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Journal

New Phytologist, 231(6)

ISSN

0028-646X

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Publication Date

2021-09-01

DOI

10.1111/nph.17558

Peer reviewed

Eco-evolutionary optimality as a means to improve vegetation and land-surface models

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Summary

1 Global vegetation and land-surface models embody interdisciplinary scientific understanding of the
2 behaviour of plants and ecosystems, and have become indispensable to project the impacts of
3 environmental change on vegetation and the interaction between vegetation and climate. However,
4 systematic errors and persistently large differences among carbon and water cycle projections by
5 different models highlight the limitations of current process formulations and point to more
6 fundamental problems in how models are developed. In this review, focusing on core plant functions
7 in the terrestrial carbon and water cycles, we show how testable unifying hypotheses derived from
8 eco-evolutionary optimality principles can provide novel, parameter-sparse representations of plant
9 and vegetation processes. New, independently tested modules could profitably be integrated into
10 modelling frameworks that account for the multiple time scales on which plants and plant
11 communities adjust to environmental change.

12

13 **Key words:** eco-evolutionary optimality, global vegetation model, land-surface model, water and
14 carbon trade-offs, stomatal behaviour, leaf economics spectrum, acclimation, plant functional
15 ecology

16 **I. Introduction**

17 The ability of land ecosystems to deliver societal benefits – including the regulation of climate, the
18 carbon cycle and water and air quality, and the provisioning of goods including food and fibre – is at
19 risk because of current rates of global environmental change (Ostberg et al., 2018). Assessing and
20 mitigating this risk requires the reliable characterization of vegetation processes, including plant
21 demography, growth and competition as well as physical land-atmosphere interactions, at multiple
22 spatial and temporal scales. Highly developed, process-based computational models now exist that
23 operate across scales, simulating photosynthesis and stomatal regulation, carbon allocation,
24 competition for light, water and nutrients, community assembly, disturbance regimes, interactions of
25 vegetation with climate and atmospheric composition, and yields of essential products including
26 crops. The two main (overlapping) categories of current models are dynamic global vegetation
27 models (DGVMs) and land surface models (LSMs). LSMs are designed for embedding in climate
28 models and represent “fast” land-atmosphere exchanges explicitly, typically with half-hourly time-
29 steps. Some LSMs treat vegetation composition and structure as static; others simulate vegetation
30 dynamics as well and are therefore also DGVMs. On the other hand, some “offline” (i.e. not coupled
31 to a climate model) DGVMs represent fast land-atmosphere exchanges implicitly, using daily time-
32 steps. Earth System Models (ESMs) are climate models that use LSMs to represent the interactions
33 of the land and ocean surfaces with the state and composition of the atmosphere. LSMs – with or
34 without dynamic vegetation – provide the means for ESMs to represent the land-atmosphere interface,
35 including impacts of atmospheric CO₂ and climate change on vegetation and feedbacks from
36 vegetation changes on CO₂ and climate.

37 DGVMs and LSMs are based on explicit formulations of a set of processes, rather than on emergent
38 empirical relationships (in contrast, for example, with forest yield tables) and they are generic in
39 terms of plant types considered (in contrast, for example, with most crop models). Process-based
40 vegetation models have been extensively applied by the climate, integrated assessment and impacts
41 modelling communities to assess the nature and impacts of projected climate change, including the
42 role of biophysical and biogeochemical feedbacks. For example, an ensemble of global vegetation
43 models is included in the Global Carbon Project’s annual update on the state of the carbon cycle
44 (Friedlingstein et al., 2020), and is widely used to assess the role of vegetation in land-atmosphere
45 interactions, such as diagnosing causes of fluctuations in the atmospheric CO₂ growth rate (Keenan
46 et al., 2016). Vegetation models in general have a wide variety of uses, from quantifying the
47 magnitude of the positive climate-carbon cycle feedback and the negative CO₂ fertilization
48 feedback to climate (e.g. Cox et al., 2013), to investigating the impact of recent climate change on
49 the hydrological cycle (e.g. Ukkola and Prentice, 2013), and projecting the impact of future climate

50 change on crop production (Inter-Sectoral Impact Model Intercomparison Project, ISIMIP2b:
51 <https://www.isimip.org/protocol/#isimip2b/>; Ostberg et al., 2018).

52 Modelling vegetation as a fully embedded component of the climate system presents major scientific
53 and computational challenges (Fisher and Koven, 2020). The many successful applications of
54 vegetation models have drawn attention away from several known systematic failures, which have
55 emerged especially when models were used to reproduce large-scale phenomena encoded in
56 atmospheric measurements. For example, both ESMs and offline DGVMs failed to reproduce the full
57 magnitude of the amplification of the high-latitude seasonal cycle of atmospheric CO₂ over the past
58 half-century (Graven et al., 2013; Thomas et al., 2016). DGVMs also failed to reproduce the observed
59 relationship between atmospheric ¹³CO₂ content and global land-atmosphere carbon exchange (Peters
60 et al., 2018). Models disagree even about the sign of the effect of global warming on primary
61 production (Ciais et al., 2013). There are large uncertainties in the modelled response of vegetation
62 to precipitation changes (Huntzinger et al., 2017) and little agreement in the simulated response to
63 CO₂ and the role of nutrient availability in modulating this response (Wieder et al., 2015). Large
64 differences in the modelled behaviour of global vegetation, which have persisted for more than two
65 decades (VEMAP 1995; Friedlingstein et al. 2006), were identified as a serious concern in the IPCC
66 Fifth Assessment Report (Ciais et al., 2013). Developments since then have not alleviated this
67 concern (Arora et al., 2020).

68 These problems suggest a need to re-assess the assumptions and processes included in current
69 vegetation models, and the modalities by which they are developed. The explosion over the last 20
70 years in the amount and variety of data available – including plant trait databases, field campaigns,
71 flux measurements, ecosystem experiments and satellite remote sensing – should greatly facilitate the
72 process of developing better models. Indeed, meta-analyses of different types of observation have
73 provided insights into universal patterns which can be used for testing general patterns in simulated
74 ecosystem responses to various drivers (Wieder et al., 2019). Finding ways to test alternative
75 hypotheses using observations will also require moving beyond meta-analysis. Large-scale field
76 experiments provide under-utilized opportunities for model evaluation (Medlyn et al., 2015), while
77 controlled-environment experiments are irreplaceable for testing general hypotheses about plant
78 function. Controlled-environment experiments could, for example, help to resolve current
79 disagreements about the impact of changes in CO₂ or nutrients on photosynthetic traits.

80 Progress, however, also requires coherent, well-motivated hypotheses to test. Franklin et al. (2020)
81 highlighted recent developments that hold promise for improving vegetation models by generating
82 such hypotheses. The concept of eco-evolutionary optimality (EEO) is one of these developments.

83 EEO invokes the power of natural selection to eliminate uncompetitive trait combinations, and
84 thereby shapes predictable, general patterns in vegetation structure and composition. The term ‘eco-
85 evolutionary’ expresses the fact that organisms adjust to their environment on both shorter
86 (ecological) timescales and longer (evolutionary) timescales. EEO hypotheses are based on
87 identifying trade-offs that organisms are required to make, for example in land plants between CO₂
88 uptake and water loss, and expressing these mathematically. At the core of modelling EEO are
89 therefore the mechanistic links between plant functional traits, their implications for resource demand
90 and acquisition and biogeochemical cycling, and their effect on the plant’s competitiveness.
91 Mechanistic ecosystem models (LSMs, DGVMs, ESMs) are suited to resolve these links and thus
92 provide a useful framework for investigating how EEO shapes global vegetation function and
93 climate-land feedbacks in the Earth system. (EEO approaches are distinct from parameter
94 optimization methods, which are widely used in the context of existing models.) EEO hypotheses
95 have shown a notable ability to predict observed patterns, providing parsimonious explanations of
96 observations at the leaf (e.g. Smith et al., 2019; H. Wang et al., 2020), plant (e.g. Farrior et al., 2013;
97 Lavergne et al., 2020a) and vegetation (e.g. Franklin et al., 2014; Baskaran et al., 2017) levels.
98 However, there is no recipe to generate a “correct” EEO criterion. EEO formulations have to be
99 assessed against data, like hypotheses in all fields of science.

100 Many modelling groups are exploring the use of EEO hypotheses to improve the representation of
101 specific processes in vegetation models (e.g. Bonan et al., 2014; De Kauwe et al., 2015; Lin et al.,
102 2015; Ali et al., 2016; Xia et al., 2017; Lawrence et al., 2019). In this review, we aim to raise
103 awareness of the broader potential for a hypothesis-testing approach based on EEO to underpin a
104 more far-reaching improvement in the robustness and reliability (*sensu* Prentice et al., 2015) of
105 vegetation models. Section II provides a perspective on the shortcomings of current models and
106 model development practice, informed by our collective experience. Section III introduces case
107 studies that exemplify how EEO can improve (and often, simplify) formulations of core processes at
108 the leaf level that are required by both LSMs and DGVMs. Section IV considers the scope for
109 applying EEO at the whole-plant and plant community levels. Section V deals with limitations of the
110 EEO concept. Section VI briefly considers the outlook for next-generation vegetation models
111 incorporating EEO principles.

112 **II. Model development: problems and solutions**

113 The origins and historical development of global vegetation models have been reviewed by Prentice
114 et al. (2007), Quillet et al. (2010), Prentice & Cowling (2013) and Fisher et al. (2014); this material
115 will not be revisited here. Current models have much in common. Processes are differentiated by

116 operational time-steps: canopy-atmosphere energy exchanges and photosynthesis are modelled in
117 LSMs typically at half-hourly time-steps; phenology, carbon allocation and growth at time-steps of
118 days to months; and vegetation dynamics and disturbance in DGVMs at time-steps of months to
119 years. The diversity of plants is still treated in many models as a limited set of PFTs, each
120 characterized by a distinct set of attributes. However, an important trend over the past decade has
121 been towards the replacement of PFTs by quantitatively varying functional traits. It is now widely
122 recognized that the assignment of fixed sets of traits per PFT is too restrictive, because (a) for most
123 quantitative plant traits, variation is greater within than between PFTs (Kattge et al., 2020), and (b) a
124 substantial fraction of the observed variation in community-mean trait values along environmental
125 gradients is linked to acclimation and adaptation within species and PFTs (Siefert et al., 2015).
126 Progress is being made towards replacing fixed PFTs in models with quantitative traits that vary
127 dynamically, mimicking acclimation and/or adaptation processes (van Bodegom et al., 2012) and
128 more realistically portraying ecosystem carbon uptake (Verheijen et al., 2015) and the dynamic
129 response of terrestrial ecosystems to climate change (Sakschewski et al., 2015).

130 A number of problems (see Box) however are slowing progress. Some of these may have arisen
131 because of the general institutional separation of model development from empirical science, and the
132 accretional nature of much of this development. Some are practical issues about the way model codes
133 are written, updated and tested. All could be mitigated by adopting different practices (see Box).
134 Several initiatives have promoted systematic data-model comparison (“benchmarking”) of land
135 models as a partial remedy for these problems (<https://www.ilamb.org/>; Collier et al., 2018). Some
136 proposed benchmark data sets – e.g. for upscaled gross primary production (Tramontana et al., 2016)
137 and plant litter properties (Pettinari and Chuvieco, 2016) – however are themselves modelled outputs,
138 which limits their usefulness. Wieder et al. (2019) drew attention to the limitations of benchmarking,
139 which (we suggest) should be considered as a necessary but by no means sufficient part of model
140 evaluation. Wieder et al. (2019) described the recent history of a leading LSM, the Community Land
141 Model (CLM), in particular its evolution from version 4 (a pioneering attempt to include the coupling
142 of ecosystem carbon and nitrogen cycles, which however greatly underestimated global land carbon
143 uptake and its sensitivity to climate) to the better-performing version 5. They made a strong case for
144 the need to use observational and experimental information as an integral part of model development
145 and improvement. Although experimental and observational studies in this field are often justified by
146 the need to improve vegetation models, the pathway by which this new knowledge is transferred is
147 indirect at best. In principle, experiments could be performed precisely to clarify how individual
148 processes should be represented in models. This is still far from being standard practice.

149 A further consequence of the current model development paradigm is the ‘complexity trap’ (Franklin
150 et al. 2020). Many recently published model “improvements” are achieved by adding complexity
151 (Fisher and Koven, 2020), but it is generally understood that this does not equate to increased realism
152 – particularly as the incorporation of new processes often increases further the number of poorly
153 known parameters that need to be specified. Moreover, developing models by accretion has inevitably
154 led to a decline in transparency (Prentice et al., 2015). In other areas of environmental modelling,
155 including climate modelling, there has been a growing realization that re-examination of basic
156 processes, reduction of complexity, and increased transparency are all necessary for progress (Held,
157 2005; Gramelsberger et al., 2020). Vegetation modelling is no exception.

158 Different Earth subsystems however pose specific challenges. The key challenge for global-scale
159 modelling of biological systems is to identify principles applicable across diverse and
160 phylogenetically distinct assemblages (Franklin et al., 2020). EEO has a key role to play because it
161 can generate coherent, testable hypotheses about plant and vegetation function that transcend
162 differences among biomes and floras. In the following section, we summarize a number of case
163 studies that demonstrate this capability, where EEO approaches have provided parsimonious
164 representations of core, leaf-level processes that are individually testable and supported by evidence.
165 Such representations are candidates to replace the (often more complex and less well-tested)
166 formulations currently in use; however, much research remains to be done to establish the most
167 effective representation for each process. Case studies are presented roughly in descending order of
168 readiness – from photosynthesis and primary production, where a proof-of-concept for
169 implementation in a LSM framework exists, to the leaf economics spectrum, which requires a novel
170 approach to account for how phylogeny and environment co-determine plant traits.

171 **III. Leaf-level optimality**

172 *Photosynthesis and primary production.* Nearly all LSMs and DGVMs simulate photosynthesis using
173 the FvCB model (Farquhar et al., 1980) or the modification proposed by Collatz et al. (1991).
174 Implementing the FvCB model requires three parameters: the maximum carboxylation rate (V_{cmax}),
175 which determines the enzymatic capacity for carbon fixation; the maximum electron-transport rate
176 (J_{max}), which determines the capacity to generate the required reducing power; and the ratio of leaf-
177 internal to ambient CO_2 ($c_i:c_a$, here denoted as χ), which relates the assimilation rate to stomatal
178 conductance. Improved understanding of how V_{cmax} and J_{max} vary with environmental conditions
179 should provide a more rigorous basis for modelling photosynthesis and primary production (Rogers
180 et al., 2017). Light use efficiency (LUE) models – widely used in remote-sensing applications –
181 simulate primary production using empirical response functions; this can limit the number of

182 parameters to be estimated but severs the link to processes. These two approaches can be reconciled
183 by applying two EEO hypotheses: the “least-cost hypothesis” that the combined costs of maintaining
184 water loss and carbon uptake pathways are minimized, allowing prediction of stomatal behaviour
185 (Dewar et al., 2018); and the “coordination hypothesis” that V_{cmax} and J_{max} acclimate to growth
186 conditions, so that the capacity of the leaf to fix carbon remains close to the level required to use the
187 available light (Field and Mooney, 1986; Chen et al., 1993; Maire et al., 2012; Quebbeman and
188 Ramirez, 2016; H. Wang et al., 2017). The result is a model applicable to all C_3 plants that has the
189 convenient mathematical form of a LUE model (H. Wang et al., 2017). It has only two new
190 parameters, both of which can be estimated from independent data sets. It yields good agreement with
191 gross primary production (GPP) data derived from eddy-covariance flux measurements (Stocker et
192 al., 2020). It also predicts a number of related physiological characteristics correctly, including the
193 global pattern of V_{cmax} in relation to light, temperature and vapour pressure deficit (VPD) (Smith et
194 al., 2019), elevational trends in photosynthetic traits and primary production (Peng et al., 2020), and
195 the response of V_{cmax} to atmospheric CO_2 . Specifically, the model predicts a decline in V_{cmax} with
196 increasing ambient CO_2 (H. Wang et al., 2017), and a steeper increase with decreasing ambient CO_2 .
197 Both have been verified experimentally (Figure 1).

198 Neglecting the adaptive adjustment of V_{cmax} to growth conditions could result in simulated PFTs
199 becoming (unrealistically) maladapted to environmental changes, and (if set too low) unrealistically
200 steep responses of photosynthesis to temperature and ambient CO_2 . The model for GPP outlined
201 above (H. Wang et al., 2017) provides an example of how EEO hypothesis formulation and testing
202 can lead not only to a more realistic representation of a key process, in the sense of being well
203 supported by observational and experimental data, but also to a less complex one. Compared to
204 conventional models, the number of parameters required as input has been dramatically reduced, by
205 two mechanisms. First, the adaptive adjustment of key quantities (V_{cmax} , J_{max} , and a variable ξ that
206 determines the response of χ to VPD) eliminates the need for these to be prescribed. Second, it
207 removes the need to provide lists of parameter values for PFTs.

208 The model is not suitable for immediate incorporation into a LSM because it works on multi-day
209 timesteps (i.e. at the time scale of leaf- and canopy-level acclimation). However, the fast responses
210 of photosynthesis and stomatal conductance to environmental variations are already well
211 characterized (Farquhar et al., 1980; Medlyn et al., 2011). All that is needed, then, is to replace fixed
212 values of V_{cmax} and J_{max} at 25°C ($V_{\text{cmax}25}$, $J_{\text{max}25}$) and ξ with slowly time-varying values that follow
213 the EEO criteria. This is straightforward in principle, and allows the same model to reproduce
214 measured daily cycles of GPP in different biomes with unchanged parameter values (Figure 2;
215 Mengoli et al., submitted). Moreover, whereas adding a new process (acclimation of photosynthetic

216 parameters) might be expected to increase model complexity, this example shows the opposite:
217 treating acclimation as an EEO process allows GPP to be represented more parsimoniously than
218 would otherwise have been possible.

219 *Dark respiration.* Leaf mitochondrial respiration supports many metabolic processes, including the
220 protein turnover required to maintain photosynthetic capacity. Leaf dark respiration (R_{dark}) is a widely
221 measured quantity. Its instantaneous temperature response is commonly represented by the Arrhenius
222 equation (Atkin et al., 2017). Many models assume that R_{dark} at 25°C ($R_{\text{dark}25}$) is proportional to
223 $V_{\text{cmax}25}$, treat this as a constant per PFT, and model the temperature-dependence of both R_{dark} and
224 V_{cmax} with Arrhenius equations. However, there is considerable spatial and temporal variability in
225 R_{dark} within PFTs as a function of environmental conditions (Atkin et al., 2015; Smith and Dukes,
226 2018). In a global analysis, H. Wang et al. (2020) showed that values of R_{dark} at current growth
227 temperature are optimized according to the need to ensure that its metabolic functions are coordinated
228 with V_{cmax} . This hypothesis predicts that acclimated values of both R_{dark} and V_{cmax} increase with
229 growth temperature – but less steeply than their instantaneous responses to temperature, which are
230 fully compensated by acclimation. These predictions are well supported by data; the conventional
231 modelling approach is not (Figure 3).

232 Neglecting the acclimation of leaf-level respiration is likely a major source of uncertainty in model
233 predictions, with serious consequences for the estimation of land carbon uptake especially in warmer
234 climates (Huntingford et al., 2017). As with photosynthesis, there is no obstacle in principle to
235 including leaf-level respiratory acclimation in DGVM or LSM frameworks. To do so requires only
236 that $R_{\text{dark}25}$ vary along with (slowly varying) $V_{\text{cmax}25}$ following H. Wang et al.’s (2020) EEO
237 hypothesis, while the fast environmental responses of R_{dark} and V_{cmax} are represented as in current
238 models (or better, for R_{dark} , via the universal temperature response reported by Heskell et al., 2016).
239 Such a scheme has not been implemented yet, as far as we are aware, in any vegetation model.

240 *Stomatal behaviour and transpiration.* Plants regulate water and energy exchanges with the
241 atmosphere by adjustment of stomatal conductance (g_s). Most current models represent g_s based on
242 the fast, experimentally observed response to VPD (Damour et al., 2010). More mechanistic models
243 have been developed (e.g. Sperry et al., 2017), but require new parameters (Drake et al., 2017). EEO
244 hypotheses, based on the trade-off between maximizing carbon gain and minimizing water loss,
245 potentially offer parsimonious solutions. One approach (Medlyn et al., 2011) is based on an
246 approximate solution to the ‘CF’ hypothesis of constant marginal water use efficiency, originally
247 proposed by Cowan and Farquhar (1977). This solution correctly predicts stomatal responses to
248 changing CO_2 and variability across environmental gradients (Medlyn et al., 2011; Medlyn et al.,

249 2013; Lin et al., 2015). It has been included in land-surface models (e.g. De Kauwe et al., 2015;
250 Oliver et al., 2018) and shown to perform as well as the empirical relationships originally used. These
251 approaches all require calibrating one ‘free’ parameter per PFT in the optimal stomatal conductance
252 formulation. To achieve a parameter-free formulation, it is possible to re-frame the CF hypothesis by
253 accounting for soil moisture dynamics in the optimization problem (Manzoni et al. 2015), but this
254 implies an assumption on how much soil water can be used by plants.

255 A different approach (Prentice et al., 2014), used in the model of H. Wang et al. (2017), is based on
256 the EEO hypothesis that leaves minimize the sum of the maintenance costs (per unit assimilation) of
257 transpiration and carboxylation. Carboxylation costs are envisaged as the respiration required for
258 Rubisco turnover, while transpiration costs are envisaged as the respiration required to maintain living
259 sapwood. This hypothesis leads to a mathematical formulation of the fast response of stomata to VPD
260 that is similar to that of Medlyn et al. (2011), while also correctly predicting the environmental
261 dependencies of the control parameter (ξ) on temperature (Lin et al., 2015) and atmospheric pressure
262 (Körner and Diemer, 1987). Each of these predicted partial dependencies of χ on environmental
263 variables can be observed in stable carbon isotope ratio ($\delta^{13}\text{C}$) measurements on leaves (H. Wang et
264 al., 2017). Lavergne et al. (2020a; Figure 4) showed they are also present in tree-ring $\delta^{13}\text{C}$
265 measurements. By providing time-series, Lavergne et al. (2020a) moreover confirmed the (weak)
266 dependency of χ on atmospheric CO_2 (Figure 4) that is predicted by this EEO hypothesis. Apart from
267 the well-known VPD response, none of these dependencies is reflected in standard LSMs, except
268 crudely, through the assignment of distinct parameter values to PFTs that occupy different climates.

269 Further alternative EEO approaches (e.g. Wolf et al., 2016; Anderegg et al., 2018; Eller et al., 2018;
270 Venturas et al., 2018; Trugman et al., 2019; Deans et al., 2020; Eller et al., 2020; Sabot et al., 2020)
271 incorporate hydraulic costs, based on the hypothesis that the short-term and long-term costs of
272 transpiration at low soil water potential contribute to the total cost of maintaining the water transport
273 pathway. The Wolf et al. (2016) formulation has been shown to be in good agreement with
274 experimental evidence for changes in stomatal conductance in response to daily and seasonal changes
275 of environmental conditions, and to perform much better than the CF model in predicting stomatal
276 responses to dry soils (Anderegg et al., 2018). Y. Wang et al. (2020) found that among eleven
277 optimality-based stomatal models, the most skilled were those taking into account the cost of stress-
278 induced hydraulic failure.

279 The development of EEO hypotheses for stomatal behaviour is a highly active research field. The
280 hypothesis of Prentice et al. (2014) provided an equation with good predictive power for the responses
281 of leaf and plant $\delta^{13}\text{C}$ to the growth environment, but its single “universal” parameter has been shown

282 to be influenced by soil moisture (Lavergne et al., 2020b) and soil pH (H. Wang et al., 2017; Paillassa
283 et al., 2020). Moreover, the variation of χ on long climatic moisture gradients appears to be
284 significantly steeper than predicted by that equation (Dong et al., 2020). Allowing variation of ξ
285 following the EEO criterion of Prentice et al. (2014) would allow stomatal acclimation to changes in
286 growth temperature (Marchin et al., 2016). This cannot happen in current models, because their
287 responses to VPD are pre-determined by PFT. However, further research is needed to determine how
288 soil influences might best be included in models. Solutions are likely to include EEO approaches to
289 explain the coordination of hydraulic and photosynthetic traits (Brodribb, 2009; Joshi et al., 2020),
290 and the influence of soil water and fertility factors on water and nutrient acquisition costs (Paillassa
291 et al., 2020).

292 *Isoprene emission.* Plant emissions of the volatile organic compound (VOC) isoprene protect
293 photosynthetic function against damage due to reactive oxygen species (ROS), which are produced
294 in leaves at high temperatures (Niinemets, 2010; Harrison et al., 2013; Lantz et al., 2019). As a result,
295 tree species that produce isoprene are competitively favoured under hot and dry conditions (Taylor
296 et al., 2018). Modelling of plant VOC emissions is important in ESMs, because these reactive
297 compounds have a significant impact on atmospheric chemistry. Many ESMs rely on a complex
298 empirical model (Guenther et al., 2006) to predict VOC emissions. More explicitly process-based
299 models of VOC emission have been devised (e.g. Pacifico et al., 2011) but still require several, poorly
300 known parameters to be specified.

301 However, the responses of isoprene emission to light, temperature and CO₂ are consistent with a
302 much simpler relationship: a linear dependency on the leaf's "energetic status", which is the
303 difference between photosynthetic electron transport and the electron requirement to support the
304 current rate of carbon fixation (Morfopoulos et al., 2013). This relationship reproduces the shapes of
305 observed responses of isoprene emission to environment, including its non-linear response to light
306 (Figure 5); its (otherwise enigmatic) decline with instantaneous increases in ambient CO₂
307 (Morfopoulos et al., 2013, 2014); and its recovery over time at high CO₂ (Sun et al., 2013). Without
308 this recovery, isoprene-emitters under high CO₂ would lose the thermo-protective benefits of isoprene
309 emission – an unlikely outcome in evolutionary terms.

310 The leaf energetic status model has the potential to simplify the representation of VOC emission in
311 ESMs, while increasing its predictive power for conditions outside those currently observed.
312 However, a key unanswered question remains, regarding the "base rate" of isoprene emission at the
313 plant-community level. The current standard approach relies (unsatisfactorily) on a fixed geographic

314 field of emission capacity. Explicit hypotheses to predict VOC emission capacity remain to be
315 formulated and tested.

316 *Leaf economics*. Leaf mass per unit area (LMA) determines how much leaf area can be produced for
317 a given total aboveground carbon allocation (Cui et al., 2019). The leaf economics spectrum (Wright
318 et al., 2004) relates LMA and leaf lifespan (LL) across all PFTs. In nature, this relationship is not
319 fixed, however, and varies with climate. Kikuzawa (1991) hypothesized that LL maximizes leaf
320 carbon gain over the lifetime of the leaf, accounting for (amortized) leaf construction costs. This EEO
321 hypothesis has recently been combined with two others. Xu et al. (2017) provided empirical support
322 for the hypothesis that the leaf ageing rate (a parameter in Kikuzawa's model) is inversely
323 proportional to LMA, and also to $V_{\text{cmax}25}$; while the coordination hypothesis, described above, predicts
324 optimal values of $V_{\text{cmax}25}$. Combining these three hypotheses leads to a theoretical prediction of the
325 leaf economics spectrum, and how it varies across environments. For winter-deciduous woody plants
326 where LL is constrained by the length of the growing season, this theory leads to a prediction of LMA
327 that is consistent with observations along an elevational and aridity gradient (H. Xu et al., 2020). For
328 evergreen plants it leads to a correct global prediction of the proportionality between LMA and LL,
329 and how this is modified by growing-season length and light (Figure 6). A changing climate will
330 inevitably alter the competitive balance among species with different LMA and LL, in ways that
331 fixed-PFT LMA schemes cannot capture.

332 In addition to affecting leaf lifespan, LMA mechanistically affects stomatal responses (Buckley et al.,
333 2015). Increasing LMA reduces the conductivity of the outside-xylem water pathway due to increased
334 path-length, and therefore causes highly negative water potentials near the stomata. This in turn may
335 necessitate a greater investment in leaf hydraulics in high-LMA species. Without such investment,
336 these species would be uncompetitive due to reduced photosynthesis rates. Thus, EEO suggests a
337 testable linkage between physiological and hydraulic traits.

338 The EEO basis for the leaf economics spectrum has not been incorporated in any vegetation model,
339 and its consequences for climate-change impacts are largely uncharted. LMA and LL, as
340 structural/morphological traits, differ from the physiological traits discussed above in showing far
341 less plasticity (or genotypic adaptation) along environmental gradients (Dong et al., 2017, 2020).
342 Therefore, their representation in models calls for a different treatment, as any change in community-
343 mean LMA and LL will depend more on species replacement (a slower process) than on physiological
344 adjustment. In addition, whereas the theory summarized above predicts environmentally conditioned
345 changes in the *relationship* between LMA and LL, it does not predict anything about their mean
346 values. Phylogenetic determinism may be helpful here. Starting from the observed global

347 distributions of these traits, it is possible to calculate how these intersect with the predicted LMA-LL
348 relationship. This will however require representing traits in models as probability distributions rather
349 than as single values.

350 **IV. Beyond the leaf level**

351 Applications of EEO concepts in vegetation modelling have mainly focused on leaf-level processes,
352 because of the short timescale of some leaf-level responses to environmental conditions. The EEO
353 framework however extends naturally to the whole-plant level, providing insightful approaches to
354 modelling processes including phenological timing (Caldararu et al., 2014; Manzoni et al., 2015) and
355 strategy, and carbon allocation to leaves, stems and roots (e.g. Valentine and Mäkelä, 2012; Kvakic
356 et al., 2020). We summarize some cases below.

357 *Carbon allocation.* Both field and experimental data show that allocation to roots increases when
358 nutrient supply is limiting, for example on infertile soils or in cold climates (Reich et al., 2014; Gill
359 and Finzi, 2016). This observation is consistent with the long-established EEO hypothesis (Rastetter
360 and Shaver, 1992; Thomas and Williams, 2014; Rastetter and Kwiatkowski, 2020) that plants,
361 requiring multiple resources, allocate effort optimally so that no one resource is limiting to growth.
362 A plant-level allocation model based on this assumption has been used to explain the contrasting
363 effects of elevated CO₂ on tree growth and nitrogen uptake and their dependence on soil nitrogen
364 availability (Franklin et al., 2009; Figure 7). An EEO approach to carbon allocation has been adopted
365 in at least one LSM (Xia et al., 2017).

366 *Soil-plant interactions and mycorrhizae.* Carbon exchanges between plants and their symbionts can
367 also be described using EEO principles. The effects of nutrient limitation are predictable based on
368 the carbon costs of nitrogen uptake via different symbionts (Terrer et al., 2018), which may in turn
369 depend predictably on soil nutrient availability (Franklin et al., 2014). The fraction of GPP allocated
370 to sustain symbionts thus becomes an outcome, rather than being imposed as an additional parameter
371 (Baskaran et al., 2017). Modelling soil-plant interactions explicitly in terms of the carbon cost of
372 nitrogen acquisition has a significant impact on modelled primary production (Brzostek et al., 2014)
373 and has been adopted in at least one LSM (Shi et al., 2016). Dynamically linking plants and microbes
374 in a terrestrial biosphere model has been shown to improve predicted carbon and nitrogen dynamics
375 across a gradient of vegetation stands varying in the abundance of trees with distinct (arbuscular and
376 ectomycorrhizal) types of mycorrhizal interaction (Sulman et al., 2017).

377 *Demography and species replacement.* Community-mean trait values are to some extent an outcome
378 of acclimation by individual plants, but also of adaptation: environmental filtering ensures that
379 genotypes and species are present only in environments that fall within their acclimation capacity.

380 The extent to which the observed variability in plant functional traits is due to phenotypic plasticity
381 (individual acclimation) or to non-plastic genotypic differentiation and species replacement (Meng et
382 al., 2015; Yang et al, 2018; Dong et al., 2017, 2020) has received little attention in the modelling
383 literature, yet it is essential information for understanding how community function and composition
384 react to rapid environmental changes. Plastic traits, such as photosynthetic capacity (Togashi et al.,
385 2018a) and the temperature optimum of photosynthesis (Kumarathunge et al., 2019; Vico et al.,
386 2019), acclimate quickly (days to weeks) within individual leaves; while other leaf traits, including
387 LMA, show only partial within-species adjustment to changes along environmental gradients (Dong
388 et al., 2017, 2020). Hydraulic traits, particularly leaf hydraulics linked to LMA and wood properties
389 in general, are also expected to show limited plasticity. Adaptive changes in the mean abundance of
390 non-plastic traits can only occur through the slower processes of community dynamics, which depend
391 on demography and competition among species. How best to represent these processes is a highly
392 active field of DGVM research (e.g. Reichstein et al., 2014; Sakschewski et al., 2015; Fisher et al.,
393 2018; Longo et al., 2019; Argles et al., 2020).

394 *Competition and coexistence.* There is a potential for optimality-models to represent not only single
395 optimal plant strategies but also coexisting strategies, which determine how resources are partitioned
396 as determined by diversity in plant function. Examples include the coexistence of different strategies
397 for coping with water shortage in dry environments (Lindh et al., 2014), and the complementarity of
398 alternative life-history strategies generating within-site heterogeneity and corresponding variation in
399 optimal strategies in resource-rich communities (Togashi et al., 2018b). Falster et al. (2017)
400 demonstrated the evolution of stable coexistence of tree species in a height-structured competition
401 model related to the demographic schemes used in DGVMs. This work offers the prospect of a wider
402 field of application for EEO-based modelling to address the origins and maintenance of species
403 diversity.

404 **V. Outstanding issues**

405 It should be abundantly clear from the discussions above that EEO, despite its utility, is not a “magic
406 bullet” that can instantly resolve problems in LSM and DGVM development. We suggest instead that
407 EEO concepts should underpin a research programme in which explicit, quantitative hypotheses play
408 a central role in data analysis and experimental design, while also providing parsimonious
409 formulations for modelling. Practitioners of this integrative approach need to be aware of the
410 limitations of EEO as well as its strengths. Some of the most important limitations are summarized
411 below.

412 *Natural selection acts on reproductive fitness, not on plant growth.* However, reproductive fitness is
413 practically impossible to measure in plants in the field. EEO hypotheses can be formulated in terms
414 of vegetative properties but it must be recognized that these are only indirectly linked to fitness. The
415 underlying assumption is that ineffective or uncompetitive trait combinations will confer low fitness
416 and be selected against.

417 It follows from the above that *no EEO hypothesis is unique.* For every trade-off considered, there are
418 alternative criteria all of which might appear to be compatible with EEO, but which make different
419 predictions. (Some examples have been discussed above.) Only empirical tests can determine which,
420 if any, of a series of alternative EEO hypotheses makes realistic predictions.

421 *The limits to optimality are a priori unknown* and can only be assessed empirically. Recent EEO
422 approaches to photosynthetic optimality illustrate pragmatic choices that have to be made in the
423 interests of parsimony. For example, it has been assumed that certain photosynthetic traits can show
424 unlimited variation, while others – such as the Michaelis constants and specificity of Rubisco, the
425 intrinsic quantum yield of electron transport, and their temperature dependencies – are constants.
426 These assumptions are supported by observations but only as an approximation; all of these properties
427 do in fact vary among plants (Ehleringer and Piercy, 1983; Dreyer et al., 2001; Singaas et al., 2001;
428 Galmés et al., 2015; Galmés et al., 2016), even if their variation is relatively modest.

429 *Optimality is approached at different rates by different processes.* In a realistically time-varying
430 environment optimality is approached rather than achieved, because the optimum is a moving target.
431 We have distinguished the fast (minutes to hours) time scales of enzyme kinetics and stomatal
432 responses to VPD from the slower time scales of leaf-level physiological acclimation (days to
433 months), carbon allocation (months to years) and species replacement (years to decades). DGVMs
434 respect these distinctions. However, the shift from a PFT to a trait basis for modelling necessitates
435 ensuring that trait shifts dependent on species replacement take place on the appropriate time scale,
436 which can be different for different traits. One method to achieve this for non-plastic traits, without
437 enumeration of actual species, is by sampling from a multidimensional trait space (Fyllas et al., 2014).

438 *The problem of absent species.* A harder issue related to time scales is how to represent dispersal and
439 migration (in other words, species replacement when the best-adapted species are not locally present)
440 in DGVMs. Attempts to do so are not reviewed here. However, we note that existing approaches will
441 remain speculative unless it can be shown that the rates of migration they predict are consistent with
442 observed rates of species replacement in response to rapid climate changes in the past (e.g. Harrison
443 and Sanchez Goñi, 2010).

444 *The importance of experiments.* Comprehensive testing of EEO hypotheses cannot rely entirely on
445 meta-analysis. Data from direct environmental manipulations are not hampered by correlations
446 between environmental variables and can therefore be used to quantify the time-scales of responses
447 (Kumarathunge et al., 2019). Controlled-environment greenhouse experiments have been used to
448 determine the rates and mechanisms underlying acclimation of photosynthetic (e.g. Scafaro et al.,
449 2017), hydraulic (e.g. Locke et al., 2013) and leaf-biochemical parameters (e.g. Dongsansuk et al.,
450 2013) to changes in the growth environment. Field experiments can scale individual to ecosystem-
451 level responses, through direct manipulations (e.g. Hoepfner and Dukes, 2012; Hovenden et al.,
452 2019), exclusion experiments (e.g. Inoue et al., 2017; Tomasella et al., 2018) or opportunistic
453 sampling strategies (e.g. Lusk et al., 2018). The increasing coordination of field experiments,
454 including experiments to examine the impacts of manipulating nutrient (e.g. NutNet; Borer et al.,
455 2013) or water supplies (e.g. DROUGHT-NET: Knapp et al., 2017), provides opportunities to
456 evaluate the role of different plant strategies for coping with environmental stresses, and such
457 networks provide key targets for model evaluation (e.g. Hilton et al., 2019).

458 *Disturbance and land use.* Theory regarding natural disturbances in vegetation is rudimentary.
459 Models for the interaction of wildfire regimes with vegetation and climate have been developed, but
460 their performance is relatively poor beyond the largest-scale geographic patterns (Forkel et al., 2019;
461 Hantson et al., 2020), and they struggle to represent the modification of natural processes by land
462 management and human settlement patterns. Even so, there is scope for EEO concepts to inform
463 research and ultimately improve models. For example, plants have evolved specific adaptations to
464 different frequencies and intensities of fire (Clarke et al., 2013; Pausas et al., 2016; Pausas, 2019)
465 which must be represented in vegetation models if they are to project the consequences of
466 environmental changes in fire-prone regions realistically. Simulating the land biosphere under direct
467 human intervention (including agriculture, pastoralism and forestry) presents further challenges.
468 Much research attention has focused on the data required to impose land-use history on vegetation
469 models (e.g. Pongratz et al., 2008; Klein Goldewijk et al., 2017). More could usefully be focused on
470 the function of modified ecosystems. Arable crops represent an extreme modification of the
471 landscape, yet crop growth conforms to the same principles as all plants and can be modelled with
472 the same EEO-based tools (Qiao et al., 2020).

473 **VI. Concluding remarks**

474 Vegetation models have shown their usefulness for projecting ecosystem productivity, vegetation
475 patterns, terrestrial carbon uptake and other ecosystem services in a rapidly changing world. These
476 projections now feed routinely into global assessments such as those being made by the IPCC, the

477 Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES) and the
478 Integrated Research on Disaster Risk project. Thus, they contribute to the evidence base necessary to
479 assess progress towards the United Nations Sustainable Development Goals. However, more reliable
480 models are required to enhance confidence in the plausibility of many of these projections. The rates
481 of expected global environmental change increases the need for models to be able to deal with
482 dynamic processes, including plant migration, adaptation, acclimation and land-use change. Global
483 changes are occurring faster than many adaptive processes and are likely to result in novel
484 environmental conditions; models must therefore be equipped to deal with non-equilibrium situations
485 and novel conditions outside the range for which they were originally developed and tested. This can
486 only be achieved by ensuring that they realistically account for acclimation and adaptation processes
487 and do not entirely rely on statistically determined, historical patterns. However, increased realism is
488 of little value if it is accompanied by over-parameterization and ever-increasing parameter
489 uncertainty. We have indicated how EEO theory can provide a means to alleviate these problems. It
490 requires clear formulation of alternative hypotheses, which in turn creates a central role for
491 observations and experiments to test and compare them.

492 There is as yet no comprehensive description of plant behaviour in terms of EEO – indeed, as some
493 of the examples above have shown, the appropriate choice of optimality criteria is an active research
494 topic in areas such as stomatal behaviour while in other areas, such as disturbance dynamics, EEO
495 concepts are in their infancy. Moreover, there is no recipe for success. EEO hypotheses must be
496 tested, and many will fail. EEO approaches are nonetheless providing robust, parsimonious and well-
497 supported representations of core processes that are represented in all vegetation models, and offer
498 promise for the development of a new generation of models.

499

500 **Acknowledgements.** We gratefully acknowledge the contribution of participants at the workshop
501 “Next generation vegetation modelling”, held at IIASA in March 2017. The idea for this review arose
502 from the insights and excitement engendered by these discussions. We thank IIASA, both for their
503 financial support of the workshop, and for continued support thereafter. We particularly thank the
504 previous Director General and CEO of IIASA, Pavel Kabat, for his support for the next-generation
505 vegetation modelling initiative. SPH acknowledges support from the ERC-funded project GC2.0
506 (Global Change 2.0: Unlocking the past for a clearer future, grant number 694481). WCr thanks the
507 Labex OTMed (grant no. ANR-11-LABX-0061) funded by the French Government Investissements
508 d’Avenir program of the French National Research Agency (ANR) through the A*MIDEX project
509 (grant no. ANR-11-IDEX-0001-02). ICP, GM and CM acknowledge support from the ERC-funded

510 project REALM (Re-inventing Ecosystem And Land-surface Models, grant number 787203). IJW
511 acknowledges Australian Research Council funding (DP170103410). HW acknowledges support
512 from the National Natural Science Foundation of China (no. 31971495) and the High End Foreign
513 Expert awards at Tsinghua University to SPH and ICP (GDW20191100161). NS acknowledges
514 funding from Texas Tech University. JP acknowledges support from the ERC-funded project
515 IMBALANCE-P (grant number 610028). AL was supported by a Marie Skłodowska-Curie
516 Individual Fellowship (ECAW-ISO, grant number 838739). OF acknowledges funding provided by
517 the Knut and Alice Wallenberg foundation. TFK acknowledges support from the Reducing
518 Uncertainties in Biogeochemical Interactions through Synthesis and Computation Scientific Focus
519 Area (RUBISCO SFA), which is sponsored by the Regional and Global Model Analysis (RGMA)
520 Program of the U.S. Department of Energy. This work is a contribution to the Imperial College
521 initiative on Grand Challenges in Ecosystems and the Environment (ICP).

522

523 **Author Contributions.** SPH, OF, WC, SP, ICP and HW developed the outline of the paper, SPH led
524 the writing process, HW, HdB, NS, AL, TK, GM, CM and OF contributed display items, and all
525 authors contributed to the final versions of the paper.

526

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Box: How land (vegetation?) models are developed: some problems and proposed solutions

Models do not encapsulate a coherent body of theory. Current models represent the accumulation of work by successive cohorts of scientists in a specific institution or collaboration. Generally, no one has a complete overview of what they contain, or why particular decisions were made in the past (Fisher and Koven, 2020). This approach is normal for operational (e.g. national weather or air-quality forecast) models, but it inhibits efforts to trace why a model behaves as it does. We suggest there is a need for a *re-design, in which the core processes are examined one-by-one* and specific hypotheses about these processes tested against relevant data. Our proposed theory-driven approach more closely resembles how models are used in most other fields of science.

Lack of clarity about hypotheses. Many aspects of plant and ecosystem function are subject to alternative interpretations, and it seems likely that some differences among the results of models originate in different hypotheses adopted for one or another process. These hypotheses are seldom explicitly stated, however. Although model codes and documentation are now commonly made public (a positive development), the models' complex history can make it hard to discern their underlying logic. There is a strong argument for greater clarity, and above all, a *re-examination of the evidence* underlying the representation of each process.

Unclear testing protocols. New process representations are often assessed by changing one component within an existing, complex model and examining the effect on model outputs. Results obtained in this way are seldom clear-cut, however, because they are potentially vitiated by errors elsewhere in the model. Instead, *model components should be tested independently of others.*

Core process representations tend to be conserved. In many vegetation models, representations of the core processes of coupled energy, carbon and water exchange have remained unchanged since their original formulation. There is, in principle, no reason why new representations of core processes should not replace existing ones. However, it is noteworthy that this has not generally happened. The "legacy" formulations were likely provisional, and might now be obsolete – yet they are seldom isolated and tested, while model "improvement" more often consists of adding new processes (Fisher et al. 2014). We propose that model development should be *re-focused on the critical analysis and evaluation of core process representations*, and new processes added only if evidence unambiguously shows that they are required.

Neglect of available observations. Model parameter estimation tends to be *ad hoc* and is frequently based on single values for 'model' species that are long outdated. Models are relatively uninformed by the wealth of currently available observations. This situation could be remedied by *closer integration of data analysis and experimental research* into model development.

1140 **Figure Captions**

1141 Figure 1. Response of photosynthetic capacity (V_{cmax}) measured in *Holcus lanatus* (C_3 grass) and
 1142 *Solanum dulcamara* (vine) to CO_2 , shown in growth-chamber experiments where ambient CO_2 and
 1143 phosphorus (P) supply were manipulated (H. J. de Boer, unpublished results). Low-P and high-P
 1144 treatments received fertilizer with nitrogen to phosphorus (N:P) ratios of 45:1 and 1:1, respectively.
 1145 The responses of V_{cmax} to sub-ambient and elevated CO_2 are consistent with the predictions of the
 1146 coordination hypothesis as modelled following the eco-evolutionary optimality formulations of Wang
 1147 et al. (2014) (P14) and Smith et al. (2019) (S19). This response is not affected by nutrient supply
 1148 (high- and low-P treatments). Experimental V_{cmax} was scaled relative to the high-P population average
 1149 under ambient CO_2 growth conditions (450 ppm), whereas modelled V_{cmax} was scaled relative to the
 1150 single 450 ppm prediction. The response to CO_2 is significant at the 99% confidence level, as is the
 1151 response to P supply, but the interaction term is non-significant indicating that the response to CO_2
 1152 is the same regardless of P supply.

1153 Figure 2. Observed (by eddy covariance, grey) and modelled (using the same parameter set, red) half-
 1154 hourly gross primary production (GPP) during one week in August 2014 at sites in a tropical (GF-
 1155 Guy) and a boreal (FI-HYY) forest. Grey bands denote the range of GPP calculated by alternative
 1156 partitioning methods in the FLUXNET2015 database (Pastorello et al., 2020).

1157 Figure 3. Field-measured (black lines) (a) leaf dark respiration rates (R_{dark}) and (b) photosynthetic
 1158 capacities (V_{cmax}) compared to their modelled responses to growing-season temperature (red solid
 1159 lines) as predicted by the coordination hypothesis (Wang et al., 2020). Both R_{dark} and V_{cmax} have been
 1160 corrected (using the Arrhenius equation, with ΔH as provided by Bernacchi et al. 2001) from the
 1161 specific measurement temperature to the growing-season average temperature for the site. The
 1162 coordination hypothesis predicts acclimated responses to temperature. The red dashed lines show the
 1163 *instantaneous* responses to temperature, i.e. the relationship assumed by most models, based on the
 1164 same Arrhenius equation. Data from the GlobResp (Atkin et al., 2015) and LCE (Smith and Dukes,
 1165 2017) data sets.

1166 Figure 4. Trends in the ratio of leaf-internal to ambient CO_2 reconstructed from stable carbon isotope
 1167 ratios in tree rings for different sites (coloured lines) and for the whole dataset (black) with respect to
 1168 (a) mean growth temperature, (b) vapour pressure deficit (VPD), (c) atmospheric CO_2 and (d)
 1169 elevation, compared to modelled trends (red line) for the whole dataset based on the least-cost
 1170 hypothesis. Data and model results from Lavergne et al. (2020a).

1171 Figure 5. Observed relationships (black lines) between leaf lifespan and (a) leaf mass per area, (b)
 1172 radiation and (c) growing-season length as a fraction of the year compared to relationships predicted

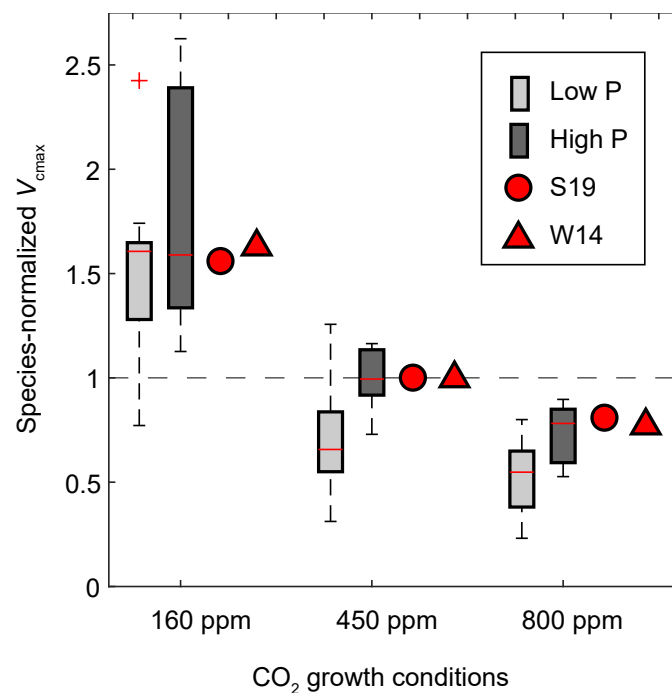
1173 (red lines) by the time-averaged maximization of leaf carbon gain through leaf lifespan after
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1179 related to leaf energetic status (Morfopoulos et al., 2013). Data are shown for an air temperature
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1188 The imposed treatment effects (independent variables) are light-saturated leaf-level photosynthetic
1189 N-use efficiency and soil N availability (maximal potential N uptake per root C). The r^2 is 0.90 for
1190 NPP, and 0.85 for N uptake (excluding circled outliers). Measured units of soil N in (a) are extractable
1191 nitrate ($0.1 \mu\text{g g soil}^{-1}$) in Duke and net N mineralization ($20 \mu\text{g g soil}^{-1} \text{y}^{-1}$) in ORNL. Data and
1192 model results from Franklin et al. (2009).

193 Figure 1. Response of photosynthetic capacity (V_{cmax}) measured in *Holcus lanatus* (C_3 grass) and
 194 *Solanum dulcamara* (vine) to CO_2 , shown in growth-chamber experiments where ambient CO_2 and
 195 phosphorus (P) supply were manipulated (H. J. de Boer, unpublished results). Low-P and high-P
 196 treatments received fertilizer with nitrogen to phosphorus (N:P) ratios of 45:1 and 1:1, respectively.
 197 The responses of V_{cmax} to sub-ambient and elevated CO_2 are consistent with the predictions of the
 198 coordination hypothesis as modelled following the eco-evolutionary optimality formulations of Wang
 199 et al. (2014) (P14) and Smith et al. (2019) (S19). This response is not affected by nutrient supply
 200 (high- and low-P treatments). Experimental V_{cmax} was scaled relative to the high-P population average
 201 under ambient CO_2 growth conditions (450 ppm), whereas modelled V_{cmax} was scaled relative to the
 202 single 450 ppm prediction. The response to CO_2 is significant at the 99% confidence level, as is the
 203 response to P supply, but the interaction term is non-significant indicating that the response to CO_2
 204 is the same regardless of P supply.

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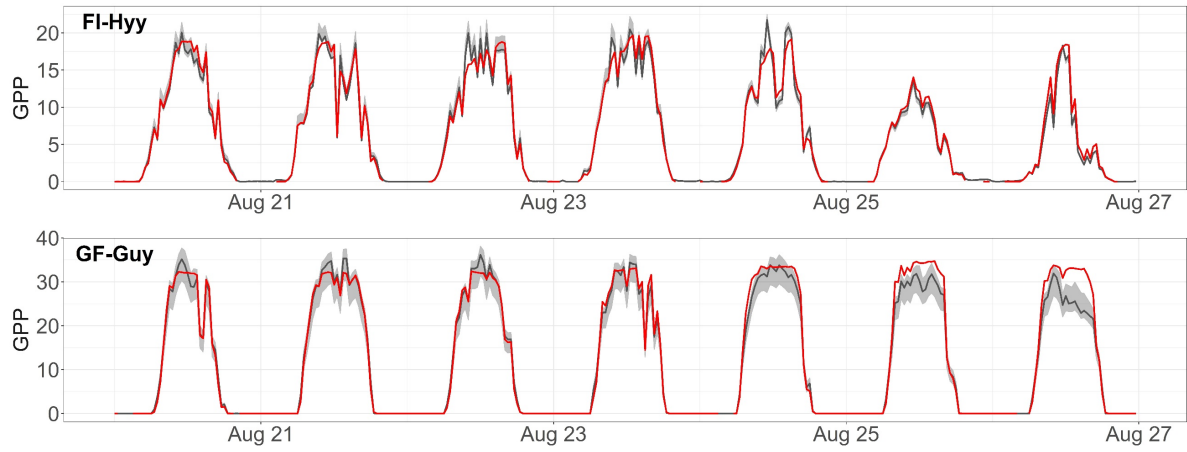
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|217 Figure 2. Observed (by eddy covariance, grey) and modelled (using the same parameter set, red) half-
|218 hourly gross primary production (GPP) during one week in August 2014 at sites in a tropical (GF-
|219 Guy) and a boreal (FI-HYY) forest. Grey bands denote the range of GPP calculated by alternative
|220 partitioning methods in the FLUXNET2015 database (Pastorello et al., 2020).

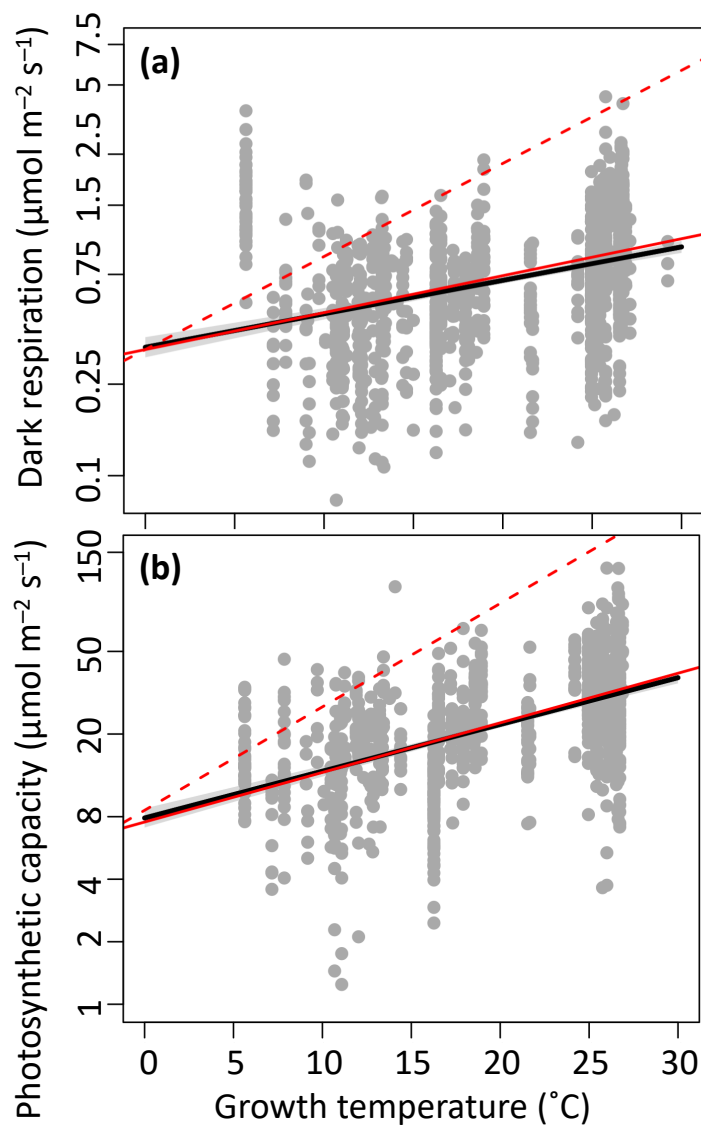
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1224 Figure 3. Field-measured (black lines) (a) leaf dark respiration rates (R_{dark}) and (b) photosynthetic
1225 capacities (V_{cmax}) compared to their modelled responses to growing-season temperature (red solid
1226 lines) as predicted by the coordination hypothesis (Wang et al., 2020). Both R_{dark} and V_{cmax} have been
1227 corrected (using the Arrhenius equation, with ΔH as provided by Bernacchi et al. 2001) from the
1228 specific measurement temperature to the growing-season average temperature for the site. The
1229 coordination hypothesis predicts acclimated responses to temperature. The red dashed lines show the
1230 *instantaneous* responses to temperature, i.e. the relationship assumed by most models, based on the
1231 same Arrhenius equation. Data from the GlobResp (Atkin et al., 2015) and LCE (Smith and Dukes,
1232 2017) data sets.
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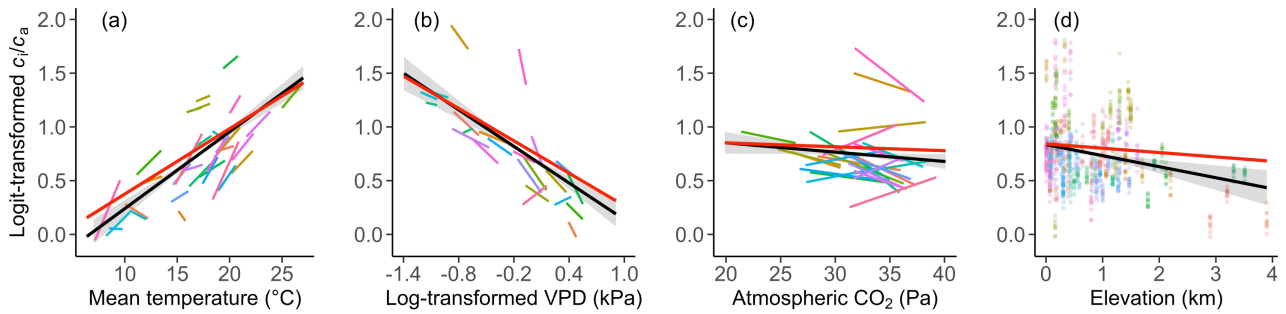
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237 Figure 4. Trends in the ratio of leaf-internal to ambient CO₂ reconstructed from stable carbon isotope
 238 ratios in tree rings for different sites (coloured lines) and for the whole dataset (black) with respect to
 239 (a) mean growth temperature, (b) vapour pressure deficit (VPD), (c) atmospheric CO₂ and (d)
 240 elevation, compared to modelled trends (red line) for the whole dataset based on the least-cost
 241 hypothesis. Data and model results from Lavergne et al. (2020a).

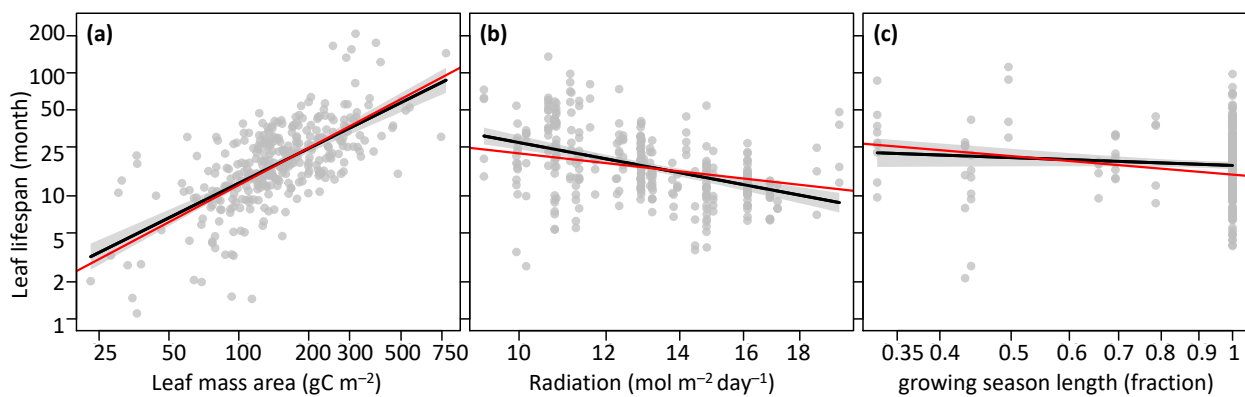
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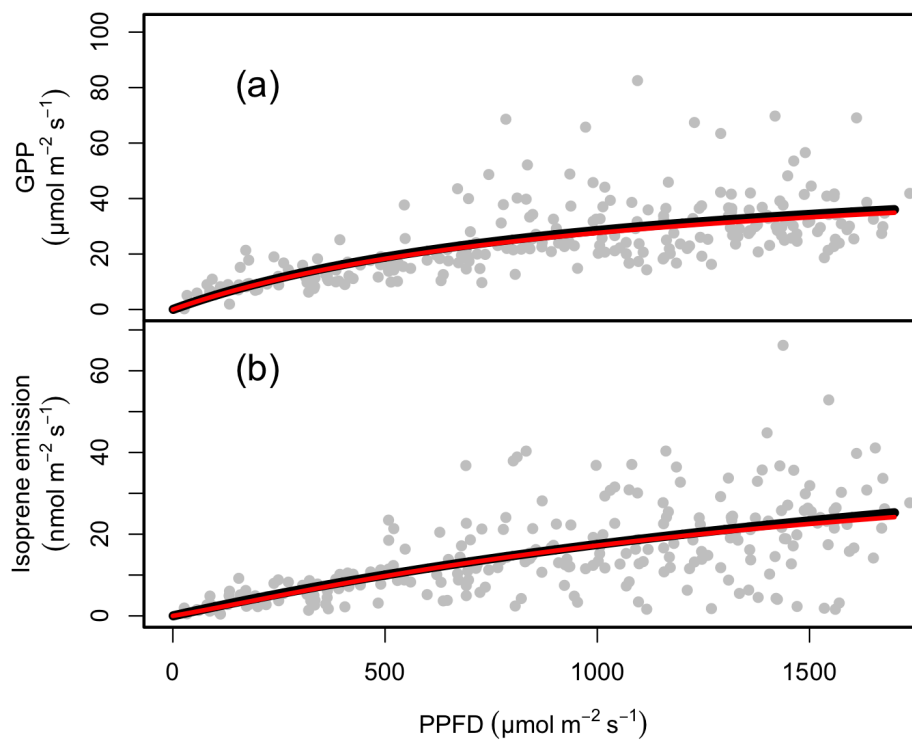
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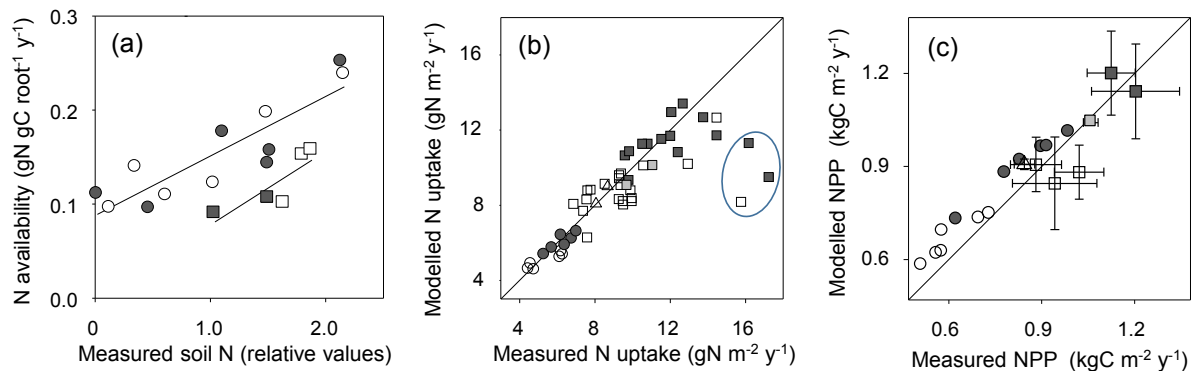
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