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et al.

Publication Date

2021-12-01

DOI

10.1016/j.scitotenv.2021.149591

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1 **Long-term leaf C:N ratio change under elevated CO<sub>2</sub> and nitrogen deposition in**  
2 **China: evidence from observations and process-based modeling**

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20 **Highlights:**

- 21 ● Leaf C:N ratio decreased in southern China, but increased in western and northern China  
22 in the past 5 decades.
- 23 ● N deposition was a dominant factor driving C:N ratio change in most areas, while  
24 elevated CO<sub>2</sub> dominated in low N deposition area.
- 25 ● Leaf C:N ratio shifts were more strongly constrained by N availability than by climate.

26

27 **Keywords:** leaf C:N ratio, data-driven modelling, process-based modelling, nitrogen deposition

28

29 **Abstract**

30 Climate change, elevating atmosphere CO<sub>2</sub> (eCO<sub>2</sub>) and increased nitrogen deposition  
31 (iNDEP) are altering the biogeochemical interactions between plants, microbes and soils, which  
32 further modify plant leaf carbon-nitrogen (C:N) stoichiometry and their carbon assimilation  
33 capability. Many field experiments have observed large sensitivity of leaf C:N ratio to eCO<sub>2</sub> and  
34 iNDEP. However, the large-scale pattern of this sensitivity is still unclear, because eCO<sub>2</sub> and  
35 iNDEP drive leaf C:N ratio toward opposite directions, which are further compounded by the  
36 complex processes of nitrogen acquisition and plant-and-microbial nitrogen competition. Here,  
37 we attempt to map the leaf C:N ratio spatial variation in the past 5 decades in China with a  
38 combination of data-driven model and process-based modeling. These two approaches showed  
39 consistent results. Over different regions, we found that leaf C:N ratio had significant but uneven  
40 changes between 2 time periods (1960-1989 and 1990-2015): a 5% ± 8% increase for temperate  
41 grasslands in northern China, a 3% ± 6% increase for boreal grasslands in western China, and by  
42 contrast, a 7% ± 6% decrease for temperate forests in southern China, and a 3% ± 5% decrease  
43 for boreal forests in northeastern China. Additionally, the structural equation models indicated  
44 that the leaf C:N change was sensitive to ΔNDEP, ΔCO<sub>2</sub> and ΔMAT rather than ΔMAP and  
45 ecosystem types. Process-based modeling suggested that iNDEP was the main source of soil  
46 mineral nitrogen change, dominating leaf C:N ratio change in most areas in China, while eCO<sub>2</sub>  
47 led to leaf C:N ratio increase in low iNDEP area. This study also indicates that the long-term leaf  
48 C:N ratio acclimation was dominated by climate constraint, especially temperature, but was  
49 constrained by soil N availability over decade scale.

50

51

52 **1 Introduction**

53 Leaf stoichiometry examines the relationships between organism structure and function  
54 (Sardans et al., 2012; Sterner and Elser, 2002). Leaf C:N ratio can be associated with important  
55 ecological processes such as litter decomposition (d'Annunzio et al., 2008), the ability to adapt  
56 to environmental stresses (Hessen et al., 2004; Woods et al., 2003), and enzymes activities of the  
57 plant photosynthesis (Evans, 1989; Evans and Poorter, 2001). Lower leaf nitrogen status (higher  
58 C:N ratio) increases non-stomatal limitations on photosynthesis through the impairment of  
59 carboxylation capacity, photochemical efficiency, and osmotic protection (Wright et al., 2005).  
60 In contrast, higher foliar N concentration (lower C:N ratio) may be associated with larger C-  
61 assimilation and growth-rate capacities (Sterner and Elser, 2002). Given the tight connection  
62 between leaf C:N ratio and productivity at the canopy scale (Reich, 2012), leaf C:N ratio is an  
63 important parameter in terrestrial ecosystem model and plays a vital role in carbon sequestration  
64 projection (Caldararu et al., 2020; Ghimire et al., 2016; Meyerholt and Zaehle, 2015).

65 Leaf C:N ratios are driven by variation in plant physiology, soil biogeochemistry, and  
66 plant community composition (Reich and Oleksyn, 2004). In the past 50 years, climate change  
67 (CLIM), elevating atmosphere CO<sub>2</sub> (eCO<sub>2</sub>) and increased nitrogen deposition (iNDEP) were  
68 likely to have changed the biogeochemical interactions between plants, microbes and soil, which  
69 further modified plant leaf C:N ratio and their carbon assimilation capability. There are mainly  
70 two hypothesizes predicting ecosystem level leaf C:N ratio acclimation under environment  
71 change: (a) the climate envelope theory, and (b) the theory of trade-off between resources  
72 acquisition and conservation. The climate envelope theory assumes that local environment  
73 constrains the species trait range, thus environmental change is likely to filter out inappropriate  
74 species or trait range, resulting in plant trait being environment selected (Atkin et al., 2015;

75 Hijmans and Graham, 2006; Kattge and Knorr, 2007; Wright et al., 2005). For example, low leaf  
76 C:N ratio means an adaptation that enhances metabolic activity and growth rates under the low  
77 temperatures. Within species, populations from colder habitats often have greater leaf N (lower  
78 C:N ratio) (Reich and Oleksyn, 2004; Weih and Karlsson, 2001; Woods et al., 2003). So, it is  
79 likely that the ecosystem level leaf C:N ratio would increase under climate warming, and there  
80 would be similar sensitivities of leaf C:N ratio change in both spatial and temporal variation to  
81 environment factors.

82 On the other hand, the theory of trade-off between resources acquisition and conservation  
83 assumes that plants will adjust relevant traits to optimize the balance between resources uptake  
84 and the cost of acquisition (Kong et al., 2014; Reich, 2014). Leaf N concentration typically  
85 reflect soil N availability (Vitousek et al., 1995), and more soil N means less cost of acquisition  
86 for plants to hold more N. This hypothesis is partly supported by short-term experiments.  
87 Warming experiments could increase leaf N concentration, which is linked to increases in N  
88 mineralization, cycling rates and availability due to enhanced soil microbial activity (Bai et al.,  
89 2013). In CO<sub>2</sub> enrichment experiments, decreasing leaf N concentration and therefore lower  
90 photosynthesis capacity were observed, especially in nitrogen limited environment (Ainsworth  
91 and Long, 2005; Crous et al., 2010; Ellsworth et al., 2004). By contrast, nitrogen fertilization is  
92 likely to restore leaf N concentration and to stimulate photosynthesis (Liu et al., 2011; McCarthy  
93 et al., 2006; Norby et al., 2003; Palmroth et al., 2006; Sigurdsson et al., 2013). Additionally,  
94 several global meta-analysis studies found that leaf C:N ratio shows a strong positive correlation  
95 to eCO<sub>2</sub> and a negative correlation to nitrogen addition for nearly any species in any locations in  
96 the world (Du et al., 2019; Huang et al., 2015; Li et al., 2013; Yang et al., 2011; Yue et al.,  
97 2017). Hence, climate warming tends to increase leaf N concentration (decrease leaf C:N ratio)

98 by accelerating the decomposition and mineralization of soil organic matter, which increases the  
99 availability of soil N. Meanwhile, eCO<sub>2</sub> could enhance plant C acquisition but increase the  
100 burden of N uptake, leading a higher leaf C:N ratio. Human induced N addition (iNDEP) could  
101 also decrease the leaf C:N ratio by increasing soil N availability. Therefore, the trade-off theory  
102 implies the possibility that environment factors sensitivities of leaf C:N ratio change in spatial  
103 and temporal scales could be different.

104         Despite its significance in understanding ecosystem dynamics, the large-scale pattern of  
105 long-term leaf C:N stoichiometry change under environment changes and its sensitivity to  
106 diverse environmental factors under natural conditions are still unclear. This uncertainty emerges  
107 from several reasons. First, as mentioned above, the mechanisms and processes related to leaf  
108 C:N ratio variation are complicated. Different hypotheses lead to opposite results. Second, the  
109 combined effects of environmental factors are difficult to predict. For instance, the opposite  
110 impacts of eCO<sub>2</sub> and iNDEP may offset the final direction of change in leaf C:N stoichiometry,  
111 making the compound response to combined influences of climate, eCO<sub>2</sub> and iNDEP weak and  
112 unclear (Yue et al., 2017). Third, synthesis are often based on small spatial scale short-term  
113 experiments, which seem to overestimate the response of long-term plant acclimation at larger  
114 spatial scales (Leuzinger et al., 2011).

115         Among the limited number of in-situ field studies, their observations were inconsistent.  
116 Natural plants in China experienced nearly 30% increasing of leaf N concentration between  
117 1980s and 2010s (Liu et al., 2013). By contrast, in Russia and Western Europe, leaf N  
118 concentration had an 8% decrease (Jonard et al., 2015; Soudzilovskaia et al., 2013), and in  
119 America about 20% decrease in the similar time period (McLauchlan et al., 2010). Nonetheless,  
120 in recent years, thanks to efforts of the whole plant science community, more plant stoichiometry

121 trait data become available, which include the TRY database (<https://www.try->  
122 [db.org/TryWeb/Home.php](https://www.try-db.org/TryWeb/Home.php)) (Kattge et al., 2011). This in conjunction with non-linear regression  
123 tools development for large-scale plant trait mapping (Butler et al., 2017; Moreno-Martinez et  
124 al., 2018) makes it possible to use data-driven methods to establish a large-scale map of leaf  
125 stoichiometry change, which would help us to understand the inconsistency in different plot scale  
126 studies under complex environment conditions, to discover the dominant environment factor of  
127 leaf C:N stoichiometry, and to test theories and hypotheses related to leaf C:N stoichiometry  
128 change. On the other hand, given the observed large sensitivity of leaf C:N ratio to environment  
129 change, more and more land surface models (LSM) are replacing the fixed stoichiometry plant  
130 formulation with a flexible CN ratio capability to improve the estimates of gross primary  
131 productivity (GPP) or net primary productivity (NPP) (Caldararu et al., 2020; Ghimire et al.,  
132 2016; Meyerholt and Zaehle, 2015). This advance in process-based modeling provides another  
133 tool to predict long-term leaf C:N ratio change at the large spatial scale under environment  
134 changes. Moreover, the sensitivity experiments by numerical model can also be used to test  
135 hypotheses related to leaf C:N ratio change.

136 In this study, we employ both data-driven modeling and process-based modeling to map the  
137 large-scale leaf C:N ratio change and analyze their sensitivities to environmental factors. We  
138 selected China as the study region, given its NDEP experienced the largest increase in the world  
139 in the past thirty years (Liu et al., 2013) and uneven distribution in different regions, e.g., a 0.5  
140  $\text{gN m}^{-2} \text{yr}^{-1}$  NDEP in Tibetan Plateau vs 5  $\text{gN m}^{-2} \text{yr}^{-1}$  NDEP in eastern China (Fig S1).  
141 Moreover, the leaf trait data in China have covered a good spatial and temporal distribution (Fig  
142 S2).

143 Our study was guided by the following questions:



- 144 1) What is the spatiotemporal pattern of leaf C:N ratio change in the past 50 years in  
145 China? Does it have large differences in different regions or for different plant  
146 functional types (PFT)?
- 147 2) What was the most sensitive environment factor for leaf C:N ratio change in China,  
148 climate change, eCO<sub>2</sub> or iNDEP? Were leaf C:N ratio change determined by climate  
149 envelope or determined by the balance between soil nitrogen supply and plant demand?
- 150 3) Does process-based modelling agree with data-driven modeling in inferred leaf C:N  
151 pattern and changes? And how well does the hypothesis in the process-based model  
152 work? What is its implication to carbon cycling modelling?

## 153 **2 Material and Methods**

154 In this study, we focus on the spatiotemporal variation of the ecosystem level leaf C:N  
155 ratio in China. We firstly collected in-situ plant traits data, and then used two methods to derive a  
156 spatiotemporal conterminous leaf C:N ratio map: (1) random forest based up-scaling of in-situ  
157 trait measurements, and (2) large scale process-based modeling using the Community Land  
158 Model version 5.0 (Lawrence et al., 2019).

### 159 2.1 In-situ plant traits data collection and data gap-filling

160 The in-situ plant trait database in this study include two variables: leaf carbon  
161 concentration (LCC in mgC g<sup>-1</sup>) and leaf nitrogen concentration (LNC in mgN g<sup>-1</sup>). These traits  
162 data consist of two parts: one is obtained from the TRY database, and the other is collected from  
163 the literature using *ISI Web of Knowledge*, *Google scholar* and *CNKI* website. Overall, we  
164 collected 28406 records worldwide, where 24529 records were contributed from the TRY  
165 database (2978 plots in China), and the other 3931 records (all in China) were collected from the

166 literature. The detail of literature sources and plot location distribution are in Table S1 and  
167 Figure S2, respectively.

168 To improve the data consistency through pre-processing, we eliminated unreasonable  
169 data and then gap-filled using the method from Moreno-Martinez et al., (2018) (see Appendix 1).  
170 Moreover, to explore the leaf trait temporal variation, we made two vital assumptions: 1) if there  
171 is not any temporal information in the original datasets, the measurement time of the plot was  
172 assigned as 2 years before publication time, following the method in Liu et al., (2013); and 2) the  
173 LNC and LCC trait measurement, mainly sampling fully expanded and hardened leaves  
174 (presumably photosynthetically more productive) from adult plants (Cornelissen et al., 2003),  
175 were representative of the decadal homeostatic status. The second assumption also helps to  
176 reduce the bias resulting from uncertainty in determining the sampling time.

177 In the gap filling process, we used the python based random forest implementation  
178 `sklearn.ensemble.RandomForestRegressor` ([https://scikit-  
179 learn.org/stable/modules/generated/sklearn.ensemble.RandomForestRegressor.html?highlight=ra  
180 ndom%20forest#sklearn.ensemble.RandomForestRegressor](https://scikit-learn.org/stable/modules/generated/sklearn.ensemble.RandomForestRegressor.html?highlight=random%20forest#sklearn.ensemble.RandomForestRegressor)) (Pedregosa et al., 2011) and  
181 segmented the predictors into 4 categories: taxonomic hierarchies, plant structure, multiple  
182 annual mean environment factors (representing long-term steady climate), and their decadal  
183 variance (representing temporal climate change) (See Table S2).

184 For climate input data, we chose the monthly 0.5x0.5 degree CRU TS 4.02 global climate  
185 dataset for the period of 1950-2015 (<https://crudata.uea.ac.uk/cru/data/hrg/>) (Harris et al., 2014).  
186 The monthly cloudiness, frost day, potential evaporation, precipitation, diurnal temperature  
187 range, mean air temperature were acquired from the datasets. A yearly 0.9x1.25 degree global  
188 nitrogen deposition (NDEP) dataset produced by atmospheric chemical transport model was

189 acquired from University Corporation for Atmospheric Research (NCAR) for the period of 1960-  
190 2015, regridded to 0.25x0.25 degree, and then corrected using site observation from Lü and Tian  
191 (2007). The NDEP spatiotemporal pattern in China from the corrected dataset is consistent with  
192 other observation studies (See Fig. S1, Zhu et al. (2015) and Yu et al. (2019)). The global spatial  
193 average CO<sub>2</sub> concentration time serial was acquired from NCAR for the period of 1960-2015.

194 R<sup>2</sup> score was used to evaluate the performance of the gap-filling processes. Fig. S3  
195 showed the performance in gap-filling with a high R<sup>2</sup> both in training process (0.68 for LNC and  
196 LCC) and in validation process (0.62 for LNC and 0.64 for LCC).

## 197 2.2 Random forest up-scaling of in-situ measurements of plant traits

198 After data pre-processing, we up-scaled the in-situ measurement to the 0.25-degree grid  
199 according to the major PFT information for a given grid, and then interpolated the results to  
200 other grids using the random forest algorithm, following Moreno-Martinez et al., (2018) (see  
201 Appendix S2 for detail). In this process, an accurate high spatial resolution land cover data is  
202 essential, and for this we used the Land Use Harmonization 2 (LUH2, [luh.umd.edu/data.shtml](http://luh.umd.edu/data.shtml))  
203 PFT datasets, because it used a robust classification method, covering over 15 PFTs at a spatial  
204 resolution of 0.25 degree with sub-grid PFT information worldwide (Lawrence et al., 2016).  
205 Here, we assumed that there was no land use/cover change (LUCC) or plant functional type  
206 change in China in the past 5 decades. This assumption is partly supported by the LUCC studies  
207 for the past 3 decades (the expansion of cultivated land was less than 40 thousand km<sup>2</sup> and the  
208 degradation of forest was less than 20 thousand km<sup>2</sup> in the past 30 years in China, which are tiny  
209 numbers compared to the total rural area 8533 thousand km<sup>2</sup> in China (Liu et al., 2010; Song and  
210 Deng, 2017), see Table S3).

211 The representativeness of climate and NDEP condition of global plot samples with  
212 respect to the China region was also tested in this study. The mean annual precipitation (MAP),  
213 mean annual air temperature (MAT), and PFT of plot samples and China region are presented in  
214 Fig. S4. The climate of plot samples has a similar distribution with the China region across  
215 different PFTs. For example, the MAT of boreal grassland ranges from -10.4 °C to 8.5°C for plot  
216 samples vs from -7.9 °C to 11.1°C for the China region. The MAP of grassland ranges from 250  
217 mm to 1550 mm for plot samples vs from 305mm to 1241 mm for the China region (See Table  
218 S5). The NDEP probability density curves across different PFTs are presented in Fig S5. The  
219 probability density curves of the plot samples are also representative of the typical distributions  
220 of China regions.

221 Fig. S6 showed the performance in up-scaling, with  $R^2$  0.64 for LNC and 0.72 LCC in the  
222 training process, and 0.35 for LNC and 0.46 for LCC in validation process, respectively. The  $R^2$   
223 for up-scaling is also comparable to other studies with a range from 0.25 to 0.56 (Table S4). Next,  
224 given the relatively low spatial resolution of the CRU dataset (0.5 degree) applied to China at  
225 and to ensure the consistency with land surface model meteorology forcing, we employed the  
226 validated model (regression relationships) with a new climate dataset of a finer spatial resolution  
227 (0.25 degree) to predict a relatively finer up-scaled map for LNC and LCC (See 2.3 for the detail  
228 of the dataset).

### 229 2.3 Process-based modelling: Community Land Model version 5.0

230 A state-of-the-art process-based model, Community Land Model version 5.0 (CLM5.0),  
231 was selected as an alternative method to analyze the leaf C:N ratio change in China. CLM5.0 is  
232 selected because 1) as a newly released simulation tool, CLM5.0 has the capability of simulating  
233 leaf carbon mass, leaf nitrogen mass, and leaf C:N ratio change with a process-based scheme of

234 flexible leaf stoichiometry; 2) CLM5.0 couples plant C, N and leaf stoichiometry explicitly, thus  
235 enabling us to analyze the impact of leaf stoichiometry on plant carbon and nitrogen cycle  
236 modelling.

237 CLM5.0 is the default land component model for the Community Earth System Model  
238 version 2 (CESM2) (<http://www.cesm.ucar.edu/models/cesm2/>). As a land surface model,  
239 CLM5.0 represents processes such as soil and plant hydrology, river routing, coupled carbon and  
240 nitrogen cycling, and crop dynamics. We refer more detail of the CLM5.0 to its full technical  
241 description ([http://www.cesm.ucar.edu/models/cesm2/land/CLM50\\_Tech\\_Note.pdf](http://www.cesm.ucar.edu/models/cesm2/land/CLM50_Tech_Note.pdf)) (Lawrence  
242 et al., 2019) and below only give a brief introduction to leaf stoichiometry or flexible CN ratio  
243 parameterization in the model. CLM5.0 assumes that plant take up nitrogen at the cost of energy  
244 in the form of carbon (Doughty et al., 2018), which thence allows plant to adjust their CN ratio  
245 by making trade-off between nitrogen uptake and associated carbon expenditure (Ghimire et al.,  
246 2016). This flexible CN ratio also results in a new nitrogen limitation scheme differing from  
247 CLM4.5 by removing the instantaneous down-regulation of potential GPP induced by  
248 insufficient soil mineral nitrogen to support plant demand. More details are described in  
249 Appendix S3.

250 We conducted CLM5.0 simulations of China at  $0.25^\circ \times 0.25^\circ$  and half-hourly  
251 spatiotemporal resolution for the period from 1960 to 2015. In order to reduce the bias in climate  
252 forcing, we combined data from two sources, 1) daily surface air temperature (mean, maximum  
253 and minimum air temperature), relative humidity, wind speed and sunshine duration from 736  
254 stations across China from 1960 to 2015 procured from China Meteorological Administrations  
255 (<http://cdc.nmic.cn>). 2) China Gridded Daily Precipitation Product with a 25 km spatial  
256 resolution across China from 1960 to 2015 obtained from China Meteorological Administrations

257 (<http://cdc.cma.gov.cn/dataSetLogger.do?changeFlag=dataLogger>). These two datasets are then  
258 used to derive a forcing data at 0.25° x 0.25° degree and 3-hourly spatiotemporal resolution in  
259 China for the period from 1960 to 2015, following the method in Lei et al. (2014). We also  
260 replaced the default soil texture data with a local version at 1 km spatial resolution (Shangguan et  
261 al., 2012). Other input datasets, including land cover, nitrogen deposition, aerosol deposition,  
262 CO<sub>2</sub> concentration, adopted the default configuration for CLM5.0 (Lawrence et al., 2019). For a  
263 consistent comparison, it is noted that we used the same climate, land cover, nitrogen deposition  
264 and CO<sub>2</sub> concentration datasets in the data-driven approach.

265 In order to separate the impact from different environmental factors, we devised  
266 simulations of four scenarios.

- 267 ● S0: static climate with all environmental factors being constant,
- 268 ● S1: climate change with other factors being constant,
- 269 ● S2: climate change and elevated atmosphere CO<sub>2</sub> concentration, and
- 270 ● S3: climate change, elevated atmosphere CO<sub>2</sub> concentration and nitrogen deposition  
271 change.

272 Based on these four simulations, the influence of climate change ('CLIM') was estimated  
273 from the difference between S1 and S0. The effect of CO<sub>2</sub> enrichment ('eCO<sub>2</sub>') was deducted by  
274 the difference between S2 and S1. And the impact of nitrogen deposition ('iNDEP') was  
275 estimated by the difference between S3 and S2. More details for the four-scenario model  
276 configurations are in Table S3.

277 Besides S0-S3, we launched another scenario simulation to test the sensitivity of carbon  
278 and nitrogen cycle to leaf C:N ratio change (by the difference of S3 and S4):

279 ● S4: model scheme with constant leaf C:N ratio, and forced by climate change, elevated  
280 atmosphere CO<sub>2</sub> concentration and nitrogen deposition change.

281 For scenario S3 was designed to simulate the real-world condition, an evaluation of  
282 simulated nitrogen and carbon cycling was conducted by comparing S3 and observation data.

283 As for nitrogen cycle validation, the net soil mineralization (16 sites), plant nitrogen  
284 uptake (24 sites) and NO<sub>x</sub> emission (27 sites) fluxes data were collected for site scale validation.  
285 Natural biological nitrogen fixation (NBNF), nitrogen deposition and denitrification fluxes at  
286 nation scale were compared to other studies to validate the gross volume of the fluxes. Overall,  
287 CLM5.0 gives a reasonable nitrogen cycle modelling, with an appropriate R<sup>2</sup> (>0.44) in site level  
288 validation (see Fig. S7) and consistent gross volume with other studies (see Table. S7).

289 For carbon cycle validation, GPP and terrestrial ecosystem respiration (TER) data in 35  
290 flux sites were collected for site scale validation, and remote sensing based MODIS C6 GPP  
291 product (Running S., 2015) and GIMMS3g LAI product (Zhu et al., 2013) were selected for  
292 spatial pattern validation. R<sup>2</sup> score was selected to evaluate the performance. Overall, CLM5.0  
293 gives a reasonable carbon cycle modelling, with a high R<sup>2</sup> (>0.65) in site level validation (see Fig.  
294 S8) and a high R<sup>2</sup> (>0.70) in spatial pattern validation (See Fig S9 and Fig S10). These  
295 performance of R<sup>2</sup> are also comparable to other carbon cycle modeling studies (See Table S8).

## 296 2.4 Analysis

297 As LCC, LNC and leaf C:N ratio are numerical ratios with significant differences  
298 between PFTs, we conducted the spatial pattern analysis by averaging over grids and PFTs rather  
299 than by over a region or whole nation. In addition, to make the modeling results more easily and  
300 simpler to understand, we rearranged the 15 PFTs classified in LUH2 into 5 PFTs, for example,  
301 by combining the 3 boreal forest PFTs (boreal evergreen needleleaf forest, boreal deciduous

302 needleleaf forest and boreal deciduous forest) into 1 PFT (boreal forest) (Fig. S11). In addition,  
 303 due to the lack of leaf trait data for crops, and the complexity of land management in crop area,  
 304 such as fertilization and seed technology improvement, we excluded crop land from this study.  
 305 Moreover, given the data-driven model can only consider one PFT for one grid, we only  
 306 analyzed the grids with major land cover proportion over 80%.

307 For temporal analysis, we separated the whole time period from 1960 to 2015 into 2  
 308 parts: 1960-1990 and 1991-2015. We defined the relative change of the variable ( $\Delta V$ ) as the  
 309 difference between the average values in these two periods (Eq 1). This time period  
 310 segmentation was determined based on 2 reasons: 1) significant environment change occurred  
 311 roughly after 1990 in China, e.g., nitrogen deposition (Liu et al., 2013) and air temperature  
 312 warming (Piao et al., 2010); 2), comparing a relatively long time period can ameliorate the  
 313 uncertainty due to insufficient leaf data samples in the early period, especially in the 1970s and  
 314 1980s.

$$315 \quad \Delta V = \frac{\overline{V_{1990-2015}} - \overline{V_{1960-1989}}}{\overline{V_{1960-1989}}} \quad (\text{Eq. 1})$$

316 For statistical analysis, we used one-way ANOVA and Tukey's test to analyze the  
 317 influences of different PFTs. In order to examine the effects of different factors, such as PFT,  
 318 MAT, MAP, NDEP and CO<sub>2</sub> concentration, we used piecewise structural equation modelling for  
 319 generalized linear regression models (Lefcheck and Freckleton, 2015), which was widely used in  
 320 plant stoichiometry studies (Hu et al., 2021; Luo et al., 2021). In order to make the regression,  
 321 boreal or temperate ecosystem was classified to 0 or 1, and grassland or forest ecosystem was  
 322 classified to 0 or 1. The initial model was constructed based on the hypothesized relationships  
 323 suggested in previous studies (Hu et al., 2021; Luo et al., 2021; Yuan and Chen, 2009). We fitted  
 324 the component models of the piecewise structural equation model as generalized linear



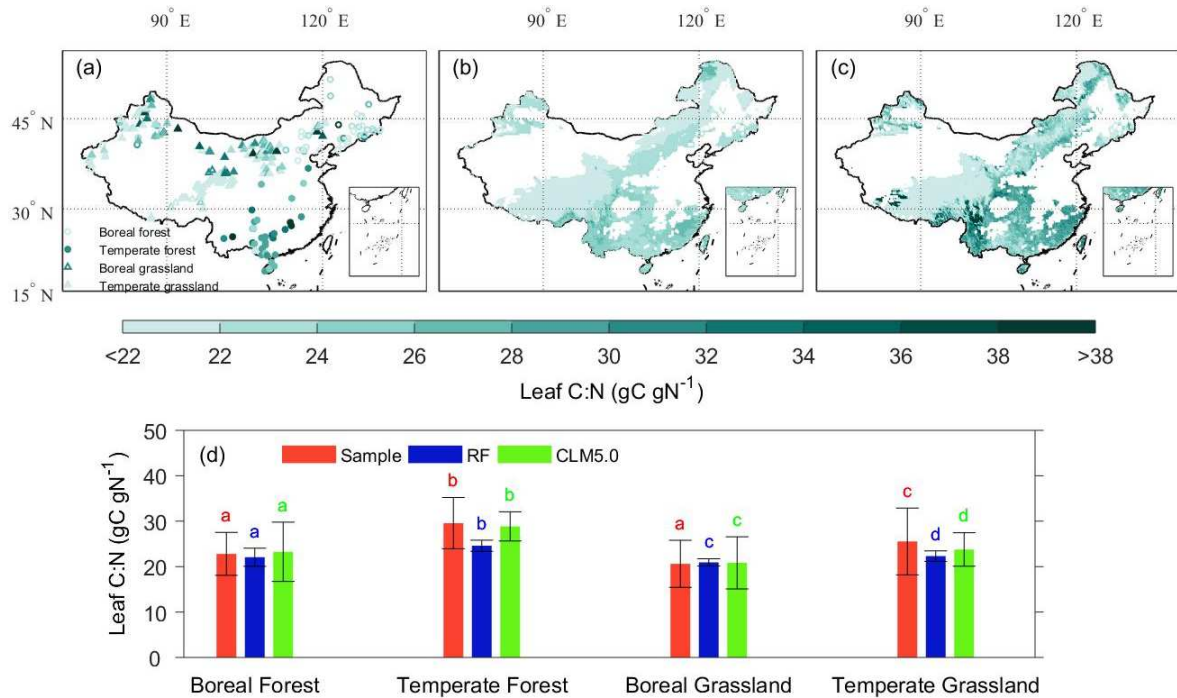
325 regression models. Goodness-of-fit for the overall model was evaluated using Fisher's C statistic,  
326 and  $p > 0.05$  showed a good fit. The standardized regression coefficient ( $\beta$ ) for each path was  
327 then estimated, and the significance was tested.

328 The whole process of piecewise structural equation modelling was performed using the R  
329 package "piecewiseSEM" (Lefcheck and Freckleton, 2015). All data analyses were performed  
330 using R v.4.0.1.

### 331 **3 Result**

#### 332 3.1 Spatial variation of multi-annual mean leaf C:N ratio

333 For spatial variation, leaf C:N ratio in southern China was larger than that in northern  
334 China in general (Fig 1a, 1b, and 1c). Leaf C:N ratio differed among different PFTs ( $p < 0.05$ ;  
335 Fig 1d). Temperate ecosystem had a higher leaf C:N ratio than boreal ecosystem, and forest  
336 ecosystem had a higher leaf C:N ratio than grassland ecosystem in China by nearly 3  
337 approaches (temperate forest in southern China  $>$  temperate grassland in northwestern China  $>$   
338 boreal forest in northeastern China  $>$  boreal grassland in the Tibetan Plateau). Among the three  
339 approaches, leaf C:N by plot sample data had the largest variation values with temperate forest  
340 ( $29.6 \pm 5.6 \text{ gC gN}^{-1}$ ), temperate grassland ( $25.5 \pm 7.4 \text{ gC gN}^{-1}$ ), boreal forest ( $22.8 \pm 4.7 \text{ gC gN}^{-1}$ )  
341 and boreal grassland ( $20.6 \pm 5.2 \text{ gC gN}^{-1}$ ), which was consistent to the results by CLM5.0 with  
342 temperate forest ( $28.9 \pm 3.2 \text{ gC gN}^{-1}$ ), temperate grassland ( $23.8 \pm 3.7 \text{ gC gN}^{-1}$ ), boreal forest  
343 ( $23.3 \pm 6.5 \text{ gC gN}^{-1}$ ) and boreal grassland ( $20.8 \pm 5.7 \text{ gC gN}^{-1}$ ). The variation obtained by  
344 random forest approach was also consistent with other two approaches but with a smaller value  
345 with temperate forest ( $24.6 \pm 1.2 \text{ gC gN}^{-1}$ ), temperate grassland ( $22.3 \pm 1.2 \text{ gC gN}^{-1}$ ), boreal  
346 forest ( $22.1 \pm 2.0 \text{ gC gN}^{-1}$ ) and boreal grassland ( $20.8 \pm 0.8 \text{ gC gN}^{-1}$ ).

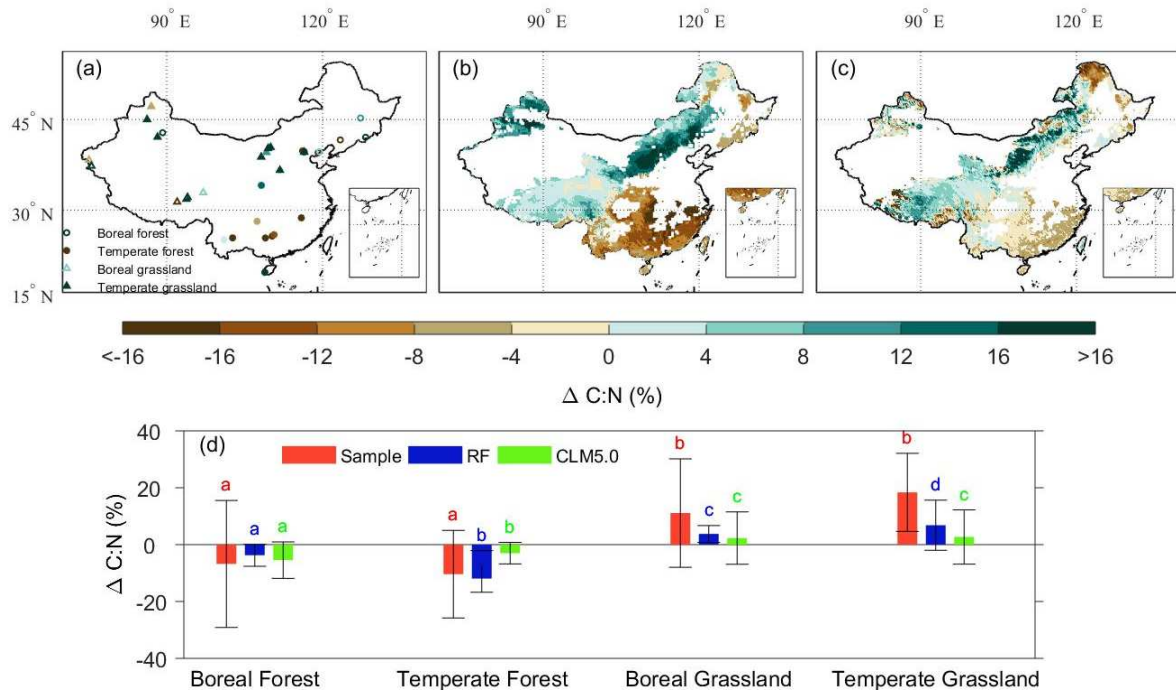


347  
 348 **Figure 1.** Spatial variation of multi-annual mean leaf C:N ratio in China (1960-2015); (a) result  
 349 by plot samples, (b) result by random forest (RF) model, (c) result by CLM5.0; (d) results from  
 350 all above 3 approaches averaged by PFT. Different lowercase letters in the same color on error  
 351 bars indicate significant differences across different PFTs at  $p < 0.05$ .

### 352 3.2 Temporal variation of multi-annual mean leaf C:N ratio

353 For temporal variation of leaf C:N ratio change in the past 5 decades, grassland in the  
 354 northwest experienced an increase in leaf C:N ratio, while forest in the southwest had a decrease  
 355 (Fig 2a, 2b, and 2c). The plot samples presented a significant difference of leaf C:N ratio change  
 356 between forest in the east ( $-9.3\% \pm 15.4\%$ ) and grassland in the west ( $+15.4\% \pm 13.7\%$ ) ( $p < 0.05$ ,  
 357 Fig 2a and 2d). The data-driven model result was consistent with respect to observed spatial  
 358 distribution both in direction and magnitude with leaf C:N ratio decreasing for forest ( $-7.7\% \pm$   
 359  $4.2\%$ ) and increasing for grassland ( $+5.4\% \pm 6.2\%$ ) (Fig 2b and 2d). Comparatively, CLM5.0  
 360 roughly captured the direction of leaf C:N ratio change in most areas in China. However, the

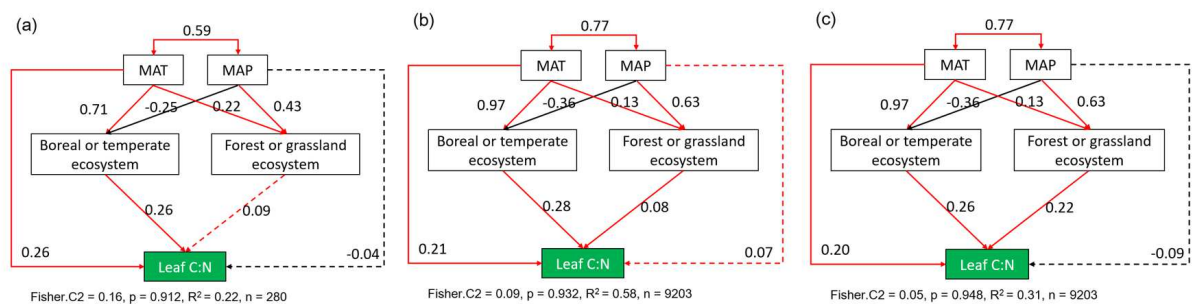
361 magnitude of change was 50% lower in CLM5.0 than in the data-driven method (Fig 2c and 2d).  
 362 If result from the data-driven model is regarded as observational benchmark, then CLM5.0  
 363 prediction with flexible CN is generally reasonable but needs to improve the leaf C:N sensitivity  
 364 to environment change. In spite of their methodological differences, these two modeling  
 365 approaches both showed an increasing leaf C:N ratio for temperate grassland in northern China  
 366 (about  $+5.3\% \pm 8.2\%$ ) and for boreal grassland in western China (about  $+3.3\% \pm 6.6\%$ ), while a  
 367 decrease for temperate forest in southern China (about  $-7.7\% \pm 6.6\%$ ) and boreal forest in  
 368 northeastern China (about  $-3.3\% \pm 5.5\%$ ) (Fig 2d).



369 **Figure 2.** Temporal variation of multi-annual mean leaf C:N ratio in China; (a) result by plot  
 370 samples, (b) result by random forest (RF) model, (c) result by CLM5.0; (d) results from all above  
 371 3 approaches averaged by PFT. Different lowercase letters in the same color on error bars  
 372 indicate significant differences across different PFTs at  $p < 0.05$ .  
 373  
 374

375 3.3 Leaf C:N ratio spatial sensitivities to environmental factors

376 According to simple linear regression models, MAT and MAP were significantly  
 377 positively correlated to leaf C:N ratio spatial variation from all 3 approaches ( $p < 0.05$ , Fig S12).  
 378 The different PFTs were also significant correlated to leaf C:N ratio spatial variation ( $p < 0.05$ ,  
 379 Fig 1). Besides, MAT and MAP dominated the PFTs spatial distribution (Fig S4). The structural  
 380 equation models were used to diagnose the causal relationships between these factors (Fig 3).  
 381 For direct effects, climate factors and PFT had comparable importance to leaf C:N variation ( $\beta$   
 382 and significance were similar). Furthermore, temperature related factors (MAT and boreal or  
 383 temperate ecosystem) had stronger impact on leaf C:N variation than precipitation related factors  
 384 (MAP and forest or grassland ecosystem). For indirect effects, MAT rather than MAP dominated  
 385 the ecosystem types (boreal or temperate), and had a further impact on leaf C:N variation. The  $\beta$   
 386 and significance of the paths from all three approaches were consistent with each other, and the  
 387 results indicate that MAT was the most important factor to the spatial variation of multi-annual  
 388 mean leaf C:N ratio.



389

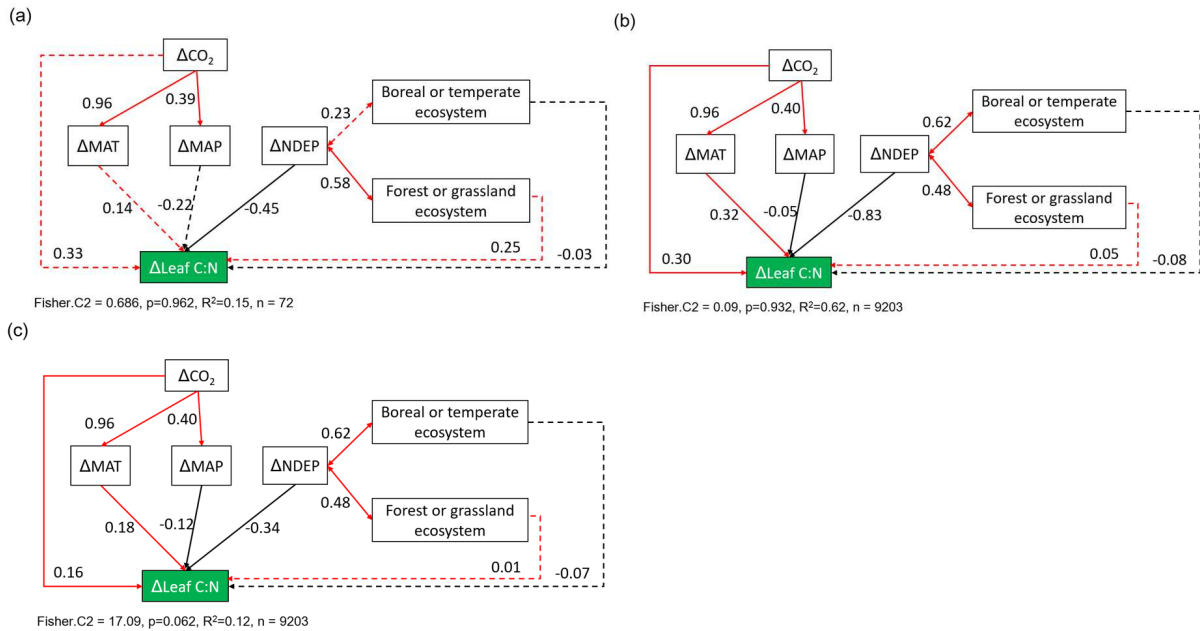
390 **Figure 3.** Structural equation models of PFT and climate as predictors of multi-annual mean leaf  
 391 C:N ratio spatial variation in China. Red lines = positive and significant; black lines = negative  
 392 and significant; dashed lines = insignificant. Standardized regression coefficients for each path

393 are given, and results for goodness-of-fit tests are also reported underneath each plot ( $p > 0.05$   
394 indicates a good fit); (a) by plot sample data, (b) by random forest model, (c) by CLM5.0

### 395 3.4 Leaf C:N ratio temporal sensitivities to environmental factors

396 According to simple linear regression models,  $\Delta$ NDEP was significantly negatively  
397 correlated to leaf C:N ratio temporal variation ( $p < 0.05$ ), and  $e\text{CO}_2$  was related to an increase in  
398 mean leaf C:N ratio from all 3 approaches (Fig S13). The leaf C:N temporal variation of plot  
399 sample data had no significant correlation with climate factors, by contrast a significantly  
400 negative correlation with  $\Delta$ MAP and a significantly positive correlation with  $\Delta$ MAT were  
401 observed from random forest model and CLM5.0. Structural equation models were used to  
402 eliminate the collinearity of the factors (Fig 4). For direct effects,  $\Delta$ NDEP had the strongest  
403 effect to reduce  $\Delta$ leaf C:N ( $\beta$  range from -0.34 to -0.83,  $p < 0.05$ ), by contrast  $\Delta$ CO<sub>2</sub> played an  
404 important role to increase  $\Delta$ leaf C:N ( $\beta$  range from 0.16 to 0.30,  $p < 0.05$ ).  $\Delta$ MAT also led to an  
405 increase in  $\Delta$ leaf C:N ( $\beta$  range from 0.18 to 0.32,  $p < 0.05$ ), but  $\Delta$ MAP had a weaker correlation  
406 to  $\Delta$ leaf C:N ( $\beta$  range from -0.05 to -0.12,  $p < 0.05$ ). Furthermore, ecosystem types had no  
407 significant relationship with  $\Delta$ leaf C:N, although different PFTs presented significant difference  
408 in  $\Delta$ leaf C:N in Fig 2. This contradiction could be explained by the colinearity of  $\Delta$ NDEP and  
409 PFTs' distribution ( $r^2$  range from 0.48 to 0.62,  $p < 0.05$ ), so that  $\Delta$ NDEP rather than ecosystem  
410 types was likely the driving factor to  $\Delta$ leaf C:N. For indirect effects,  $\Delta$ CO<sub>2</sub> significantly  
411 correlated to  $\Delta$ MAT ( $\beta = 0.96$ ,  $p < 0.05$ ) and  $\Delta$ MAP ( $\beta$  range from 0.39 to 0.40,  $p < 0.05$ ), and  
412 had a further impact on  $\Delta$ leaf C:N. The  $\beta$  values and significance of the paths from random forest  
413 model and CLM5.0 were consistent with each other, by contrast the  $\Delta$ leaf C:N relationships with  
414  $\Delta$ CO<sub>2</sub> and climate were not significant by plot samples, probably due to the tiny sample size ( $n$

415 =72). Generally, the results indicate the importance of  $\Delta$ NDEP,  $\Delta$ CO<sub>2</sub> and  $\Delta$ MAT in leaf C:N  
 416 ratio temporal variation.



417  
 418 **Figure 4.** Structural equation models of PFT, climate, nitrogen deposition (NDEP) and CO<sub>2</sub>  
 419 concentration as predictors of multi-annual mean leaf C:N ratio temporal variation in China. Red  
 420 lines = positive and significant; black lines = negative and significant; dashed lines =  
 421 insignificant. Standardized regression coefficients for each path are given, and results for  
 422 goodness-of-fit tests are also reported underneath each plot (p > 0.05 indicates a good fit); (a) by  
 423 plot sample data, (b) by random forest model, (c) by CLM5.0.

#### 424 4. Discussion

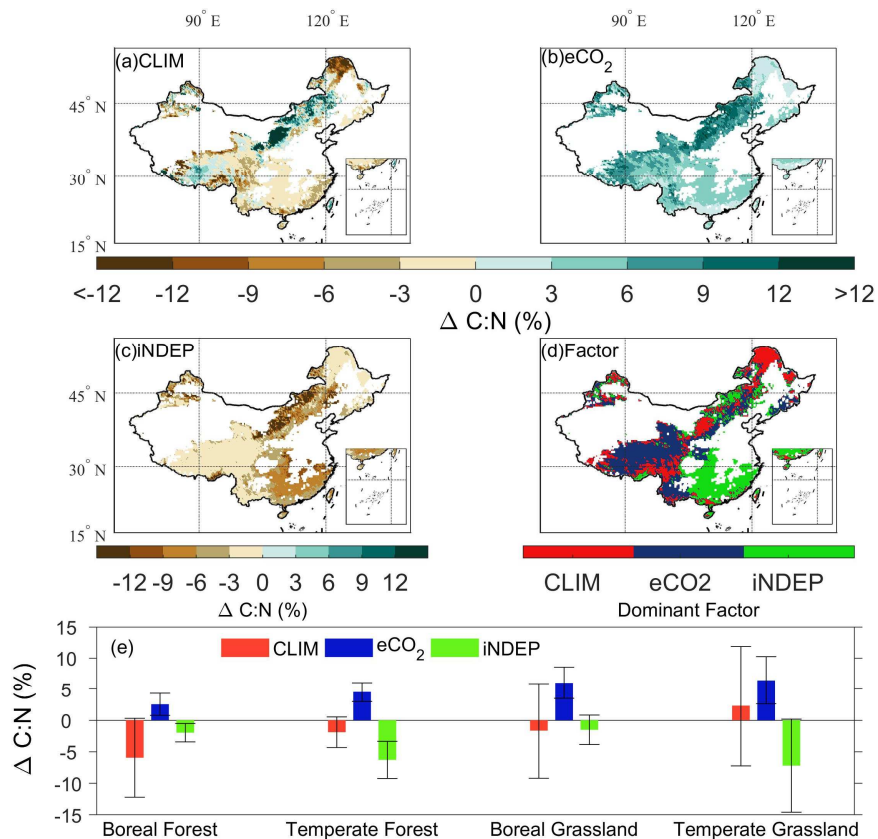
##### 425 4.1 Difference in leaf C:N spatial and temporal variation sensitivities to environmental factors

426 In our study, we find that the spatial variation of multi-annual mean leaf C:N ratio has a  
 427 strong correlation with long-term mean climate (especially MAT) and ecosystem types by plot  
 428 sample data, random forest model and CLM5.0 (Fig 1 and Fig3). The spatial variation of leaf

429 C:N follows the climate envelop theory. The ecosystem in a colder environment tends to have a  
430 lower leaf C:N ratio, and MAT is likely to be a driving factor to influence ecosystem types and  
431 finally leaf C:N ratio. These results were consistent with the findings at global scale (Hu et al.,  
432 2021; Reich and Oleksyn, 2004) and regional scale (Han et al., 2005; Luo et al., 2021; Tang et al.,  
433 2018).

434 However, the temporal variation sensitivities to environment factors are not consistent  
435 with the spatial variation sensitivities. We find that temporal  $\Delta$ leaf C:N was more sensitive to  
436  $\Delta$ NDEP,  $\Delta$ CO<sub>2</sub> and  $\Delta$ MAT than to  $\Delta$ MAP and ecosystem types with larger  $\beta$  and more  
437 significant path (Fig 4). This inference was also supported by scenario simulations using  
438 CLM5.0 (Fig 5). From CLM5.0, we inferred that the decreasing leaf C:N ratio in southern China  
439 was dominated by iNDEP, with very high  $\Delta$ NDEP in the past 5 decades (from 1.5 gN m<sup>-2</sup> yr<sup>-1</sup> to  
440 5 gN m<sup>-2</sup> yr<sup>-1</sup>) (Fig 5c, 5d and S1). eCO<sub>2</sub> also induced significant leaf C:N ratio increase,  
441 especially in southwestern China where NDEP was lower (0.3 gN m<sup>-2</sup> yr<sup>-1</sup>) (Fig 5b and 5d).  
442 Overall in most area in China,  $\Delta$ leaf C:N was determined by the relative strength between  
443 iNDEP and eCO<sub>2</sub> (Fig 5d and 5e). These results are consistent with observations in Europe,  
444 reporting that tree leaf nitrogen concentration became lower in low nitrogen deposition area in  
445 north Europe, and by contrast, in central Europe where nitrogen deposition was higher, leaf  
446 nitrogen concentration did not decrease (Jonard et al., 2015; Mellert and Göttlein, 2012).

447



448

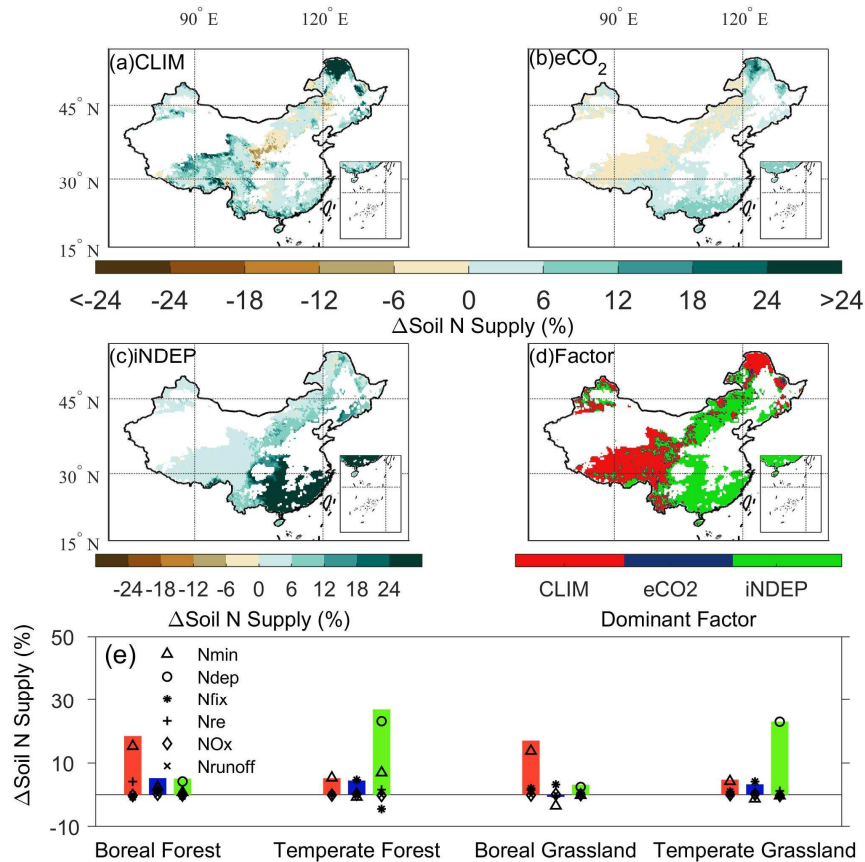
449 **Figure 5.** Leaf C:N ratio sensitivities to environment factors by numerical experiments with  
 450 CLM5.0; (a) climate change; (b) CO<sub>2</sub> enrichment; (c) nitrogen deposition increase; (d) dominant  
 451 factor in different region; (e) sensitivities from CLM5.0 averaged by PFT.

452 The differences between spatial and temporal sensitivities also suggest that the climate  
 453 envelop mechanism (leaf C:N constrained by climate and its change) may be more appropriate  
 454 for long-term relationship between leaf traits and environment, when plant groups or species are  
 455 given sufficient time to replace one another along environmental gradients (Yang et al., 2019),  
 456 but not at the first order to influence the leaf C:N change in decade scale. By contrast, soil N  
 457 uptake and supply in soil-microbe-plant biogeochemical processes are likely to dominate the leaf  
 458 C:N change in decade scale. Soil nutrient availability dominating the maintenance of plant  
 459 stoichiometric homeostasis was widely observed empirically (Luo et al., 2021; Schreeg et al.,



460 2014; Yan et al., 2014; Yu et al., 2010). In this study, with CLM5.0 nitrogen cycle modelling  
461 results, soil N available under environmental change was deduced to be dominated by iNDEP  
462 and warming. iNDEP was found as the most important nitrogen source that changed soil mineral  
463 nitrogen supply in China, especially in temperate ecosystems (50% contribution for temperate  
464 forest, 70% contribution for temperate grassland Fig 6d and Fig 6e), and was also the dominant  
465 factor to soil mineral nitrogen pool change (Fig S14a). In boreal ecosystems, nitrogen release  
466 from soil organic matter decomposition (net nitrogen mineralization) was the primary supplier to  
467 soil nitrogen, contributing 55% for boreal forest and 80% for boreal grassland (Fig 6e). The  
468 importance of atmospheric N deposition to soil mineral N was widely reported (Geng et al., 2021;  
469 Marty et al., 2017; Mgelwa et al., 2020), indicating that iNDEP increasing soil N availability  
470 could be a dominating mechanism to influence leaf C:N ratio change in China. Besides, our  
471 results also confirm previous studies that reported a dilution effect under eCO<sub>2</sub> with a lower leaf  
472 N concentration (Deng et al., 2015; Sardans et al., 2017). This probably was related to  
473 progressive N limitation, PNL (Luo et al., 2004; Reich et al., 2006), and caused by the unbalance  
474 of soil N uptake demand and supply.

475



476

477 **Figure 6.** Soil mineral nitrogen supply temporal variation simulated by CLM5.0; (a) climate  
 478 change; (b) CO<sub>2</sub> enrichment; (c) nitrogen deposition increase; (d) dominant factor; (e) result  
 479 averaged by PFT. Nmin is for net nitrogen mineralization; Ndep is for nitrogen deposition; Nfix  
 480 is for biological nitrogen fixation; Nre is for nitrogen translocation, NO<sub>x</sub> is for NO<sub>x</sub> (NO and  
 481 N<sub>2</sub>O) emission in denitrification process, Nrunoff is for soil nitrogen leaching loss.

#### 482 4.2 Implication to carbon cycle modelling

483 Recently, two major changes are being made to improve carbon cycle models, one is  
 484 enabling flexible plant trait coupling (Fisher et al., 2015; Peaucelle et al., 2019; Scheiter et al.,  
 485 2013), the other focuses on incorporating nutrient limitation processes (Shi et al., 2016; Wang et

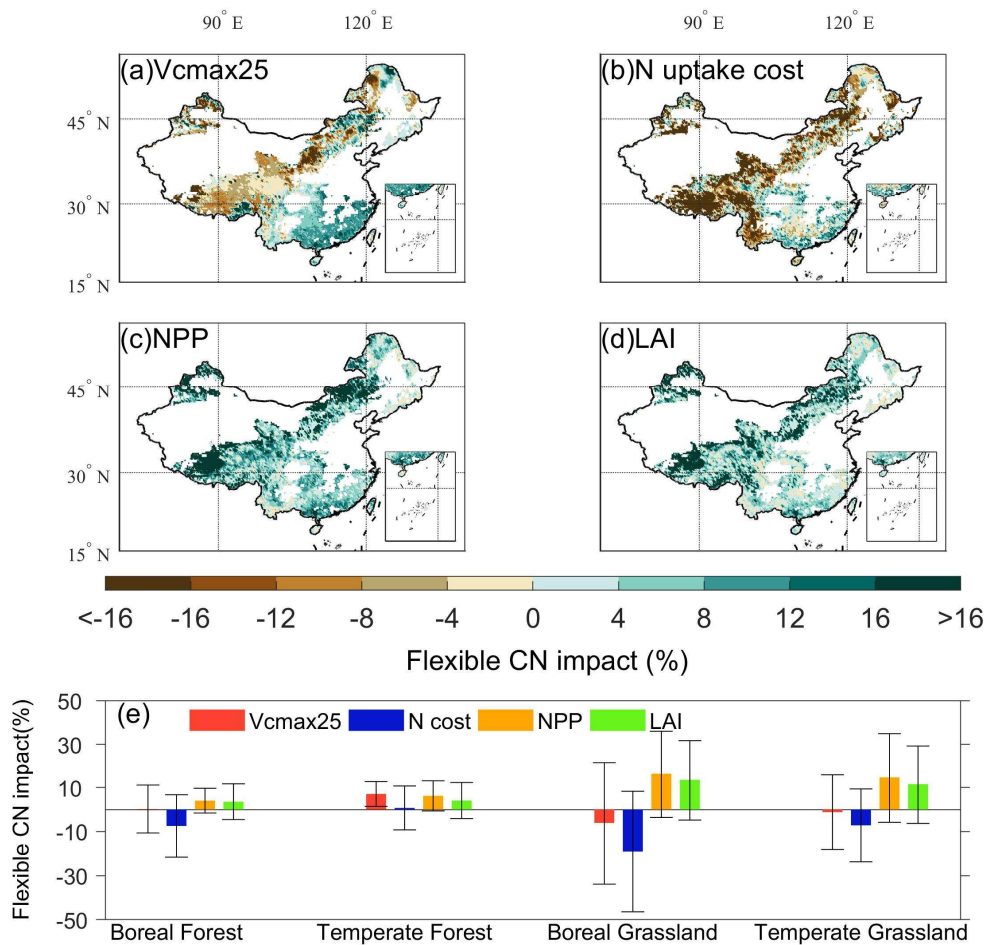
486 al., 2010; Zaehle et al., 2014; Zhu et al., 2019). Improving leaf stoichiometry and representing its  
487 flexibility is at the intersection of these two changes.

488 First, we infer that the climate envelope theory of trait-environment interaction is likely  
489 not suitable for modeling leaf stoichiometry at decade scale. Therefore, we should be cautious on  
490 employing simple trait and environment regression relationships in dynamic vegetation models  
491 or land surface models (Verheijen et al., 2015; Walker et al., 2017). By contrast, results from the  
492 process-based model CLM5.0 which assumes nutrient acquisition to be balanced with metabolic  
493 cost are consistent with relationships inferred from both plot data and data-driven model (Fig 2),  
494 indicating the trade-off theory is likely more reasonable.

495 The FUN model used in CLM5.0 is one of the first to represent the interaction between  
496 plant and microbe nutrient uptake interaction (Fisher et al., 2010; Shi et al., 2016), which is a  
497 possible scheme to reflect plant trade-off between resources acquisition and conservation. The  
498 scheme represented plant-microbial symbiosis explicitly, which was observed in more and more  
499 filed studies (Terrer et al., 2016; Terrer et al., 2018). Even though the process representation is  
500 very simple and empirical, it likely has improved the model performance by incorporating these  
501 complex interactions between plant, microbe and soil.

502 Second, we corroborate previous studies that flexible plant CN ratio is likely to have a  
503 significant impact on modeled carbon cycle (Caldararu et al., 2020; Peaucelle et al., 2019;  
504 Walker et al., 2017).  $V_{\text{cmax}25}$  is observed to be highly correlated to leaf C:N ratio change and  
505 presented an opposite direction to leaf C:N ratio change. Here CLM5.0 predicts that  $V_{\text{cmax}25}$  in  
506 northwestern China has decreased about 10%, while in southeastern China has increased about  
507 10% caused by leaf C:N ratio change (Fig 7a). The increase of  $V_{\text{cmax}25}$  was often higher than N  
508 uptake cost (Fig 7b). Therefore, the flexible CN scheme tends to increase plant carbon

509 assimilation for most areas, leading to higher NPP and LAI (Fig 7c and 7d). Further, CLM5.0  
 510 predicts that grassland ecosystem responded more strongly in NPP and LAI than forest  
 511 ecosystem (about  $16\% \pm 20\%$  vs  $5\% \pm 8\%$ )(Figure 7e). Hence, given the continuously  
 512 increasing atmospheric CO<sub>2</sub> concentration and N deposition in the world, the flexible plant  
 513 stoichiometry scheme is essential in carbon cycle modelling and projection.



514  
 515 **Figure 7.** Flexible CN ratio impact on carbon cycle modelling in CLM5.0; (a) V<sub>cmax25</sub>; (b)  
 516 nitrogen acquisition cost; (c) net primary productivity (NPP); (d) leaf area index (LAI); (e)  
 517 results average by PFT.

518 4.3 Limitation and prospection

519 We noticed that the data-driven model and the process-based CLM5.0 were making  
520 different predictions. Generally, the leaf C:N ratio sensitivities to both eCO<sub>2</sub> and iNDEP in  
521 CLM5.0 are lower than those in the data-driven model. The increasing leaf C:N ratios in low  
522 NDEP area predicted by CLM5.0 are around 50% lower than in the data-driven model, and the  
523 decreasing leaf C:N ratios in high NDEP area are around 70% lower than in the data-driven  
524 model (Fig 2). The experimental manipulations of CLM5.0 also reported these weak LNC  
525 sensitivities (Wieder et al., 2019).

526 Here, we reckon two possible reasons for these weak sensitivities. First, the meaning and  
527 representativeness of leaf C:N ratios in data-driven model and CLM5.0 are slightly different. The  
528 measurement data mainly sampled photosynthetic productive leaves of plant, which are likely to  
529 have a stronger LNC variation within canopy. Field measurement reported that the bottom  
530 canopy layer had a stronger LNC seasonal variation than upper canopy during the growing  
531 season (Coble et al., 2016). Leaf spectroscopy by remote sensing also revealed higher variation  
532 in sun-lit leaf LNC (Yang et al., 2016). The vertical LNC pattern within canopy is constrained by  
533 hydraulic conductance, bottom canopy leaves tend to have a higher LNC within canopy (Coble  
534 et al., 2016; Niinemets, 2012). These observations thus indicated that the higher canopy leaves  
535 formed by stimulated carbon assimilation under eCO<sub>2</sub> have a larger chance to decrease LNC due  
536 to the coupling between nitrogen and water transport. By contrast, the process-based model  
537 CLM5.0 uses a big leaf model and assumes all leaves as one plant organ of the same nutrient  
538 variation, and is thus unable to represent the within canopy LNC difference.

539 Second, there could be missing mechanisms in the LNC variation representation in  
540 CLM5.0. For example, field measurement revealed that elevated CO<sub>2</sub> could limit nitrogen

541 transportation from soil to leaf with less transpiration (McDonald et al., 2002; McGrath and  
542 Lobell, 2013), by contrast, iNDEP was likely to enhance transpiration (Zhou et al., 2017) to  
543 transport more nitrogen to canopy. Therefore, by lacking a linkage between nutrient-mass flow  
544 and transpiration, CLM5.0 tends to underestimate the LNC response to eCO<sub>2</sub> and iNDEP.  
545 Furthermore, in field manipulation experiments, leaf C:N ratio was found to be most sensitive to  
546 both eCO<sub>2</sub> and nitrogen addition, by contrast, woody structure tissues had the lowest sensitivity  
547 (Du et al., 2019; Xia and Wan, 2008; Yang et al., 2011). These observations indicate that plant  
548 carbon and nitrogen tissue allocation strategies could be very different. For woody plants, root  
549 and leaf tend to have less increase in biomass but larger increase in nitrogen concentration in  
550 nitrogen addition experiments, by contrast the stem and branch nitrogen concentrations increase  
551 at a rate 30% lower, but biomass is three times faster (Wang et al., 2017; Xia and Wan, 2008).  
552 Beyond lacking this mechanism, CLM5.0 is also incapable of capturing the different C:N ratio  
553 sensitivities across different plant tissues. In contrast to empirical observations, CLM5.0 shows a  
554 very similar carbon and nitrogen increasing response across different tissues to both eCO<sub>2</sub> and  
555 iNDEP, thus the leaf C:N ratio sensitivity could be underestimate by this relatively fixed C and  
556 N allocation strategy (Fig S15, Fig S16).

557         In order to make the modeling results more consistent with leaf trait observations, we  
558 recommend improvements in three aspects: 1) more carbon and nitrogen concentration trait  
559 measurements within different canopy layers and across different tissues should be acquired.  
560 They will help reveal how different the C:N ratio response to environment change within canopy  
561 layers and across tissues, and the mechanism behind them; 2) the process-based LSMs should  
562 focus on explicit representation of canopy structure and vertical leaf trait dynamics, and the big  
563 leaf model with static canopy scaling parameters was improper compared with observations

564 (Rogers et al., 2017). Fortunately, there are efforts underway to address this model insufficiency  
565 (e.g.,(e.g., Fisher et al., 2018; Koven et al., 2019)). 3) Further studies and quantitative relations  
566 should be explored on the linkage between transpiration, hydraulic conductance and plant  
567 nitrogen concentration profile, and the mechanism about difference in carbon and nitrogen  
568 allocation strategies across plant tissues.

## 569 **5 Conclusion**

570 In this study, we employed both a data-driven model and a process-based model CLM5.0  
571 to map the leaf C:N ratio spatiotemporal variation in China in the past 5 decades. We found that  
572 leaf C:N ratio had a temporally significant but spatially uneven pattern of change. We found an  
573 increasing leaf C:N ratio for temperate grassland in northern China (about  $+5.3\% \pm 8.2\%$ ) and  
574 for boreal grassland in western China (about  $+3.3\% \pm 6.6\%$ ), by contrast a decrease for  
575 temperate forest in southern China (about  $-7.7\% \pm 6.6\%$ ) and for boreal forest in northeastern  
576 China (about  $-3.3\% \pm 5.5\%$ ). In addition, the structural equation models indicated that the  
577 temporal leaf C:N change was sensitive to  $\Delta\text{NDEP}$ ,  $\Delta\text{CO}_2$  and  $\Delta\text{MAT}$  rather than to  $\Delta\text{MAP}$  and  
578 ecosystem types. These relationship were supported by CLM5.0 scenario analysis, where  
579 nitrogen deposition was the dominant factor that led to leaf C:N ratio decreasing in southern  
580 China. Elevating  $\text{CO}_2$  is likely to increase leaf C:N ratio in low NDEP area. Climate change has  
581 stronger impact in boreal area than in temperate area with more net nitrogen mineralization  
582 increases in boreal area under climate warming.

583 Furthermore, the study also found the differences between leaf C:N spatial and temporal  
584 variation sensitivities, suggesting that the climate envelope theory was more suitable to predict  
585 long-term variation but had limitation to model trait-environment change in leaf C:N ratio at  
586 decade scale, and highlight the important role of nitrogen acquisition cost and soil nitrogen

587 availability in determining leaf C:N ratio. Finally, we inferred that there is significant impact of  
588 flexible leaf C:N ratio on plant and ecosystem carbon cycling, and it is important for models to  
589 incorporate plant and microbial interaction and their carbon cost during plant nitrogen uptake.  
590 The leaf C:N ratio response predicted by two models agree with field observation, but CLM5.0  
591 showed weaker magnitude. The poor performance of CLM5.0 is likely due to insufficient  
592 representation of within canopy leaf trait dynamics, linkage between transpiration, hydraulic  
593 conductance and plant nitrogen concentration profile, and differences in carbon and nitrogen  
594 allocation strategies across plant tissues in CLM5.0, all of which should be further explored in  
595 future studies.

596

## 597 **Acknowledgments**

598 In this research, M. Y. Sheng and D.W. Yang are supported by the National Natural Science  
599 Foundation of China (Project No. 42041004). J.Y. Tang is supported by the Director, Office of  
600 Science, Office of Biological and Environmental Research of the US Department of Energy  
601 under contract No. DE-AC02- 05CH11231 as part of RUBISCO project under the Regional and  
602 Global Climate Modeling (RGCM) Program. CLM5.0 was acquired from NCAR  
603 ([www.cesm.ucar.edu/models/clm/](http://www.cesm.ucar.edu/models/clm/)).

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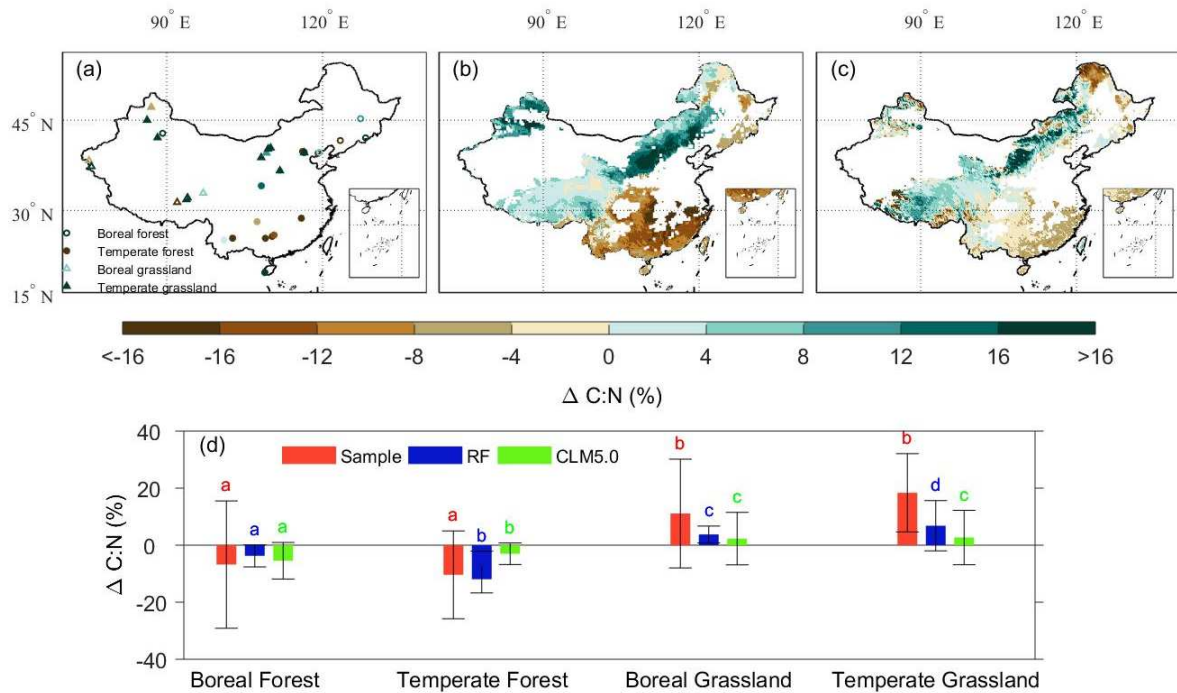
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3 **Figure.** Temporal variation of multi-annual mean leaf C:N ratio in China (the difference between  
 4 2 time period from 1960-1989 to 1990-2015); (a) result by plot samples, (b) result by random  
 5 forest (RF) model, (c) result by CLM5.0; (d) results from all above 3 approaches averaged by  
 6 PFT. Different lowercase letters in the same color on error bars indicate significant differences  
 7 across different PFTs at  $p < 0.05$ .