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Larval ecology and bionomics of *Anopheles funestus* in highland and lowland sites in western Kenya

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Abstract

Background

An. funestus is a major Afrotropical vector of human malaria. This study sought to investigate the larval ecology, sporozoite infection rates and blood meal sources of *An. funestus* in western Kenya.

Methods

Larval surveys were carried out in Bungoma (Highland) and Kombewa (lowland) of western Kenya. Aquatic habitats were identified, characterized, georeferenced and carefully examined for mosquito larvae and predators. Indoor resting mosquitoes were sampled using pyrethrum spray catches. Adults and larvae were morphologically and molecularly identified to species. Sporozoite infections and blood meal sources were detected using real-time PCR and ELISA respectively.

Results

Of the 151 aquatic habitats assessed, 62/80 (78%) in Bungoma and 58/71(82%) in Kombewa were positive for mosquito larvae. Of the 3,193 larvae sampled, *An. funestus* larvae constitute 38% (1224/3193). Bungoma recorded a higher number of *An. funestus* larvae (85%, 95%, CI, 8.722–17.15) than Kombewa (15%, 95%, CI, 1.33–3.91). Molecular identification of larvae showed that 89% (n = 80) were *An. funestus*. Approximately 59%, 35% and 5% of *An. funestus* larvae co-existed with *An. gambiae s.l, Culex spp* and *An. coustani* in the same habitats respectively. Of 1,221 *An. funestus s.l* adults sampled, molecular

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identifications revealed that *An. funestus* constituted 87% (n = 201) and 88% (n = 179) in Bungoma and Kombewa, respectively. The *Plasmodium falciparum* sporozoite rate of *An. funestus* in Bungoma and Kombewa was 2% (3/174) and 1% (2/157), respectively, and the human blood index of *An. funestus* was 84% (48/57) and 89% (39/44) and for Bungoma and Kombewa, respectively.

Conclusion

Man-made ponds had the highest abundance of *An. funestus* larvae. Multiple regression and principal component analyses identified the distance to the nearest house as the key environmental factor associated with the abundance of *An. funestus* larvae in aquatic habitats. This study serves as a guide for the control of *An. funestus* and other mosquito species to complement existing vector control strategies.

Introduction

Malaria is still the most devastating vector-borne disease in sub-Saharan Africa, contributing approximately 215 million cases in 2019, which accounted for about 94% of all global cases [1]. Anti-vectorial programmes consisting mainly of long-lasting insecticide-treated nets and indoor residual spraying have contributed immensely towards reducing malaria incidence and mortality in malaria-,endemic areas of sub-Saharan Africa, which are characterized with having high entomological inoculation rates (EIR, infective bites per person per year) [1–3]. Kenya is noted among 17 countries estimated to have attained a reduction in malaria incidence in 2020 compared to 2015 [1]. Despite this unprecedented success in the fight against malaria, the Global Technical Strategy (GTS) 2020 milestone for reducing mortality and morbidity has not been achieved globally [1].

Currently, there is no "silver bullet" to successfully achieve the elimination and eradication goal outlined by the GTS. An important component of malaria vector control that needs reconsideration in the malaria elimination and eradication agenda is larval source management. Historical records have shown that a key component of malaria eradication efforts in Israel, Italy, and the United States of America was source reduction through larval habitat modifications [4]. In malaria-endemic areas of Africa, the use of insecticide-treated net, when combined with larval control, has been predicted to reduce the number of adult mosquito emergence by 50% and subsequently decrease the entomological inoculation rate (EIR) up to 15 to 25-fold [5]. However, the use of larval control strategies requires adequate knowledge of the larval ecology of the vectors, as well as better characterization of their breeding habitats in different ecological settings [6]. Targeting the most productive breeding habitats for larval control can be cost-effective, depending on the anopheline species, heterogeneity of aquatic habitats, and proximity to human dwellings [7,8].

Anopheles funestus sensu stricto (s.s) (hereafter, An. funestus) is a major Afrotropical vector of human malaria, exhibiting anthropophilic, and endophilic behaviours [9,10]. In western Kenya, An. funestus is one of the principal vectors of human malaria, owing to its high resistance to pyrethroids used for bed net impregnation, high sporozoite rate, and persistence in indoor malaria transmission [11,12]. Historically, western Kenya witnessed a decline in An. funestus populations after the introduction of insecticide-based control tools [13] and this reduction was observed for some time until a resurgence was reported a decade ago [12,14]. While this resurgence might be influencing malaria transmission in western Kenya, very few

studies have examined and characterized the larval habitats of *An. funestus*. Unlike reports on *An. gambiae* s.s. and *An. arabiensis* in this region, few studies have found *An. funestus* larvae in large permanent habitats with thick aquatic vegetation and algae [15,16].

Anopheles mosquitoes breed in a range of aquatic habitats and assessing findable habitats is key in controlling immature stages of vectors. However, locating breeding habitats of *An. funestus* is difficult, as it infrequently breeds in notable habitats with other malaria vectors. Hence, a better understanding and characterization of the breeding habitats and ovipositional behavioural patterns of *An. funestus* in endemic areas is crucial before larval source management can make malaria elimination and eradication feasible. Moreover, knowledge on *Plasmo-dium falciparum* infection rate and the feeding preference of *An. funestus* can help in predicting the intensity of malaria transmission. In this study, we characterized the aquatic habitats and the larval abundance of *An. funestus*. We also examined *Plasmodium falciparum* sporozoite infection rates and blood meal sources of indoor resting adults' mosquitoes in highland and lowland areas.

Materials and methods

Study sites

This study was conducted in two distinct locations, a highland town (Bungoma) and a lowland town (Kombewa), situated about 55 km apart in western Kenya (Fig 1).

Bungoma [00.54057°N, 034.56410°E, 1386–1,545m above sea level (asl)] is characterized by a perennial malaria outbreak. The mean annual rainfall and temperatures are 150 mm and 22.5°C respectively [17]. The area previously had *Plasmodium falciparum* malaria episodes > 45% with a hospitalization rate up to 55% [18]. Agricultural production of crops (tobacco, cereals, sugar cane, onions and other vegetables) and raising of farm animals (cattle, sheep and goat) and poultry form the backbone of the rural economy of Bungoma. The principal vectors of human malaria in Bungoma are *An. funestus*, *An. gambiae* and *An. arabiensis* [19,20].

Kombewa (34⁰ 30'E, 0⁰ 07'N, 1150–1300 m asl) is located in Kisumu County, a lowland area with slow water drainage, in the Lake Victoria basin. This area experiences two rainy seasons: a long season from March to May characterized by peak malaria transmission and a short season between October and November. The mean annual rainfall and temperatures are 1200 mm-1300 mm and 20°C—35°C respectively [21]. There is, however, yearly variation in the rainfall pattern in the region. It is a hyperendemic malaria zone with a *P. falciparum* parasitemia rate of 57.5% [12] and EIR of 31.1 infective bites per person per year [11]. *An. funestus* has been reported to be the principal malaria vector predominating in this study area [11,22].

Larval sampling

Larval surveys were carried out monthly from November 2019 to November 2020. Breeding sites were identified within 2km of the study villages. Larvae were sampled using standard dippers (350 ml) and 10L bucket [23]. A maximum of 20 dips was taken at each habitat to identify the presence or absence of larvae and aquatic predators. Mosquito larvae were sampled along the edges of the habitats after waiting for 3–5 minutes for larvae to rise to the water surface if there were any. *An. funestus sensu lato (s.l)* larvae were preserved in absolute ethanol for subsequent identification using molecular analysis. Larvae were morphologically identified following referenced keys [24,25]. All breeding sites were georeferenced using a global positioning system installed in ODK software in Android Samsung tablet (Version SAM-T380). Larval habitats variables were characterized following a questionnaire that was imputed using ODK software.

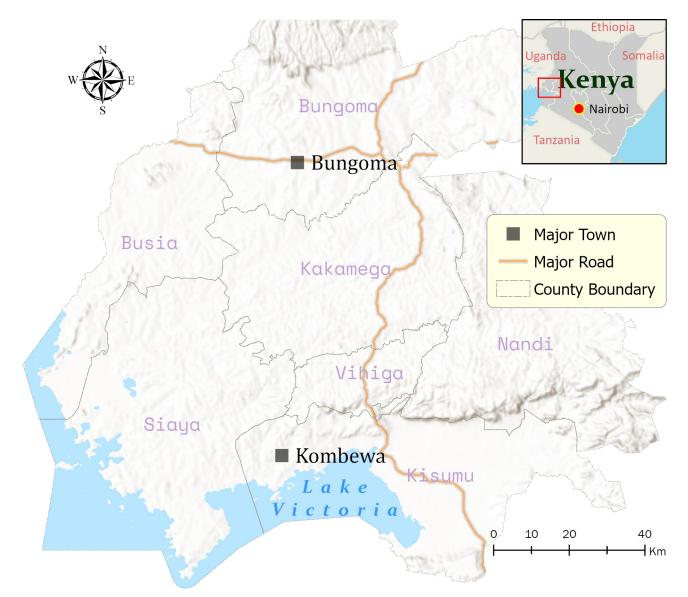


Fig 1. Map of the study areas in western Kenya. The map was generated using ArcGIS Pro 2.6 software. Map source: ESRI, CGIAR, and USGS (available at: www.esri.com).

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Characterization of aquatic habitats

All potential breeding sites for mosquito larvae were identified and classified as: man-made pond, natural pond/rain pool, drainage ditch, swamp/marshes and tyre tracks. Man-made ponds/pits were made purposely for moulding clay pots in Bungoma, whereas in Kombewa, they were dug for sand winning. Aquatic habitats were classified as permanent if they could hold water for more than three weeks. Environmental variables including the presence of vegetation, category of vegetation, the height of vegetation, vegetation coverage, substrate type, and distance to a nearest house, habitat dimensions (length, width and depth), water physics (flow status and clarity) and present or absent of aquatic predators were recorded for each water body. In addition, the surrounding land use type (cultivated land/cropland, grassland/pasture, wetland/swamp and road) were also recorded.

The vegetation cover was classified as: emergent, freely floating, submerged, and no vegetation. The height of vegetation coverage was measured using a meter stick and grouped into < 5, 6-10, and > 10 m. The distance to the nearest dwellings was estimated and grouped as (A) < 100 m, (B) 100–200 m, and (C) 201–500 m. The types of plants and predators in the habitats were identified using picture charts. The dimensions (length, width and depth) of each habitat were measured using a meter stick, and the average depth was recorded. An Aquafluor[™] meter (Model: 8000–010, Turner Design, San Jose, CA, USA) was used to measure cyanobacteria (blue-green algae) levels in water samples from the aquatic habitats.

The clarity of the water was observed and classified as (A) clear (transparent like a glass), (B) opaque (not transparent and impenetrable to light), (C) cloudy (normally appeared white in color) and (D) muddy/brownish (brown in color due to disturbance stirring its deposits). The substrate type was classified as (A) mud/dirty, (B) sand and (C) stone. Land-use types for the surrounding habitats were described as (A) cultivated land/cropland (farmland use for cultivation of food crops and other crops), (B) grassland/pasture (lands with suitable grasses for grazing animals) and (C) wetland/swamp (surroundings characterized by mostly aquatic plants species covered by water/ low-lying ground not cultivated and covered with water) and (D) road (land meant for passage of vehicles and people).

Adult sampling

To find out whether larval abundance correlated with the adult *Anopheles* density, adult mosquitoes were collected indoors using the pyrethrum spray catches [26] from randomly selected houses near the aquatic habitats from November 2019 to August 2020. The adult collection was carried out every two months during the sampling period. There was an equal distribution of houses in the study sites so sixty (60) houses were randomly selected near the larval habitats at each study site during each field visit. Houses were selected based on the presence of residents in the house, permission to get access into the indoor living rooms and proximity to nearby aquatic habitats. Collections were carried out in the morning hours between 06:30 to 10:00 h [27]. The physiological status of the anophelines was visually classified into blood-fed (abdomen is dilated and bright red), unfed (empty abdomen/no blood meal), half gravid (the abdomen is whitish posteriorly and dark reddish anteriorly) and gravid (dilated and whitish abdomen) [28]. Samples were stored in 1.5 ml Eppendorf tubes containing silica gel desiccant and cotton wool at -20° C in the Sub-Saharan Africa International Centre of Excellence for Malaria Research, Tom Mboya University College, Homa Bay, Kenya, for further analysis.

DNA extraction and molecular identification of species

Adult mosquitoes were cut into two parts to separate the head and thorax from the abdomen. DNA was extracted from the head and the thorax using the Chelex[®]-100 method [29], while the abdomen was preserved for blood meal analysis. DNA was extracted from the larvae of *An*. *funestus s.l.* using the ethanol precipitation method [30]. Molecular identification of sibling species of the *An*. *funestus* group was performed using multiplex polymerase chain reaction (PCR) by amplifying the polymorphic ITS2 region of ribosomal DNA using species-specific primers for *An*. *funestus*, *An*. *rivulorum*, *An*. *vaneedeni*, *An*. *parensis*, *An*. *leesoni*, *An*. *rivulorum-like* by following already developed protocols [31,32]. A sub-sample of 641 (551 adults and 90 larvae) were identified using PCR. Of this number, 20% (n = 110) of the adult mosquitoes and 11% (n = 10) of the larvae failed to amplify. Of the *An*. *funestus s.l* specimens that failed to amplify by PCR after three attempts, 65 (60 adults and 5 larvae) were randomly selected and sent to the University of California, Irvine, for sequencing using the Sanger sequencing method. The ITS2 region of nuclear ribosomal DNA was amplified using the

forward primer ITS2A (TGTGAACTGCAGGACACAT) and the reverse primer ITS2B (TATGCTTAAATTCAGGGGGT); amplicons were sequenced using ABI Big Dye Terminator Cycle Sequencing Kits, as described by Daibin et al., [33].

Plasmodia species genotyping using multiplex RT-PCR

A modified TaqMan assay was used to detect Plasmodia species (P. falciparum, P. ovale and P. malarae) infections in the DNA samples as previously described [34,35]. Plasmodium falciparum species-specific 18S ribosomal RNA primers (P. falciparum forward primer 5-ATTGCT TTTGAGAGGTTTTGTTACTTT-3 and reverse primer 5-GCTGTAGTATTCAAACACAATGAA CTCAA-3, P. malariae forward primer 5-AGTTAAGGGAGTGAAGACGATCAGA-3 and reverse primer 5-CAACCCAAAGACTTTGATTTCTCATAA-3 and P. ovale forward and reverse primer 5-AACCCAAAGACTTTGATTTCTCATAA-3 and 5-CCGACTAGGTTTTGGATGAAAGATTTT T-3) and probes (FAM- CATAACAGACGGGTAGTCAT-MGB for P. falciparum, VIC-ATGAGT GTTTCTTTTAGATAGC-MGB for P. malariae and (NED- CGAAAGGAATTTTCTTATT-MGB for P. ovale) were used. A final volume of 12µl containing 2 µl of sample DNA, 6 µl of PerfeCTa® qPCR ToughMix[™], Low ROX[™] Master mix (2X), 0.5 µl of each probe, 0.4 µl of each forward primers (10 μ M), 0.4 μ l of each reverse primers (10 μ M) and 0.1 μ l of double-distilled water was loaded in QuantStudio[™] 3 Real-Time PCR System. The thermal profile used was set as follows; 50°C for 2 min, (95°C for 2 min, 95°C for 3 sec and 58°C for 30 sec) for 45 cycles. The real-time PCR was performed in the QuantStudio 3 real-PCR instrument (Applied Biosystems, Thermo Fisher Scientific). The *Plasmodium* infection was analyzed by comparing melt curves with the positive controls using QuantSudio TM Design and Analysis Desktop software v1.5.1.

Blood meal analysis

Direct enzyme-linked immunosorbent assays (ELISA) was used to detect the origin of blood meal in the abdomen of blood-fed *Anopheles* mosquitoes following existing protocol [36]. Briefly, 50µl of phosphate-buffered saline (PBS) was added to the abdomen of each mosquito specimen and was incubated overnight. Following grinding, 950µl of PBS was added to each sample for washing and 50µl of each sample was loaded into each of the wells of the ELISA plate and incubated for two hours. Hosts specific positive controls and negative control from unfed lab-raised mosquitoes were also added. Following incubation and washing, anti-host specific conjugates (antibodies) against human, goat, chicken and dog proteins were added to the wells. ABTS (2, 2'-Azinobis [3-ethylbenzothiazoline-6-sulfonic acid]-diammonium salt) peroxidase substrate for each human, goat, chicken and dog blood meal source was added to each of the wells. The samples were incubated for thirty minutes for the reaction to occur. The non-reacting samples were tested subsequently using bovine immunoglobulin G. All ELISA results were read visually [37].

Ethical statement

This study was approved by Maseno University's Ethics Review Committee (MUERC/00778/ 19). Verbal consent was sought and granted from heads of households and owners of farmlands where the adult mosquitoes and larvae were sampled.

Data analysis

Data were analyzed using SPSS software (Version 21 for Windows, SPSS Inc., Chicago, IL) and Graph Pad Prism V.8.0.1. The relative abundance of each species was expressed as a percentage of the number of larvae per species divided by the total number of larvae collected for

all species combined per habitat type. The relative abundance of *An. funestus* was calculated as the number of *An. funestus* larvae from a specific habitat divided by the total number of *An. funestus* larvae in all samples from a habitat type.

The types of breeding sites, the number of mosquito larvae sampled and species, and the number of aquatic predators were presented in tables and figures. The Kruskal-Wallis test was used to compare the number of larvae in different habitat types and for samples having more than two groups: habitat type (man-made pond, natural pond/rain pool, drainage ditch, swamp/marshes, tyre tracks), distance to the nearest house (<100, 100-200, 201-500), water clarity (clear, opaque, brownish/muddy, cloudy), aquatic plant species in a habitat [Pennisetum purpureum (elephant grass), Schoenoplectus californicus, others/unknown], and land-use type (cultivated land/cropland, grassland/pasture, wetland/swamp, road). The Mann-Whitney U test was used to compare samples with two variables: vegetation (present or absent), category of vegetation (emergent or non-emergent), water flow status (stagnant or flowing water), and predators (present or absent). Multiple regression and principal components analyses were used to identify environmental variables associated with the abundance of An. funestus in the aquatic habitats. The human blood index (HBI) was calculated as the percentage of Anopheles mosquitoes that fed on humans over the total number of blood-fed Anopheles for which the blood meal origins were determined. The sporozoite rate of P. falciparum was calculated as the proportion of Anopheles tested for sporozoites over the total genotyped. Spearman's correlation (r_s) was used to find the relationship between the adult An. funestus population and the larval density at each study site. The results were considered statistically significant at P < 0.05. For the sequence data analysis of the An. funestus s.l. specimen that failed to amplify by PCR, the de novo assembly of reads was performed using geneious software [38]. Basic Local Alignment Search Tool (BLAST) was used to identify sequence similarities against sequences in GeneBank. Sequences with high identity scores or low E-value were retrieved and used in the construction of a phylogenetic tree to identify the unknown. Evolutionary analyses were conducted in MEGA X after basic alignment using ClustalW algorithm [39]. All sequences of ITS2 are available at GenBank under accession numbers MZ435355-MZ435414.

Results

Distribution of *An. funestus* larvae and other mosquito larvae in Bungoma and Kombewa

A total of 151 potential mosquito aquatic habitats were assessed. Of these, 62/80 (78%; 95% CI: 0.68–0.87) and 58/71 (82%; 95% CI: 0.73–0.91) in Bungoma and Kombewa were positive for mosquito larvae, respectively. The number of aquatic habitats with *An. funestus* larvae in Bungoma was 55/80 (69%; 95% CI; 0.58–0.79), whereas 23/71 (32%; 95% CI: 0.21–0.43) were in Kombewa. In all, *An. funestus* larval habitats constituted 65% (n = 78) of the mosquito-positive habitats, whereas *An. gambiae s.l, An. coustani*, and *Culex spp* positive habitats made up of 57% (n = 69), 14% (n = 17) and 36% (n = 43), respectively.

A total of 3,193 mosquito larvae (39% *An. funestus*, 30% *An. gambiae* s.l., 28% *Culex* spp. and 3% *An. coustani*) were collected from the various habitats in Bungoma and Kombewa. Bungoma had a higher number of *An. funestus* larvae (85% 95%, CI, 8.722–17.15) than Kombewa (15% 95%, CI, 1.33–3.91). Approximately 59% of *An. funestus* larvae were collected from breeding sites with co-existing *An. gambiae* s.l larvae in the same habitats (Fig 2). Similarly, 35% and 5% of *An. funestus* larvae were found co-existing with *Culex spp.*, and *An. coustani* larvae, respectively (Fig 2). No pupae were identified during the sampling. Molecular results on the sibling species of the *An. funestus* group larvae revealed that 89% (n = 80) were *An. funestus*, 11% (n = 9) were *An. rivulorum* and 1% (n = 1) was *An. sp.9*.

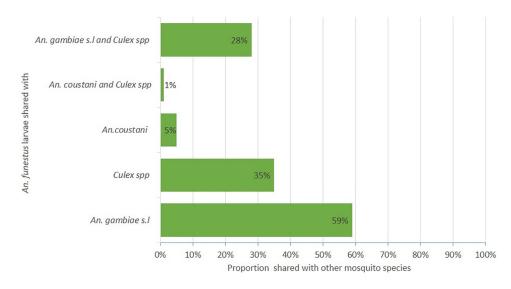


Fig 2. Proportion of *An. funestus* larvae shared with other mosquitoes in the larval habitats.

https://doi.org/10.1371/journal.pone.0255321.g002

Characteristics of mosquito larval habitats

An. funestus larvae were found in various habitats, with or without vegetation: man-made ponds, natural ponds/rain pools, drainage ditches, swamp/marshes, and tyre tracks. There were no significant differences in *An. funestus* larval density between the various habitat types of *An. funestus* ($\chi 2 = 8.917$, df = 4, *P* = 0.063). However, man-made ponds comprised the highest number of *An. funestus* positive habitats (36%, n = 28) and had the highest proportion of larvae with *An. funestus* in Bungoma (53%, n = 553) and in Kombewa (61%, n = 115) (Table 1). The larval abundance of *An. funestus* in man-made ponds, natural ponds/rain pools and drainage ditches was significantly different between the study sites (*P*< 0.05) (Table 1). Field observations showed that man-made ponds constituted the main permanent aquatic habitat type in the study areas. This was followed by swamp/marshes, natural ponds/rain

Study site	Habitat		Larval Counts per Mosquito Species (n (%))					
	Туре	N	An. funestus s.l	An. gambiae s.l	An. coustani	Culex spp		
Bungoma	Man-made ponds	26	553 (53) ^a	81(37)	0(0)	15(47) ⁱ		
	Natural ponds/rain pools	12	100 (10) ^c	2(1) ^g	0(0)	0(0)		
	Drainage ditches	19	374(36) ^e	120(54)	0(0)	3(9) ^k		
	Swamps/marshes	21	8(1)	18(8)	0(0)	14(44)		
	Tyre tracks	2	0(0)	0(0)	0(0)	0(0)		
Kombewa	Man-made ponds	20	115(61) ^b	67(9)	33(32)	162(19) ^j		
	Natural ponds/rain pools	11	25(13) ^d	249(34) ^h	24(23)	188(21)		
	Drainage ditches	17	40(21) ^f	286(39)	47(45)	489(56) ¹		
	Swamp/marshes	22	9(5)	137(18)	0(0)	34(4)		
	Tyre tracks	1	0(0)	0(0)	0(0)	0(0)		
Bungoma	Total	80	1035	221	0	32		
Kombewa	Total	71	189	739	104	873		

Table 1. Number of different mosquito species collected from aquatic habitats in Bungoma and Kombewa in Kenya in the months of November 2019 to November 2020.

N is the number of habitats and n is the number of mosquitoes of different species in each habitat type. a and b, c and d, e and f, g and h, i and j, k and l superscripts indicate statistical significance at P<0.05 between the number of mosquito larvae in each habitat type between the study sites.

https://doi.org/10.1371/journal.pone.0255321.t001

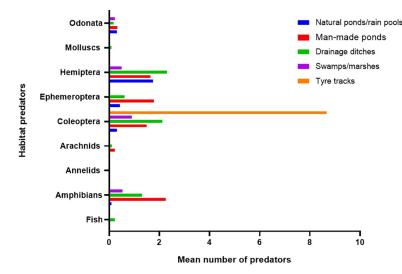


Fig 3. Distribution of aquatic predators in the various habitats. Odonata: damselfly/dragonfly, Molluscs: snails/ slugs/mussels, Hemiptera: backswimmer/giant water bugs/cacidas, Ephemeroptera: mayfly, Coleoptera: water beetles/ weevils, Arachnids: spiders/ticks/mites, Annelids: Segmented worms, Amphibians: frogs/toads/tadpoles, Fish: tilapia.

https://doi.org/10.1371/journal.pone.0255321.g003

pools, drainage ditches and tyre tracks, at frequencies of 27%, 19%, 15%, and 3%, respectively. There was no significant difference in the means among the various predators found in the larval habitats (P = 0.05). Fig 3 shows the mean distribution of the various predators.

Association between environmental variables and *An. funestus* larval abundance

The various environmental variables in the larval habitats associated with the presence of *An*. *funestus* larvae and other mosquito species are summarized in Table 2.

Multiple regression analysis models revealed that distance to the nearest house (P = 0.0122) was the best predictor of *An. funestus* larval density in the habitats (Table 3). The F-ratio in the ANOVA table revealed that, the *An. funestus* larval density was significantly associated statistically with key habitat variables (habitat type, land use type, vegetation coverage, vegetation height, distance to house, habitat size, water depth, water clarity, algae abundance and predator counts) at the time of larval sampling (F = 2.06, d.f = 10, 114, P = 0.033, $R^2 = 0.153$). Furthermore, principal component analysis identified the distance to the nearest house as the key environmental factor associated with the abundance of *An. funestus* larvae (S1 Fig).

The Mann–Whitney U and Kruskal–Wallis tests showed that *An. funestus* larval density had no significant difference between present and absent of vegetation (U = 1192, *P* = 0.978) (Fig 4A), emergent and non-emergent vegetation (U = 1369, *P* = 0.166) (Fig 4A), stagnant and flowing water (U = 287, *P* = 0.136) (Fig 5C), present and absent of aquatic predators (U = 2215, *P* = 0.162) (Fig 5B), among clear, opaque, cloudy and muddy/brownish water (χ 2 = 7.316, df = 3, *P* = 0.062) (Fig 4C), or with the presence of *Pennisetum purpureum*, *Schoenoplectus californiscus* and other aquatic plant species in the habitats (χ 2 = 2.671, df = 2, *P* = 0.263) (Fig 4B) respectively. However, *An. funestus* larval density showed statistically significant differences associated with distances to the nearest house (<100, 100–200 and 201– 500 m) (χ 2 = 25.138, df = 2, *P*<0.0001) (Fig 5A) and land-use types (cultivated land/cropland, grassland/pasture, wetland/swamp and road) (χ 2 = 29.197, df = 3, *P* = 0.000) (Fig 4C). Our field observation showed that *An. funestus is* most abundant in habitats surrounded by cultivated land compared to those in the grassland areas, wetlands, and roads.

Environmental variable	All aquatic breeding sites n (%)	Potential habitats without mosquito larvae n (%)	Aquatic habitats with <i>An. funestus</i> n (%)	Aquatic habitats with An. gambiae s.l n (%)	Aquatic habitats with An. coustani n (%)	Aquatic habitats with <i>Culex spp</i> n (%)
Habitat type		. ,				
Man-made pond	46 (31)	6 (19)	28(36)	13(19)	6(35)	12(28)
Natural pond/rain pool	23(15)	3 (10)	15(19)	13(19)	4(24)	6(14)
Drainage ditch	36(24)	12(39)	12(15)	19(27)	1(6)	8(19)
Swamp/marshes	43(28)	10(32)	21(27)	22(32)	6 (35)	16(37)
Tyre tracks	3(2)	0(0)	2(3)	2(3)	0(0)	1(2)
Vegetation						
Present	133(88)	28 (90)	67(86)	63(91)	17(100)	41(95)
Absent	18(12)	3 (10)	11(14)	6(9)	0(0)	2(5)
Category of vegetation						
Emergent	101(76)	19(68)	48(72)	53(84)	14(82)	34(83)
Free floating	29(22)	9(32)	16(24)	8(13)	3(18)	5(12)
Submerged	3(2)	0(0)	3(4)	2(3)	0(0)	2(5)
Water Flow Status						
Stagnant/standing water	145(96)	31(100)	74(95)	64(93)	17(100)	40(93)
Flowing/Disturbed Water	6(4)	0(0)	4(5)	5(7)	0	3(7)
Predators						
Present	100(66)	12(39)	54(69)	56(81)	16(94)	35(81)
Absent	51(34)	19(61)	24(31)	13(19)	1(6)	8(19)
Distance to Nearest House (m)						
<100	39(26)	17(55)	12(15)	11(16)	4(24)	10(23)
100-200	76(50)	13(42)	38(49)	34(49)	8(47)	25(58)
201-500	36(24)	1(3)	28(36)	24(35)	5(29)	8(19)
Water Clarity						
Clear	43(28)	11(35)	26(33)	17(25)	6(35)	11(25)
Opaque	69(46)	13(42)	39(50)	27(39)	9(53)	18(42)
Cloudy	30(20)	4(13)	10(13)	21(30)	2(12)	12(28)
Muddy/Brownish	9(6)	3(10)	3(4)	4(6)	0(0)	2(5)
Land-use type						
Cultivated land/ cropland	82(54)	18(58)	56(72)	36(52)	0(0)	13(30)
Grassland/Pasture	61(40)	11(36)	18(23)	31(45)	16(94)	27(63)
Wetland/swamp	4(3)	1(3)	2(2.5)	2(3)	0(0)	0(0)
Road	4(3)	1(3)	2(2.5)	0(0)	1(6)	3(7)

Table 2. Characteristics of aquatic habitats of An. funestus and other mosquito species at Bungoma and Kombewa western Kenya (November 2019-November, 2020).

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Relationship between larval abundance and adults mosquitoes sampled indoors

The results of the Pearson correlation test showed that there was a statistically significant but weakly positive linear relationship between *An. funestus* larval abundance and the presence of adult mosquitoes collected indoors in Bungoma (rs = 0.178, *P* = 0.026). However, there was no

Habitat variable	Beta	Std.Err. of Beta	В	Std.Err. of B	t(114)	P-level
Habitat Type	-0.0876	0.0984	-0.0613	0.0688	-0.8902	0.3752
Land use type	-0.1816	0.1047	-0.2464	0.1420	-1.7348	0.0855
Vegetation coverage	0.0957	0.0997	0.0024	0.0025	0.9598	0.3392
Vegetation Height	-0.0219	0.0890	-0.0058	0.0236	-0.2466	0.8056
Distance to house	0.2355	0.0925	0.0016	0.0006	2.5468	*0.0122
Habitat size	-0.0594	0.0929	-0.0006	0.0009	-0.6391	0.5240
Water depth	0.1088	0.0894	0.2402	0.1975	1.2163	0.2264
Water clarity	0.0171	0.1065	0.0033	0.0207	0.1605	0.8728
Algae abundance	-0.1588	0.0980	-0.0014	0.0008	-1.6193	0.1081
Predator counts	-0.0568	0.0909	-0.0045	0.0072	-0.6254	0.5329

Table 3. Multiple regression analysis of factors associated with An. funestus larval density.

Beta standardized coefficient, B the unstandardized coefficient value, * significant at P<0.05.

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significant relationship between *An. funestus* larval abundance and adult mosquitoes sampled indoors in Kombewa (rs = 0.003, P = 0.972).

Composition of indoor resting Anopheles mosquitoes

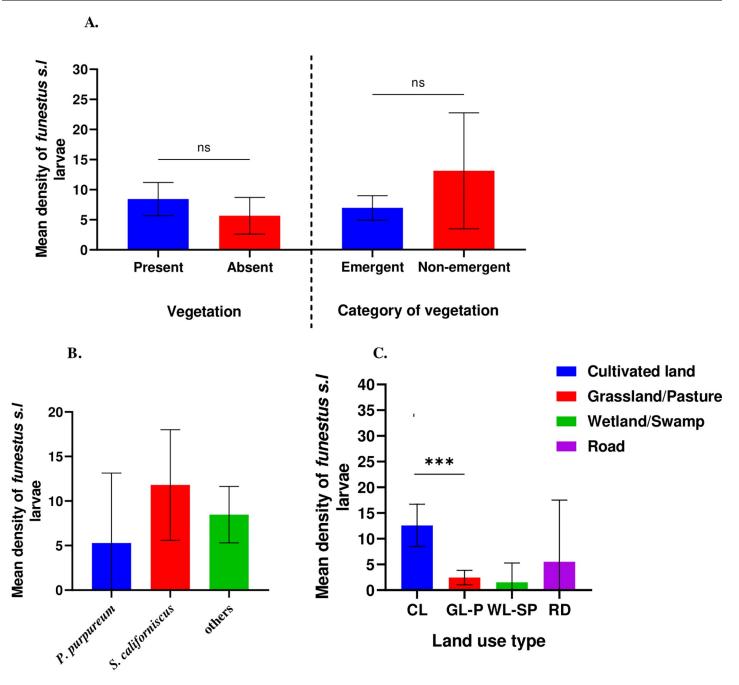
A total of 1,221 *An. funestus s.l* and 195 *An. gambiae s.l* were collected indoors in Bungoma and Kombewa. Of the 1,221 *An. funestus s.l* sampled, 46% (n = 565) and 54% (n = 656) were collected from Bungoma and Kombewa respectively. For molecular identification, a sub-sample of 551 *Anopheles* comprising 380 *An. funestus s.l* and 171 *An. gambiae s.l* from the study sites were analysed. Of the *An. funestus s.l* analysed, 201 and 179 were from Bungoma and Kombewa, respectively. Our results revealed that *An. funestus* predominated in the two study areas: *An. funestus* constituted 87% (n = 201) and 88% (n = 179) in Bungoma and Kombewa, respectively.

Of the 171 *An. gambiae* s.l. analysed, 76 and 95 were from Bungoma and Kombewa, respectively. *An. gambiae* was the main species identified in Bungoma (86%, n = 76) and Kombewa (81%, n = 95). *An. arabiensis* made up 15% (n = 76) and 19% (n = 95) in Bungoma and Kombewa, respectively. The evolutionary relationship of the unamplified sequenced data is shown in Fig.6.

The evolutionary history was inferred using the Neighbor-Joining method [40]. The optimal tree with the sum of branch length = 3.05389016 is shown (Fig 6). The percentage of replicate trees in which the associated taxa clustered together in the bootstrap test (1000 replicates) are shown next to the branches [41]. The evolutionary distances were computed using the Kimura 2-parameter method and are in the units of the number of base substitutions per site. The rate variation among sites was modelled with a gamma distribution (shape parameter = 1). This analysis involved 15 nucleotide sequences. All ambiguous positions were removed for each sequence pair (pairwise deletion option). There was a total of 815 positions in the final dataset. Evolutionary analyses were conducted in MEGA X [42].

Plasmodium falciparum (Pf) sporozoite infection rate

All the sub-samples (n = 551) subjected to molecular identification were genotyped to detect sporozoite. The *Pf* sporozoite rate of *An. funestus* in Bungoma and Kombewa was 2% (3/174) and 1% (2/157), respectively (Table 4). However, none of the *An. rivulorum* found in Bungoma and Kombewa was positive for *Pf* sporozoite (Table 4). Only one *An. gambiae* mosquito from



Aquatic plant species in habitat

Fig 4. Association of habitat variables with An. funestus density (A) vegetation (present and absent) and category of vegetation (emergent and non-emergent), (B) aquatic species in the habitats, (C) land use type.

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Kombewa (1%, 1/77) tested positive for *Pf* sporozoite. However, sporozoites were not detected in *An. arabiensis* in the study sites.

Anopheles blood meal origins

A total of 208 samples (Bungoma, n = 114 and Kombewa, n = 94) were analysed for the origin of the mosquito blood meals. The HBI of *An. funestsus* was 84% (48/57) and 89%

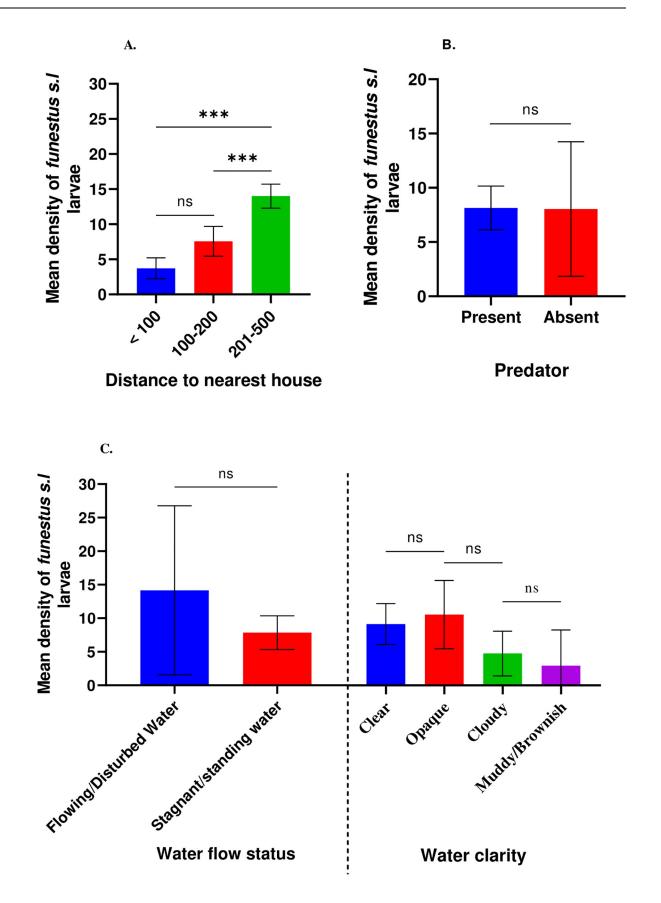


Fig 5. Association of habitat variables with *An. funestus* larval density (A) distance to the nearest house, (B) predator (present and absent in habitats) and (C) water flow status and water clarity.

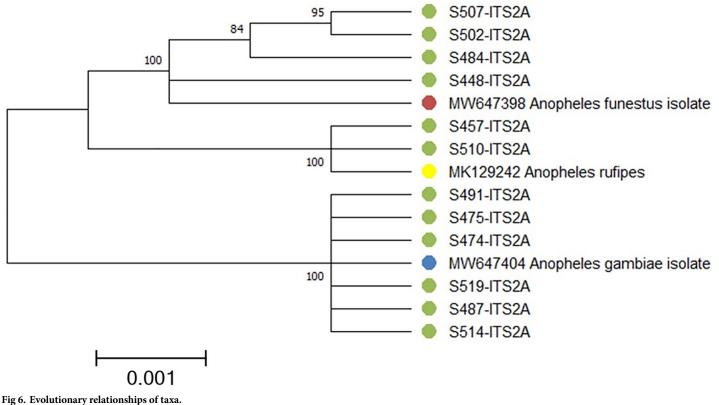
https://doi.org/10.1371/journal.pone.0255321.g005

(39/44) for Bungoma and Kombewa, respectively (Table 5). Table 5 shows the HBI for all species tested.

Discussion

The great diversities of anopheline larval habitats in addition to their inaccessibility makes larval ecology studies of malaria mosquitoes methodologically cumbersome [15]. The presence of quality larval habitats is significant in determining the abundance and distribution of adult mosquitoes. This study was designed to add to the limited amount of information on the larval ecology of *An. funestus* in western Kenya. Two study areas were selected, a highland site (Bungoma) and a lowland site (Kombewa), and their aquatic habitats were examined and characterized to determine if there have been changes in the breeding habitats of *An. funestus* in the village sites. This study revealed that *An. funestus* is a major vector influencing malaria transmission in the region, confirming a previous report that *An. funestus* has re-emerged and could be responsible for malaria transmission in western Kenya [12].

Our findings revealed that *An. funestus* larvae thrive in a wide range of aquatic habitats and co-breeds with other malaria vectors in the same habitats. Although there were no significant differences observed in the various habitats types, man-made ponds had the highest proportion of *An. funestus* larvae. Man-made ponds, created mostly for making clay pots and sand winning, contributed remarkably to the proportion of *An. funestus* habitats and larvae



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Site	An. gambiae s.l	No. tested	<i>Pf</i> +ve (%)	An. funestus s.l	No. tested	<i>Pf</i> +ve (%)
Bungoma	An. gambiae	65	0 (0)	An. funestus	174	3 (2)
	An. arabiensis	11	0 (0)	An. rivulorum	27	0 (0)
	Total	76			201	
Kombewa	An. gambiae	77	1 (1)	An. funestus	157	2 (1)
	An. arabiensis	18	0 (0)	An. rivulorum	22	0 (0)
	Total	95			179	

Table 4. Sporozoite rate of Anopheles mosquitoes sampled from indoors in Bungoma and Kombewa, November 2019 to November 2020.

Numbers in the bracket indicate sporozoite rate; Pf, Plasmodium falciparum, +ve, positive.

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abundance in the study areas. This corroborates previous findings in western Kenya where man-made habitats accounted for an increase in populations of *An. gambiae* [43,44]. Field observations have shown that man-made ponds are permanent habitats that hold water for a longer period compared to other habitat types. This suggests that malaria transmission in the study areas is partly man-made, and thus, proper environmental management specifically through habitat manipulation could curtail malaria transmission by major vectors of human malaria. This study confirms how anthropogenic modification of ecosystems can contribute greatly to the abundance and distribution of malaria vectors [45].

We found that more than 50% of *An. funestus* larvae co-existed in aquatic habitats with *An. gambiae s.l* larvae. Moreover, *An. funestus* shared the same habitats with *Culex spp* and *An. coustani*. Previous studies in neighbouring countries, Tanzania [46] and Uganda [47] have reported that *An. funestus* shared habitats with other *Anopheles* and *Culex spp.*, indicating that any larval control programme targeting *An. funestus* would have a profound effect in controlling other equally important vectors of human malaria and other mosquito-borne diseases.

Hitherto, it has been reported that *An. funestus* prefers breeding in aquatic habitats with thick vegetation [15,16] but this study revealed that *An. funestus* can breed in habitats with aquatic vegetation or without vegetation. The presence of vegetation has been noted to be an important environmental variable associated with *Anopheles* mosquito larvae density [48]. For example, aquatic macrophytes play important role in the oviposition, larval survival and development of anophelines as they serve as a food source, protection for the larval stages and provide enabling environment for mosquito breeding [49–52]. However, this study revealed that

Table 5. Blood meal sources of Anopheles mosquitoes sampled from indoors in Bungoma and Kombewa, November 2019 to November 2020.

Site	Species	No. tested		Blood meal sources		HBI
			human	bovine	goat	
Bungoma	An. gambiae	43	21	15	7	49
	An.arabiensis	6	4	2	0	67
	An. rivulorum	8	6	2	0	75
	An. funestus	57	48	7	2	84
	Total	114	79	26	9	
Kombewa	An. gambiae	34	20	10	4	59
	An. arabiensis	3	3	0	0	100
	An. rivulorum	13	10	3	0	77
	An. funestus	44	39	3	2	89
	Total	94	72	16	6	

HBI, human blood index.

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An. funestus can breed in habitats with or without aquatic vegetation. Our data showed that there was no significant difference in the means between habitats with aquatic vegetation and habitats without aquatic vegetation.

Multiple linear regression analysis showed that the abundance of algae in the habitat was not a predicting factor for the density of *An. funestus* in this study. However, Gimnig et al [15] noted that algae abundance was positively correlated with *An. gambiae* density in western Kenya and also an important factor predicting the abundance of *An. pseudopunctipennis* in the Americas [53] and crucial for the development of *An. pretoriensis* in the Rift Valley province of Ethiopia [54].

While Aklilu et al [54] noted that algae abundance was an important factor, distance to the nearest house was not an important component associated with *An. gambiae s.l* and *An. pretoriensis* abundance in their study. However, the proximity of productive larval habitats to human or animal habitation to obtain a blood meal can determine the density of adult mosquitoes [55]. In western Kenya, a previous study reported that the distance to the nearest house was significantly associated with the abundance of *An. gambiae* [56]. Among the environmental variables assessed in our study, principal component and multiple linear regression analyses identified the distance to the nearest house as a major predictor of *An. funestus* abundance in habitats, in agreement with Minakawa et al [56] for *An gambiae*. Our findings suggest that larval source management targeting *An. funestus* aquatic habitats located near houses could reduce the adult mosquito population. The implementation of integrated vector management will also help to control both the aquatic stages and adults vectors in both study sites.

Aquatic predators are well known to influence the abundance of mosquito larvae in breeding environments and are considered beneficial biological control agents of mosquito larvae [57– 60]. Notwithstanding, there was no significant difference in *An. funestus* larval density between aquatic habitats with predators and habitats without predators. A similar study by Ndenga et al [61] noted that the presence of predators was not significantly associated with the low density of *An. gambiae s.l* larvae. Conversely, a previous study in Tanzania [62] and central Sudan [63] reported that most predators were identified in habitats with fewer densities of mosquito larvae. However, we witnessed that there was no reduction in the density of *An. funestus* larvae in the presence of predators in shared habitats. This could be ascribed to the presence of other prey in larval habitats. Kumar *et al* [64] documented that in the presence of alternative prey, the habitats, the larval consumption-ability of predators was significantly reduced in the habitats.

This study revealed that *An. funestus* was the predominant indoor resting vector corroborating the findings of previous investigations in Bungoma [19,20] and Kombewa [12,22]. The relative abundance, high sporozoite rate and HBI of *An. funestus* suggest that it is the main vector mediating malaria transmission in the study areas. We speculate that the adaptation of *An. funestus* to breed in warmer, open sunlit habitats may significantly reduce the developmental time of larval stages and increase the adult population of this species.

We acknowledge, however, the following limitations of our study: first, this study did not integrate detailed water chemistry analysis as part of the variable in assessing the larval ecology of *An. funestus*. Hence, further studies should be undertaken to assess the physicochemical characteristics of aquatic habitats that allow the co-existence of *An. funestus* with other mosquito species. Second, we were unable to examine the productivity of *An. funestus* aquatic habitats and their ability to facilitate the development of larvae to emerged adults.

Conclusion

An. funestus was found breeding in a variety of aquatic habitats and co-existing with larvae of other mosquito species and aquatic predators. An. funestus was found in permanent and semi-

permanent aquatic habitats, with or without aquatic vegetation, slow-moving/disturbed or stagnant water that was clear, opaque, cloudy and brownish. The only significant factor predicting the abundance of An. funestus in the aquatic habitats was the distance to the nearest house. Thus, larval control programmes should aim at targeting aquatic habitats near human dwellings to reduce the abundance of adult An. funestus. This study serves as a guide for the control of aquatic stages of An. funestus using larval source management or larviciding to complement existing vector control strategies. The relative abundance, high sporozoite rate, and HBI also confirm the importance of An. funestus in malaria transmission and the need for continuous vector surveillance before implementing vector control interventions.

Supporting information

S1 Fig. Principal component analysis indicating distance to nearest house as a predictor of *An. funestus* larval abundance. (TIF)

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