}} < 0$) to stronger ($\kappa_{\text{fixed}} > 0$) cascades along a productivity gradient does not depend on omnivory strength. Instead, omnivory strengths near the extinction boundaries attenuate the discrepancy between cascades, such that omnivory cascades are weakest when productivity is low and ω approaches values leading to omnivore exclusion, and strongest when productivity is high and omnivores have nearly excluded herbivores (Fig. 1).

To explain the non-intuitive result of stronger cascades with fixed omnivory, we examined the relationship between primary productivity and the optimal foraging effort that would lead to the highest per capita growth rate by omnivores at equilibrium, $\Omega_{\text{flexible}}^*$ (Appendix S1: Table S1). The gray curve in Fig. 1a illustrates $\Omega_{\text{flexible}}^*$ as a function of ρ ; the growth rate-maximizing strategy monotonically approaches pure herbivory as productivity increases, recapitulating results that coexistence occurs over a wider range of omnivory strengths when omnivores forage flexibly (Křivan and Diehl 2005). Precisely at $\rho > \rho_{\text{crit}}$, the fixed omnivore is no longer able to achieve the optimal foraging strategy (Fig. 1a). Intuitively, this indicates that strong trophic cascades are induced by omnivores when their foraging effort toward producers is guaranteed to be energetically suboptimal.

We next sought to determine whether the presence of a critical productivity, or a switch from weaker to stronger cascades with fixed omnivory, depends on other model parameters. In Fig. 2, we show that ρ_{crit} (i.e., the location of the vertical dashed line in Fig. 1a along the x -axis) is

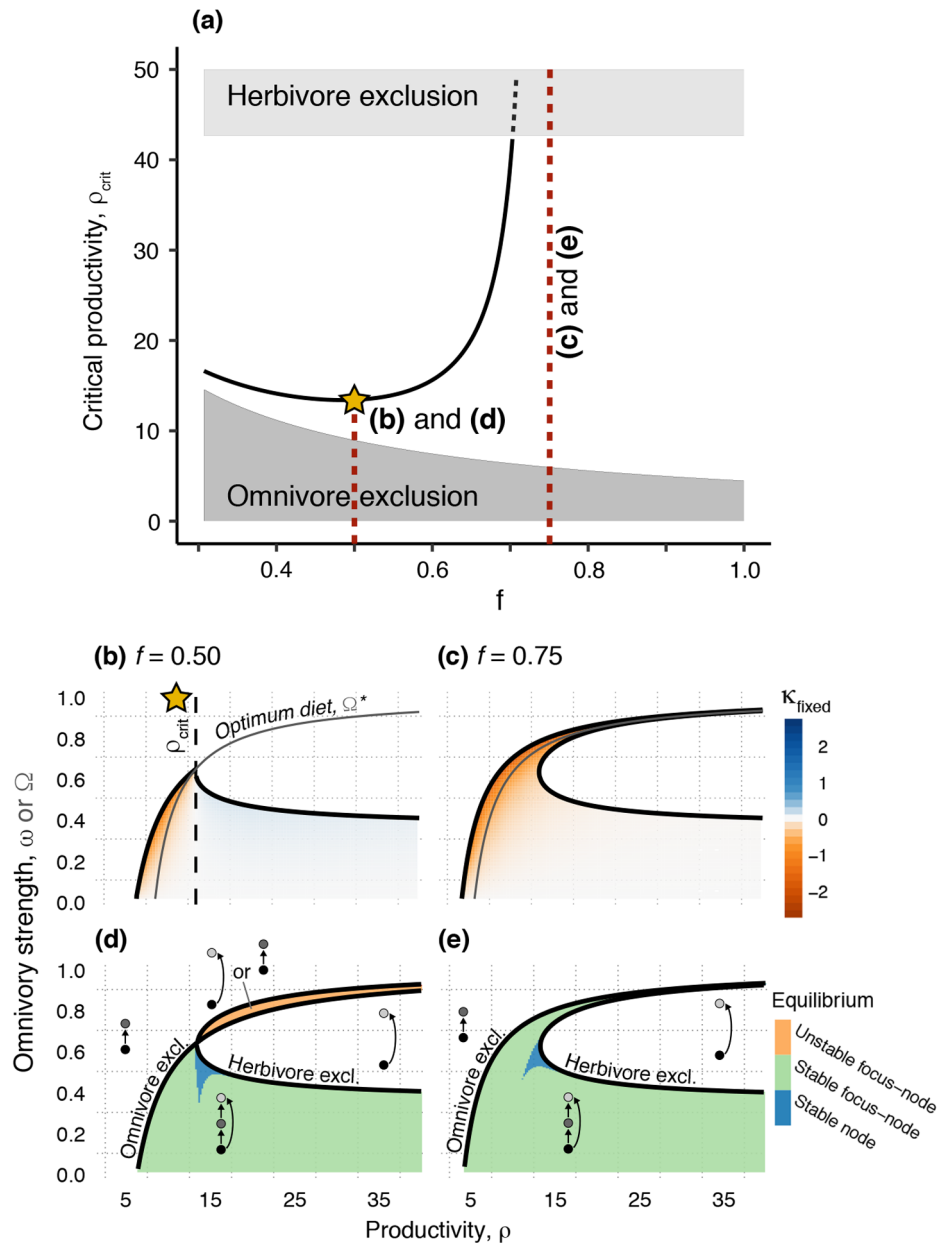


Fig. 2. (a) Relationship between critical productivity ρ_{crit} and efficiencies f (Table 1). The curve is solid if the critical productivity lies in the coexistence region, and dashed otherwise. The light and dark gray shaded regions mark the extinction of herbivores and omnivores, respectively. The left- and right-hand red dashed lines correspond to panels b, d and c, e, respectively. Parameter values are the same as in Fig. 1. (b, c) Relationships between relative cascade strength κ_{fixed} , productivity ρ , and omnivory strength ω . Colors show κ_{fixed} values within the three-species coexistence region. Blue represents stronger cascades with omnivory, and orange represents weaker cascades with omnivory. Black curves mark extinction boundaries for either the omnivore or herbivore species (see panels d and e). The gray curve shows the equilibrium foraging strategy for the flexible omnivore, $\Omega^*_{flexible}$. A vertical dashed line marks the critical productivity ρ_{crit} . (d, e) Bifurcation curves identify stability boundaries separating steady states with different dynamics. Colors mark different types of equilibria. Non-coexistence regions are labeled and identified by the inset networks of producers (black), herbivores (dark gray), and omnivores (light gray).

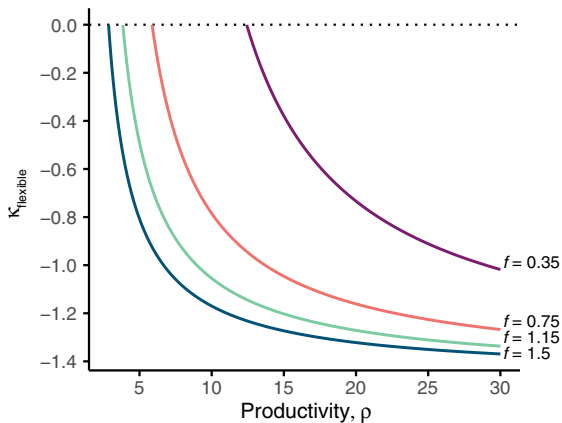


Fig. 3. Numerical solutions relating relative cascade strength κ_{flexible} , scaled productivity ρ , and scaled resource profitability f in the flexible model. Colors represent associated values of f in the margin. Parameter values are $\alpha = 7.5$, $\beta = 5.5$, $\delta = 2$, and $v = 1.05$.

sensitive to the conversion efficiency parameter, f . Recall that f is the product of the producer-to-herbivore conversion efficiency, $\hat{e}_{r,n}$, and the ratio of herbivore- and producer-to-omnivore conversion efficiencies (i.e., omnivore rewards), $\hat{e}_{n,p}/\hat{e}_{r,p}$ (Table 1). Model-based and experimental studies have suggested that $\hat{e}_{r,n} > \hat{e}_{r,p}$ and $\hat{e}_{n,p} > \hat{e}_{r,p}$ are realistic conditions for omnivores in nature (Diehl and Feißel 2000, Křivan and Diehl 2005). Satisfying these conditions, we would generally expect small f values when herbivores are only slightly more rewarding than producers to omnivores (for instance, if $\hat{e}_{n,p} = \hat{e}_{r,p} + \epsilon$ where ϵ is a small number), and large f values when rewards from eating herbivores are much higher than for producers, $\hat{e}_{n,p} \gg \hat{e}_{r,p}$. For large enough values of f , the curve of ρ_{crit} enters a non-coexistence region (Fig. 2a). Thus, the potential for strong omnivory cascades is lost as f increases, regardless of other population- or community-level properties; an example of when this happens is shown in Fig. 2c, e. We examine the sensitivity of these results to other model parameters in Appendix S1: Fig. S1. Briefly, the critical productivity shifts to the right (i.e., larger ρ_{crit} values lead to smaller parameter regions with strong omnivory cascades) as α and δ increase, and shifts toward zero (i.e., smaller ρ_{crit} values lead to an expansion of the parameter region with stronger omnivory cascades) as β increases (Appendix S1: Fig. S1).

Flexible omnivores never generate stronger trophic cascades

Unlike fixed omnivores, flexibly foraging omnivores can never induce cascades that are stronger than in the analogous food chain. We show analytically that at a positive equilibrium solution, $\kappa_{\text{flexible}} < 0$. At the interior equilibrium (Appendix S1: Table S1), if $\phi := -\alpha\delta + \delta f(\alpha - \rho) + \beta f(\rho - 1)$ then

$$p_{\text{flexible}}^* = \frac{\phi}{f\beta^2}$$

$$\Omega_{\text{flexible}}^* = \frac{\alpha\delta + f\rho(\delta - \beta)}{-\phi}.$$

The ratio of flexible omnivory to linear chain trophic cascade strengths,

$$\kappa_{\text{flexible}} = \log_2 \frac{r_{\text{flexible}}^*}{r_{\text{chain}}^*} = \log_2 \frac{\frac{\delta}{\beta}}{1 - \frac{\alpha\delta}{\beta f \rho}}.$$

As $p_{\text{flexible}}^* > 0$ and evidently $f\beta^2 > 0$, we must have $\phi > 0$. Moreover, since $0 < \Omega_F^* < 1$, we must also have $\delta f \rho < \beta f \rho - \alpha\delta$. That is,

$$\alpha\delta + f\rho(\delta - \beta) < 0. \quad (7)$$

Combining Eq. 7 with the solution $\Omega_{\text{flexible}}^*$ (Appendix S1: Table S1) shows that for positive equilibria, $\kappa_{\text{flexible}} < 0$, since $\phi < 0$ cannot be true for a biological system. Thus, consistent with conceptual models of trophic cascades (Strong 1992, Pace et al. 1999), cascades in systems with flexibly foraging top omnivores are bounded in strength by those in their analogous food chains. Numerical results confirm these analytical expectations and illustrate how increasing consumer reward ratios (i.e., increasing f) attenuates this result but does not alter the qualitative relationship between κ_{flexible} and ρ (Fig. 3).

DISCUSSION

Intuition suggests that trophic cascades will not occur when top predators additionally feed on primary producers (Polis and Strong 1996, Pace et al. 1999, Duffy et al. 2007, Shurin et al. 2010, Kratina et al. 2012, Wootton 2017), but our results predict that strong cascades will emerge under a wider range of foraging types than

previously expected. We identified many cases in which omnivores are indeed likely to generate weak cascades, although we have shown that this should not be a uniform expectation. Particularly, in high productivity systems in which foraging rewards do not strongly differ between producers and herbivores (Figs. 1a, 2a), fixed omnivores are capable of generating stronger cascades than would be expected if they did not consume producers at all. This is due to suboptimal omnivore foraging and the additional source of herbivore population losses in models of fixed omnivory, in which the herbivore must compete with its own consumer for resources (Diehl and Feiel 2000). This result provides at least one general explanation for the weak (Finke and Denno 2005, Denno and Finke 2006), comparable or indistinguishable (Borer et al. 2005), and strong (Okun et al. 2008, France 2012) cascades that have now been observed with omnivorous top predators: They largely depend on primary productivity and the types of omnivory. It is not surprising that a more comprehensive catalogue of foraging behaviors will improve predictions of trophic cascades, but our model-based results indicate that this knowledge may be especially important when species consume resources across trophic levels.

Comparisons of fixed and flexible models showed that omnivores were capable of generating strong cascades only when consuming an energetically suboptimal level of primary producers could be guaranteed (Figs. 1a, 2b, c). This leads to the question: How common is this type of fixed foraging in food webs? Empirical evidence for approximately fixed foraging exists for groups as diverse as protists, arthropods, and mammals (Clark 1982, Diehl and Feiel 2001, Mooney and Tillberg 2005). Fixed omnivory may also manifest in other ways, for example, when organisms forage in a way that is suboptimal in terms of pure energetics but is otherwise required to maintain nutritional or stoichiometric balances (Berthoud and Seeley 1999, Remonti et al. 2016, Zhang et al. 2018). Suboptimal foraging has also been observed in heavily disturbed or human-altered systems where consumer behaviors are not adapted to current resource conditions, or when changes in habitat structure alter the ability to efficiently locate preferred food sources (Walsh et al. 2006).

Allometric scaling relationships between species' demographic rates and body masses have helped identify biological constraints on the strengths of trophic cascades in food chains with top predators (DeLong et al. 2015), but body mass may have additional implications for cascades that are generated by species facing complex foraging decisions. The prevalence of dynamical or adaptive foraging behaviors, like those represented by our flexible model, across the tree of life has shown associations with organismal brain sizes and body masses by proxy (Eisenberg and Wilson 1978, Rooney et al. 2008, Edmunds et al. 2016). Body mass distributions may also influence cascades that are induced by species with size-mediated ontogenetic shifts from herbivory to carnivory (Pace et al. 1999), wherein average population-level foraging behaviors could be characterized as "omnivory" and would reflect intraspecific size structures. Future empirical work and simulation-based analyses of more complex models will be key for uncovering additional relationships between species' body masses and trophic cascades in food webs and to develop a coherent understanding of when foraging behavior drives deviations from predictions of cascades from simple tritrophic food chain models. In many of these cases, omnivory could appear as an average population-level behavior and not necessarily at the level of the individual.

Our analysis focuses on models characterized by type I functional responses that relate resource biomass to consumer growth. Alternate nonlinear functional responses (e.g., Holling type II; Holling 1959) may modulate the effects of omnivory on trophic cascades. Preliminary analyses show that closed-form equilibrium solutions similar to those in Appendix S1: Table S1 can also be obtained for type II functional responses. The predictive power of these solutions in cases where the model shows oscillatory behavior remains an open question. We conjecture that, for mild instabilities, the oscillatory behavior introduced by saturating consumption would result in similar qualitative outcomes predicted by equilibrium values when cascades are measured as time-averaged quantities (Fox 2007). However, for larger-amplitude oscillations cascade strengths will likely depend strongly on the time scale over which they are measured,

potentially yielding a mechanism for the observation that cascade strengths are related to experimental duration (Borer et al. 2005). Models that formally extend the concept of trophic cascades to cases of nonstationary equilibria will be an important direction for future analyses.

Fig. 1 suggests an interesting analytical question for future study. Namely, multiple qualitative changes are observed precisely at the phase transition for strong omnivory cascades, $\kappa_{\text{fixed}} > 0$, which is indicated by the vertical dashed line $\rho = \rho_{\text{crit}}$. Also occurring at this point are the phase boundaries for species coexistence at stable equilibrium given by the curves separating different non-coexistence regions (Fig. 1b); the gray curve showing the optimal omnivory strength, $\Omega_{\text{flexible}}^*$, as a function of ρ passes into a region that is unattainable by the fixed forager (Fig. 1a), and two saddle-node bifurcation curves intersect (Fig. 1b). It remains to understand why these curves all intersect at a single point, and how this relates to qualitative changes in trophic cascade strengths.

Examples from agroecosystems and disturbed natural habitats indicate that trophic cascade theories can directly inform applied management problems and efforts to mitigate human alteration of ecosystems (Schmitz 2006, Estes et al. 2011). Our comparative analyses together with the ubiquity of omnivory in nature (Arim and Marquet 2004, Kratina et al. 2012, Wootton 2017) suggest that omnivores may contain promise for such applications of cascade theory. For instance, nutrient inputs to agricultural systems that lead to artificially enriched communities are exactly the conditions where we expect a potential for strong omnivorous cascades. If management goals include reducing the density of agricultural pests in enriched systems through integrated strategies that manipulate top trophic levels, then, counterintuitively, top omnivores with certain features may warrant additional consideration (Agrawal et al. 1999). Achieving these outcomes in practice may prove challenging (Cortez and Abrams 2016).

CONCLUSIONS

Omnivory has long been cited as a reason for why trophic cascades are less frequent or weaker than expected, although empirical data on the role

of omnivory have been equivocal (Borer et al. 2005, Shurin et al. 2010, Kratina et al. 2012, Wootton 2017). Our theory generally agrees with the prediction of omnivory in weakening cascades, but also demonstrates where these predictions are weak or even where they exhibit unexpected changes. Thus, these predictions generate a framework for future investigation that can focus expectations on when and where omnivory effects might occur in more complex ecosystems. At the least, our models help elucidate the mixed support for an intuitive ecological prediction.

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