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1 Evapotranspiration methods compared on a Sierra Nevada
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1 **Abstract**

2
3 Evapotranspiration, a major component in terrestrial water balance and net primary
4 productivity models, is difficult to measure and predict. This study compared five models of
5 potential evapotranspiration (PET) applied to a ponderosa pine forest ecosystem at an AmeriFlux
6 site in Northern California. The AmeriFlux sites are research forests across the United States,
7 Canada, Brazil, and Costa Rica with instruments on towers that measure carbon, water, and
8 energy fluxes into and out of the ecosystems. The evapotranspiration models ranged from
9 simple temperature- and solar radiation-driven equations to physically-based combination
10 approaches and included reference surface and surface cover-dependent algorithms. For each
11 evapotranspiration model, results were compared against mean daily latent heat from half-hourly
12 measurements recorded on a tower above the forest canopy. All models calculate potential
13 evapotranspiration (assuming well-watered soils at field capacity), rather than actual
14 evapotranspiration (based on soil moisture limitations), and thus overpredicted values from the
15 dry summer seasons of 1997 and 1998. A soil moisture function was integrated to estimate actual
16 evapotranspiration, resulting in improved accuracy in model simulations. A modified Priestley-
17 Taylor method performed well given its relative simplicity.

18
19 *Keywords:* Evapotranspiration; AmeriFlux; FLUXNET; Biosphere-atmosphere interactions;
20 Shuttleworth-Wallace; Penman-Monteith; McNaughton-Black; Penman; Priestley-Taylor

1 1. Introduction

2
3 Over the entire land surface of the globe, rainfall averages around 750 mm-year⁻¹, of which
4 some two-thirds is returned to the atmosphere as evapotranspiration, making evapotranspiration
5 the largest single component of the terrestrial hydrological cycle (Baumgartner and Reichel,
6 1975). Carbon dioxide (CO₂)-induced greenhouse warming has accelerated the necessity to
7 understand the hydrologic cycle and climate change (GCIP-GEWEX, 1993; Houghton et al.,
8 1990; IGP-BAHC, 1993; Kaczmarek et al., 1996; Watson, 1995). Evapotranspiration and CO₂
9 uptake by vegetation are intrinsically coupled, leading to links and feedbacks between land
10 surface and climate that have only begun to be explored (Hutjes et al., 1998). Forests can
11 strongly influence the global hydrologic and carbon cycles and thus climate (Musselman and
12 Fox, 1991). A limited number of canopy-scale eddy covariance studies have shown that drought
13 stress plays a significant role in net ecosystem exchange (e.g., Baldocchi, 1997). The
14 development of models describing climate-landscape relationships, such as water and carbon
15 fluxes at ecosystem levels, is a necessary step in understanding and predicting the effects of
16 changes in climate on landscape and on water resources (Kite, 1998).

17 Evapotranspiration model estimates and field measurements vary widely. Differences in the
18 treatment of evapotranspiration are prominent among both climate and terrestrial ecosystem
19 models (Shuttleworth, 1991; VEMAP, 1995). Evapotranspiration has always been difficult to
20 measure, especially on an ecosystem or watershed spatial scale. Methods have been developed
21 to measure evapotranspiration at the leaf level, the tree level, and the stand level. At the stand
22 level, instruments mounted on a tower above the canopy are routinely used to measure humidity
23 and wind velocities at high frequency, with water fluxes out of the forest canopy calculated by
24 the eddy covariance method. Inasmuch as the majority of moisture supplied by precipitation
25 returns to the atmosphere as evapotranspiration, and because evapotranspiration is one of the
26 most difficult processes to evaluate in hydrologic analysis, estimates are generally considered to
27 be a significant source of error in streamflow simulation (Burnash, 1995). Nonetheless, effective
28 characterization of the evapotranspiration process is critical for completing the water balance in
29 terrestrial ecosystems, and accurately predicting the effects of global climate and land use
30 change. A process-based understanding of evapotranspiration is needed to quantify likely
31 changes in evapotranspiration due to climate and land surface change (Choudhury and
32 DiGirolamo, 1998; Hutjes et al., 1998). Therefore, we find it necessary to evaluate various
33 evapotranspiration methods employed in these modeling efforts.

34 The current modeling approach for estimating evapotranspiration is to calculate potential
35 evapotranspiration (PET) using methods driven by meteorological data and/or vegetation
36 characteristics, and to scale this estimate down to actual evapotranspiration (AET) based on
37 limitations in available water (i.e., soil moisture) (Federer et al., 1996; Stannard, 1993;
38 Vörösmarty et al., 1998). PET has been used to describe the evapotranspiration that would occur
39 given an adequate water supply at all times (Linsley et al., 1958). However, the term PET is
40 somewhat ambiguous, because the upper limit to evapotranspiration is dependent on vegetation
41 type as well as soil water and climatic conditions (Burman and Pochop, 1994). The historical
42 development of the PET concept has led to a variety of both PET definitions and methods
43 (Federer et al., 1996; Shuttleworth, 1991). Following the nomenclature of Shuttleworth (1991),
44 we examined two types of PET methods: reference-surface PET methods and surface-dependent
45 PET methods. Reference-surface evapotranspiration is defined as evapotranspiration that would
46 occur from a land surface specified as a “reference crop” (usually a short, uniform, green plant

1 cover such as alfalfa or grass) under designated weather conditions and soil at field capacity
2 (also termed well-watered soil) (Federer et al., 1996; Shuttleworth, 1991). Reference-surface
3 methods generally focus on an empirical relationship between temperature and PET, but neglect
4 vegetation. Surface-dependent evapotranspiration is defined as evapotranspiration that would
5 occur from a specified land surface, and the methods generally include a combination of
6 vegetation and soil characteristics.

7 Several methods for estimating evapotranspiration have been introduced in the literature, and
8 our study includes many of them. Vörösmarty et al. (1998) compared 11 methods on all the
9 watersheds of the continental US using a water-balance model. Other studies compared PET
10 methods at the fetch scale for sparsely vegetated rangeland (Stannard, 1993), wildland vegetation
11 in semiarid rangeland (Di, 1993), partial canopy/residue-covered fields (Farahani and Ahuja,
12 1996), maize with bare soil (Farahani and Bausch, 1995), and barley (Tourula and Heikinheimo,
13 1998). Federer et al. (1996) compared PET methods at seven locations, but did not compare the
14 PET estimates with measurements, because the purpose was to inform global modeling efforts.
15 McNaughton and Black (1973) performed energy balance measurements of evapotranspiration in
16 a Douglas fir forest with soil at field capacity, resulting in a PET method used in our study. Few
17 studies have analyzed evapotranspiration dynamics in forest ecosystems not only because of the
18 general focus on agriculture, but also because of the difficulty of obtaining evapotranspiration
19 measurements in forests.

20 The primary objective of our study is to compare a subset of PET models used by
21 Vörösmarty et al. (1998), but at a smaller spatial scale that uses calculated evapotranspiration in
22 a forest plantation based on micrometeorological measurements and eddy covariance methods at
23 our study site. The models include surface-dependent methods developed by Shuttleworth and
24 Wallace (1985), Penman and Monteith (1965), Priestley and Taylor (1972), and McNaughton
25 and Black (1973), and a reference-surface method by Penman (1948); the models were chosen
26 because they are commonly used in water balance (e.g., Arnell and Reynard, 1996; Willmott et
27 al., 1985a) and terrestrial ecosystem (net primary production) models (e.g., Melillo et al., 1993;
28 Parton et al., 1993; Running and Hunt, 1993). While scale and water-limitations at our site are
29 important to this study, the critical factor driving our study is the use of ecosystem scale flux
30 data. The flux measurements at the Blodgett Forest site, as part of AmeriFlux and the larger
31 FLUXNET network of towers across the world, measure water, carbon, and energy fluxes as
32 well as meteorological variables above the forest ecosystem canopy (e.g., Goldstein et al., 2000).
33 Although the first modeling and analysis of forest evapotranspiration was done in the 1970s
34 (e.g., Spittlehouse et al., 1979), the flux data from the tower are novel and only recently have
35 researchers used such data to drive larger-scale ecosystem models.

36 We assess the PET methods at this stand-scale using data from a tower at the Blodgett Forest
37 Research Station in California. We analyzed data acquired continuously over two climatically
38 different growing seasons: 1997 was drier than the climatic mean, and 1998 was cooler and
39 wetter than the climatic mean (influenced by El Niño). Evapotranspiration rates (measured
40 evapotranspiration) were derived from eddy covariance measurements, and environmental
41 variables were measured that are known to influence evapotranspiration rates, such as net
42 radiation, wind speed, air temperature, relative humidity, and soil moisture.

43 44 45 **2. Methods and materials**

2.1. Site description

Data were gathered during the summers of 1997 and 1998 in the Sierra Nevada mountains at Blodgett Forest Research Station (38°53'42.9"N, 120°37'57.9"W, 1315 m), a research forest of the University of California, Berkeley (Goldstein et al., 2000). The forest was planted in 1990 and was dominated by ponderosa pine trees (*Pinus ponderosa* Doug. E. Laws), the most common conifer species in North America. The canopy also included individuals of Douglas fir (*Pseudotsuga menziesii*), white fir (*Abies concolor*), giant sequoia (*Sequoiadendron giganteum*), incense-cedar (*Calocedrus decurrens*) and California black oak (*Quercus kelloggii*). The major understory shrubs were manzanita (*Arctostaphylos* spp.) and *Ceanothus* spp. (Xu, 2001a). In 1997, about 25% of the ground area was covered by shrubs, 30% by conifer trees, 2% by deciduous trees, 7% by forbs, 3% by grass and 3% by stumps. The forest area was in a stage of rapid growth, as exhibited by the 10% increase in leaf area index (LAI) between the 1997 (2.9-4.2) and 1998 (3.2-4.5) growing seasons. The site is characterized by a Mediterranean climate with an average annual precipitation of 163 cm (180 cm in 1997 and 117 cm in 1998), the majority of which falls between September and May, and almost no rain in the summer. The soil is a fine-loamy, mixed, mesic, ultic haploxeralf in the Cohasset series whose parent material was andesitic lahar (Goldstein et al., 2000).

2.2. Measurements

Infrastructure for the canopy scale flux measurements consisted of a 10 m measurement tower (Upright Inc.). From 1 June to 10 September 1997 and from 1 May to 30 October 1998, fluxes of CO₂, H₂O, and sensible heat were measured by the eddy covariance method. Environmental parameters such as wind speed and direction, air temperature and humidity, net and photosynthetically active radiation, soil temperature, soil moisture, soil heat flux, rain, and atmospheric pressure were also monitored. A system to measure the vertical profiles of CO₂ and H₂O was added in 1998. The data acquisition system was separated in two parts: (1) a fast response system which monitored data at high frequency (up to 10 Hz) used to calculate eddy covariance, with raw data stored in 30 min data sets; and (2) a slow response system which monitored environmental parameters and stored 30 min averaged data (Goldstein et al., 2000).

Wind velocity and temperature were measured at 10 Hz with a three-dimensional sonic anemometer (ATI Electronics Inc., Boulder, CO) mounted 5 m above the canopy to obtain an accurate reading of air above the canopy. The height will vary in other ecosystems depending on the vegetation characteristics, but the purpose is to measure representative wind characteristics. CO₂ and H₂O mixing ratios were measured with an infrared gas analyzer (IRGA, LICOR model 6262, Lincoln, NE). Fluxes of CO₂, H₂O, and sensible heat between the forest and the atmosphere were determined by the eddy covariance method (Goldstein et al., 2000). This method quantifies vertical fluxes of scalars between the forest and the atmosphere from the covariance between vertical wind velocity and scalar fluctuations averaged over 30 min periods (e.g., Baldocchi et al., 1988; Moncrieff et al., 1996; Shuttleworth et al., 1984; Wofsy et al., 1993). Environmental parameters were recorded on a CR10X datalogger (Campbell Scientific Inc., Logan, UT). After initially determining that spatial variability in soil water content is low, soil moisture probes were buried horizontally at 10 and 20 cm depth. Brandes (1998) has also shown, through principal components analysis, that lateral spatial variability contributes only a small portion (<10 percent) of the total variance of a soil moisture data set. The soil moisture

1 measurements at 20 cm likely reflected the moisture in the rooting zone; the soil moisture
2 measurements at 10 cm were sensitive to variability in surface conditions, therefore we used the
3 20 cm depth data. Total (all-sided) LAI was estimated using two techniques that resulted in
4 similar estimates, (1) the LI-2000 (Li-Cor Inc., Lincoln, NE), and (2) an allometric method that
5 scaled up from leaf-level determination using the measured geometry of trees.

6 Systematic errors associated with the eddy covariance method include time lags between
7 wind and scalar data due to travel through sampling tube and instrument response time, damping
8 of high frequency fluctuations by the closed-path IRGA and travel through the sampling tube,
9 sensor separation between wind and scalar measurements (Rissman and Tetzlaff, 1994), and
10 inability of the sonic anemometer to resolve fine-scale eddies in light winds (Goulden et al.,
11 1996; Moncrieff et al., 1996). Generally, these types of errors result in the underestimation of
12 flux (Leuning and King, 1992). The inability of the sonic anemometer to resolve the vertical
13 wind occurs mainly at night as the fluctuations become dominated by small, high frequency
14 eddies (Goulden et al. (1996) use $u^* < 0.17 \text{ m s}^{-1}$ as the threshold for reliable measurements).
15 The inability of the sonic anemometer to resolve fine-scale eddies in light winds (e.g., during
16 night) produced systematic errors in the sensible heat flux to correct the CO₂ and H₂O fluxes.
17 Thus, although daytime turbulence was strong enough to produce reliable measurements, the
18 calmer conditions during night rendered the nighttime flux measurements less reliable (Goldstein
19 et al., 2000). For the purpose of this study, we corrected for outliers (greater than three standard
20 deviations from the mean) and missing data points (via interpolation or backup sensors), and
21 evaluated the evapotranspiration models using daytime (5am – 9pm) averages because nighttime
22 measurements were unreliable. Sample size in 1997 was 87 daytime averages based on 3900
23 measurements, and sample size in 1998 was 149 daytime averages based on 8700 measurements.
24
25

26 2.3. *Evapotranspiration models*

27

28 Five PET models of increasing complexity were tested under two classes of land surface
29 speciation (Federer et al., 1996; Shuttleworth, 1991). Reference-surface evapotranspiration is
30 defined as evaporation that would result from a specific land surface, referred to as a “reference
31 crop” (Vörösmarty et al., 1998). Surface-dependent evapotranspiration is defined as the
32 evaporation that would occur from any of a variety of designated land surfaces. A summary of
33 the parameters and units used in each method is presented in Table 1.

34 The Priestley-Taylor model (Priestley and Taylor, 1972), the simplest of the five, is defined
35 as:

$$36 \lambda E = \alpha \Delta A / (\Delta + \gamma)$$

37

38 where λE is total potential evapotranspiration (in flux units of W/m^2 , for example), Δ is the
39 derivative of saturated vapor pressure versus temperature, A is total available energy (net
40 radiation minus soil heat flux), and γ is the psychrometric constant. Δ and γ are functions of air
41 temperature (all models analyzed here are thus functions of air temperature). Priestley and
42 Taylor (1972) determined an average value of 1.26 for α based on measurements of
43 evapotranspiration from a variety of well-watered vegetated and water surfaces (i.e., PET). To
44 estimate AET, α has been redefined to be a function of soil moisture (Flint and Childs, 1991).
45

46 The McNaughton-Black model (McNaughton and Black, 1973) is defined as follows:

$$\lambda E = c_p \rho D / \gamma r_{cs}$$

where c_p is specific heat at constant pressure, ρ is air density, D is vapor pressure deficit, and r_{cs} is bulk stomatal resistance of the canopy.

The Penman model (Penman, 1948) is defined as follows:

$$\lambda E = (\Delta A + 73.64 \rho \gamma (1 + 0.54u) D) / (\Delta + \gamma)$$

where u is wind speed, and 73.64 is 2.6 times the latent heat of vaporization (units converted).

The Penman-Monteith model (Monteith, 1965) expanded upon the Penman model:

$$\lambda E = (\Delta A + c_p \rho D / r_{aa}) / (\Delta + \gamma + \gamma (r_{cs} / r_{aa}))$$

where r_{aa} is the aerodynamic resistance above the canopy, and r_{cs} is stomatal resistance of the canopy.

For the Shuttleworth-Wallace model (Shuttleworth and Wallace, 1985), the evaporation from the soil, λE_s , and the transpiration from the canopy, λE_c , are derived from the Penman-Monteith combination equations:

$$\lambda E_s = (\Delta A_s + \rho c_p D_0 / r_{sa}) / (\Delta + \gamma (1 + r_{ss} / r_{sa}))$$

$$\lambda E_c = (\Delta (A - A_s) + \rho c_p D_0 / r_{ca}) / (\Delta + \gamma (1 + r_{cs} / r_{ca}))$$

where A_s is available soil energy, and D_0 is vapor pressure deficit in the canopy; r_{sa} is the aerodynamic resistance between the substrate and canopy source height, r_{ca} is the boundary layer resistance of the vegetation, and r_{ss} is soil resistance. The aerodynamic resistance above the canopy (r_{aa}) and the aerodynamic resistance between the substrate and canopy source height (r_{sa}) are functions of leaf area index, eddy diffusivity decay constant, roughness length of the vegetation (function of vegetation height), zero plane displacement (function of vegetation height), a reference height above the canopy where meteorological measurements are available, wind speed, von Karman's constant, and roughness length of the substrate. D_0 is derived from the Ohm's law electrical analog for the vapor pressure and temperature difference between the canopy and the reference height above the canopy where fluxes out of the vegetation are measured. D_0 is a function of the measurable vapor pressure deficit at the reference height, D :

$$D_0 = D + (\Delta A - r_{aa} \lambda E_c (\Delta + \gamma)) / \rho c_p$$

and D can thus be substituted for D_0 into the combination equations. The total evaporation from the crop, λE , for the Shuttleworth-Wallace model is the sum of the Penman-Monteith combination equations with D substituted in for D_0 :

$$\lambda E = C_c P M_c + C_s P M_s$$

1 where PM_c describes evaporation from the closed canopy, and PM_s describes evaporation from
 2 the bare substrate. The new Penman-Monteith equations have the form:

$$3 \quad PM_c = \frac{(\Delta A + (\rho C_p D - \Delta r_{ca} A_s) / (r_{aa} + r_{ca}))}{(\Delta + \gamma(1 + r_{cs} / (r_{aa} + r_{ca})))}$$

$$4 \quad PM_s = \frac{(\Delta A + (\rho C_p D - \Delta r_{sa} (A - A_s) / (r_{aa} + r_{sa})))}{(\Delta + \gamma(1 + r_{ss} / (r_{aa} + r_{sa})))}$$

5
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9
10 The coefficients C_c and C_s are resistance combination equations:

$$11 \quad C_c = 1 / (1 + R_c R_a / (R_s (R_c + R_a)))$$

$$12 \quad C_s = 1 / (1 + R_s R_a / (R_c (R_s + R_a)))$$

13
14
15 where

$$16 \quad R_a = (\Delta + \gamma) r_{aa}$$

$$17 \quad R_s = (\Delta + \gamma) r_{sa} + \gamma r_{ss}$$

$$18 \quad R_c = (\Delta + \gamma) r_{ca} + \gamma r_{cs}$$

19
20
21
22 The evapotranspiration models described above calculate potential evapotranspiration rather
 23 than actual evapotranspiration. Potential evapotranspiration is defined as the evapotranspiration
 24 flux from the ecosystem under well-watered soil conditions (i.e., soil at or close to, field
 25 capacity). We derived actual evapotranspiration from potential evapotranspiration using a
 26 simple soil moisture function, $f(\phi)$ (Saxton et al., 1986):

$$27 \quad \lambda E_{\text{actual}} = f(\phi) * \lambda E$$

28
29
30 where $\lambda E_{\text{actual}}$ is the actual evapotranspiration and the soil moisture function is a dimensionless
 31 variable estimated by a simple linear model:

$$32 \quad f(\phi) = M / \text{Field capacity}$$

33
34
35 where M is soil volumetric moisture at 20 cm depth (at rooting zone). Field capacity was
 36 determined as 39% based on Saxton et al. (1986) and compared well to maximum soil moisture
 37 observed after rain events. Brandes and Wilcox (2000) have shown that simple linear models of
 38 the evapotranspiration/soil moisture process are appropriate for hydrologic modeling. Soil
 39 moisture models have been developed with increasing complexity to better represent soil
 40 physics, such as the module in WATFLOOD that includes permanent wilting point, saturation,
 41 and a root fraction to simulate non-linear features of moisture extraction by vegetation from soil
 42 (Soulis et al., 2000). However, we chose the simple, physically-based soil moisture function for
 43 three reasons: 1) the data requirements and modeling demands are greatly increased for more
 44
45

1 complex soil moisture functions; 2) these PET methods are often used in larger spatial scale
2 models where detailed soil physics may not be possible to calculate accurately; and 3) our main
3 goal is to assess the relative accuracy of the PET methods, rather than evaluate the merit of
4 various approaches to modeling soil physics.

5 6 7 **3. Results**

8 9 *3.1. Potential versus measured evapotranspiration*

10
11 For all potential evapotranspiration models, simulated PET compared reasonably well with
12 measured evapotranspiration at the beginning of the summer season (April-May). However, as
13 the soil moisture decreased through the summer, all models tended to overpredict
14 evapotranspiration because the PET models were designed for well-watered soil conditions
15 rather than natural summertime Mediterranean drought conditions (Figure 1). 1997 was drier
16 than 1998, and greater evapotranspiration was observed in 1998. The summer of 1997 was also
17 substantially windier than 1998—this fact influenced the wind-sensitive Penman method, which
18 predicted unrealistically high amounts of evapotranspiration due to the fast winds. In fact, the
19 Penman method predicts an increase of well over $100 \text{ W}\cdot\text{m}^{-2}$ in evapotranspiration for every 0.5
20 $\text{m}\cdot\text{s}^{-1}$ increase in wind speed, holding all other variables constant. The specific increase depends
21 on the values of the fixed variables; VPD in particular effects that increase.

22 PET results from the Shuttleworth-Wallace, Penman-Monteith, and McNaughton-Black
23 models had similar trends and magnitudes; McNaughton-Black tended to give the highest
24 estimates followed by Penman-Monteith and Shuttleworth-Wallace, respectively. Penman-
25 Monteith approximated Shuttleworth-Wallace in the dry season of 1997. The Priestley-Taylor
26 model nearly approximated the measured evapotranspiration in both years, especially in the
27 higher soil moisture year of 1998.

28 29 *3.2. Actual versus measured evapotranspiration*

30
31 Actual evapotranspiration was derived from potential evapotranspiration for each model by
32 applying the soil moisture function. The actual evapotranspiration calculations provided good
33 approximations of measured evapotranspiration. With the soil moisture function, Shuttleworth-
34 Wallace ($r^2 = 0.46$ in 1997; $r^2 = 0.69$ in 1998), Penman-Monteith ($r^2 = 0.43$ and 0.66), and
35 McNaughton-Black ($r^2 = 0.37$ and 0.62) all performed well with similar trends and magnitudes.
36 Penman-Monteith and McNaughton-Black approximated Shuttleworth-Wallace through both
37 seasons, though McNaughton-Black began to more severely overpredict AET than the other
38 models late in both seasons. Priestley-Taylor ($r^2 = 0.73$ and 0.58) significantly underpredicted
39 measured evapotranspiration with the soil moisture function. Although the Penman simulations
40 were improved, the Penman model still significantly overpredicted measured evapotranspiration.
41 Flint and Childs (1991) state that the assumptions and simplifications used by the Penman model
42 to model the aerodynamic components of evapotranspiration make the Penman model useful
43 only for calculating potential evapotranspiration. The primary modified version of the Penman
44 model, the Penman-Monteith model, allows for calculation of actual evapotranspiration given
45 values for resistances. The soil moisture function performed better across the relatively wet
46 season of 1998 than the dry season of 1997; the models tended to underpredict measured

1 evapotranspiration in 1997 (Figure 2 and 3). In addition, there was smaller scatter, but more
2 sample points, in 1998.

3 4 3.3. Modified α for Priestley-Taylor model

5
6 Whether or not use of a soil moisture function is appropriate for the Priestley-Taylor model
7 has been answered, in part, by numerous studies on α . It has become increasingly common to
8 redefine α based on soil moisture, rather than add on a soil moisture function to the original α
9 value of 1.26. Priestley and Taylor originally proposed that α be reduced when soil water
10 content falls below some critical soil moisture value where soil water supply limits
11 evapotranspiration. Spittlehouse and Black (1981) state that the Priestley-Taylor model is in
12 error and no longer appropriate to use if α is fixed. Based on the regression analysis prescribed
13 by Flint and Childs (1991), our redefined α is:

$$14 \qquad \qquad \qquad \alpha = 0.84M + 0.72$$

15
16 where M is soil volumetric moisture of the top 20 cm (rooting zone). The average value for α
17 across both years was 0.87, although that value was slightly lower for the drier 1997 and higher
18 for the wetter 1998. It should be noted that although this equation follows the work of Flint and
19 Childs (1991), the parameters for both this equation and for Flint and Child's redefined α are not
20 physically based and should thus be looked at critically in comparisons with other models.
21 However, our newly calculated values for α approximate the actual measured values for α at
22 similar sites, as tabled by (Flint and Childs, 1991); we append our value in Table 2.
23 Determination of α based on either the regressed soil moisture function or measurements done at
24 similar sites resulted in a greatly improved Priestley-Taylor model ($r^2 = 0.74$ and 0.85). With the
25 new α value, Priestley-Taylor AET estimates were not significantly different from measured
26 evapotranspiration across both years. Again, the Priestley-Taylor model performed well despite
27 its relative simplicity.
28
29
30

31 4. Discussion

32
33 Shuttleworth-Wallace, Penman-Monteith, and McNaughton-Black resulted in similar
34 simulations due to the common theoretical basis of their equations—the Penman model.
35 McNaughton-Black, which excludes the radiation budget and any effect from the soil, is a
36 simplification of Penman-Monteith, whereas Shuttleworth-Wallace adds a soil layer to the
37 Penman-Monteith model. The simulations revealed that Penman-Monteith tended to give an
38 intermediate result between these three models. Shuttleworth-Wallace is specifically designed
39 for sparse crops where vegetation is not densely distributed and the soil surface may contribute
40 significantly to evapotranspiration, which is representative of the Blodgett site. Nonetheless, the
41 substrate does not significantly contribute to evapotranspiration because of low soil moisture,
42 particularly in 1997. Thus, the Shuttleworth-Wallace model reduced back to the Penman-
43 Monteith model and gave only slightly better results. In the relatively wet season of 1998,
44 Shuttleworth-Wallace resulted in a more accurate simulation than in 1997 because the increased
45 soil moisture lead to greater soil evaporation. Still, the soil evaporation was not a significant
46 factor at this site and thus the McNaughton-Black model, which neglects the soil as an

1 evaporation source, yielded similar results. Shuttleworth-Wallace has performed well in the
 2 literature as well (e.g., Di, 1993; Farahani and Bausch, 1995; Iritz et al., 1999; Vörösmarty et al.,
 3 1998), but the main drawback is the difficulty and extensiveness of the parameter estimation
 4 (Farahani and Ahuja, 1996). Because our study area is intensively measured as a research site,
 5 the parameters for Shuttleworth-Wallace were available for this analysis; other sites or large
 6 scale modeling efforts may not be so fortunate.

7 The upper bound to potential evapotranspiration should be the net radiation (under relatively
 8 stable conditions). According to energy balance models, incoming net radiation is partitioned
 9 into latent heat, sensible heat, and heat absorbed by the ground—therefore, latent heat, as a
 10 fraction of net radiation, should not exceed net radiation. The McNaughton-Black model is not a
 11 function of net radiation and depends heavily on the accuracy of its other input parameters. An
 12 erroneously low stomatal resistance of the canopy (r_{cs}), for instance, can cause the McNaughton-
 13 Black model and other PET models that are function of r_{cs} to overpredict evapotranspiration.

14 For Shuttleworth-Wallace, Penman-Monteith, and McNaughton-Black, we used a constant r_{cs}
 15 throughout both seasons derived from the available minimum and maximum values measured at
 16 the site. The models are highly sensitive to r_{cs} and simulated evapotranspiration differed by as
 17 much as 26% at the minimum r_{cs} and 20% at the maximum r_{cs} . We evaluated and propagated
 18 error in r_{cs} , along with aerodynamic resistance above the canopy (r_{aa}), bulk boundary layer
 19 resistance of the vegetation (r_{ca}), aerodynamic resistance for the substrate and canopy (r_{sa}), and
 20 surface resistance of the substrate (r_{ss}) in the Shuttleworth-Wallace model via Gaussian error
 21 propagation (the final uncertainty for total evapotranspiration is equal to the square root sum of
 22 squares of the partial derivative of total uncertainty with respect to each resistance multiplied by
 23 the standard deviation of each resistance, respectively):

$$24 \quad S_{\lambda E} \approx [((\partial \lambda E / \partial r_{ca}^c) (Sr_{ca}^c))^2 + ((\partial \lambda E / \partial r_{sa}^c) (Sr_{sa}^c))^2 + ((\partial \lambda E / \partial r_{ss}^s) (Sr_{ss}^s))^2 + ((\partial \lambda E / \partial r_{aa}^a) (Sr_{aa}^a))^2]^{0.5}$$

25
 26
 27 The model is nonlinear, and, for the purposes of the uncertainty analysis, we treated the variables
 28 as uncorrelated with one another. Gaussian error propagation is suitable here since it allows for
 29 examination of total uncertainty derived from the uncertainties in the parameters of the model.
 30 We found that uncertainty in r_{cs} contributes to 53% of the total uncertainty in the Shuttleworth-
 31 Wallace model. Based on this finding, much of the overprediction among the PET models may
 32 be attributed to uncertainty in r_{cs} . Again, a relatively more complex evapotranspiration model
 33 may not be more accurate than a simpler model because the complex model is more difficult to
 34 parameterize and results are vulnerable to large error resulting from the propagation of error
 35 through some of its parameters.

36
 37 The success of Priestley-Taylor, given its relative simplicity, shows that this simple model
 38 may be more preferable than the complex models by Shuttleworth-Wallace, Penman-Monteith,
 39 and McNaughton-Black for a partially-closed canopy under water-limitation. We add to the
 40 tabled values of the Priestley-Taylor coefficient, α , for different surface conditions. The table
 41 has the potential of playing a crucial role in the integration of the Priestley-Taylor method for
 42 estimating AET into larger scale ecosystem models. If most surface conditions have an
 43 associated α value under varying water stress conditions, or include α as a function of soil
 44 moisture, then it may be possible that ecosystem models could use those values with the method
 45 to accurately assess AET given relatively few input parameters.

1 The soil moisture function plays a key role in deriving actual evapotranspiration from
2 potential evapotranspiration. PET assumes that soil water availability is not the limiting factor
3 for AET (soil is at field capacity) and will thus overpredict evapotranspiration under drier soil
4 conditions. In both years, the PET models performed well at the beginning of the summer
5 season when soil moisture was still high from spring rainfall and residual moisture from
6 snowmelt. But, as soil moisture declined throughout the summer, PET subsequently
7 overpredicted measured evapotranspiration. It is crucial that our study included summertime
8 drought conditions in a Mediterranean environment: it is under these, and similar, conditions that
9 the assumptions of a given PET model can lead to inaccurate results. Analysis of PET models
10 under environmental conditions that they were not initially designed for may be inappropriate.
11 Many ecological models use PET functions on continental and global scales, and are subject to
12 the same overprediction of actual evapotranspiration because of such assumptions (e.g., Raich et
13 al., 1991; Running and Coughlan, 1988; Sellers et al., 1996; Thornton et al., 1997).

14 Thus, the relationship between potential and actual evapotranspiration must be addressed in
15 these and future models. Furthermore, these large-scale models would benefit from a simple, but
16 accurate, evapotranspiration component that would be relatively easy to parameterize. Within
17 widely used general circulation models (GCMs) and numerical weather prediction models
18 (NWP), hydrological components such as evapotranspiration, precipitation and runoff have yet
19 to be estimated with great accuracy. Models of the atmosphere, such as GCMs and NWP, can
20 be used to predict the impact of changes in the composition of the atmosphere using gas
21 properties and the fundamental equations of energy transfer, mass conservation and atmospheric
22 motion into changes in wind, temperature and moisture content. Despite the rigor of the
23 atmospheric descriptions in such models, the atmosphere/land surface interaction is usually
24 modeled quite crudely (Kite, 1998). Most GCMs (e.g., Dickinson and Kennedy, 1991; Raich et
25 al., 1991; Sellers et al., 1996) use only a few hydrological parameters lumped over large grid
26 squares, and contain no run-on/runoff transfer between grid squares. Such models generally
27 overestimate precipitation and, as the vertical water balances in each grid square are independent
28 of surrounding squares, they generally underestimate evapotranspiration, leading to an
29 overestimate of water available for runoff (Kite, 1998). Information on land surface
30 evapotranspiration is very important in the understanding of climate change. For example, the
31 reduction in evapotranspiration (and the change in surface energy balance) associated with the
32 removal of vegetation in the Sahel has been shown, via GCMs, to produce a reduction in rainfall
33 (e.g., Charney, 1975; Cunnington and Rowntree, 1986). Therefore, there is a great need for
34 process-based evapotranspiration models than can characterize different vegetation to allow us to
35 better understand and predict any links between land-use change and climate change (Wallace,
36 1995). We suggest that the Priestley-Taylor method may be most applicable to models run at
37 large spatial scales because it is easier to parameterize than the widely used Penman-Monteith
38 method, although further research is needed to confirm this suggestion.

39 Factors not taken into account that may affect the relationship between simulated and actual
40 evapotranspiration include vegetative quality and other environmental variables. For example,
41 ozone deposition, grazing of insects on leaves, the influence of animals such as cows on the
42 environment, and disease are not taken into account when modeling evapotranspiration. Aside
43 from systematic errors associated with the eddy covariance method, possible bias in the data and
44 models include assumed values for three Shuttleworth-Wallace variables—surface resistance of
45 the substrate, roughness length of bare substrate, and extinction coefficient of the crop for net
46 radiation. We halved, doubled, and multiplied each parameter by a factor of 10 to test for

1 sensitivity; nonetheless, Shuttleworth-Wallace is not highly sensitive to these parameters.
2 Simulated evapotranspiration differed by less than 5% given the changes in these parameters. A
3 major environmental phenomenon influencing the data was the occurrence of an El Niño event
4 before the summer of 1998 that caused the vegetation to grow significantly in 1998; the
5 heterogeneity across seasons allowed for ideal comparisons of the same site under different
6 environmental conditions to see how robust the evapotranspiration models were.
7
8

9 **5. Conclusions**

10
11 Shuttleworth-Wallace, Penman-Monteith, McNaughton-Black, Priestley-Taylor, and Penman
12 models for estimating evapotranspiration were compared using data from AmeriFlux tower
13 measurements at a ponderosa pine ecosystem. Vörösmarty et al. (1998), in comparing these
14 models on a global scale, found that the Shuttleworth-Wallace method performed best. In our
15 study, Shuttleworth-Wallace, Penman-Monteith, and McNaughton-Black all yielded similar
16 results, although Shuttleworth-Wallace performed slightly better than Penman-Monteith and
17 McNaughton-Black; this similarity was because these models are derived from the Penman
18 model, and because of the insignificant effect of the substrate on evapotranspiration at our site.
19 Priestley-Taylor, with an appropriately defined α value, performed remarkably well, especially
20 given its relative simplicity. The Penman model was very sensitive to wind speed in our study.
21 When applying PET models, one must be aware of soil moisture conditions so that potential and
22 actual evapotranspiration are differentiated. Integration of data from all the FLUXNET sites
23 across the globe will be critical in determining the best possible evapotranspiration model to use
24 at global scales for predicting changes in land-surface exchange due to climate change.
25
26

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28
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35

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37

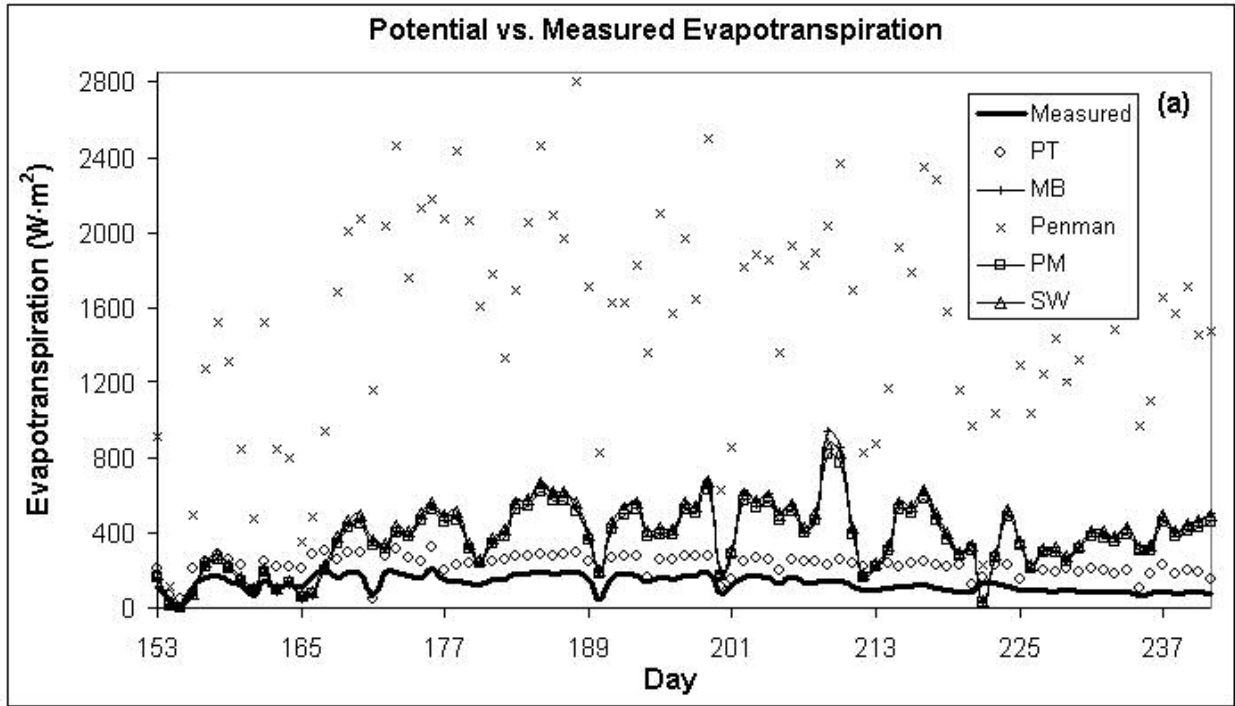
1 **Figure Captions**

2
3 Figure 1. Potential evapotranspiration without soil moisture function for (a) 1997 and (b)
4 1998. Shuttleworth-Wallace (SW), Penman-Monteith (PM), and McNaughton-Black (MB) all
5 have similar trends and magnitudes, and Priestley-Taylor (PT) comes closest to measured
6 evapotranspiration (Measured ET) in both years with $\alpha = 1.26$. The Penman method was highly
7 sensitive to wind speed in 1997.

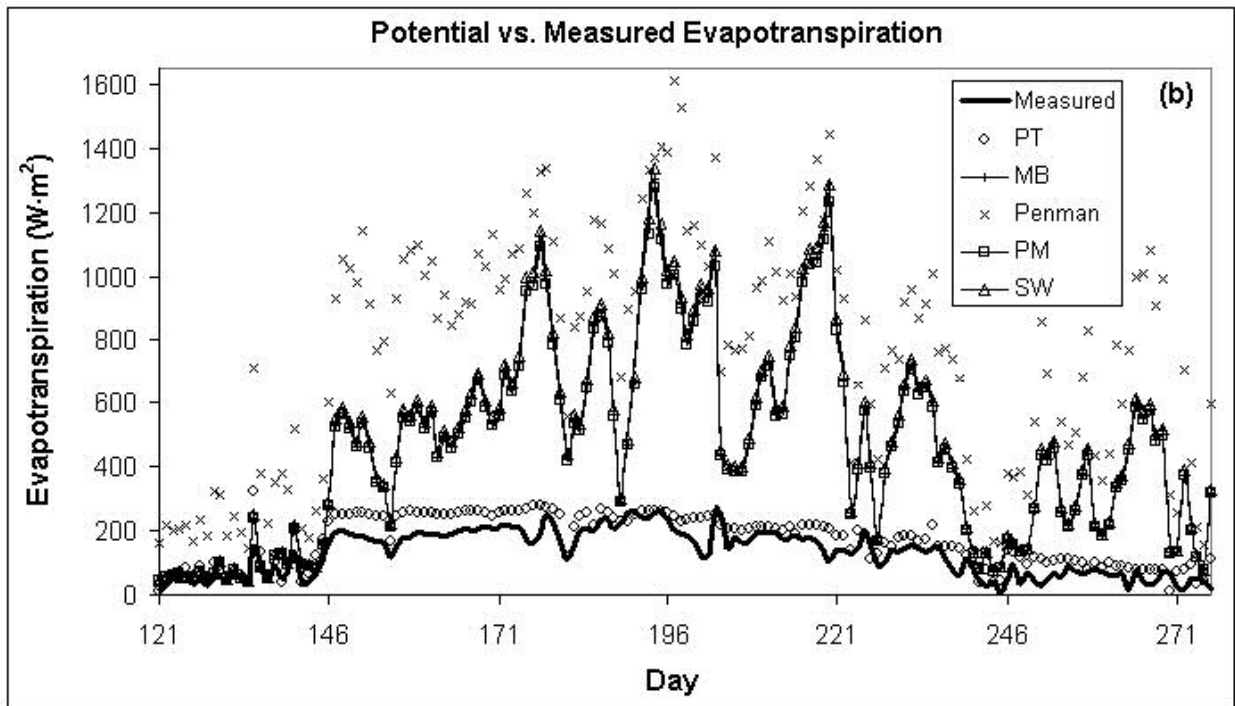
8
9 Figure 2. Simulated versus measured (actual) evapotranspiration during (a) 1997 and (b)
10 1998. The soil moisture function is integrated in Shuttleworth-Wallace (SW), Penman-Monteith
11 (PM), and McNaughton-Black (MB). Penman-Monteith and McNaughton-Black excluded due to
12 similarity to Shuttleworth-Wallace; while PM approximates SW through both seasons, MB
13 begins to diverge with overprediction late in both seasons as soil moisture decreases. The
14 Priestley-Taylor (PT) graph is shown with $\alpha = 0.73$ in 1997 and 0.94 in 1998. The soil moisture
15 function brought the simulations down to good approximations of measured evapotranspiration
16 (Measured ET). Penman method excluded due to continued overprediction.

17
18 Figure 3. Simulated versus measured evapotranspiration during (a) 1997 and (b) 1998. The
19 soil moisture function is integrated in Shuttleworth-Wallace (SW), Penman-Monteith (PM),
20 McNaughton-Black (MB), Penman, and Priestley-Taylor (PT with Soil Moisture function). The
21 second Priestley-Taylor (PT) graph includes the modified α . The thick diagonal line is the 1:1
22 line, and the thin diagonal line is the actual difference between the means. Points below the 1:1
23 line represent over-prediction, and points above the 1:1 line represent under-prediction. Sample
24 size in 1997 was 87 daytime averages based on 3900 measurements, and in 1998 was 149
25 daytime averages based on 8700 measurements.

26

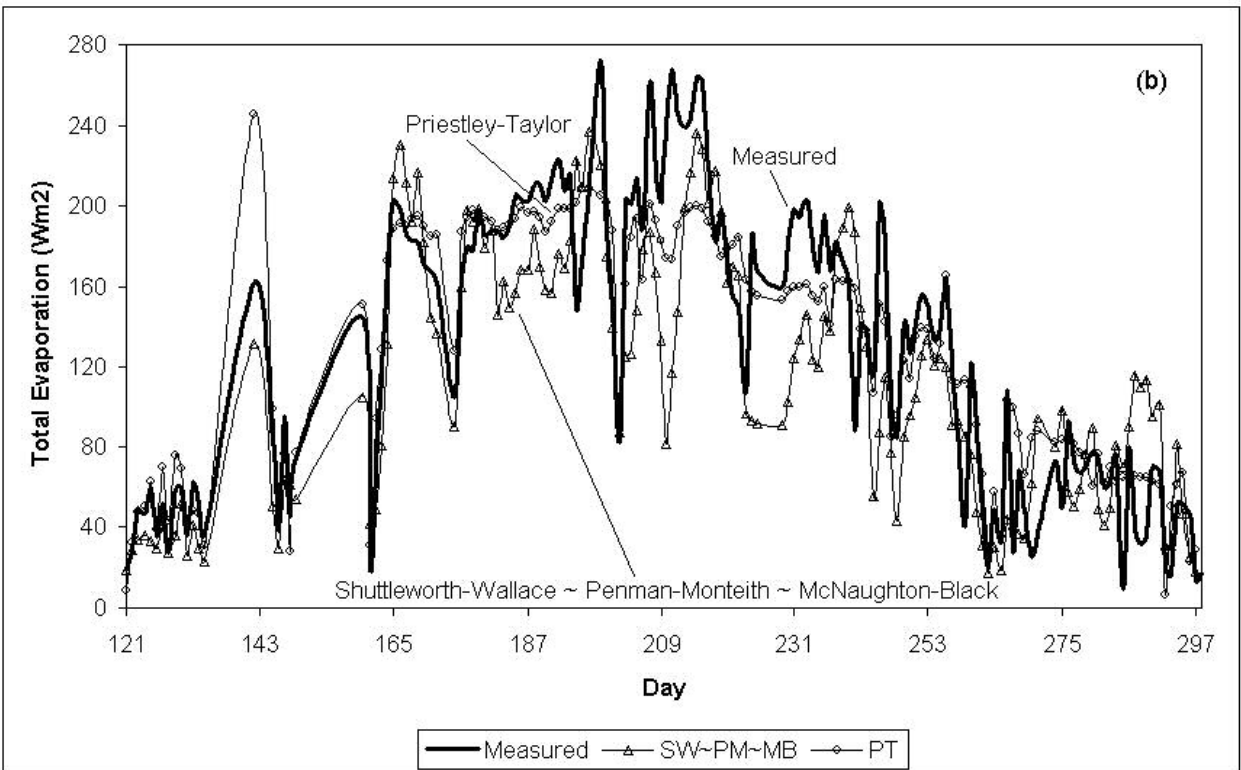
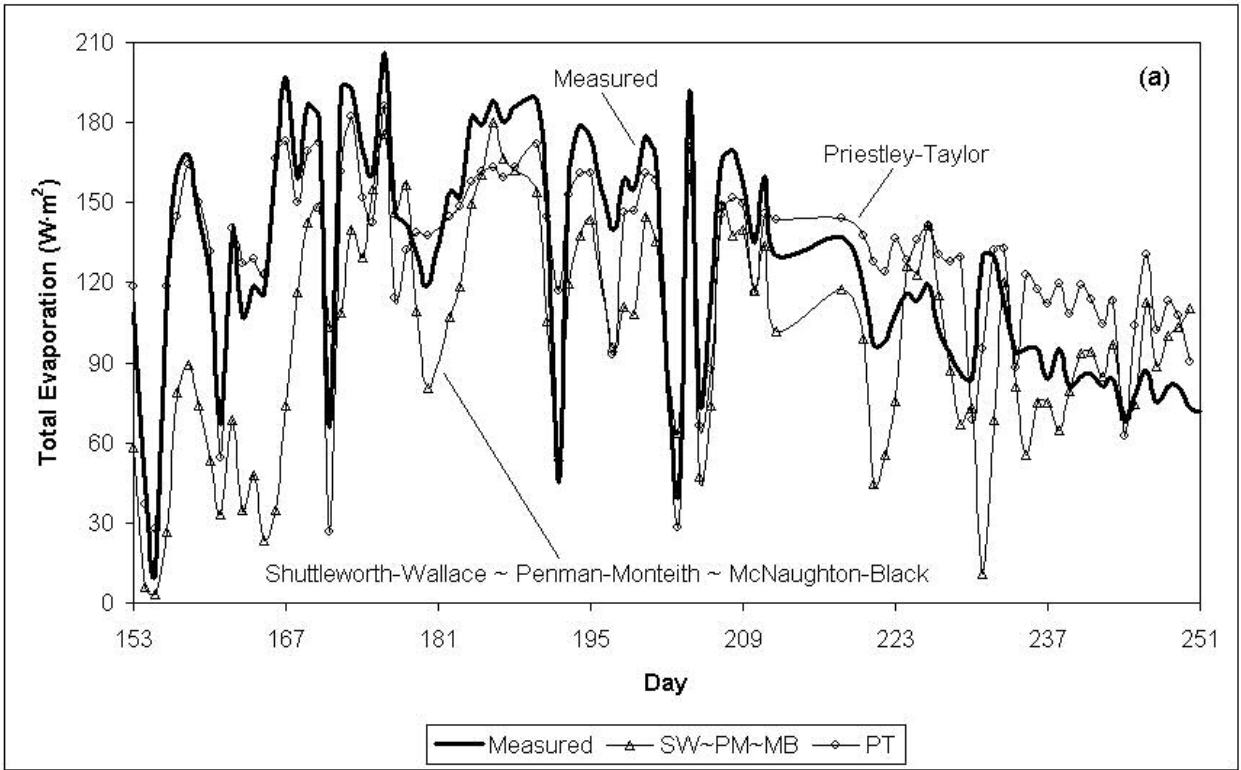


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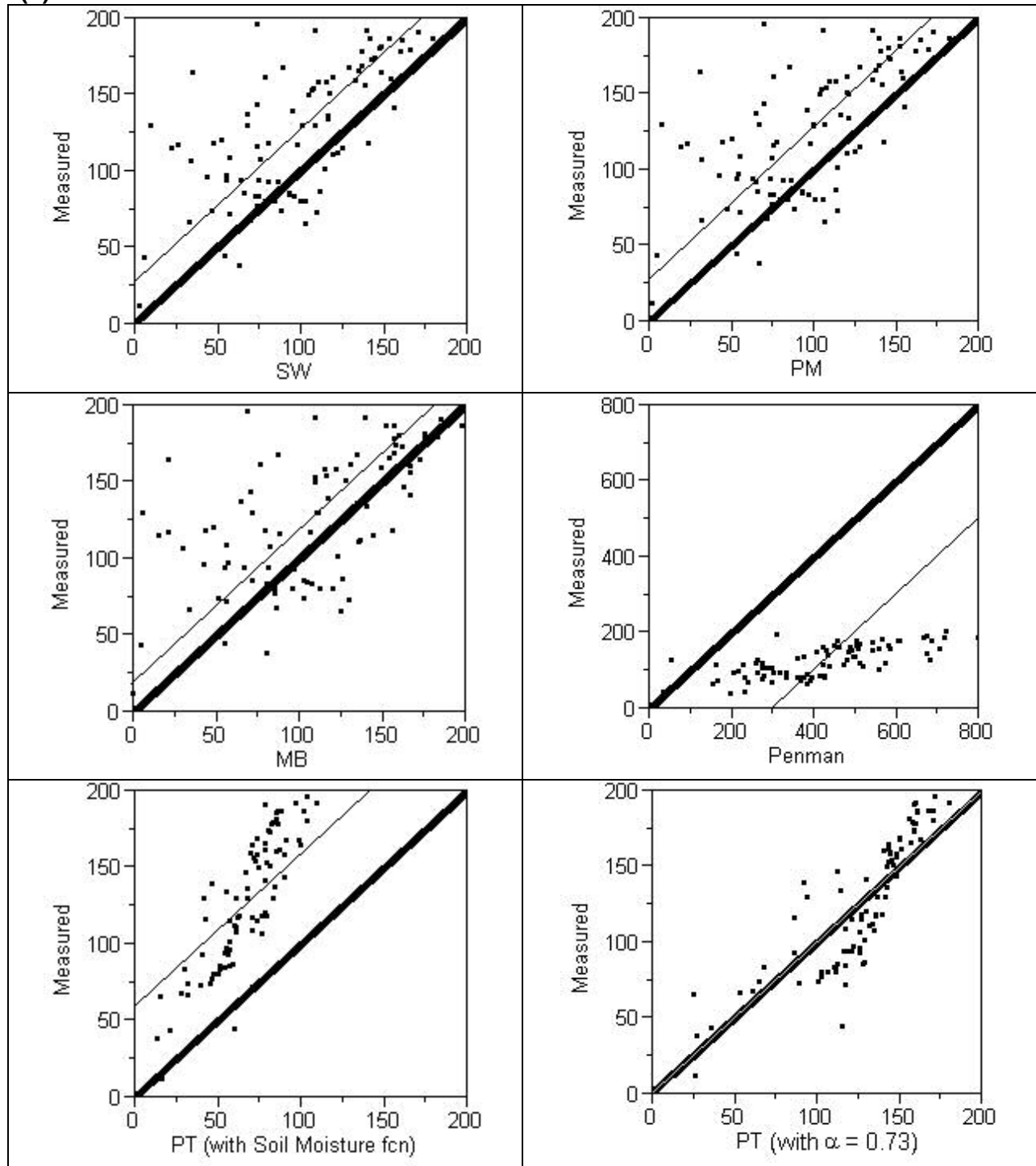
Figure 1.



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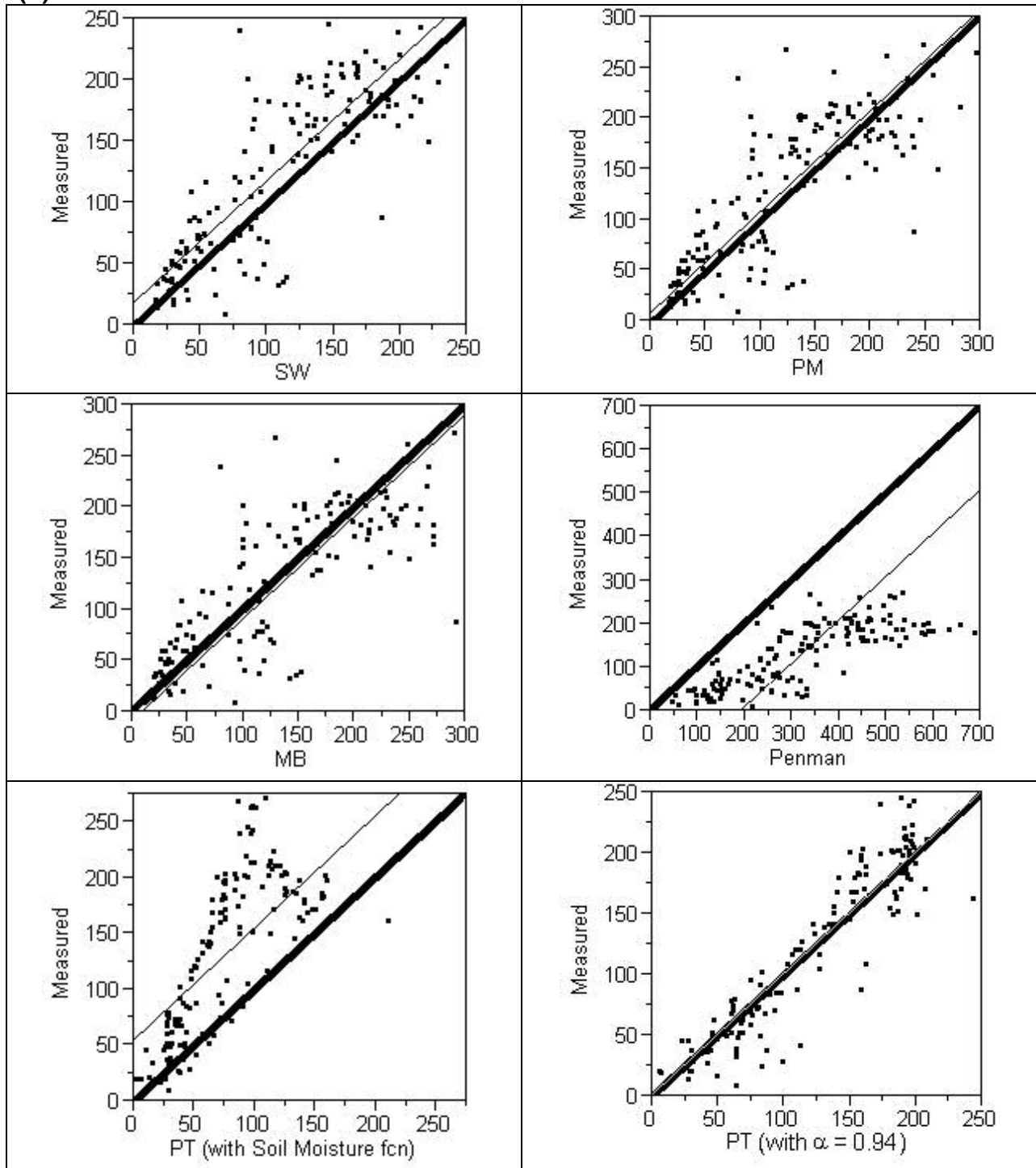
Figure 2.

1 (a) 1997



2
3

1 (b) 1998



2
3 Figure 3.
4

1 Table 1
 2 Comparison of the increasing complexity of the models in terms of number of parameters required. PT is
 3 Priestley-Taylor, MB is McNaughton-Black, PM is Penman-Monteith, and SW is Shuttleworth-Wallace.
 4

Parameter	Symbol	Units	PT	MB	Penman	PM	SW
Rate of change of vapor pressure with temperature	Δ	$\text{kPa}\cdot\text{K}^{-1}$	√		√	√	√
Total available energy	A	$\text{W}\cdot\text{m}^{-2}$	√		√	√	√
Psychrometric constant	γ	$\text{kPa}\cdot\text{K}^{-1}$	√	√	√	√	√
Air temperature	T_a	$^{\circ}\text{C}$	√	√	√	√	√
Specific heat at constant pressure	c_p	$\text{J}\cdot\text{kg}^{-1}\cdot\text{K}^{-1}$		√		√	√
Air density	ρ	$\text{kg}\cdot\text{m}^{-3}$		√	√	√	√
Vapor pressure deficit	D	kPa		√	√	√	√
Bulk stomatal resistance of the canopy	r_{cs}	$\text{s}\cdot\text{m}^{-1}$		√		√	√
Wind speed	u	$\text{m}\cdot\text{s}^{-1}$			√	√	√
Aerodynamic resistance above the canopy	r_{aa}	$\text{s}\cdot\text{m}^{-1}$				√	√
Bulk boundary layer resistance of the vegetation	r_{ca}	$\text{s}\cdot\text{m}^{-1}$					√
Aerodynamic resistance for substrate and canopy	r_{sa}	$\text{s}\cdot\text{m}^{-1}$					√
Surface resistance of the substrate	r_{ss}	$\text{s}\cdot\text{m}^{-1}$					√
Available soil energy	A_s	$\text{W}\cdot\text{m}^{-2}$					√

5
 6

1 Table 2
 2 Measured values of the Priestley-Taylor coefficient, α as tabled by Flint and Childs (1991) with our
 3 value.
 4

α	Surface conditions	Reference
1.57	Strongly advective conditions	(Jury and Tanner, 1975)
1.29	Grass (soil at field capacity)	(Mukammal and Neumann, 1977)
1.27	Irrigated ryegrass	(Davies and Allen, 1973)
1.26	Saturated surface	(Priestley and Taylor, 1972)
1.26	Open-water surface	(Priestley and Taylor, 1972)
1.26	Wet meadow	(Stewart and Rouse, 1977)
1.18	Wet Douglas-fir forest	(McNaughton and Black, 1973)
1.12	Short grass	(De Bruin and Holtslag, 1982)
1.05	Douglas-fir forest	(McNaughton and Black, 1973)
1.04	Bare soil surface	(Barton, 1979)
0.90	Mixed reforestation (water limited)	(Flint and Childs, 1991)
0.87	Ponderosa pine (water limited, daytime)	<i>This study</i>
0.84	Douglas-fir forest (unthinned)	(Black, 1979)
0.80	Douglas-fir forest (thinned)	(Black, 1979)
0.73	Douglas-fir forest (daytime)	(Giles et al., 1984)
0.72	Spruce forest (daytime)	(Shuttleworth and Calder, 1979)

5